

21 **Abstract**

22 China's shift from plant-based to animal-based protein has raised concerns over its
23 health and environmental consequences. Drawing dietary data from nine provinces
24 (1997-2011), this study quantifies the environmental impacts of changing protein
25 consumption and highlights trade-offs between improved diet quality and
26 environmental sustainability. Notable disparities emerge across regions and
27 socioeconomic groups: high-income, male, and well-educated individuals tend to
28 overconsume protein, while older adults, low-income groups, and people with obesity
29 often face protein deficits. Scenario analysis suggests that curbing overconsumption
30 could offset the environmental costs of addressing deficiencies. While adopting the
31 Chinese Dietary Guidelines may increase environmental pressures, these can be
32 mitigated by replacing red meat with poultry or plant-based protein. Our findings call
33 for integrated dietary policies that align health and environmental objectives. Future
34 guidelines should account for regional dietary cultures and the nutritional needs of
35 vulnerable groups to promote more sustainable and equitable food system.

36

37 **Keywords:** protein transition, environmental impact, health-environment nexus,
38 socio-economic disparities, dietary scenario analysis

39 1. Introduction

40 Dietary protein consumption connects human health with planetary health. As
41 the main source of amino acids, protein is fundamental for human health (Moughan,
42 2021; Willett et al., 2019). Adequate protein intake, particularly high-quality protein,
43 is essential for the physical and cognitive development of all age groups (El Bilali et
44 al., 2019). However, the global transition in protein consumption patterns is highly
45 uneven and shaped by socio-economic disparities.

46 Global protein consumption patterns are diverging. In many developing
47 countries, including China, rapid economic growth and urbanization have driven a
48 shift from traditional, plant-based staples toward increased consumption of animal-
49 sourced foods (El Bilali et al., 2019; Guo et al., 2022; Willett et al., 2019). In contrast,
50 many developed countries, especially European countries are undergoing a reverse
51 transition towards plant-based diets and alternative proteins (Aiking and de Boer,
52 2020; Duluins and Baret, 2024). **These divergent trajectories underscore that the**
53 **“protein transition” is not a simple one-to-one food substitution, but rather a broader**
54 **food-system transformation requiring careful attention to implementation pathways**
55 **and potential unintended consequences (Jenkins et al., 2024).**

56 Protein-energy malnutrition contributes significantly to the diet-related health
57 burden globally (Blakstad et al., 2021; Han et al., 2022; Moughan, 2021). In
58 developed countries, excessive protein consumption—particularly from red and
59 processed meats—poses a significant health challenge (Aiking and de Boer, 2020; He
60 et al., 2019; Wu et al., 2014). Average protein intake in many Western nations
61 exceeds recommended levels by 50-100% (Aiking and de Boer, 2020). Protein
62 overconsumption is associated with greater health risks of chronic non-communicable
63 diseases, including cardiovascular diseases, stroke, hypertension, and diabetes (Clark
64 et al., 2019; Guo et al., 2022; Tilman and Clark, 2014; Willett et al., 2019).
65 Meanwhile, many developing countries are experiencing a dual burden of
66 malnutrition with the coexistence of undernutrition and obesity (Behrens et al., 2017;
67 Blakstad et al., 2021; He et al., 2019). Protein deficiency remains a key driver of
68 undernutrition in many developing countries, especially among children under five
69 years old (Aiking and de Boer, 2020; Wang et al., 2023).

70 The global protein transition also carries significant environmental
71 consequences. Intensive production of animal-sourced protein, particularly red meat,
72 requires substantially more resources and generates higher environmental impacts
73 than plant-based alternatives (Aiking and de Boer, 2020; Henchion et al., 2017;
74 Tilman and Clark, 2014; Willett et al., 2019). For instance, greenhouse gas (GHG)
75 emissions per gram of protein from ruminant meat are approximately 250 times
76 higher than those from legumes (Tilman and Clark, 2014). Environmental impacts
77 also vary within animal-based foods: pork, poultry, eggs, dairy, and fish typically
78 have much lower emissions per gram of protein than ruminant meats (Ambikapathi et
79 al., 2022; Poore and Nemecek, 2018; Tilman and Clark, 2014). With the global
80 population projected to reach nearly 10 billion by 2050, current dietary trajectories are
81 likely to intensify pressures on both human health and planetary boundaries (Clark et
82 al., 2019; Duluins and Baret, 2024; Henchion et al., 2017; Wu et al., 2014).

83 Yet the environmental implications of protein transition remain under-explored
84 in developing countries. Cross-country evidence suggests that the health–environment
85 efficiency of diets follows a nonlinear trajectory: it often deteriorates during the phase
86 of rising animal-source food consumption before improving in later stages of
87 development (He et al., 2024). China is currently in this critical middle phase, making
88 it an important case for understanding how dietary shifts can be steered toward more
89 sustainable pathways.

90 In this study, we investigate the nutrition-environment nexus from the
91 perspective of protein consumption in China between 1997 and 2011—a period
92 marked by rapid dietary transition. Using the latest available data from the China
93 Health and Nutrition Survey (CHNS), we estimate individual-level daily protein
94 intake and quantify associated environmental footprints, including GHG emissions,
95 land use, acidification, eutrophication, and freshwater consumption, based on life
96 cycle assessment (LCA). We develop four alternative healthy diet scenarios and
97 compare their environmental impacts to assess the synergies and trade-offs of dietary
98 changes, particularly in addressing the dual burden of protein malnutrition across
99 different population subgroups. Our findings offer practical insights into healthy and
100 sustainable dietary interventions for diverse subpopulations in China and other

101 developing countries undergoing similar protein transition trajectories distinct from
102 those in Western nations.

103 **2. Literature Review**

104 Recent studies addressing the diet–environment–health trilemma have largely
105 focused on developed countries, where shifting diets away from animal-based
106 foods—especially red meat—towards more plant-based options can yield both human
107 health and environmental benefits (Ambikapathi et al., 2022; Clark et al., 2019;
108 Duluins and Baret, 2024; Humpenöder et al., 2022; Rubio et al., 2020; Tilman and
109 Clark, 2014). However, these synergies do not apply universally. Behrens et al.
110 (2017) found that nationally recommended diets could substantially reduce GHG
111 emissions, eutrophication, and land use in high-income nations, but increase
112 environmental impacts in lower-middle-income countries. Similarly, Blakstad et al.
113 (2021) showed that while plant-source foods yield greater health and environmental
114 benefits for adults, animal-source foods remain essential for children’s development
115 in contexts such as urban Ethiopia. **Recent global projections further indicate that**
116 **emerging and developing economies may face higher trade-offs—particularly for**
117 **water use and food affordability—during the initial phases of dietary transitions**
118 **toward healthier patterns (Deng et al., 2025).**

119 The EAT-Lancet Commission proposed a global reference diet emphasizing
120 plant-source foods with limited red and processed meat (Willett et al., 2019).
121 Although useful for advocacy, such global benchmarks often lack sensitivity to local
122 food cultures and consumption patterns, limiting their effectiveness in guiding
123 national and regional dietary transitions (Blakstad et al., 2021; Miller et al., 2022b).
124 Given that dietary patterns are shaped by regional and socioeconomic contexts (Clark
125 et al., 2019; Miller et al., 2022b; Tilman and Clark, 2014), few countries met the
126 EAT-Lancet targets as of 2018 (Miller et al., 2022a).

127 China has experienced rapid dietary transition characterized by declining intake
128 of staple grains and increasing consumption of animal products, fruits and vegetables
129 (Han et al., 2022; He et al., 2019; He et al., 2018; He et al., 2021; Wang, X.X. et al.,
130 2022; Zhang et al., 2022). Between 1982 and 2012, average per capita calorie intake
131 decreased by 12.9%, primarily due to reduced physical labor, while nutritional status

132 improved significantly—the undernourishment rate dropped from 15% in 2002 to
133 11.4% in 2012 (Han et al., 2022; He et al., 2019; He et al., 2021). However, excessive
134 intake of red meat, fats and added sugar has led to rising rates of overweight and
135 obesity (Du et al., 2004; Su et al., 2015; Wang, L. et al., 2021; Wang, L. et al., 2022;
136 Wang, X.X. et al., 2022).

137 China currently faces a dual burden of malnutrition. On one hand, the adult
138 obesity rate increased from 7.1% to 11.9% between 2002 and 2012, with household
139 consumption of grains, meat, and cooking oil exceeding nutritional requirements (Cai
140 et al., 2024; He et al., 2018; Wang, L. et al., 2022). On the other hand, intake of dairy,
141 eggs, seafood, fruits, vegetables, soybeans, and nuts remains insufficient, and protein
142 contributes only 13.2% of daily energy—below the recommended 15–16% (Cai et al.,
143 2024; Jiang et al., 2021). Vulnerable groups such as children under five and adults
144 over 70 continued to experience protein deficiency as of 2015 (Han et al., 2022).

145 Alongside nutritional shifts, diet-related environmental burdens in China have
146 increased across all demographic groups (He et al., 2018; Song et al., 2015). He et al.
147 (2021) found that while reduced calorie intake lowered diet-related GHG emissions
148 by 21%, increased consumption of animal-source foods led to a 25% increase in
149 emissions. Shifting toward healthier diets may thus generate complex environmental
150 synergies and trade-offs (Guo et al., 2022; He et al., 2021; Wu et al., 2022), and the
151 specific types of animal products consumed can make a substantial difference (Cai et
152 al., 2024).

