A modelling framework for evaluation of the hydrological impacts of nature-based approaches to flood risk management, with application to in-channel interventions across a 29km² scale catchment in the United Kingdom

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Abstract

Nature-based approaches to flood risk management are increasing in popularity. Evidence for the effectiveness at the catchment scale of such spatially-distributed, upstream measures is inconclusive, however. It also remains an open question whether, under certain conditions, the individual impacts of a collection of flood mitigation interventions could combine to produce a detrimental effect on run-off response.

A modelling framework is presented for evaluation of the impacts of hillslope and in-channel natural flood management interventions. It couples an existing semi-distributed hydrological model with a new, spatially-explicit, hydraulic channel network routing model.

The model is applied to assess a potential flood mitigation scheme in an agricultural catchment in North Yorkshire, U.K., comprising various configurations of a single variety of in-channel feature. The hydrological model is used to generate subsurface and surface fluxes for a flood event in 2012. The network routing model is then applied to evaluate the response to the addition of up to 59 features. Additional channel and floodplain storage of

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approximately 70,000m³ is seen with a reduction of around 11% in peak discharge. While this might be sufficient to reduce flooding in moderate events, it is inadequate to prevent flooding in the double peaked storm of the magnitude that caused damage within the catchment in 2012. Some strategies using features specific to this catchment are suggested in order to improve the attenuation that could be achieved by applying a nature-based approach.

Keywords: Natural Flood Risk Management, flood hydraulics, semi-distributed hydrological models, nature-based solutions

1 Introduction

Since the Second World War more intensive agricultural practices, improved field drainage and changes to land management in the U.K. have led to a significant decrease in many catchments' capacity to retain storm run-off (Wheater et al., 2008; Wheater & Evans, 2009). It has been suggested that this has contributed to the occurrence and severity of flooding downstream of such catchments (O'Connell et al., 2007).

In the period 1980 to 2010 there were 563 individual flood events across 37 European countries, and between 1998 and 2013 flooding caused damage estimated at €54 billion (EEA, 2016). The European Floods Directive (EU 2007), was implemented in November 2007 and requires member states to evaluate the extent and risk of flooding and to take action to mitigate those risks. Amongst other recommendations it emphasises the need for "natural water retention" for flood risk mitigation.

The Pitt Review (Pitt, 2008) that followed extensive floods in England in 2007 contained a recommendation to better "work with natural processes". The U.K. Environment Agency and other bodies with interests in flood management responded positively to the Review's recommendations (Environment Agency, 2012), and the approach was included in the Flood and Water Management Act (2010).

In the years since these pieces of legislation were enacted flood mitigation approaches known variously as Natural Flood (Risk) Management (NFM/NFRM), Natural Water Retention Measures (NWRMs, see http://www.nwrm.eu), Nature-based Solutions (NBS), or Working With Natural Processes (WwNP) have gained popularity across Europe. They aim to increase interception and infiltration, slow overland and channel flows and add catchment storage by introducing changes to land use and surface roughness and networks of "soft" engineered features constructed mainly from natural and immediately sourced materials (SEPA, 2012; Quinn et al., 2013). The principle is to restore and improve the catchment's natural ability to retain storm run-off and to release it slowly, leading to attenuation of downstream flood peaks, whilst retaining or enhancing its ecosystem services (Barber & Quinn, 2012; Maclean et al., 2013).

The Floods Directive is envisaged to be implemented in coordination with the Water Framework Directive (WFD; EU 2000), which requires member states to implement river basin management plans that ensure their good ecological and chemical status. It is recognised that nature-based approaches to river basin flood-risk management may also target the objectives of the WFD (Wharton & Gilvear, 2007; EEA, 2016).

Conventional flood protection schemes employ engineered structures and measures, often in combination as a whole system, with two main objectives. Some measures such as raised walls or flood banks (levees) and dredging either increase channel conveyance to reduce water levels locally or simply hold water back from spilling onto the floodplain. Others, such as dams and overflow basins aim to attenuate the input signal of an upstream flood wave. The hydrology of flood mitigation schemes utilising artificial reservoir storage is well understood and the discharge characteristics are largely controllable. Reservoirs are regulated in the U.K. by the Reservoirs Act (1975), amended by the Flood and Water Management Act (2010), which now applies to features with storage capacity greater than 10,000m³.

Nature-based approaches, in contrast, include such techniques as afforestation of hill slopes to increase permeability and downslope transmissivity and so reduce saturated surface run-off and shelterbelts to intercept such flows (Wheater et al., 2008). Tree cover also reduces effective precipitation input through increased canopy interception and evaporation losses. Bosch and Hewlett (1982) reviewed 94 international studies and concluded that, on average, water yield reduced by between 10 and 40mm for every 10% of catchment reforested, evergreens providing the most effect.

Introduction to the channel of wooden screens or barriers, engineered log-jams (ELJs) or large woody debris (LWD) adds friction and reduces flow velocities (Thomas & Nisbet, 2012; Quinn et al., 2013; Dixon et al., 2016). One significant impact of such in-channel interventions is to create a backwater effect (Quinn et al., 2013),. This may lead to reconnection of the flood plain with the channel at storm flows, and the effect can be substantially enhanced if combined with riparian tree-planting to increase flood plain roughness (Thomas & Nisbet, 2007; Nisbet & Thomas, 2008).

Careful positioning of features such as low earth bunds can disconnect fast overland flow pathways from the channel (Quinn et al., 2013). Offline storage areas can also be used to retain flood water diverted from the channel (Nicholson et al., 2012). Typical capacities are in the range of 200m³ to 1000m³ (Quinn et al., 2013), which means they are unlikely to become subject to the Flood and Water Management Act.

Across Europe many NWRM schemes have piloted, for example in areas affected by the Central European floods of 2010 (Skublics & Rutschmann, 2015). In the U.K. there are now over 150 schemes in place (see the online map at http://naturalprocesses.jbahosting.com, JBA Trust, 2016). One of the earliest was in the Belford Burn catchment in Northumberland (see Wilkinson et al., 2010, Nicholson et al., 2012). Significant flooding of the Burn affected the town of Belford, most recently in 2007. A traditional "hard" engineered approach to flood

mitigation was rejected due to cost and the relatively few properties benefiting (Wilkinson et al., 2010). Instead, a nature-based approach was proposed that made use of distributed, unsupervised, on-and off-line features. A feasibility study using a simulated pond network with an aggregate capacity of approximately 20000 m³ showed a 15-30% reduction in peak flows. The storage equates to a total run-off of 3.5mm, or around 1mm/hr over the duration of the smallest storm that caused flooding (Nicholson et al., 2012). Initially 35 of these features were installed, adding approximately 9000-10000 m³ of static storage. The effective storage would, however, be greater due to backwater effects of in-channel features (Quinn et al., 2013). Construction costs were estimated at between £70,000 and £100,000 (Quinn et al., 2013). A further 20 features were subsequently added by the U.K. Environment Agency. As a bottom-up approach, NFRM presents many opportunities for stakeholder engagement, and an essential element is the participation of local stakeholders from an early stage. This was the case of the Ryedale scheme upstream of Pickering in North Yorkshire (Lane et al., 2011; Nisbet et al., 2011). Here a partnership including residents, local authorities and Forest Research were able to install a total of 167 woody debris dams within the channels and 187 bale dams blocking gullies in the upper moorland areas (Nisbet et al., 2011). This was estimated to provide capacity sufficient to protect the affected areas from a 1 in 25 year event. Significant funding subsequently became available and the original approach evolved into a hybrid design with the addition of a £2m, 120,000m³, engineered flood detention basin. Since the Belford scheme was established Nicholson et al. (2012) report a reduction in the magnitude of storm flows. Evidence for the effectiveness of NFRM applied at larger scales is inconclusive (Blanc et al., 2012) and a generalised model to assess its impacts has, up to now, been considered impractical. Individual features can be shown to provide benefits on a local scale. Ghimire et al. (2014) applied hydrodynamic modelling to a single storage feature and demonstrated a reduction of 9% in peak flows immediately downstream of the feature,

achieved by diverting flood discharge into a pond with a capacity of 27,000 m³. Thomas and Nisbet (2012) estimated a reduction in flow velocities of up 2.1 m/s achieved through the restoration of five woody debris dams in a 0.5 km reach and a 15 minute retardation in the downstream flood peak. Thomas and Nisbet (2007) simulated flood flows thorough new riparian woodland along a 2.2km reach and demonstrated a best case 50% reduction in velocity and delay of 140 minutes in the time to peak.