153 Despite these advances, existing nexus studies of China’s dietary change have
154 focused primarily on calorie intake rather than protein consumption (Cai et al., 2024;
155 He et al., 2018; Song et al., 2015). This gap is notable because protein sources are
156 commonly used to classify dietary patterns and carry distinct environmental footprints
157 (Willett et al., 2019). Recent modeling studies indicate that healthy and sustainable
158 diets in China can generate substantial long-term co-benefits, but outcomes depend on
159 feasibility constraints and on bundling measures across consumption and production
160 domains (Cai et al., 2025; Wang, X. et al., 2025). Policy analyses further emphasize
161 that dietary guidelines should integrate sustainability with affordability and cultural
162 acceptability to be actionable (Zhang et al., 2025).

163 Socio-demographic and socioeconomic factors play a critical role in shaping
164 dietary patterns and associated environmental impacts in China (Cai et al., 2024; He
165 et al., 2018; He et al., 2021). Urban, high-income, and highly educated populations
166 generally exhibit better dietary quality but higher environmental burdens (Cai et al.,
167 2024). Greater affordability among high-income groups enables increased
168 consumption of resource-intensive foods such as meat and dairy, exacerbating
169 inequalities in environmental impacts across income groups (Kou et al., 2024).
170 Geographic, climatic, and cultural differences also significantly influence household
171 food choices and related environmental impacts (Xiong et al., 2022). However,
172 current nutrition-environment nexus studies often overlook these heterogeneous
173 contexts, limiting their relevance for targeted policy interventions. A thorough
174 understanding of the linkage between protein consumption and environmental
175 outcomes at the sub-national level remains missing, particularly for developing
176 countries.

177 In light of these gaps, we examine the nutrition-environmental nexus through the
178 lens of protein transition in China. Specifically, we compare nutritional quality and
179 related environmental impacts across socio-economic and demographic groups at the
180 sub-national level, focusing on East, Central, West and Northeast China. We then
181 develop four healthy diet scenarios based on the latest Chinese Dietary Guidelines
182 and historical protein consumption patterns to ensure practical relevance. By
183 leveraging micro-level data, this study captures the implications of protein transitions
184 across subgroups and regions over time.

185 **3. Methods**

186 **3.1 Quantifying Individual Protein Intake Status**

187 Individual protein intake was calculated based on dietary data from the China
188 Health and Nutrition Survey (CHNS), an ongoing longitudinal cohort study jointly
189 conducted by the University of North Carolina at Chapel Hill and the Chinese Center
190 for Disease Control and Prevention. Six waves of surveys were conducted during our
191 study period (1997, 2000, 2004, 2006, 2009 and 2011)¹. The survey covers nine

¹ Although CHNS surveys were also conducted in 2015 and 2019, detailed dietary intake data beyond 2011 are not publicly available, making 2011 the most recent wave accessible for this analysis.

192 provinces representing four major regions of mainland China: East China (Shandong
193 and Jiangsu), Central China (Henan, Hubei, and Hunan), West China (Guizhou and
194 Guangxi) and Northeast China (Heilongjiang and Liaoning)².

195 The nine CHNS provinces account for 41.4% of mainland China's population
196 (2010 census) and encompass diverse socioeconomic, demographic, and cultural
197 contexts across regions (Cai et al., 2024; He et al., 2021). CHNS offers several key
198 advantages for studying the nutrition-environment nexus in China. First, its
199 longitudinal design enables tracking of dietary transitions over an extended period.
200 Second, its detailed individual-level socioeconomic data—including income,
201 education, and urbanization—allow for granular subgroup analyses that most dietary
202 surveys cannot support. Third, CHNS has been widely used in prior studies
203 examining regional protein consumption patterns in China, providing a consistent
204 basis for comparison (He et al., 2018; He et al., 2021; Cai et al., 2024).

205 To improve the representativeness of the sample, we stratified individuals into
206 352 subgroups based on region, age, gender, and urbanization level, and reweighted
207 them using distributions from the closest Chinese Population Census, following He et
208 al. (2021). We validated this approach by comparing the reweighted 2011 per capita
209 daily protein intake (64.9 g) with the 2012 national benchmark (65.9 g) from *the*
210 *Scientific Research Report on Dietary Guidelines for Chinese Residents* (Chinese
211 Society of Nutrition, 2021), showing close alignment. Detailed procedures are
212 provided in the Supporting Information.

213 Despite these strengths, certain limitations should be noted. The CHNS does not
214 include megacities such as Beijing and Shanghai, where higher incomes, greater
215 dietary diversity, and more food-away-from-home consumption may be associated
216 with higher intake of animal-source foods and dairy. As a result, our estimates may
217 understate protein intake among high-income urban residents in the most developed
218 metropolitan areas. Nevertheless, the data remain informative for understanding
219 dietary transitions and subgroup disparities across the four broad regions covered by
220 the survey.

² Although CHNS expanded its coverage to include three additional megacities after 2011, we retained the original nine provinces in our analysis to ensure temporal consistency over our study period.

221 CHNS records household dietary data over three consecutive days using a
222 weighting method, alongside individual intake data collected via 24-hour recall.
223 Following Du et al. (2004), we triangulated more accurate individual daily food
224 intake by combining household weighing and individual self-reported data (see the SI
225 for details). All food items were coded using the *Chinese Food Composition Tables*
226 (Yang, 2004) and categorized into 14 main protein source groups: beef, lamb, pork,
227 eggs, poultry, fish, milk, seafood, other animal sources, grains, legumes, nuts, other
228 plant-based foods, and other foods. Consistent with He et al. (2018), individuals with
229 intake levels exceeding four standard deviations above the group mean were treated
230 as outliers.

231 Individual protein intake was calculated by multiplying food intake by protein
232 density, based on data from the *Chinese Food Composition Tables* (Yang, 2004).
233 Recommended Nutrient Intake (RNI) values vary by age, gender, body weight and
234 physical activity status. We computed each individual's recommended daily protein
235 intake according to China's *Dietary Reference Intakes Guidelines*. Details are
236 provided in the Supporting Information. To evaluate protein intake adequacy, we
237 developed a protein deviation index for each individual:

$$Protein\ deviation_i = \frac{Protein\ intake_i - Protein\ RNI_i}{Protein\ RNI_i} \quad (1)$$

238 where $Protein\ deviation_i$ indicates how much each individual's protein intake
239 level deviates from the RNI; $Protein\ intake_i$ is the actual protein intake of
240 individual i , and $Protein\ RNI_i$ is the recommended protein intake for individual i . A
241 positive value indicates adequate or excessive intake, while a negative value suggests
242 protein deficiency.

243 **3.2 Assessing Dietary Environmental Impacts**

244 Life Cycle Assessment (LCA) coefficients were drawn primarily from the
245 harmonized meta-analysis by Poore and Nemecek (2018), supplemented by Clark et
246 al. (2019) for a small number of food categories where needed (see details in Table S7
247 in the SI). Poore and Nemecek (2018) compiled product-specific impacts across five
248 indicators using a globally reconciled LCA database and reported impacts along the

249 supply chain. This LCA dataset has been widely used to assess the sustainability of
250 various diets globally (Bianchi et al., 2022; Green et al., 2021; Jung et al., 2024).

251 In this study, dietary environmental footprints are estimated using a cradle-to-
252 retail boundary, which captures upstream production and supply-chain stages up to
253 retail, including agricultural inputs and on-farm processes (e.g., fertilizer use and
254 irrigation), as well as post-farm stages including processing, packaging, transport, and
255 retail. This boundary also incorporates transportation-related processes and losses
256 occurring before retail.

257 However, the boundary does not include consumer-stage activities after
258 purchase—such as household cooking, food preparation, and post-purchase food
259 waste—because these stages are not covered in the underlying meta-analytic
260 coefficients. For the few food categories supplemented from Clark et al. (2019), the
261 coefficients represent impacts of production, including the manufacture and use of
262 agricultural inputs and equipment, but exclude transport, processing, retail, and food
263 preparation. We therefore treat these coefficients as production-stage estimates and
264 acknowledge that any remaining boundary mismatch may render our cradle-to-retail
265 estimates slightly conservative for those items.

266 To integrate the datasets, we aggregated CHNS food items into 47 LCA food
267 categories based on weight equivalence (see details in the SI). We focus on five types
268 of dietary environmental footprints: greenhouse gas (GHG) emissions, land use,
269 freshwater use, acidification, and eutrophication. The calculation formula is:

$$X_i = \sum_{m=1}^n (\text{standard weight}_m \times f_m) \quad (2)$$

270 where X_i is the environmental impact of individual i 's diet; standard weight_m
271 is the adjusted weight of food item m ; f_m is the corresponding environmental impact
272 coefficient.

273 We further conducted correlation analyses to examine how socioeconomic and
274 physical characteristics influence individual-level protein malnutrition and dietary
275 environmental impacts.

276 3.3 Potential Healthy Dietary Scenarios

277 To explore the synergies and trade-offs of the nutrition-environment nexus, we
278 constructed four practical healthy dietary scenarios based on China's Dietary
279 Guidelines (CDG) and observed protein consumption patterns across subpopulations.
280 The 2022 CDG provides recommended intake ranges for major broad food groups
281 (e.g., cereals, tubers, fruits, vegetables, animal foods, beans and nuts, and milk and
282 dairy products) and specifies balanced dietary patterns by energy requirement levels
283 (see Table S6) (Chinese Society of Nutrition, 2023).

284 Based on the CDG, we estimated each individual's energy needs by age and
285 gender and derived a corresponding recommended dietary pattern and protein intake
286 structure (see details in the SI). However, the CDG does not differentiate among types
287 of animal foods, despite their wide variation in environmental impacts (Cai et al.,
288 2024). To address this, we developed four dietary scenarios incorporating adjustments
289 within and between food groups to evaluate the environmental synergies and trade-
290 offs of meeting each individual's protein RNI.