Despite these localised studies, there remains a need for a modelling approach that can assess the impacts of NFM on realistic catchment scales and the interactions between individual interventions and subcatchments. The effectiveness of riparian measures for flood risk mitigation is considered to be due primarily to desynchronisation of subcatchment flood peaks (Thomas & Nisbet, 2007; Nisbet & Thomas, 2008; Dixon et al., 2016). Dixon et al. (2016) estimated a 19% potential reduction in peak flows, mainly due to this effect, downstream of a catchment in which 20-40% of the area had been reforested. There remains uncertainty whether the effects of interventions within individual subcatchments could in fact combine to synchronise previously asynchronous peaks (Blanc et al., 2012). The effects of overflow, or even cascading failure, of in-channel structures could also have a detrimental effect on the response (Nicholson et al., 2012).

We demonstrate here a coupled hydrological hillslope run-off and hydraulic channel model developed to evaluate the impact on storm run-off of a variety of interventions at scales up to that of a small catchment (<100km²). It is able to take into account the effects of antecedent conditions, noted to have a significant impact on a scheme's performance (Blanc et al., 2012). It allows examination of the water level and discharge throughout the channel network which allows the local impacts of in-channel interventions to be estimated. In order to reduce data demands the model utilises by default simplified, but realistic, channel geometries but allows the use of empirical geometries determined from ground surveys.

The study applies the model to evaluate the impact on storm run-off response of a small, intensively cultivated catchment to emplacement of various configurations of in-channel features. A single variety of such feature will be investigated in this paper. We will however discuss how the model could be used to evaluate the impacts of other types of feature and hillslope interventions including enhanced flood plain roughness and hydrological alterations introduced by land use change such as afforestation.

2 Modelling approach

A fully-distributed representation of a realistic catchment and of the processes affected by NFM measures would be both practically and computationally infeasible. In our approach the spatial complexity of the problem is significantly reduced by first applying a semi-distributed hydrological model to simulate hillslope run-off into the channel network. This aggregates similar areas together and treats these as single units within the simulation. Broad-scale measures such as tree-planting and more localised, but greater than grid-scale, features such as run-off detention areas can be included as discrete units in the aggregation. Modifications to those units' parameters and structure to reflect changes introduced when the measures are applied will allow examination of their effects on the response, both at the scale of the unit or on the catchment as a whole.

Lacking a spatially-explicit channel network representation, the hydrological model alone is unable to model the local effects of sub-grid scale, in-channel, features such as debris dams. Network routing is therefore handled by a 1D hydraulic scheme employing a spatiallyexplicit channel representation, which receives as input the surface and subsurface run-off predicted by the hillslope component. The presence of run-off attenuation features can be simulated in the hydraulic model by altering the stage-discharge relationship applied to route discharge through successive sub-reaches of the reaches comprising the network.

Thus the hybrid model can not only handle wider interventions whose effects become noticeable only when applied on the regional scale, but can include sufficient spatial detail to capture features that have greatest impact locally.

The modelling approach is shown in Figure 1. Due to space limitations full details of the schemes are given in the Electronic Supplement included with the paper.



Figure 1. Overview of the coupled hydrological hillslope- hydraulic network routing modelling framework employed in the study. The hydrological component combines landscape layers to provide an simplified hillslope representation as Hydrological Response Units (HRUs). A realisation applies HRU parameters and meteorological inputs to generate input to the channel network of the hydraulic routing component. This module combines the reach inputs with channel geometries into a realisation and outputs water levels and discharges across the network which allows, for example, examination of the effects of in-channel interventions. The two components generate complementary estimates for the catchment outlet discharge

2.1 Hillslope run-off component

Hillslope run-off into the channel network is estimated using the semi-distributed hydrological model Dynamic TOPMODEL. Details of the principles and use of Dynamic TOPMODEL are given in Beven and Freer (2001) and Metcalfe et al. (2015). The implementation used in this study is that by Metcalfe et al., (2016), which employs the opensource R language and environment and is freely available from the Comprehensive R Archive Network (CRAN) archive (https://cran.r-project.org/).

The model groups the catchment into hydrological response units (HRUs) according to landscape characteristics such as topography, land cover or soil type. These units are not necessarily spatially contiguous but they and their time-varying states, averaged over their areas, may be mapped back into space. Each unit is treated as a separate store, the downslope discharge out of which is determined by a suitable storage-discharge relationship. Subsurface flows between the units are distributed according to a flow distribution matrix *W* estimated from surface slopes used as proxies for the direction of maximum hydraulic gradient. This, along with the units' storage-discharge relationships, leads to a kinematic wave formulation for downslope flow out of each unit, with a fixed wave speed characteristic to it and its parameters. The hydrological component has relatively few parameters (see Table 1), making the simple to configure and run and allows rapid identification of behavioural model realisations against observed discharges. Each unit may, however, take a distinct set of parameters. This allows the response to spatially-distributed landscape intervention that may alter hydrological characteristics to be investigated.

Parameter	Description	Units	Value
m	Form of exponential decline in	m	0.002
SrZ _{max}	Max root zone storage	m	0.1
srz ₀	Initial root zone storage	%	0.99
$ln(T_0)$	Lateral saturated transmissivity	log(m²/hr)	18
<i>sd_{max}</i>	Max effective deficit of saturated zone	m	0.5
t_d	Unsaturated zone time delay	hr/m	230
V _{chan}	Channel routing velocity	m/hr	1000
Vof	Overland flow routing velocity	m/hr	50
<i>pr_{fact}</i>	Precipitation weighting factor	-	1.1

 Table 1. Dynamic TOPMODEL parameters (values are those calibrated against November 2012 storm event, or defaults if not included in calibration)

Saturated excess overland flow is routed to the channel downslope through the HRUs using a surface flow distribution matrix W_{of} similar to that applied to the subsurface, and the mean overland flow wave velocity parameter v_{of} ([L]/[T]) whose value can be specified separately for each unit. The effects of introducing surface roughness to slow saturated overland flow, for example through afforestation, can be approximated by changing the value of v_{of} in the appropriate unit. To the simulate behaviour of features to intercept overland flow the appropriate elements of W_{of} can be changed to restrict the downslope drainage out of areas associated with those features.

The flow distribution matrices are also applied to route hillslope subsurface and surface runoff to the channel, represented as a single lumped unit. An estimate for discharge at the catchment outlet is obtained at each time step by routing the channel input in that interval using a time delay histogram derived from the network flow distances. A fixed channel wave velocity v_{chan} ([L]/[T]) is applied throughout the network.

2.2 Channel routing

The simple treatment of channel routing in the hydrological model does not allow access to water levels and flow velocities throughout the river network. These will be required in order to assess the local effects of in-channel interventions, in particular their effect in reconnecting

the channel with the flood plain. A spatially-explicit, 1D hydraulic channel routing scheme has therefore been implemented.

There exist numerous hydraulic models that allow detailed routing of channel discharge. The model employed by Thomas and Nisbet (2007), for example, is HEC-RAS (Brunner, 1995). This solves the steady state Energy Equation or St Venant unsteady flow equations through successive channels sections. It models the effect of a structure impinging the channel such a bridge as a head loss component additional to friction and contraction / expansion losses. Addition of features and specifications of channel profiles and reaches with these detailed modelling packages is, however, a complex task, and not amenable to automation or rapid modification and calibration. The structures that can be evaluated are generally limited to those implemented by the software, and may not correspond to those we may wish to evaluate as part of an NFM scheme.

Our simpler routing model allows rapid specification of a channel network using spatial vector data and parametric channel and overbank geometries. It enables programmatic definition of in-channel features and their insertion at arbitrary locations across the network. Given suitable constraints on the degrees of freedom allowed in their definition, it can be used to calibrate channel and overbank geometries and roughness. The use of simplified, but realistic, channel geometries allows our model to be run quickly and cheaply where detailed morphological data are not available.

The model employs a composite channel section, similar to that used by HEC-RAS, where flow is portioned between overbank and in-channel components and combined to produce the total cross-sectional discharge. This approach has limitations in non-prismatic channels with high sinuosity and where there are significant interactions between the floodplain and channel

flow. It is however generally adequate for flood routing problems where only predictions of discharges and water levels are required (Knight, 2005).

A separate roughness coefficient may be applied to channel and overbank areas. The effect of introducing roughness to the floodplain to slow overbanked flow can therefore be investigated by altering the coefficient applied to the overbank component, albeit there may be variations in the effective roughness due to hydrodynamic effects of flow around obstacles such as fallen trees (Thomas & Nisbet, 2007).