291 • **Scenario 1 (S1): Meeting RNI with current diet structure.**

292 S1 assumes that individuals meet their recommended nutrient intake (RNI) for
293 protein while maintaining their original protein intake structure across food groups.
294 This scenario is designed to illustrate the environmental impacts of achieving
295 adequate protein intake without altering existing dietary habits.

296 • **Scenario 2 (S2): Meeting RNI with balanced diet.**

297 In S2, we assume that each individual's diet meets both the protein RNI and the
298 balanced dietary pattern recommended by the CDG. To reflect typical eating habits,
299 we maintain each individual's original protein distribution within food groups.
300 Comparing S2 with S1 highlights the environmental implications of transitioning to
301 more nutritionally balanced diets.

302 • **Scenario 3 (S3): Partial red meat substitution with poultry.**

303 Building on S2, S3 adjusts the composition of animal protein by replacing red
304 meat with poultry—up to 50% of an individual's total meat-based protein—unless
305 their existing poultry intake already exceeds this level. Poultry provides high-quality

306 protein and typically has substantially lower environmental impacts than ruminant
307 meat (Tilman and Clark, 2014). Comparing S3 with S2 isolates the mitigation
308 potential of within-animal substitution while holding overall diet balance constant.

309 The 50% replacement rate is designed as a policy-relevant, moderate-intensity
310 change. It is broadly consistent with the EAT-Lancet reference diet, which allocates a
311 much smaller amount of red meat (14 g/day) than poultry (29 g/day), implying a
312 strong shift away from red meat within animal-source foods (Willett et al., 2019). For
313 Chinese consumers whose animal-source protein is currently dominated by red meat,
314 a 50% substitution represents a substantial but achievable shift in this direction. This
315 choice also reflects evidence that incremental substitutions are more likely to be
316 adopted and sustained than radical dietary changes (Aiking and de Boer, 2020).
317 Finally, it aligns with recent Chinese policy signals that encourage moderating red
318 meat consumption and diversifying animal proteins, with poultry often promoted as a
319 practical alternative (Chinese Society of Nutrition, 2022).

320 • **Scenario 4 (S4): Partial substitution of animal protein with plant protein.**

321 Also based on S2, S4 modifies the protein-source composition by replacing 20%
322 of animal-based protein with plant-based protein. Comparing S4 with S2 isolates the
323 potential mitigation benefits of moving from animal- to plant-based protein while
324 keeping the overall diet aligned with the CDG-recommended balanced pattern.

325 The 20% substitution rate is set as a conservative but meaningful adoption level.
326 First, it is consistent with the Chinese Dietary Guidelines emphasis on plant foods as
327 the foundation of healthy diets and with the National Nutrition Program (2017-2030)
328 “Double Protein” initiative that encourages greater intake of soy and other legumes.
329 Second, it lies within the range of existing dietary practices observed in our sample:
330 individuals in the upper quartile of plant-protein consumption already obtain over half
331 of their protein from plant sources, suggesting that a 20% shift is behaviorally
332 plausible for a broader population. Third, available evidence supports the health and
333 nutritional feasibility of partial substitution: replacing some animal protein with plant
334 protein is associated with lower mortality risk (Song et al., 2016), and well-designed
335 plant-based protein mixes can meet essential amino acid requirements (Mariotti and
336 Gardner, 2019).

337 Across scenarios, S1 preserves observed protein-source shares, whereas S2 to S4
338 are anchored to the CDG-balanced pattern with composition adjustments introduced
339 only where specified above (details in the SI). S3 and S4 are designed as illustrative,
340 policy-relevant scenarios rather than optimization outcomes. Alternative replacement
341 intensities would scale impacts accordingly; assessing such sensitivity is a useful
342 direction for future work when more recent micro-level dietary data become
343 available.

344 **3.4 Uncertainty Analysis**

345 We conducted Monte Carlo simulations ($n = 100$ iterations) to quantify
346 uncertainty in the estimated environmental impacts. Food-specific environmental
347 impact coefficients (Land use, GHG emissions, freshwater use, acidification, and
348 eutrophication) were modeled as normally distributed variables. For most food
349 categories, we adopted estimates from Poore and Nemecek (2018), who assumed a
350 normal distribution for environmental impacts and reported the 5th and 95th
351 percentiles. We derived the standard deviation for each food item from these
352 percentiles using the empirical rule for normal distributions. For the few food
353 categories covered in Clark et al. (2019), where only mean values were reported, we
354 imputed standard deviations using the average standard deviation of all food items
355 from Poore and Nemecek (2018). This approach allowed us to propagate uncertainty
356 stemming from variability in life cycle assessment data across food items. The mean
357 and standard deviation of environmental impact coefficients by food category are
358 summarized in Table S7.

359 **4. Results**

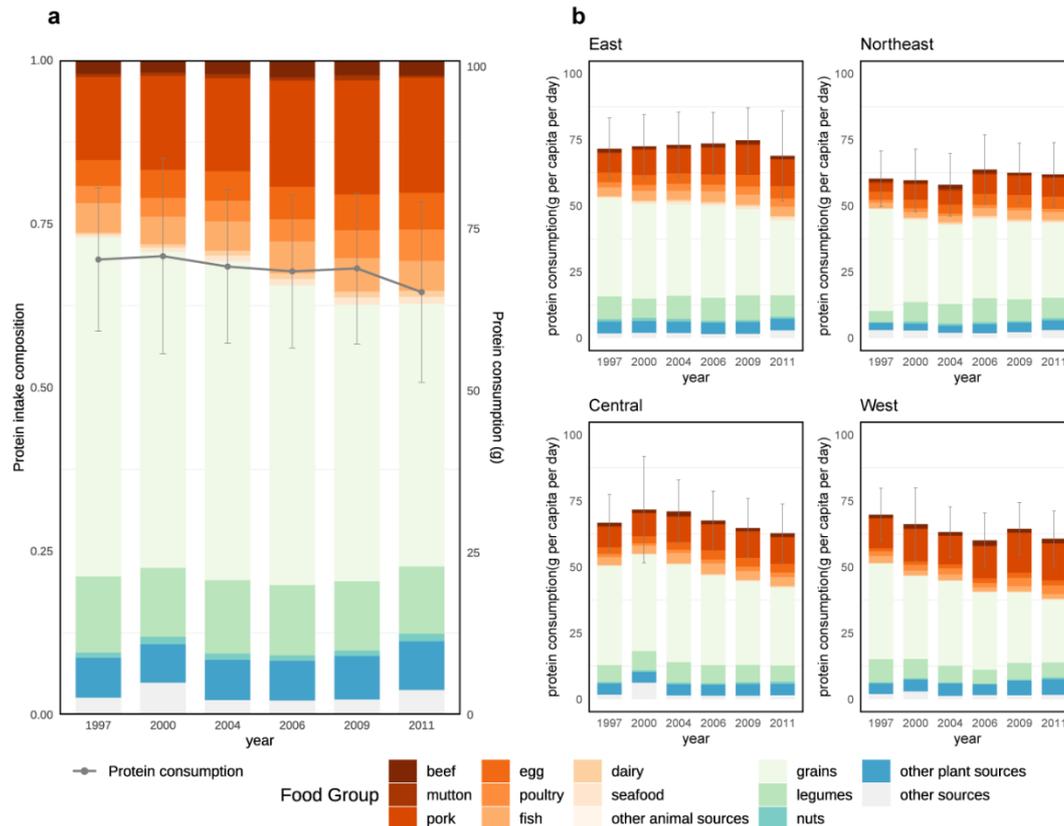
360 **4.1 A Changing Dietary Pattern of Protein Consumption**

361 Chinese residents' per capita protein consumption levels decreased significantly
362 during the study period. Between 1997 and 2011, the average daily protein intake for
363 adults aged 18-65 declined from 70.2 g to 64.9 g, paralleling a decrease in daily
364 energy intake from 2,104.7 kcal to 1,876.0 kcal (see Figure 1a). Despite this decline,
365 the 2011 average remained 11.3% above the RNI of 58.3 g. Although varied among
366 regions, all four regions' average per capita protein intake exceeded the recommended

367 levels, with East China 14.6% higher, followed by Central China (7.1%), West China
368 (4.9%), and Northeast China (3.9%).

369 Consistent with prior studies, our findings indicate a dietary shift among Chinese
370 residents toward increased consumption of animal-based proteins (FAO, 2023; Li et
371 al., 2023; National Bureau of Statistics, 2009). While plant-based foods remained the
372 primary source of dietary protein, their share declined from 70.3% in 1997 to 59.0%
373 in 2011. Over the same period, the contribution of animal-based foods rose
374 significantly from 27.1% to 37.1%. In 2011, cereals were the largest source of dietary
375 protein (40.4%), followed by pork (17.6%) and soybeans (10.0%). Eggs, poultry, and
376 fish contributed 5.4%, 4.7%, and 4.6% of average protein consumption, respectively.
377 However, dairy products contributed just 0.9% of total protein in 2011, with per
378 capita daily intake at only 0.6 g—far below the 9.0 g recommended by the CDG.

379 Protein-source composition differed markedly across regions (Figure 1b). In
380 2011, East China had the highest per capita daily protein intake (69.4 g) and the most
381 diversified structure, with relatively higher contributions from legumes (11.5%) and
382 aquatic foods (fish and seafood, 6.8%) and a lower reliance on pork (14.3%) than
383 other regions. This pattern is consistent with the “Eastern Healthy Dietary Patterns”
384 described in national dietary guidance (Chinese Society of Nutrition, 2021). By
385 contrast, West China recorded the lowest total protein intake (60.3 g/day) yet the
386 highest dependence on red meat, which accounted for 26.3% of total protein—about
387 twice that in Northeast China (13.6%). Central and Northeast China remained more
388 plant-protein dominant, with plant sources accounting for 65.7% and 65.1% of total
389 protein, respectively, compared with 60.5% in East and 59.1% in West China. These
390 regional disparities likely reflect the combined effects of dietary culture, food
391 availability, and socioeconomic development (Zhang et al., 2023), and they
392 underscore that protein-source choices (e.g. red-meat-heavy patterns) may lead to
393 disproportionately higher environmental pressures even when total protein intake is
394 relatively low.