In the solution scheme each river reach is divided into approximately equally-sized subreaches. At each time step a system of differential equations for sub-reach storages through every reach in the channel network is solved using an iterative, upwinded implicit numerical scheme (described in the Electronic Supplement). Boundary conditions at the upstream inlet to each reach are determined from a channel distribution matrix calculated from the detailed river network. Total discharge between successive sub-reaches Q is then calculated according to the common Manning Equation for open channel flow

$$Q = \frac{AR^{\frac{2}{3}}\sqrt{S_0}}{n} \tag{1}$$

where S_0 is the local bed slope, *n* the Manning roughness, *R* the hydraulic radius from the and *A* the flow area perpendicular to the principal downstream flow. Although not used here a correction coefficient could be applied to take into account a non-uniform velocity profile through the channel section. Combined with the mass conservation equation (1) results in a Diffusive Wave approximation to the St Venant equations for open channel discharge. With a known channel cross sectional geometry *A* and *R* can be determined in terms of water depth. For parametric morphologies such as a trapezoid these will be analytic expressions, for those obtained from empirical data they will be in the form of a look-up table. This yields stage-discharge relationships for subcritical flow through the unobstructed sub-reaches. Åkesson et al. (2015) demonstrated that for flood discharge prediction, routing and hydraulic model performance may be more important than channel morphology. Thus by default a simple channel geometry is used in the network routing model. This is a trapezoidal channel section with gently sloped straight floodplain areas extending indefinitely on either side. Such geometries can be parametrised by the channel base width w, the bank slopes s_b , floodplain gradient s_{ob} and bank-full level, D (see Figure 2).



Figure 2. Definition sketch for composite trapezoidal channel profile employed by the routing model. *D* is bank-full depth. $s_b = D/x_b$; $s_{ob} = z_{ob}/x_{ob}$

The input supplied by the hillslope run-off component to its lumped channel unit is distributed between the reaches according to their length and relative upslope areas. This provides a lateral recharge term for each reach, applied uniformly along its length. The solution scheme outputs, for each time step in a simulation, water levels, flow areas and total discharges through those areas at the mid-point of every sub-reach in the network.

2.3 Representation of in-channel features

The addition of run-off attenuation features to the channel network will alter the stagedischarge relationship for the sub-reaches in which they are placed. By replacing the default relationship with those appropriate to the types of features inserted, the local and aggregated effect of an NFM intervention comprising a configuration of in-channel features can be evaluated.

A feature such as weir or barrier may be used that introduces a hydraulic jump. In this case associated energy loss must be taken into account in the relationship. The feature, jump and return to subcritical flow should also be completely contained within a sub-reach. A suitable discharge function, such as a Weir-type equation, can be defined to deal with situations where the feature is overtopped.

Suggested stage-discharge and overflow relationships for some channel features are presented in the Electronic Supplement. Although the geometries considered are highly simplified, Hailemariam et al. (2014) demonstrated that similar idealisations performed adequately against observed data within a simulation of a flood event in a low-lying agricultural catchment in the Netherlands.

3 Study area

The Brompton catchment (Figure 3) lies in the Swale, Ure Nidd and Upper Ouse WFD management catchment, North Yorkshire, U.K., part of the Humber River Basin District. The 29.3 km² area upstream of the village of Water End (-1.416976, 54.36302) is predominately well-drained, undulating arable land with a mean elevation of 68m AMSL. Brompton Beck becomes North Beck downstream of the village before joining the Wiske in Northallerton. Throughout the catchment superficial deposits of sandy clay glacial till with gravel and boulders are overlain on a mudstone bedrock. Rainfall averaged 624 mm p.a. in the period 2008-2014. Convective storms are seen in the summer and synoptic rainfall dominates in the autumn and winter. Water End suffered severe flooding in 2000, with further flooding in September and November 2012. The earlier event had estimated Annual Exceedance Probability (AEP) of 1% and those in 2012 an AEP of 1.3%.



Figure 3. Overview of the Brompton study catchment and its two subcatchments, North Yorkshire, United Kingdom, showing regional (b) and national (a) context within the Humber River Basin District (RBD). Shown are the positions of the hypothetical in-channel features whose influence on the storm response are the main subject of the study, along with an indication of the batch in which they were applied. The position of the rail embankment crossing the main channel of Ing Beck, discussed in the text is shown. Locations of the outlet gauge in Water End and the two nearby rain gauges at Leeming and Topcliffe also provided.

A scheme similar to that implemented in Belford has been suggested for Brompton, but there are significant differences between the catchments. Brompton is intensively farmed, with 95% of its area classed as arable or improved grassland. In Belford only the lower half is in this classification, with rough pasture and upland grazing making up the higher reaches

(Nicholson et al., 2012). There are few areas of woodland in Brompton, which largely precludes the use of low-cost, locally-sourced woody debris dams employed in the wooded riparian area at Belford (Wilkinson et al., 2010, Nicholson et al., 2012).

The Belford scheme was determined to require at least 20,000 m³ of detention storage within the 5.7 km² catchment area (Nicholson et al., 2012). Brompton is approximately 5 times the plan area of Belford and the storage requirements of an effective scheme will be commensurately greater. Quinn et al. (2013) consider that an areal contribution of 1 to 10% would be required to add sufficient storage to significantly attenuate the storm hydrograph. Given arable land prices of up to £18000/ha (RICS, 2016) dedicated artificial storage area could be prohibitively expensive. To avoid this, Quinn et al.. (2013) suggest placing features within the channel or in areas of marginal land in steep-sided banking around it. Storage areas will come into operation comparatively rarely and a complementary approach would be to compensate landowners for damage due to periods of inundation. Guidance is available on appropriate rates for various grades of agricultural land subject to different drainage conditions. For example, a week-long flood is estimated to cause damage of £650/ha to extensive arable under good drainage (Penning-Rowsell, 2013).

In contrast to the largely natural channels within Belford, the Brompton network is heavily modified. There are many enlarged and artificial ditches increasing land-channel connectivity; density is 1203m per km² and there is evidence of extensive and well-maintained subsurface field drainage that connects directly to this ditch network.

3.1 Available data

A Digital Elevation Model (DEM) of the catchment was built from ground-scraped elevation data at 2 metre spatial resolution sourced from the U.K. Environment Agency. Catchment boundaries above the level gauge at Water End were determined using SAGA GIS. The detailed river network (DRN) obtained from the EDINA DigiMap service was "burnt" into the DEM to a maximum depth of 2m, with a small graduated buffer to impose a consistent hydraulic gradient and flow direction within the channels. Thirty-five river reaches were identified with a median length of 674m.

Hourly rainfall data for nearby weather stations were obtained from the BADC MIDAS repository (Met Office, 2006). The nearest station is at Leeming, about 10km to the SW and at a similar elevation; there is also a gauge at Topcliffe, 17km south of the catchment (see Figure 3b). Using the evapotranspiration module provided with the Dynamic TOPMODEL package, a time series of potential evapotranspiration was generated to give a total actual roughly equivalent to a typical yearly water balance of 230mm.

There is a single gauge at the catchment outlet in Water End, recording stage data at 15minute intervals. Data for the period 2002-2013 were obtained from the U.K. Environment Agency.

In 2005 a feasibility study for a flood mitigation scheme upstream of Northallerton was undertaken by JBA Consulting. As part of this a HEC-RAS project was set up for North Beck which connects Brompton with the Wiske. Detailed channel profiles and rating curves are supplied for every reach, including one in which the gauge is found. The first data point was at 1.1m, a level that was exceeded in less than 10% of the September– December 2012 period studied in the subsequent analysis. The rating was extrapolated from this first level to zero by back-solving for a roughness coefficient *n* in the Manning Equation (1) used to estimate discharge in the routing scheme. The local bed slope from the HEC-RAS data was applied and flow area and hydraulic radius for the 1.1m stage were estimated from the channel profile provided. The corresponding discharge was substituted and a value n = 0.03 obtained. The resulting rating curve allowed a time series of reconstructed discharges to be obtained for the entire study period.

Given the uncertainty introduced by the simplified DRN vector data in determining accurate elevations for channel cells, it was felt that the simplicity of applying an overall bed slope for the entire network would more than compensate for the possible improvements in model accuracy from using a slope calculated for each reach. A value was therefore estimated from the elevation range of DEM cells containing the main channel of Ing and Brompton Becks divided by its total length.