395

396 Figure 1. Per capita daily protein intake of adults aged 18-65 from 1997 to 2011. a,
 397 National average protein consumption weighted by population; b, Regional protein
 398 consumption and source composition. Error bars show one standard deviation from
 399 the mean of per capita daily protein intake (16th and 84th percentiles).

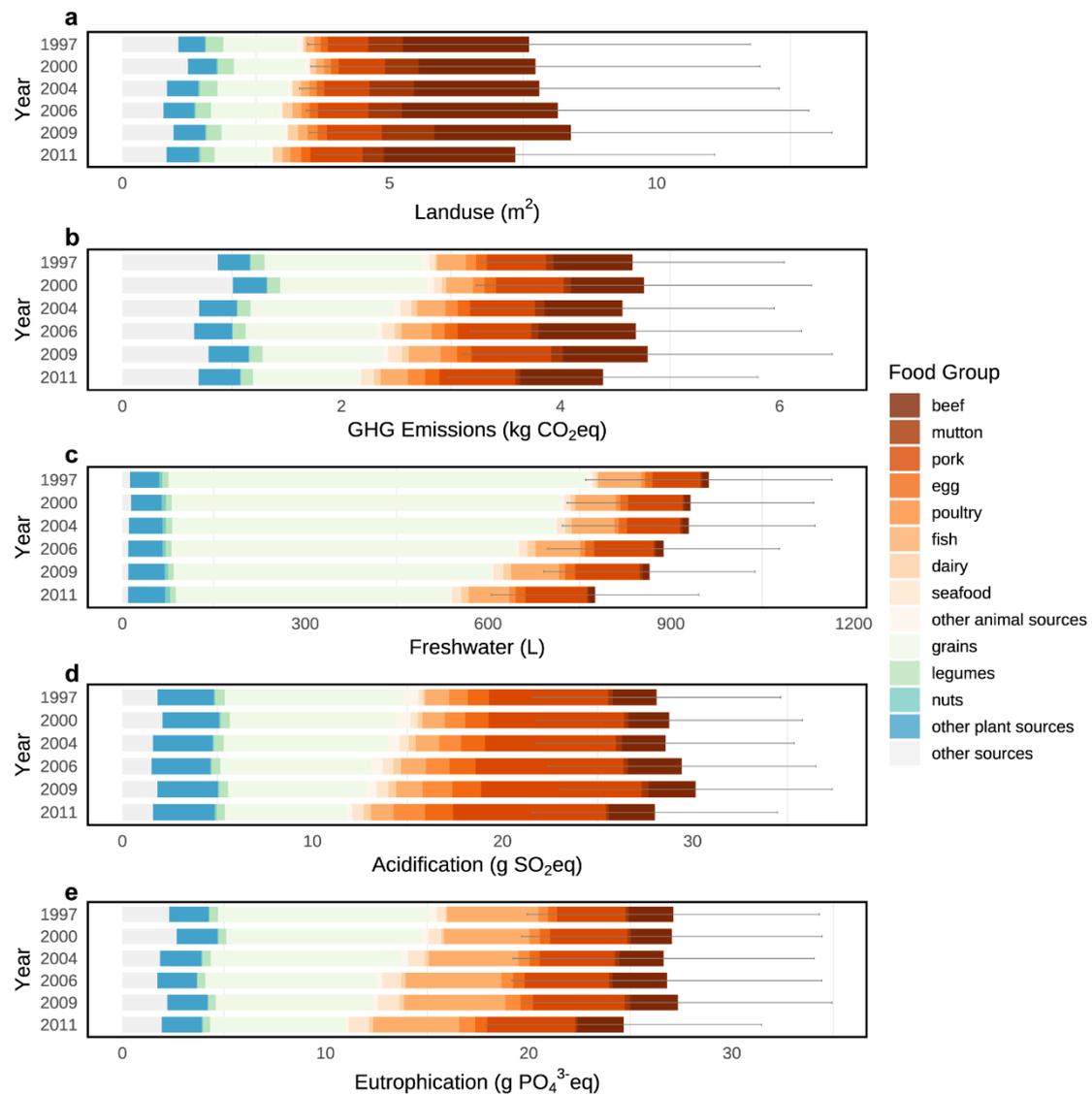
400 4.2 Environmental Impacts of Protein Consumption

401 Figure 2 illustrates the environmental impacts of Chinese residents' protein
 402 consumption from 1997 to 2011. Per capita land use rose from 7.61 m²/day in 1997 to
 403 a peak of 8.39 m²/day in 2009, before declining to 7.35 m²/day in 2011. The increased
 404 land use footprint from 1997 to 2009 was primarily driven by the rapid rise in beef
 405 and pork consumption. Diet-related GHG emissions fluctuated between 4.57 and 4.80
 406 kg CO₂e/day until 2009 and then decreased by 8.5% to 4.39 kg CO₂e in 2011.

407 Acidification impacts were mainly driven by pork consumption, which accounted
 408 for 22.3% to 28.7% of total diet-related acidification across the study period. As pork
 409 intake increased, acidification effects peaked at 30.2g SO₂eq/day in 2009, then
 410 declined slightly due to reduced cereal consumption. Both eutrophication and
 411 freshwater consumption showed a downward trend, largely attributed to the decreased

412 consumption of grains. Eutrophication emissions fell from 27.1 to 24.7 g PO₄³⁻eq
413 /day, and freshwater use declined from 962.2 to 775.7 L/day.

414 Changes in both the quantity and composition of protein intake have jointly
415 shaped China's dietary environmental impacts. While the overall decline in protein
416 consumption helped alleviate some environmental pressures, the rising share of
417 animal-based sources—especially red meat—kept overall impacts relatively high.
418 Notably, West China, despite having the lowest protein intake among the four
419 regions, showed the highest diet-related environmental impacts. This was largely
420 driven by higher consumption of beef and pork, as well as a heavier reliance on rice
421 as a staple food, which contributed 47.5% of its plant-based protein intake. Since rice
422 has higher environmental intensity per gram of protein, its prominence in the diet
423 contributed to greater impacts. In Northeast China, higher lamb consumption also led
424 to increased land use impacts. Detailed regional patterns are shown in Figure S1.



425

426 Figure 2. Environmental impacts of per capita daily protein consumption of adults
 427 aged 18-65 in China (1997-2011). a, Land Use; b, GHG Emissions; c, Freshwater use;
 428 d, Acidification; e, Eutrophication. Error bars show one standard deviation from the
 429 mean of total environmental impacts, corresponding to the 16th and 84th percentiles
 430 from the 100 Monte Carlo simulation trials.

431 4.3 Subgroup Variation in Protein Intake and Associated Environmental Impacts

432 Protein intake levels and related environmental footprints varied considerably
 433 across demographic and socio-economic subgroups in each region (see Figure 3).
 434 Correlation analyses indicate that gender, age, income, education, urbanization level,
 435 and BMI are all significantly associated with both nutritional health and diet-related
 436 environmental impacts (see Table S8).

437 On average, men exhibit higher levels of protein overconsumption and greater
438 environmental footprints than women. Individuals with higher income and education
439 levels also tended to consume more protein, particularly from animal-based sources,
440 resulting in larger environmental impacts. The nutrition-environment relationship
441 between urban and rural residents is more complex. While rural residents consumed
442 1.2% more protein, they generated lower environmental footprints (except for
443 freshwater use), largely due to urban residents' higher preference for animal-based
444 protein.

445 While protein overconsumption was common among most Chinese individuals,
446 certain subgroups remain at risk of protein deficiency, especially older adults, low-
447 income groups, and individuals with overweight or obesity. Adults over 65 had an
448 average protein intake 15.6% below the RNI, with a relatively high reliance on plant-
449 based sources (64.4% of total intake). Low-income individuals (lowest 25%) across
450 all regions fell short of the RNI by an average of 2.3%. People with low education
451 levels or excess body weight also showed protein deficits in some regions.

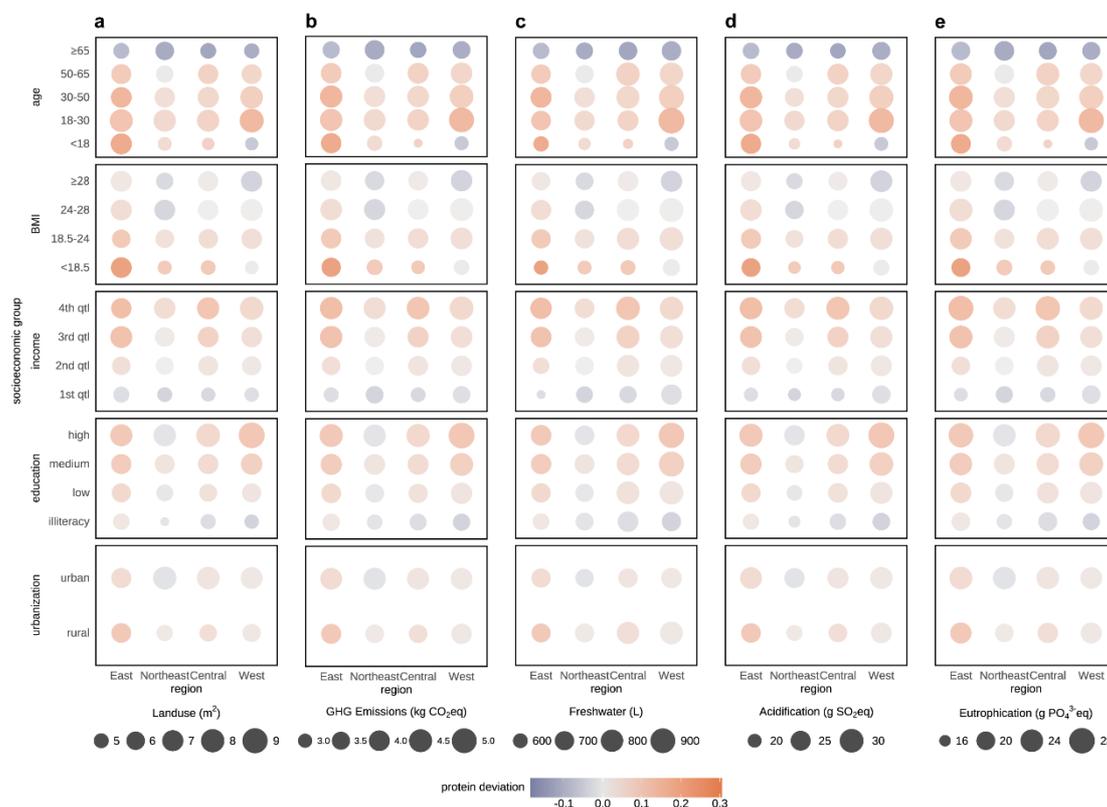
452 These findings are consistent with earlier studies highlighting protein deficiencies
453 among older adults in China, who face increased nutritional needs due to age-related
454 metabolic and physiological changes and higher health risk of sarcopenia (Liu et al.,
455 2018; Rafii et al., 2015; Xu et al., 2015). Improving the quality and quantity of
456 protein intake for older adults remains critical.

457 Among individuals under 18, patterns of protein malnutrition differ by region. In
458 East China, children and adolescents generally had protein overconsumption issues,
459 with an average protein intake 27% above the RNI. In contrast, youth in West China
460 faced a protein deficiency issue, with average intake 9% lower than the RNI. These
461 findings are consistent with previous research documenting malnutrition and stunting
462 among children in West China (Wang et al., 2009; Zeng et al., 2003).

463 Overweight ($24 < \text{BMI} < 28$) and obese ($\text{BMI} > 28$) individuals in Northeast and
464 West China, also exhibited signs of insufficient protein intake. On average, their
465 consumption was 5.0% and 1.7% below the recommended levels respectively in these
466 two regions. This group requires both higher total energy and greater protein intake to
467 meet their nutritional needs. Like older adults, obese individuals relied more heavily

468 on plant-based proteins, which accounted for 62.0% of total protein intake, mainly
 469 due to greater consumption of refined grains commonly found in processed and
 470 convenience foods.

471 In general, individuals with protein deficiencies consume more plant-based and
 472 fewer high-quality animal proteins than those with adequate protein intake (see Table
 473 S9). Addressing these nutritional deficiencies may lead to increased environmental
 474 burdens due to the higher impacts associated with animal-based protein sources.
 475 Additionally, some subgroups, such as urban residents and highly educated
 476 individuals in Northeast China, had average protein intake levels slightly below the
 477 RNI but disproportionately high environmental footprints, indicating inefficient
 478 dietary structures from a sustainability perspective.



479

480 Figure 3. Disparities in Protein Intake and Environmental Impacts in 2011. a, Land
 481 Use; b, GHG Emissions; c, Freshwater use; d, Acidification; e, Eutrophication. Circle
 482 size indicates the level of environmental impact, and the color represents the deviation
 483 of protein intake from the RNI. Red indicates sufficient or excessive intake, while
 484 purple indicates insufficient intake.

485 **4.4 Environmental Impact Assessment of Healthy Protein Intake Scenarios**

486 We designed four healthy dietary scenarios to evaluate the environmental impacts
487 of addressing both protein deficiency and overconsumption across different
488 subgroups. Figure 4 shows the changes in total environmental impacts under each
489 scenario relative to the current diet in 2011. Dietary intake structures for each
490 scenario are detailed in Figure S2.

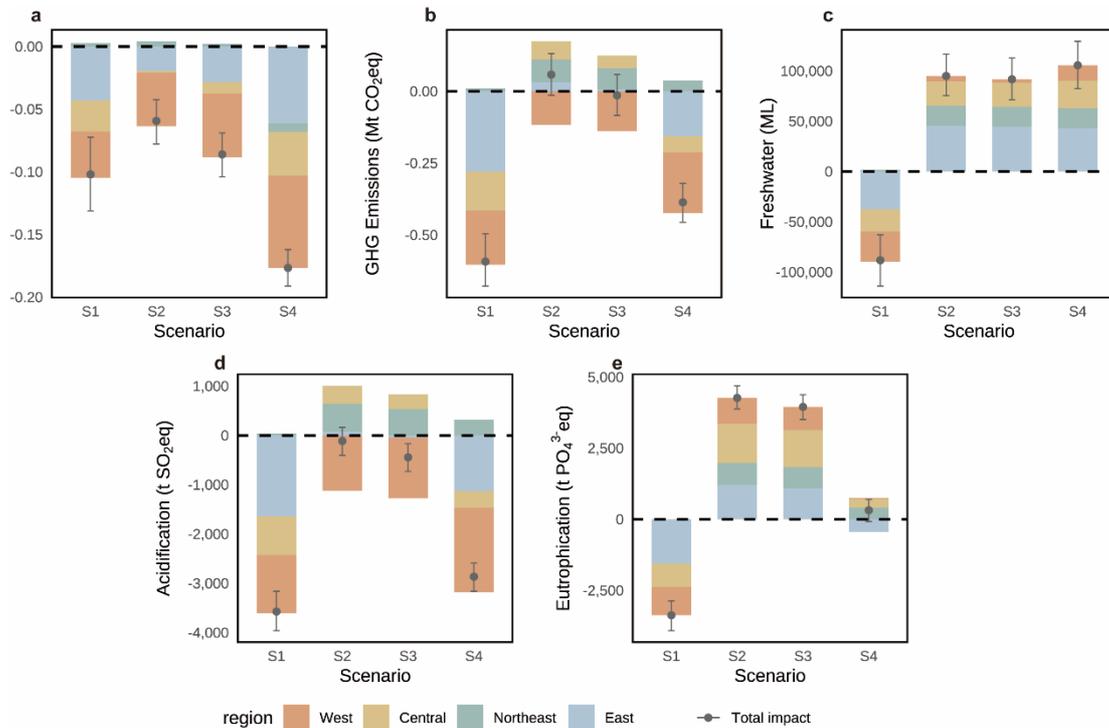
491 Scenario 1 (S1) assumed that all individuals meet their protein RNI while
492 maintaining their existing dietary structure. Since most Chinese residents consume
493 more protein than required, this adjustment leads to nationwide reductions in all five
494 environmental impact categories. However, in Northeast China and among older
495 adults—groups with prevalent protein deficiency—S1 resulted in 18-22% increase in
496 land use, GHG emissions, freshwater use, acidification, and eutrophication due to
497 increased protein intake.

498 Scenario 2 (S2) adopts a more balanced diet recommended by the Chinese
499 Dietary Guidelines, with higher intakes of dairy and fish and lower consumption of
500 cereals and red meat (Chinese Society of Nutrition, 2022). Although designed for
501 improved health across age groups, this shift increases environmental impacts by 5.4-
502 26.4% compared to S1, largely due to greater reliance on resource-intensive animal
503 products. However, in West China, S2 leads to reduced environmental burdens,
504 primarily due to a substantial decline in red meat consumption (Figure S3).

505 Scenario 3 (S3) explores the substitution of 50% of red meat with poultry. This
506 results in modest environmental improvements due to poultry's lower impact per unit
507 of protein. Scenario 4 (S4) replaces 20% of animal-based protein with plant-based
508 alternatives. S4 shows the greatest environmental benefits, reducing land use, GHG
509 emissions, acidification, and eutrophication by 7.0-17.9%. However, S4 is associated
510 with a slight rise in freshwater use.

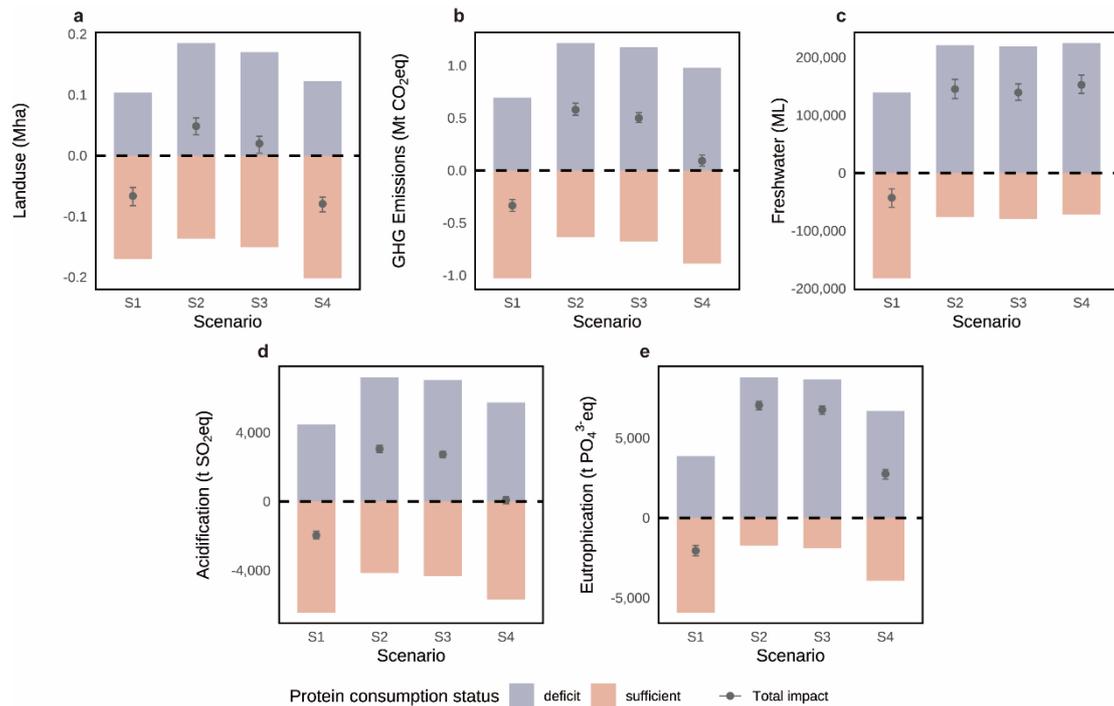
511 We further disaggregated the environmental impacts of eliminating protein
512 under-consumption and overconsumption (Figure 5). Under the current dietary
513 structure, meeting the protein requirements of deficit groups would annually increase
514 land use by 0.104 million hectares, GHG emissions by 0.69 million metric tons
515 CO₂eq, freshwater use by 139.5 billion liters, acidification by 4,473.5 metric tons