3.2 Field drainage

Subsurface field drainage discharging directly into the channels is apparent in many areas of the catchment. They are mostly less than a metre below the surface, constructed from clay, metal or terracotta and up to 30cm in diameter. According to local farmers some date from the late 19th Century and others were installed as a result of subsidies for field drainage in the 1980s; recent work in the NW of the catchment has used plastic piping of a smaller diameter. Dynamic TOPMODEL by default utilises an exponential transmissivity profile. This is parameterised by T_0 ([L]^{2/}[T]), the limiting (saturated) transmissivity and m ([L]) a parameter controlling the decline of conductivity with depth. Such a form is not required and any suitable profile can be used. A discontinuous transmissivity profile was tried to take into account the effect of the additional capacity introduced at the depth where the drains were found. This, however, seemed to only give marginally closer results to observed hydrographs and introduced extra parameters requiring calibration. The effects of the field drainage appeared to be adequately simulated by favouring parameter sets with high T_0 to reflect higher downslope throughput rates introduced by artificial drainage and small absolute values of m to reflect conductivity that declines rapidly beneath the level of the drains.

3.3 Taking account of tunnels beneath the railway line

At (-1.4082, 54.37606) Ing Beck is crossed by the Northallerton–Middleborough railway, which is carried by an embankment around 30m wide (marked as a cross in Figure 3c). The beck flows through an arched concrete-reinforced tunnel within a brick viaduct, with soffit approximately 4m above the channel bed. Guidelines suggest these features to be considered bridges rather than culverts (Ackers et al., 2015).

It was noted that the area immediately upstream of the crossing provides one of the largest areas of marginal riparian land in the catchment. This is probably due to relatively frequent inundation at storm flows resulting from backwater effects introduced by the tunnel. One of the hypothetical measures considered in the following analysis was to close off the tunnel under the railway line with an engineered sluice gate such as those utilised in flood storage basins. An empirical storage-depth relationship was deduced from the elevation data for the riparian area for the reach approaching the tunnel and applied to the representation of that reach in the flow routing scheme. The results of the simulated intervention are presented in Section 5.3.

4 Storm events, September and November 2012

A wet summer in 2012 led to the soil moisture content and water table in early autumn being higher than normal. In the early hours of the 25th of September intense rain led to a rapid rise in flow rate and flooding within the village later that day. Reconstructed peak flow through the level gauge was 19.2 m³/s at 3:15pm. Two months later another series of storms caused further flooding (see Figure 4).



Figure 4. Reconstructed specific discharges for the series of storms described in the text in the period September-November 2012.

A minor storm on the 22nd with maximum intensity of 3mm/hr apparently saturated the soil but did not flood the village. After a more prolonged rainstorm of the same maximum intensity a larger peak discharge of 11.2 m³/s is seen in the hydrograph on Sunday 25th, but local residents reported water levels just below that which would cause flooding. There was an overnight recession followed by another storm that began in the early hours of the following day, with rainfall intensity peaking at 4.4mm/hr around 10am. Rated flow through the gauge peaked at 14.0 m³/s around 5pm. A number of houses were flooded on this occasion. This suggests that for a scheme to prevent flooding in an event of this magnitude it would have to reduce peak flows by the difference in the two maxima, i.e. by approximately 2.8 m³/s, equivalent to a specific run-off of 0.38mm/hr or around 20% of the peak.

4.1 Calibration of hydrological and hydraulic models

Dynamic TOPMODEL was run using a time step of 15 minutes for the double-peaked storm event of $25^{th} - 27^{th}$ November 2012. Rainfall data from the Leeming AWS (see Figure 3b) were used and applied evenly across the catchment area. Data from Topcliffe showed similar timings and quantities, indicating that this event was a synoptic event typical of winter rainfalls in this area. A small scaling factor was applied to reconcile the water balance between input rainfall and observed discharges across the event. Dynamic TOPMODEL does allow for individual rainfall inputs and / or scaling factors for individual HRUs within the catchment model. Given the small extent of the catchment, around 6.5 km square, and the nature of the event, however, it was considered that a uniform rainfall input was adequate for the study.

The model parameters shown in Table 1 were calibrated by running through approximately 5000 realisations with parameters selected at random from the ranges given in the table and applying a performance metric (the Nash Sutcliffe Efficiency, NSE) to the simulated and reconstructed flows at the outlet with a weighting that took into account the amount of saturated overland flow predicted. Simulations with lower amounts of overland flow were favoured with this weighting in order to reflect the subsurface drainage in evidence. Predicted surface and subsurface hillslope run-off a well-fitting simulation were distributed between the reaches of the channel network and applied to the hydraulic channel routing model.

The default symmetrical, trapezoidal cross-sectional geometries were applied throughout the channel network as they agreed qualitatively with morphologies observed from catchment walkovers. Although noting the potential for differences across the network, in order reduce the degrees of freedom an identical parametrisation was used throughout. The response to variations in the channel geometry parameters and Manning roughness values was then used to calibrate the routing model. The timings of both flood peaks were matched to within 15

minutes of those observed. Around 1500 realisations were analysed, and the parameters selected are presented in Table 2.

Parameter	Description	Units	Value
D	Bank-full depth	m	1.8
<i>s</i> _b	Bank slope	-	2.2
Sob	Floodplain slope	-	0.01
n	Manning roughness value for ch	annel	0.03
n _{ob}	Manning roughness value for flo	odplain m	1.7
w	Channel width at base	m	2

Table 2. Hydraulic model parameters calibrated for November 2012 storm event

4.2 Selection and sensitivity analysis of flood mitigation interventions

The intensively-farmed nature of the study catchment means there are few options for widespread tree planting and little marginal land in which to site off-line-storage features. Inchannel features such as rubble and debris barriers that operate at lower flow stages were also not thought suitable interventions. The Swale and Ure Internal Drain Board (IDB) manages much of the catchment and installation of such features would run counter to the Board's remit of keeping the channels clear of debris.

Overland flow barriers or bunds are used to disconnect fast surface flow pathways from the channel and retain the run-off for the duration of an event. Catchment walkovers suggest that much of the main channel is already partially disconnected from the floodplain by low levees formed from dredgings generated during channel maintenance undertaken by the IDB. In addition, most behavioural model simulations weighted to favour higher transmissivities to reflect subsurface field drainage suggested little overland flow. Most of this was generated in areas immediately beside the channel which are likely to have been inundated across much of the event. Overland flow interception barriers were therefore not considered an effective or practical intervention for this catchment.

Barriers or screens with an opening beneath allow unobstructed flow at normal levels but impinge on storm flows that exceed the underside clearance. They slow these higher flows and introduce channel storage through their backwater effect. These structures are similar to the underflow sluice described in Chow (1959) and their hydraulic characteristics are outlined in the Electronic Supplement. Impermeable, rather than "leaky", barriers would be most effective in attenuating open channel discharge, although potentially subjected to high hydraulic stresses. The height of the opening could be configured to meet the levels expected for events of a given return period. In practice the geometry of such features is likely to be constrained by compliance with IDB regulations and environmental legislation such as that to allow fish passage (see for example Baudoin et al., 2014).

The storm peak during the simulated November 2012 event arrived about 30 minutes earlier at the outlet of the Winton Beck subcatchment than at the outlet of Ing Beck. Assuming that the rainfall was not a localised convective event, this suggests that measures to slow the combined catchment response should concentrate on Ing Beck as delaying Winton Beck's response could result in the two peak flows coinciding. There are in addition access issues preventing measures being deployed around Winton Beck.

A sensitivity analysis was undertaken to determine the effect of adding increasing numbers of the in-channel features on the catchment response to the flood event. 59 potential sites for underflow ditch barriers were identified along the 4.7 km length of Brompton and Ing Becks and their unnamed tributaries (see Figure 3c). Features were added in batches of 10, each separated from the nearest upstream barrier by at least 300m, from the highest reaches downstream until the available network was filled. Initially a configuration where all barriers shared the same clearance of 30 cm above the channel bed was considered. An approach aimed at maximising storage utilisation was then tried, where the barriers' clearances were decreased on tributaries and increased on the main channel.

Features were sized according to the bank-depth of 1.8m with a small upper clearance to allow overflow to drain downstream over the feature rather than into neighbouring fields, as stipulated by the IDB. The barrier tops were set to 1.6m above the base of the channel. As features were added to the network, the functional relationship for the underflow barrier was applied to reaches discharging through a feature, and the routing algorithm run using the modified relationship applied for the corresponding element of the input discharge vector. In addition to discharge, water level and any overflow were recorded for each feature.

5 Results

5.1 Storm simulation

Figure 5 shows the discharge simulated for the storm event of $25^{th} - 27^{th}$ November using the parameters for the hydrological and hydraulic models given in Table 1 and Table 2, respectively.



Figure 5. Simulated hydrograph for a storm event that occurred in November 2012 within the Brompton catchment. Discharges reconstructed from observed water levels shown in green, simulated values in blue. Uncertainty bounds of $\pm 5\%$ could be applied to the reconstructed flows.

Reconstructed flows from the observed water levels are also shown. While noting that the observed discharges may be rather uncertain in the way they have been reconstructed (see earlier) the calculated NSE of the simulated discharges was 0.95. The time of the first peak observed at the stage gauge was 11.30am on Sunday 25th November and the corresponding simulated peak was at 11:45am. Time at peak for simulated flows was 16:30 and for the reconstructed flows 17:00 on Monday 26th November. The variation in observed discharges around the storm's peak during the hours of 4pm and 6pm was less than 0.25% of the total. This timing discrepancy was therefore considered to be within the range that could be accounted for by measurement uncertainty.

5.2 Response tests

Results for the batches 40, 50 and 59 barriers in the locations shown in Figure 3, using an opening h_b =0.3m and maximum height of 1.6m, are summarised in Table 3 and displayed graphically in Figure 6. This shows the times of peak discharge, difference between the maximum storage in catchment with and without the barriers, and an overall utilisation factor for the scheme *util*. At any one time the utilisation factor for an individual barrier is the proportion of the potential flow area above the barrier opening that it is intercepting. When the flow is unobstructed the factor is zero, when the water level is above the barrier opening but below its top it is 100%; when the feature starts overflowing the factor starts to falls as the total flow area exceeds the barrier area:

$$util(h) = \begin{cases} 0 & h < h_b \\ 1 & h_b \le h \le h_{\max} \\ \frac{(A - A_0)}{A_b} & h > h_{\max} \end{cases}$$
(2)

Where A_0 is the opening area beneath the barrier, A_b is the barrier area above the opening, A the total channel flow area, h the water level, h_b is the barrier opening height and h_{max} its maximum height above the channel bed.



Figure 6. Absolute (top) and relative discharges for the various configurations employing a barrier clearance of 30cm for all barriers. Maximum attenuation of 0.35 mm/hr is seen within the rising limb of the main storm although largest attenuation of the peak is 0.21mm. Times at peak for the unaltered network are shown by the dotted lines. The maximal configuration delays the main storm peak by 2 hours 45 minutes.

Table 3. Summary of catchment response to adding up to 59 barriers, each with a clearance of 30cm, distributed at 300m intervals along the length of Ing Beck and its tributaries.

No. barriers	Time at peak 26 th Nov	Delay to peak (hours)	Peak discharge (mm/hr)	Reduction in peak discharge (mm/hr)	% reduction of peak	Peak channel storage (m ³)	Max additional channel storage (m ³)	Max utilisation (%)
0	16:30		1.98			74846		
40	16:30	0	1.98	0.00066	0.03	75891	1045	2.5
50	16:45	0.25	1.96	0.018	0.93	90418	15572	21
59	19:15	2.75	1.77	0.21	10.64	170458	95612	27.4

There was virtually no effect on the response until the 40 barrier case and so results for fewer barriers are omitted. Examination of water levels for the 59 feature case shows that at the height of the storm just 20 barriers, mostly located on the main channel, were in operation; utilisation peaked at around 27%. Of the operational barriers 19 were actually overflowing at this point. The greatest attenuation in the peak discharge, seen with 59 barriers, is

approximately 0.21 mm/hr or 10.7%. At about 0.35 mm/hr, the largest attenuation in discharge is, however, observed in the rising limbs of both storm peaks. This suggests that the storage capacity of the scheme is filled before both storm peaks. The maximal case delays the arrival of the main storm peak by 2 hours 45 minutes.

The impact increases rapidly as the final 9 barriers are added to the downstream reaches of Brompton Beck, suggesting that most of the effect is due to lower barriers. The disproportionate effect of these features appears to be due to their effect in diverting flow onto the floodplain, where the much higher roughness reduces flow velocities by a factor of 50. The lowest barrier, for example, diverts overland 45% of the flow from the sub-reach it drains, compared to just 0.3% for the corresponding sub-reach in the unobstructed channel.

The scheme seemed to be operating sub-optimally due to under-utilisation of the higher barriers and those downstream reaching capacity and overflowing. In an attempt to improve the impact a second configuration was applied that lowered the clearance of barriers on tributary reaches to just 10cm and raised barriers on the main channel to 80cm. The intention was to improve the utilisation of higher barriers whilst preventing those lower downstream from running out of capacity. The results are shown in Table 4.

 Table 4. Summary of catchment response to adding up to 59 barriers with clearances of 80cm for those on main channel and of 10cm on its tributaries.

No. barriers	Time at peak 26 th Nov	Delay to peak (hours)	Peak discharge (mm/hr)	Reduction in peak discharge (mm/hr)	% reduction of peak	Peak storage (m ³)	Max additional storage (m ³)	Max utilisation (%)
0	16:30		1.98			74846		
30	16:30	0	1.97	0.0065	0.33	77585	2739	83.1
40	16:30	0	1.94	0.036	1.84	82528	7682	87
50	16:45	0.25	1.91	0.063	3.2	94442	19596	77.1
59	18:30	2	1.79	0.19	9.39	144459	69613	70.3

The 30 barrier case now has some effect and, due to many more of the barriers coming into operation at peaks flows, utilisation improves considerably; up to 50 are used at some point.

With 40 barriers in place, utilisation reaches a maximum of 87%. In contrast to the first configuration the utilisation then actually starts to decrease as further barriers are added. This is a consequence of the lower barriers operating for shorter periods of time due to their much increased clearance. All of the barriers on the main channel still overflow at some point, however.

Despite the higher utilisation, at 9.4% the peak attenuation in the 40 barrier case is actually lower than for 30 barriers and the delay in the main storm peak is reduced to 2 hours. It appears that some of the features are still draining after the first storm and that the remaining capacity is exhausted more quickly when the next storm arrives. Storage retained from earlier in the storm will contribute to the later storm flows. These barriers "run out" of capacity sooner on the rising limb of the second storm peak than the first, leading to a lower impact at the peak. Brim-full reaches will respond almost as quickly as open channels, leading to a smaller delay in the arrival of the peaks.

5.3 Discussion

Although the first configuration reduced peak flows by almost 11%, it would have not been sufficient to prevent flooding. The overall utilisation for the second configuration was larger but it could not provide any greater capacity to absorb the storm peak. It may be that a further approach with slightly higher upstream and lowered downstream clearances would have avoided this effect, and the simplicity of the routing model allows rapid set up and analysis of this and any other configuration. However, any conclusions drawn are likely to be predicated on the type of event considered. A single-peaked event, such as that commonly used in the assessment of flood scheme performance, might have produced markedly different conclusions. The scheme's performance would in this case be constrained only by its absolute storage capacity, rather than its ability to recover between events.

There are other options that would provide significant storage potential sufficient to retain the run-off of the entire event. For example, the railway embankment could be utilised as an "almost" NFM intervention by installing an engineered sluice across the tunnel conveying Ing Beck beneath the line. This could be lowered to reduce the maximum flow rate at storm flows and allow significant quantities of storage to build up behind the embankment during a storm event. The effect of this single intervention was modelled as for the underflow barriers used in the previous analysis but the maximum height h_{max} was set at 5m. Any overbank flow generated behind the embankment was intercepted and contributed to the buildup of storage. The response to applying sluice clearances of 1m and 0.5 m is shown in Figure 7 and summarised in Table 5.



Figure 7. Theoretical attenuation of storm hydrograph achieved by installing a sluice across the railway viaduct tunnel and lowering its clearance to 1m and 0.5m. The smallest clearance attenuates the peak to under the discharge that would give rise to flooding at Water End. In this case storage retained behind the viaduct peaks at 168000m³ and is completely drained by the evening of 28th November, approximately 50 hours after the main storm peak.

Table 5. Summary of catchment response to installation of an underflow sluice across the tunnel conveying Ing Beck under the railway line.

Sluice clearance (m)	Time at peak 26th Nov	Delay to peak (hours)	Peak discharge (mm/hr)	Reduction in peak discharge (mm/hr)	% reduction of peak	Peak storage (m ³)	Max additional storage (m ³)
	16:30		1.98	()		74846	
0.5	15:30	-1	1.42	0.56	28.24	241710	166864
1	16:45	0.25	1.67	0.3	15.4	115400	40554

Applying a 0.5m diameter reduces the second peak by 25% to below the level that would have caused flooding. Peak storage is 168,000m, equivalent over the approximately 72 hour duration of the event to 0.3mm/hr falling across the 8 km² area draining through the viaduct. Maximum water depth of 3m immediately behind the gate was seen some hours before the flood peak and an area of approximately 20 ha flooded to an average depth of about 1m. The area is completely drained by the evening of the 28th, 36 hours after the storm peak. The smallest clearance brings the storm peak forward by one hour. This is simply due to its much smaller magnitude coinciding with a point midway up the rising limb of the unattenuated case.