516 SO₂eq, and eutrophication by 3,857.0 metric tons PO₄³⁻eq. In contrast, reducing
 517 protein overconsumption yields greater environmental benefits across all impact
 518 categories. Nevertheless, achieving adequate and balanced protein intake for deficit
 519 populations would still result in higher GHG emissions, freshwater use, and
 520 eutrophication than curbing excess consumption.



521

522 Figure 4. Changes of environmental impacts under four healthy diet scenarios
 523 compared to the 2011 baseline at the national level (individuals aged 18-65). a, Land
 524 Use; b, GHG Emissions; c, Freshwater use; d, Acidification; e, Eutrophication. Bar
 525 colors represent regions; gray dots denote combined changes in environmental
 526 impacts. Error bars represent one standard deviation from the mean of total
 527 environmental impacts, corresponding to the 16th and 84th percentiles from the 100
 528 Monte Carlo simulation trials.



529

530 Figure 5. Changes in environmental impacts from eliminating protein
 531 overconsumption and deficiency across all ages under four scenarios. a, Land Use; b,
 532 GHG Emissions; c, Freshwater use; d, Acidification; e, Eutrophication. Gray dots
 533 denote combined changes in environmental impacts. Error bars represent one standard
 534 deviation from the mean of total environmental impacts, corresponding to the 16th and
 535 84th percentiles from the 100 Monte Carlo simulation trials.

536 5. Discussion

537 5.1 Regional Nutrition-Environment Nexus in China

538 China's diverse food culture has led to significant regional variation in dietary
 539 patterns, which in turn affects both health outcomes and environmental sustainability
 540 (Song and Cho, 2017). In East China, abundant access to fish, shrimp and seafoods,
 541 coupled with higher urbanization levels, has fostered a more balanced and diversified
 542 protein intake (Li et al., 2023; Zhang et al., 2021). This region also exhibits a higher
 543 Healthy Eating Index and lower rates of chronic diseases. Reflecting these features,
 544 the Chinese Dietary Guidelines have described it as representative of "Eastern
 545 Healthy Dietary Patterns" (Cai et al., 2024; Chinese Society of Nutrition, 2021).
 546 Consequently, East China faces fewer nutritional or environmental trade-offs when
 547 shifting toward recommended healthy diets.

548 In contrast, West China has the lowest per capita protein intake but the highest
549 share of animal-based proteins, particularly from red meat, which accounted for
550 26.3% of total protein consumption, a level comparable to Western Europe (FAO,
551 2023). A transition towards a healthier diet in this region could reduce meat-sourced
552 protein by nearly half, leading to substantial environmental benefits. Compared to
553 other nations at similar development stages, China has a more abundant protein
554 supply, marked by higher intake of vegetables and oil crops but very limited milk
555 consumption (Schmidhuber et al., 2018). These findings highlight the need for region-
556 specific dietary recommendations that balance nutritional goals with environmental
557 sustainability.

558 **5.2 Socioeconomic Disparities in Protein Intake and Environmental Impact**

559 A key contribution of this study is its socioeconomic subgroup analysis of the
560 environmental implications of addressing both protein overconsumption and
561 deficiency. While protein overconsumption is common, certain groups—such as older
562 adults, low-income individuals, and those with overweight or obesity—still face
563 inadequate protein intake. These conditions often overlap. For instance, low-income
564 urban populations in China show higher rates of overweight and obesity, and adults’
565 mean BMI has steadily increased since the 1980s (Wang, L.M. et al., 2021). As China
566 now has the world’s largest population of individuals with overweight and obesity
567 individuals (Chen et al., 2023; Wang, Y.F. et al., 2021), it is increasingly important to
568 ensure adequate intake of high-quality protein without increasing health risks related
569 to obesity.

570 Aging presents another challenge. The share of the population over age 65
571 increased from 8.9% in 2010 to 13.5% in 2020. Protein deficiency has become a
572 major health concern among older adults. According to the *China Older Adults*
573 *Nutrition and Health Report*, nearly half of older adults were in poor nutritional status
574 in 2018. Adjusting their diets to meet recommended health standards would require a
575 21.2% increase in total protein intake, a 26.4% reduction in plant-based proteins, and
576 a 52.0% rise in animal-based protein. While these changes could reduce chronic
577 disease-related economic costs by \$84.1 billion, they would also lead to a 41.4-50.9%
578 rise in environmental burdens (Chai et al., 2016).

579 Our findings highlight the need for policymakers to consider socio-economic
580 diversity when addressing the nutrition-environment nexus. Beyond promoting
581 sustainable diets in developed areas, more targeted guidance is needed for the elderly,
582 people with obesity, and rural low-income populations. Public awareness of healthy
583 and sustainable diets remains limited among these populations, highlighting the need
584 for more effective outreach through local media and other accessible platforms.

585 **5.3 Synergies and Trade-offs between Nutrition and Environment**

586 Previous studies have often concluded that aligning diets with the Chinese
587 Dietary Guidelines (CDG) can reduce environmental burdens, typically based on
588 national-average data or specific substitution assumptions. Our results complement
589 this literature by showing that China's food system remains largely plant-based, yet
590 animal-source foods have continued to expand during the dietary transition. In this
591 context, moving toward a CDG-aligned balanced diet can improve nutritional
592 adequacy but may increase environmental impacts when it entails higher consumption
593 of animal-source foods, particularly for certain regions and subgroups (Deng et al.,
594 2025; He et al., 2018; He et al., 2024; Wu et al., 2022). Importantly, these trade-offs
595 are not inevitable: they can be substantially mitigated by adjusting which protein
596 sources expand, especially by limiting red meat and shifting toward lower-impact
597 alternatives, and by curbing protein overconsumption.

598 Our scenario results illustrate this source-specific leverage. Relative to S2, S4
599 delivers larger reductions in environmental impacts, reducing land use by 14.2%,
600 GHG emissions by 17.9%, acidification by 15.3%, and eutrophication by 12.8%. In
601 contrast, S3 yields more modest but still meaningful reductions (-8.7%, -11.2%, -
602 9.4%, and -7.6%, respectively). Freshwater use shows the opposite pattern: S4
603 slightly increases freshwater use (+3.2%) due to the higher water requirements of
604 certain legume crops, while S3 reduces freshwater use (-2.1%). Overall, these
605 findings align with global evidence that shifting part of protein demand from animal-
606 to plant-based sources generally produces larger environmental gains than within-
607 animal-source substitutions (Poore and Nemecek, 2018; Clark et al., 2019).

608 Both S3 and S4 maintain adequate total protein intake by design, but they differ
609 in practical nutritional considerations. S3 preserves high biological-value protein

610 through poultry. S4 requires attention to protein quality for some vulnerable groups
611 (e.g., children and older adults), yet a 20% substitution is unlikely to pose protein-
612 quality risks for most adults when implemented through diverse plant sources—
613 particularly soy and legumes—and within an overall balanced diet.

614 Implementation feasibility also differs. S3 is likely to face lower behavioral
615 resistance because it retains animal-protein-based meal structures and builds on
616 culturally familiar consumption patterns. By contrast, S4 may encounter greater
617 barriers in regions with strong meat-eating traditions and limited access to high-
618 quality plant-protein options. For example, West China combines relatively high red-
619 meat reliance with lower legume intake, suggesting that plant-protein substitution
620 may require complementary measures to improve availability and acceptability.

621 Taken together, our findings support a staged or blended pathway: promoting
622 feasible within-animal shifts (e.g., replacing red meat with poultry and fish) as a near-
623 term, lower-resistance option, while expanding plant-protein substitution over the
624 longer term through public procurement, nutrition education, and product innovation.
625 Such an approach could help the CDG evolve toward more operational guidance that
626 integrates environmental considerations and differentiates protein sources by their
627 environmental intensity. This pathway aligns with recent China-wide assessments that
628 highlight strong spatial heterogeneity in diet-related footprints and the value of
629 bundling dietary shifts with broader food-system measures (Cai et al., 2025; Wang, X.
630 et al., 2025; Wang, Y. et al., 2025), while underscoring affordability and cultural
631 acceptability as key prerequisites for successful implementation (Zhang et al., 2025).

632 **5.4 Strengths and Limitations**

633 To our knowledge, this study is among the first to examine China’s nutrition-
634 environment nexus through the lens of protein transition using micro-level dietary
635 data. Leveraging CHNS, we quantify individual protein intake and associated
636 environmental footprints across four regions and diverse subgroups (gender, age,
637 income, education, urban/rural status, and BMI). This design allows us to identify
638 distributional patterns that are typically obscured in national averages and to estimate
639 the implications of addressing both protein overconsumption and deficiency at the
640 national level. By assessing five environmental impact categories and evaluating

641 scenarios grounded in observed consumption habits, we provide a realistic assessment
642 of potential synergies and trade-offs in shifting toward healthier diets.