It is important to recognise that the scenario involving flow restriction through the tunnel has been included solely as an illustration of the potential to gain significant additional storage capacity through a combination of topography and existing infrastructure. Such opportunities might be available in other catchments. In practice, there are many other considerations relating to the safety and operation of infrastructure, and it is highly unlikely that the owners would allow their asset to be subjected to hydraulic loading in this manner. The additional storage capacity introduced could also lead the intervention to become subject to the Reservoirs Act, requiring a detailed geotechnical survey, much higher design specifications and significantly greater capital cost.

Similarly, the constraints imposed by the IDB on the channel features, such as the requirement that they did extend into riparian areas or cause flooding here, clearly limited their effectiveness. Even so, in the modelled scheme downstream barriers appeared to be diverting large quantities of water onto the floodplain: the lowest retained up to 5000m³ in the channel and floodplain immediately upstream, compared to 1500m³ for the same sub-reach for the unobstructed case. This suggests that more relaxed design constraints could allow for a distributed solution that introduced the required storage whilst avoiding the regulatory and cost implications of a large single storage area. If combined with enhancement of riparian roughness to reduce overbank velocities and retain storage on the flood plain, such as was the case in the Ryedale scheme (Nisbet et al., 2011), this could provide significant impacts on storm flows.

Clearly, channel features extending into the flood plain to promote overbank flow would undergo significant hydraulic stress in the course of an event and would have to be constructed to high standards to prevent failure. Well-fixed sturdy impermeable wooden

screens might be able to withstand the hydrodynamic stresses, but the channel bed and banking around them would be subject to scour. A regular maintenance regime would have to be established to ensure that the features' performance did not degrade across events and that their structural integrity remained intact.

6 Conclusions and further developments

This study has demonstrated that, with care, a distributed Natural Flood Risk Management scheme could be implemented in the study catchment to reduce peak flows, but that inchannel features alone would not provide sufficient attenuation to prevent recent flood events, at least given realistic constraints in their number and dimensions. It has shown that even an extensive scheme could be substantially underutilised and provide little or no attenuation to the storm hydrograph. Barriers furthest downstream contribute relatively more attenuation with the attendant risk of failure due to large hydraulic stresses.

A scheme utilising a reduction in the capacity of the railway tunnel feature specific to the catchment could, however, deliver the required response. However this might not be considered an acceptable solution for other reasons, such as the potential for damage at storm flows. In addition, the volumes involved mean that if a structure of equivalent storage capacity were built it would become subject to regulation by the U.K. Reservoirs Act.

We have shown that that there can be a marked contrast between the potential attenuation provided by a network of features and the actual capacity utilised. Designing such schemes to maximise the utilisation is likely to be difficult, and there is always the potential that an optimal configuration for one type of event may in fact prove to be less effective in others. It appears that a double-peaked storm event of the type considered, although not uncommon, can cause problems in distributed schemes. Their aggregated storage is theoretically able to retain much of the storm run-off but during the course of a storm much of this storage is

either unused or overloaded. The available storage capacity may become saturated in intermediate events and not recover sufficiently quickly in order to provide capacity for later storms.

Much of the attenuation is due to the lower barriers, mainly through their effect of reconnecting the floodplain with the channel and diverting large quantities of overbanked flow through the rougher riparian area. It is suggested that schemes should concentrate on encouraging this effect using fewer, larger barriers further downstream in preference to many smaller barriers in the upper reaches. This approach, however, would also introduce much greater hydraulic stresses and correspondingly greater rating specifications and maintenance requirements, albeit that this would be reduced due to the fewer structures. The potential of cascading failures should be considered, not least due to the potential for blockage of downstream structures such as tunnels.

A coupled hillslope run-off – channel network routing model was developed to evaluate the scheme proposed for the catchment. Although applied here to features located entirely within the channel network, it provides a flexible framework within which many types of NFM intervention, including those across hillslopes, can be evaluated. Further studies could develop and extend this framework to apply it to measures such as those described in the introduction. It could be applied to larger catchments with more heterogeneous land-use, and with a wider variety of storm events. This will provide a more robust evidence base on which to assess the applicability, design and effectiveness of flood management schemes across a range of catchments and scenarios. Given the significant uncertainties inherent in modelling systems of this nature, and lack of evidence on how best to simulate changes in the processes affected by NFM measures, uncertainty estimation should be an integral part of such an expanded modelling framework.

The (hypothetical) measures that provided the greatest effect were the throttling of the flow through the railway tunnels and emplacement of barriers downstream to divert water onto the flood plain, both of which would involve the inundation of significant areas of productive agricultural land. This introduces much potential for conflict with land owners and regulatory bodies such as the IDB, whose priorities in terms of channel and run-off management are likely to diverge from those of NFM practitioners. Experience from the Belford scheme where a pilot site was established before the main scheme was begun (Wilkinson et al., 2010), and the stakeholder collaborative approach adopted in Ryedale (Lane et al., 2011) suggest ways that resistance to the adoption of such new approaches to flood risk management may be overcome.

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Supporting information

Electronic Supplement 1: Hydraulic channel routing model

Electronic Supplement 2: Hydraulic characteristics of run-off attenuation features

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Supplementary material can be found in the supporting information included with this paper.

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7 Electronic Supplement :

7.4 Hydraulic channel routing scheme

The depth-averaged one-dimensional St Venant equations for open-channel flow in a prismatic channel with an arbitrary profile, expressed in terms of flow area A, water level h and total discharge Q through the area, are (see Henderson, 1966; Cunge et al., 1980; Knight, 2006; Beven, 2012 and many others)

$$\frac{\partial A}{\partial t} + \frac{\partial Q}{\partial x} = r \tag{S1}$$

$$\frac{\partial Q}{\partial t} + \frac{\partial}{\partial x} \left(\beta \frac{Q^2}{A} \right) + gA \left(\frac{\partial h}{\partial x} + s_f - s_0 \right) = 0$$
(S2)

x is the distance measured in the downstream direction and *r* the lateral recharge per unit length of the channel. Recharge is the sum of specific subsurface base flow q_{bf} and any overland flow q_{of} and is supplied by the hydrological component described in the main text. The total channel input at each time step is distributed between the reaches according to a weighting matrix derived from the surface topography, similar to that used to route base flows between landscape units in Dynamic TOPMODEL. For reach *i* the recharge r_i is then applied uniformly along its length.

The channel bed slope, assumed constant over the reach is s_0 , β is a momentum correction coefficient to account for variation of flow velocity across the flow area, which, in the absence of further information, can be taken as unity. Given these assumptions the mean channel velocity is taken as V = Q/A.

 s_f is the head loss due to friction against the bed per unit length of downstream flow. For uniform flow the friction slope can be approximated using the Manning relationship as

$$s_f = \frac{V|V|n^2}{R^{4/3}}$$
(S3)

with n the roughness and R the hydraulic radius calculated from the wetted perimeter and flow area. If gradually-varied and subcritical flow is assumed, the first two terms in (S2), representing the temporal and advective acceleration respectively can be neglected, leading to a diffusive wave approximation for open channel flow:

$$\frac{\partial h}{\partial x} + s_f - s_0 = 0 \tag{S4}$$

Substituting (S3) into (S4) and rearranging gives an expression for the mean flow velocity. Multiplication by the flow area then leads to an analytical expression for the discharge:

$$Q = \frac{AR^{2/3}}{n} \sqrt{s_0 - \frac{\partial h}{\partial x}}$$
(S5)

Given a channel network comprised of M reaches, a numerical scheme is now constructed as follows in order to solve for channel flows at discrete time steps across a simulation.

In order to reduce the problem from a system of partial differential equations in two independent variables x and t to a coupled system of ordinary differential equations (ODEs) the spatial dimension x is discretised. This is known as the Method of Lines (MOL, see Hamdi et al., 2007). Reach *i* is subdivided into N_i segments, each of length Δx_i . If the flow out of segment *j* is Q_j , its end point at downstream position $x_j = j\Delta x_i$, then mass continuity expressed by (S1) is approximated in this segment by

$$\left. \frac{dA}{dt} \right|_{x=x_j} \approx r_i - \frac{\left(Q_j - Q_{j-1}\right)}{\Delta x_i}, j = 1, N_i$$
(S6)

Water levels h_j are calculated at the boundary of all segments using the chosen channel geometry and the flow area at the current time step. For a rectangular channel of width w the water level is simply A/w but any profile may be specified, including ones where a shallow floodplain is defined. This allows the water surface gradient $\frac{\partial h}{\partial x} - s_0$ to be estimated for the segment. A "downwind" scheme is used so that the effects of afflux behind constrictions and backwater effects at confluences can be propagated upstream, hence

$$\left. \frac{\partial h}{\partial x} \right|_{x=x_i} \approx \frac{\left(h_{j+1} - h_j \right)}{\Delta x_i} \tag{S7}$$

The channel is divided into overbank and in-channel components that may take distinct geometries and roughness values. If the flow remains in channel the appropriate values for the channel may be simply substituted into (S5) in order to obtain the overall discharges. Where water goes overbank the discharges in and out of channel are calculated separately and added to give the overall flow through the subreach. The overbank area is deemed to be that lying above the floodplain; the area above the channel but above the bankfull depth D is allocated to the in-channel flow. In a typical trapezoidal channel this results in the overbank area making a broad-based triangle on each bank with the in-channel component a hexagon. The values for R and A are calculated in each of the areas and they and the appropriate Manning n and the water surface gradient are substituted into (S5) and summed.

In expression (S6) the input, Q_0 , into the first segment must be determined by the upstream inputs, if any, and likewise in (S7) for the downstream water level. This is accomplished as follows.

Reaches are defined strictly between the entry points of tributaries. The channel network is formalised as a directed graph whose edges correspond to the channel reaches and the vertices to springs, confluences and the catchment outlet. A flow direction, or adjacency, matrix F is constructed from the graph, describing how flow is routed downstream out of the reaches. Its elements F_{ij} are equal to 1 if reach j flows into reach i, zero otherwise.

If flows out of all reaches is held in the vector \underline{Q}_{out} and the upstream inputs (zero for source reaches with no upstream input) by \underline{Q}_0 , then mass and momentum conservation is enforced by setting $\underline{Q}_0 = F\underline{Q}_{out}$.

In channels with arbitrary profiles mass conservation implies that flow areas, rather than water levels, are additive. That is, if reaches *i* and *j* with final flow areas A_i and A_j converge to form reach *k* then, assuming no other inputs, the flow area at the first segment of reach *k* is $A_k = A_i + A_j$. For source reaches A_0 will be zero. Assuming quasi-steady flow for the duration of the time step, if $\underline{A'}_0$ is the vector of flow areas at the start of reaches that have an upstream input, given by the vector \underline{A}_{out} , and F' the adjacency matrix for just those reaches then the upstream flow areas can be estimated as $\underline{A'}_0 = F'^{-1}\underline{A}_{out}$.

The corresponding water level will be calculated according to the relationship for the channel profile applied. This allows the water surface gradient calculated in (S7) to be determined at the end of each reach. At the outlet as there is no downstream reach supplied the water surface gradient is carried through from the previous segment. The resulting system of $\sum_{i=1}^{M} N_i$ ODEs is now solved to the end of the time step. A variable time step is used internally within the algorithm; convergence is not predicated on the external time step employed. Details of the runoff dynamics could

be lost at longer time intervals and a step of 15 minutes is typically used. We employ the Livermore Solver for Ordinary Differential Equations (lsode, Petzold, 1993) to solve the system. It is open-source compiled FORTRAN optimised for both stiff and non-stiff systems. It automatically employs an implicit backwards Euler scheme in periods of non-linearity and a forward (explicit) approach when flows are more stable. Segment lengths must be chosen that are short enough to determine the flux gradient reasonably accurately but long enough to prevent it being over-estimated and giving rise to numerical instabilities in front of a flood wave. Samuels (1989) suggests a minimum segment length given by the expression $\Delta x_{min} = \frac{0.15D}{s_0}$; 50m has been used in the analysis presented in the main text.

7.5 Representation of in-channel features

Given an appropriate stage-discharge relationship for a particular type of runoff attenuation, such as those presented below, the above scheme may be modified to take into account the effect of adding features with the channels.

Features may be sited within at the end of any subreach, index f, say, within the network. The value for the output discharge, Q_f , calculated in (S5) is replaced by the appropriate value calculated from the stage-discharge relationship for the feature. If the feature vertical extent is below the bankfull level then overbank flow is assumed to bypass the feature and is calculated as before. An additional overflow component is added to the in-channel flow through the feature, calculated as a function of the water level above its top. This is typically a weir-type equation (ISO, 1980):

$$Q_{over} = \frac{2}{3} C_{\rm w} w_b \sqrt{2g} (h - h_{max})^{3/2}$$
(S8)

where w_b is the dam's width h_{max} along its top side, its maximum height above the channel bed, and C_w a weir coefficient, which must be empirically determined. A typical value is 0.68 for a sharp, non-contracted rectangular weir but will differ for other geometries. When the water overtops the feature, additional flow is calculated by its overflow function applied across the top width.

Approximate storage-discharge relationships for typical runoff attenuation features are presented in the following section of this Electronic Supplement. These include bunds, large woody debris (LWD) barriers, underflow ditch barriers or screens and overflow storage basins.

8 Hydraulic characteristics of runoff attenuation features

As previously described, in order to incorporate runoff attenuation features in the routing scheme it is required to specify a stage-discharge relationship appropriate to that structure. For most the downstream water will have some impact on the discharge through the feature. In the routing scheme described in the previous section this is approximated by the level at the start of the time step.

8.6 Bunds

A simple feature to add storage is a small earth dam (also known as a bund) placed across the channel or in an overland flow pathway. In the case of a bund placed in a channel with maximum height h_{max} above its bed, with the outlet treated as one or more smooth pipes close to the base each with diameter $d < h_{max}$ the discharge may simply be calculated by consideration of conservation of energy.

The Torricelli equation gives the discharge velocity, $v = \sqrt{2g\Delta h}$ where Δh is the head difference calculated by the difference between upstream and downstream water levels, *v* the outlet cross sectional average velocity and *g* the gravitational constant.

For smooth outlet pipe(s) of total cross-sectional area a_p the outlet discharge is thus $Q = a_p \sqrt{2g\Delta h}$. For realistic pipes a coefficient in the range 0.5 to 1 could be applied to account for friction loss.

The above assumes that the water immediately behind the feature is effectively stationary. In practice the approach velocity will affect upstream level, as the velocity head will be translated into potential energy. Equating the upstream and downstream heads:

$$h_2 = \frac{v_1^2}{2g} + h_1 \tag{S9}$$

where v_1 is the approach velocity, h_1 the water level upstream of the dammed, stationary subreach, and h_2 the water level immediately behind the feature. A typical approach velocity of around 1m/s would therefore result in an additional water depth of around 5cm.

For a dam placed across a rectangular channel of width *w* filled to a level *h* the specific storage ([L]³/[L]) immediately upstream is simply *hw*. A more realistic profile would be a symmetrical inverted trapezoidal prism. The channel forms its shortest width *w*, also taken as the width of the base of the dam. The banks to the top of the dam are assumed approximately straight with slope s_b . The specific storage per reach length given a water level *h* at the upstream edge of a dam feature placed within this profile can be shown to be $s = h\left(\frac{h}{s_b} + w\right)$. This allows the dammed water level *h* for a given specific storage *s* to be calculated as

$$h = \frac{s_b}{2} \left(-w + \sqrt{w^2 + \frac{4s}{s_b}} \right) \tag{S10}$$

A functional relationship between storage and discharge can now be established.

When the dam is overtopped the weir equation (ISO, 1980) can be used to predict the overflow discharge:

$$Q_{over} = \frac{2}{3} C_{\rm w} w_b \sqrt{2g} (h - h_{max})^{3/2}$$
(S11)

where w_b is the dam's width, h_{max} along its top side, its maximum height above the channel bed, and C_w a weir coefficient, which must be empirically determined. A typical value is 0.68 for a sharp, non-contracted rectangular weir but will differ for other geometries. When the water overtops the feature, flow calculated by this function is added to the discharge through the outlet pipe(s) to obtain the total discharge through the feature.

In practice an earth or wooden bund would be losing water both through infiltration and seepage through its walls. In this case it may be more realistically modelled as a woody debris dam with tightly-spaced members (see 8.8).

8.7 Ditch barriers

Barriers or screens, typically wooden, can be placed across the channel a small distance above the normal water level. These allow unobstructed drainage at for levels below the top of the barrier opening flows but increasingly impede discharge as the water level rises above this. They are hydraulically similar to underflow sluices (gates) such as those employed in irrigation networks or canals, but in a natural flood management scheme are unlikely to be manually operated.