643 Several limitations should be acknowledged. First, the most recent CHNS wave
644 with publicly available individual-level dietary intake data required for our analysis is
645 2011. Although later CHNS rounds exist, the necessary micro-level dietary modules
646 are not fully accessible for the post-2011 period. Importantly, available evidence
647 suggests that the broad patterns captured here remain relevant. For example, the 2015
648 China Nutrition and Chronic Diseases Report reports a protein-source composition—
649 46.9% from cereals, 5.9% from legumes, and 35.2% from animal-based foods—that
650 is broadly consistent with our 2011 estimates (Chinese Society of Nutrition, 2021).
651 National statistics also indicate a continued rise in animal-source protein in
652 subsequent years. These comparisons support the interpretation of our results as
653 capturing structural features of China’s protein transition, while we also recognize
654 that the absolute levels and some subgroup patterns may have evolved with ongoing
655 urbanization and dietary modernization.

656 Second, we focus on protein intake and its environmental impacts, without jointly
657 assessing other macronutrients (fats and carbohydrates) or micronutrients (vitamins
658 and minerals). While protein is widely used as a basis for dietary classification, this
659 scope may overlook nutritional trade-offs that emerge when optimizing diets across
660 multiple nutrients. Future research integrating multiple nutrients would help evaluate
661 broader health-environmental interactions.

662 Third, our LCA coefficients follow a cradle-to-retail boundary and therefore
663 exclude consumer-stage processes such as household cooking, storage, and post-
664 purchase waste. As a result, our estimates should be interpreted as pre-consumer diet-
665 related footprints rather than full cradle-to-plate impacts. This boundary is well suited
666 for comparing relative impacts across protein sources because upstream production
667 dominates most categories; for example, Poore and Nemecek (2018) show that farm-
668 stage processes account for the majority of impacts (approximately 61% of GHG
669 emissions, 79% of acidification, and 95% of eutrophication), whereas packaging,
670 transport, and retail typically contribute only a small share for most products.
671 Extending the framework to incorporate consumer-stage impacts—particularly in

672 China, where cooking practices and food waste may vary substantially across regions
673 and socioeconomic groups—would further improve the completeness of such
674 assessments.

675 **5.5 Post-2011 Dietary Trends and Contemporary Policy Relevance**

676 While our micro-level analysis is based on CHNS data from 1997 to 2011,
677 subsequent trends indicate that China’s protein transition has continued after our
678 study period, reinforcing the contemporary relevance of our conclusions.

679 National and international statistics confirm the ongoing shift toward animal-
680 based protein in China. According to FAO Food Balance Sheets, China's daily per
681 capita protein supply increased by 15.8 g between 2010 and 2021, reaching 124.6 g—
682 a level that now slightly exceeds that of the United States (124.3 g) and substantially
683 surpasses the global average (FAO, 2023). Although plant sources still account for
684 approximately 60% of total protein supply at the national level, city-level evidence
685 reveals a more pronounced transition: in Beijing, the share of protein from animal-
686 based foods rose from 29.2% in 2010 to 48.0% in 2022, while the contribution from
687 cereals declined from 43.8% to 31.7% over the same period (Chen et al., 2025). Over
688 the same period, meat consumption continued to rise, with poultry consumption
689 increased more rapidly between 2017 and 2021, while pork remained the dominant
690 meat category (USDA ERS, 2023). Recent provincial-scale assessments further
691 document strong spatial heterogeneity in diet-related footprints and identify red-meat
692 reduction as a major mitigation lever in future dietary strategies (Wang, Y. et al.,
693 2025).

694 These dietary shifts carry significant environmental implications. Recent
695 syntheses estimate that China’s food system contributes a substantial emissions
696 burden, with non-CO₂ gases from livestock production playing a major role (Liu et
697 al., 2023). In this context, the key policy-relevant insight from our scenario analyses
698 is not merely whether protein intake increases or decreases, but rather which protein
699 sources change. Our results demonstrate that reducing overconsumption can create
700 environmental “space” to address protein deficits among vulnerable groups, and that
701 adjusting protein composition—by substituting red meat with poultry (S3) or with
702 plant-based protein (S4)—can substantially mitigate environmental pressures. This

703 source-specific logic is increasingly aligned with recent policy signals, including the
704 2022 Chinese Dietary Guidelines' emphasis on dietary diversity, higher intake of
705 legumes and dairy, and moderation of red and processed meat (Chinese Society of
706 Nutrition, 2022), as well as broader national agendas linking nutrition to green
707 development and climate commitments.

708 In sum, although our empirical estimates end in 2011, the direction of China's
709 protein transition has continued, implying that the environmental stakes of aligning
710 nutrition goals with protein-source choices are likely even higher today. The
711 mitigation pathways assessed in this study therefore remain directly applicable to
712 contemporary policy design, particularly for tailoring interventions by region and
713 subgroup and for prioritizing feasible shifts away from red-meat-intensive dietary
714 patterns.

715 **6. Conclusion**

716 China's protein consumption has undergone notable changes in both quantity and
717 composition, marked by a clear shift towards animal-based sources. Despite a modest
718 decline in total protein intake, diet-related environmental impacts remain high, driven
719 primarily by increased consumption of animal-based protein.

720 This study reveals significant disparities across regions and socio-economic
721 subgroups. High-income, male, and more highly educated individuals tend to
722 overconsume protein and generate greater environmental burdens, while older adults,
723 low-income groups, and individuals with obesity often face protein deficiencies. Our
724 scenario analysis demonstrates that reducing protein overconsumption could offset the
725 environmental costs of addressing these deficiencies. Moreover, although adopting
726 the balanced dietary pattern recommended by the Chinese Dietary Guidelines may
727 increase environmental pressures, such trade-offs can be substantially mitigated by
728 adjusting protein sources—specifically, reducing red meat intake and promoting
729 poultry or plant-based alternatives.

730 These findings underscore the need for integrated dietary policies that
731 simultaneously address health and environmental objectives. Future guidelines should
732 account for regional dietary cultures and the nutritional needs of vulnerable
733 populations, thereby supporting a more balanced, equitable, and environmentally

734 responsible food system. Given the continued post-2011 rise in animal-source protein
735 demand, the environmental stakes of aligning nutrition goals with protein-source
736 choices are likely even higher today, reinforcing the timeliness and relevance of the
737 targeted substitution and overconsumption-reduction strategies proposed in this study.

738 Reference

- 739 Aiking, H., de Boer, J., 2020. The next protein transition. *Trends in Food Science & Technology*. 105,
740 515-522. <https://doi.org/10.1016/j.tifs.2018.07.008>.
- 741 Ambikapathi, R., et al., 2022. Global food systems transitions have enabled affordable diets but had less
742 favourable outcomes for nutrition, environmental health, inclusion and equity. *Nature Food*. 3(9),
743 764-779. <https://doi.org/10.1038/s43016-022-00588-7>.
- 744 Behrens, P., et al., 2017. Evaluating the environmental impacts of dietary recommendations. *Proc. Natl.*
745 *Acad. Sci. U. S. A.* 114(51), 13412-13417. <https://doi.org/10.1073/pnas.1711889114>.
- 746 Bianchi, F., et al., 2022. Replacing meat with alternative plant-based products (RE-MAP): a randomized
747 controlled trial of a multicomponent behavioral intervention to reduce meat consumption. *The*
748 *American Journal of Clinical Nutrition*. 115(5), 1357-1366. <https://doi.org/10.1093/ajcn/nqab414>.
- 749 Blakstad, M.M., et al., 2021. Life expectancy and agricultural environmental impacts in Addis Ababa
750 can be improved through optimized plant and animal protein consumption. *Nature Food*. 2(4), 291-
751 298. <https://doi.org/10.1038/s43016-021-00264-2>.
- 752 Cai, H., et al., 2024. How do regional and demographic differences in diets affect the health and
753 environmental impact in China? *Food Policy*. 124, 102607.
754 <https://doi.org/10.1016/j.foodpol.2024.102607>.
- 755 Cai, H., et al., 2025. The multiple benefits of Chinese dietary transformation. *Nature Sustainability* 8(6),
756 606-618. <https://doi.org/10.1038/s41893-025-01560-6>.
- 757 Chai, P., et al., 2016. A study of the economic burden of disease in old age malnutrition in China(in
758 Chinese). *Chinese health economy*. 35(03), 13-16.
- 759 Chen, K., et al., 2023. Prevalence of obesity and associated complications in China: A cross-sectional,
760 real-world study in 15.8 million adults. *Diabetes Obes. Metab.* 25(11), 3390-3399.
761 <https://doi.org/10.1111/dom.15238>.
- 762 Chen, N., et al., 2025. Trends in diet structural composition and quality among adults in Beijing, China
763 (2010-2022). *Frontiers in nutrition*, 12, 1610823. <https://doi.org/10.3389/fnut.2025.1610823>.
- 764 Chinese Society of Nutrition, 2021. *The Scientific Research Report on Dietary Guidelines for Chinese*
765 *Residents*. People's Medical Publishing House, Beijing.
- 766 Chinese Society of Nutrition, 2022. *Dietary Guidelines for Chinese residents*. People's Medical
767 *Publishing House*, Beijing.
- 768 Chinese Society of Nutrition, 2023. *Reference Intake of Dietary Nutrients for Chinese Residents*.
769 *People's Medical Publishing House*, Beijing.

770 Clark, M.A., et al., 2019. Multiple health and environmental impacts of foods. *Proc. Natl. Acad. Sci. U.*
771 *S. A.* 116(46), 23357-23362. <https://doi.org/10.1073/pnas.1906908116>.