Consider a horizontal barrier whose top is at a height h_{max} above the channel bed with a clearance of the underside of the barrier from the channel bed of *a* and an upstream water level at a particular time of h_0 . As can be seen from Figure S.1 there is a section of supercritical flow immediately downstream of the barrier. After a short

distance and hydraulic jump the flow returns to subcriticality and the water level recovers. The level of the critical flow section is h_1 and the subcritical downstream, or tailwater, level is h_2 .



Figure S 1. Definition sketch of underflow barrier (Swamee, 1992) The modes of operation of the feature are (Swamee, 1992):

- 1. $h_0 < a$ no impediment to flow due to barrier;
- 2. $h_0 \ge a$ barrier in operation with free discharge;
- 3. $h_1 \ge a$ barrier in operation with submerged discharge;
- 4. $h_0 \ge h_{max}$ overflow discharge in addition to modes 2 or 3.

Error! Reference source not found. The standard depth-discharge relationship for the operational underflow sluice is (Chow, 1959)

$$Q = C_d w_b a \sqrt{2ga} \tag{S12}$$

where w_b is the width of the barrier at its base. This can be seen as analogous with the Torrecelli Equation but with the addition of a energy gain/loss factor C_d , termed the Coefficient of Discharge. C_d incorporates upstream velocity head, friction and energy loss through the hydraulic jump. The development of C_d with h_0 and h_2 must be empirically determined.

For the first mode the open-channel routing approach described in the main paper is applied. For the operational barrier with free discharge (mode 2) Swamee (1992) identifies an approximate functional relationship between C_d and h_0 through experimental results due to Henry (1950):

$$C_{d,free}(h_0) = 0.611 \left(\frac{h_0 - a}{h_0 + 15a}\right)^{0.072}$$

He also derives a condition to determine the existence of free discharge

$$h_0 \ge h_{thresh}(h_2) = 0.81. h_2 \left(\frac{h_2}{a}\right)^{0.72}$$

He then suggests for submerged discharge (mode 3) the following relationship

$$C_{d,sub}(h_0, h_2) = \frac{C_{d,free}(h_0)\Delta h^{0.7}}{(0.32. (h_{thresh} - h_0)^{0.7} + \Delta h^{0.7})}$$
(S13)

where $\Delta h = h_0 - h_2$

Overflow when the feature is overtopped is calculated as for a bund, (S13).

The incorporation of this type of feature into the channel network for flood mitigation purposes is tested in the main paper. When a feature was incorporated in the routing scheme it was found necessary to define a short transitional region for water levels just above the barrier. This is to prevent a discontinuity between discharges in modes 1 and 2 or 3 which prevents the scheme from converging. In the transitional zone the discharge is linearly interpolated between its value in mode 1 immediately below the barrier at in mode 2 at the top of the region. This region could be thought of corresponding to the regime where the barrier is just impinging the flow, with orifice flow starting to become apparent.

8.8 Large woody Debris (LWD) or "leaky" dams

A barrier comprised of tree trunks, large branches, timber lengths or fallen trees placed across the channel adds a significant resistive component to flow. Unlike the impermeable barriers described in the previous section, where water behind the barrier is treated as being effectively stationary, the velocity through the barrier will be nonnegligible and the resistance will be proportional to this velocity. This approximation will become inaccurate with increasing approach velocity and the velocity head attains a similar magnitude to the hydrostatic head. The head loss across the barrier can be estimated by the density and shape of the timber pieces. This structure could be seen as analogous to a trash screen such as those employed to trap debris upstream of structures such as culverts or before the intakes to water treatment works and power plants.

Kirschmer's equation (Kirschmer, 1926) may be used to estimate the proportion of upstream velocity head lost through the structure. Equating total head upstream and downstream of the barrier allows a head loss coefficient, c, to be calculated ,such that

$$c = \beta \left(\frac{s}{b}\right)^{4/3} \sin(\delta)$$

Where *b* is the distance between the bars, *s* their thickness, δ the angle of inclination to the horizontal and β a coefficient reflecting the flow resistance due to the shape of the bars making up the dam. Setting δ to reflect a fairly steep angle, e.g. $\pi/3$, and using the value β =1.79 given for round bars and openings between one quarter and three quarters the size of the timber gives an approximate range 3 < c < 10.

Equating total head upstream and downstream of the barrier allows the discharge given an upstream velocity head and downstream head to be calculated, giving the required stage-discharge relationship.

$$v_{down}^2 = g\Delta h + (1-c)v_{up}^2$$
$$Q = Av_{down} = A\sqrt{g\Delta h + (1-c)v_{up}^2}$$

8.9 Brush and rubble weirs

A matrix of brush or loose rubble placed in the channel also provides a resistance to flow that reduces its velocity and potentially adds some upstream storage at high flows. It could be modelled as a reach of very high Manning roughness or as a finelyspaced trash screen.

Alternatively, discharge through the weir could be treated as a steady state with a power law relationship between velocity v and water surface slope through the dam:

$$\nu = fL \left(\frac{dh}{dx}\right)^n \tag{S14}$$

where *x* is the distance from the upstream side, *h* is the local water level, $f([T]^{-1})$ a frictional resistance factor per unit width of the dam material, assumed constant, and *L* the width through the dam. For smooth laminar flow *n*=1 and for completely turbulent flow *n*=2. At normal levels flow is likely to be partially turbulent so 1 < n < 2

Substituting the discharge per unit channel width $q([L]^3/[L]) = vh$. into (S14) and rearranging to separate variables:

$$\frac{q}{fL} = h \left(\frac{dh}{dx}\right)^n$$
$$\left(\frac{q}{fL}\right)^{1/n} dx = \left(h^{1/n}\right) dh$$

where h1 and h2 are defined as before. Integration through the dam gives

$$\left(\frac{q}{fL}\right)^{1/n} \int_0^L dx = \int_{h_{up}}^{h_{down}} \left(h^{1/n}\right) dh$$
 with $m = \frac{1}{n} + 1 = \frac{n+1}{n}$ Thus

$$Q = fL \left(\frac{h_3}{mL}\right)^n$$

where $h_3 = (h_1^m - h_2^m)$.

There are considerable unanswered questions about the suitability of such features in mid and lower reaches of a catchment. Here they may be subject to significant hydraulic loads during a storm event with the potential for failure and the resulting debris causing damage to structures further downstream.

8.10 Culverts and bridges

A culvert is a tunnel to convey a stream through an embankment or underground, for example under a road or railway. Culverts are hydraulically complex due to their distinct flow modes (Chow, 1959, Environment Agency, 2015). Bridges are similar to culverts except that their openings are larger and the longitudinal dimension is typically much smaller relative to the size of the opening. A structure is generally regarded as a bridge if the opening width is > 2m or the ratio of its width with respect to downstream flow versus the height of the openings is < 5 (Environment Agency, 2015).

Culverts may be utilised as "ready-made" online RAFs by placing a screen or gate on one end to restrict its throughput capacity and so retain storm runoff.

One way of modelling such a feature is as an impermeable dam drained by a smooth semi-circular pipe. When the culvert is partially full the cross sectional flow area *A* is calculated as

$$A = r^{2}acos(1 - h'/r) - (r - h')\sqrt{2rh' - h'}$$

where r is the culvert diameter and h' the clearance of the water level below the soffit. Discharge can then be estimated either using a linear routing algorithm assuming a fixed celerity discharging through the area, or, if more detailed information of the culvert geometry were known, the Manning equation could be applied to give a mean velocity across the profile.

When the culvert is full the pipe-flow derivation given in S8.6 may be employed, albeit that in some modes both entrances may be submerged but the barrel itself not completely filled (Chow, 1959).

In the main paper this approach is applied to the tunnel under a railway line. The effect of applying 0.5m and 1m diameter opening is simulated and the backwater effects cause an extensive area of marginal land and arable pasture to be inundated. In an urban area may cause problematic flooding and the approach is not recommended in these environments.

8.11 Overflow storage basins

Flood flow may be diverted into a storage area close to the channel, either actively by opening a sluice gate or passively by lowering the bank to allow localised overflow into the basin. Likewise, egress from the basin may be controlled by a manually-operated gate or by a channel in the basin side whose base is at the height of the

desired level. Reinforcement to prevent scouring in both cases will be important (Wilkinson & Quinn, 2010).

In an NFMS an unsupervised approach is more likely to be considered. This could be modelled by inserting additional reaches within the DRN: one for the section next to the inlet, one for the basin itself, and one in the main channel at the outlet. Excess flow from the first reach, calculated by the weir equation (S11) would be entirely routed into the "basin" reach

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