772 Deng, Z., et al., 2025. Transitioning to healthy and sustainable diets has higher environmental and
773 affordability trade-offs for emerging and developing economies. *Nature Communications* 16(1),
774 3948. <https://doi.org/10.1038/s41467-025-59275-3>.

775 Du, S.F., et al., 2004. Rapid income growth adversely affects diet quality in China particularly for the
776 poor! *Social Science & Medicine.* 59(7), 1505-1515.
777 <https://doi.org/10.1016/j.socscimed.2004.01.021>.

778 Duluins, O., Baret, P.V., 2024. A systematic review of the definitions, narratives and paths forwards for
779 a protein transition in high-income countries. *Nature Food.* 5, 28-36.
780 <https://doi.org/10.1038/s43016-023-00906-7>.

781 El Bilali, H., et al., 2019. Food and nutrition security and sustainability transitions in food systems. *Food*
782 *and energy security.* 8(2), e00154. <https://doi.org/10.1002/fes3.154>.

783 FAO, 2023. FAOSTAT Statistics Database of the Food and Agricultural Organization of the United
784 Nations. <https://www.fao.org/faostat/en/#data/FBS>. (accessed Feb 26, 2026).

785 Green, A., et al., 2021. Reconciling regionally-explicit nutritional needs with environmental protection
786 by means of nutritional life cycle assessment. *Journal of Cleaner Production.* 312.
787 <https://doi.org/10.1016/j.jclepro.2021.127696>.

788 Guo, Y.X., et al., 2022. Environmental and human health trade-offs in potential Chinese dietary shifts.
789 *One Earth.* 5(3), 268-282. <https://doi.org/10.1016/j.oneear.2022.02.002>.

790 Han, L.Y., et al., 2022. Burden of Nutritional Deficiencies in China: Findings from the Global Burden of
791 Disease Study 2019. *Nutrients.* 14(19), 3919. <https://doi.org/10.3390/nu14193919>.

792 He, P., et al., 2019. Environmental impacts of dietary quality improvement in China. *Journal of*
793 *Environmental Management.* 240, 518-526. <https://doi.org/10.1016/j.jenvman.2019.03.106>.

794 He, P., et al., 2018. The environmental impacts of rapidly changing diets and their nutritional quality in
795 China. *Nature Sustainability.* 1(3), 122-127. <https://doi.org/10.1038/s41893-018-0035-y>.

796 He, P., et al., 2021. Drivers of GHG emissions from dietary transition patterns in China: Supply versus
797 demand options. *Journal of Industrial Ecology.* 25(3), 707-719. <https://doi.org/10.1111/jiec.13086>.

798 He, P., et al., 2024. Health–environment efficiency of diets shows nonlinear trends over 1990–2011.
799 *Nature Food* 5(2), 116-124. <https://doi.org/10.1038/s43016-024-00924-z>.

800 Henchion, M., et al., 2017. Future Protein Supply and Demand: Strategies and Factors Influencing a
801 Sustainable Equilibrium. *Foods.* 6(7), 53. <https://doi.org/10.3390/foods6070053>.

802 Humpenöder, F., et al., 2022. Projected environmental benefits of replacing beef with microbial protein.
803 *Nature.* 605(7908), 90-96. <https://doi.org/10.1038/s41586-022-04629-w>.

804 Jenkins, W.M.N., et al., 2024. Will the protein transition lead to sustainable food systems? *Global Food*
805 *Security* 43, 100809. <https://doi.org/10.1016/j.gfs.2024.100809>.

806 Jiang, H., et al., 2021. Energy intake and energy contributions of macronutrients and major food sources

807 among Chinese adults: CHNS 2015 and CNTCS 2015. *European Journal of Clinical Nutrition*. 75(2),
808 314-324. <https://doi.org/10.1038/s41430-020-0698-0>.

809 Jung, S., et al., 2024. Sustainable dietary patterns and all-cause mortality among US adults. *Int. J.*
810 *Epidemiol.* 53(1). <https://doi.org/10.1093/ije/dyad176>.

811 Kou, J., et al., 2024. Income-based environmental effects of family food consumption and the
812 affordability towards healthy diets. *Sustainable Production and Consumption*. 51, 371-384.
813 <https://doi.org/10.1016/j.spc.2024.09.019>.

814 Li, Y.C., et al., 2023. Structural transition of protein intake in urban China: Stage characteristics and
815 driving forces. *Agribusiness*. 39(S1), 1559-1577. <https://doi.org/10.1002/agr.21860>.

816 Liu, G., Zhang, F., Deng, X., 2023. Half of the greenhouse gas emissions from China's food system occur
817 during food production. *Communications Earth & Environment*. 4(1), 161.
818 <https://doi.org/10.1038/s43247-023-00809-2>.

819 Liu, Z., et al., 2018. Prevalence of Undernutrition and Related Dietary Factors among People Aged 75
820 Years or Older in China during 2010-2012. *Biomed. Environ. Sci.* 31(6), 425-
821 437. <https://doi.org/10.3967/bes2018.056>.

822 Mariotti, F., Gardner, C.D., 2019. Dietary protein and amino acids in vegetarian diets—a
823 review. *Nutrients* 11(11), 2661. <https://doi.org/10.3390/nu11112661>.

824 Miller, V., et al., 2022a. Global, regional, and national consumption of animal-source foods between
825 1990 and 2018: findings from the Global Dietary Database. *Lancet Planet. Health*. 6(3), E243-E256.
826 [https://doi.org/10.1016/S2542-5196\(21\)00352-1](https://doi.org/10.1016/S2542-5196(21)00352-1).

827 Miller, V., et al., 2022b. Global dietary quality in 185 countries from 1990 to 2018 show wide differences
828 by nation, age, education, and urbanicity. *Nature Food*. 3(9), 694-702.
829 <https://doi.org/10.1038/s43016-022-00594-9>.

830 Moughan, P.J., 2021. Population protein intakes and food sustainability indices: The metrics matter. *Glob.*
831 *Food Secur.-Agric.Policy*. 29, 100548. <https://doi.org/10.1016/j.gfs.2021.100548>.

832 National Bureau of Statistics, 2009. Daily Nutrient Intake Per Capita in Urban and Rural Areas.
833 https://www.stats.gov.cn/zt_18555/ztsj/hstjnj/sh2009/202303/t20230303_1926632.html. (accessed
834 Feb 26, 2026).

835 Poore, J., Nemecek, T., 2018. Reducing food's environmental impacts through producers and consumers.
836 *Science*. 360(6392), 987-992. <https://doi.org/10.1126/science.aag0216>.

837 Rafii, M., et al., 2015. Dietary Protein Requirement of Female Adults >65 Years Determined by the
838 Indicator Amino Acid Oxidation Technique Is Higher Than Current Recommendations. *J. Nutr.*
839 145(1), 18-24. <https://doi.org/10.3945/jn.114.197517>.

840 Rubio, N.R., et al., 2020. Plant-based and cell-based approaches to meat production. *Nature*
841 *Communications*. 11(1), 6276. <https://doi.org/10.1038/s41467-020-20061-y>.

842 Schmidhuber, J., et al., 2018. The Global Nutrient Database: availability of macronutrients and
843 micronutrients in 195 countries from 1980 to 2013. *Lancet Planet. Health*. 2(8), E353-E368.

844 [https://doi.org/10.1016/S2542-5196\(18\)30170-0](https://doi.org/10.1016/S2542-5196(18)30170-0).

845 Song, F.F., Cho, M.S., 2017. Geography of Food Consumption Patterns between South and North China.
846 *Foods*. 6(5), 34. <https://doi.org/10.3390/foods6050034>.

847 Song, G.B., et al., 2015. Food consumption and waste and the embedded carbon, water and ecological
848 footprints of households in China. *Science of the Total Environment*. 529, 191-197.
849 <https://doi.org/10.1016/j.scitotenv.2015.05.068>.

850 Song, M., et al., 2016. Association of animal and plant protein intake with all-cause and cause-specific
851 mortality. *JAMA internal medicine*, 176(10), 1453-1463.
852 <https://doi.org/10.1001/jamainternmed.2016.4182>.

853 Su, C., et al., 2015. Epidemics of overweight and obesity among growing childhood in China between
854 1997 and 2009: Impact of Family Income, Dietary Intake, and Physical Activity Dynamics(in
855 Chinese). *Chinese Medical Journal*. 128(14), 1879-1886. [https://doi.org/10.4103/0366-](https://doi.org/10.4103/0366-6999.160648)
856 [6999.160648](https://doi.org/10.4103/0366-6999.160648).

857 Tilman, D., Clark, M., 2014. Global diets link environmental sustainability and human health. *Nature*.
858 515(7528), 518-522. <https://doi.org/10.1038/nature13959>.

859 U.S. Department of Agriculture, Economic Research Service, 2023. China's Meat Consumption: Growth
860 Potential. Economic Research Report No. 320.
861 https://ers.usda.gov/sites/default/files/_laserfiche/publications/106999/ERR-320_summary.pdf
862 (accessed Feb 26, 2026).

863 Wang, L., et al., 2021. The co-benefits for food carbon footprint and overweight and obesity from dietary
864 adjustments in China. *Journal of Cleaner Production*. 289, 125675.
865 <https://doi.org/10.1016/j.jclepro.2020.125675>.

866 Wang, L., et al., 2022. Exploring the environment-nutrition-obesity effects associated with food
867 consumption in different groups in China. *Journal of Environmental Management*. 317, 115287.
868 <https://doi.org/10.1016/j.jenvman.2022.115287>.

869 Wang, L.M., et al., 2021. Body-mass index and obesity in urban and rural China: findings from
870 consecutive nationally representative surveys during 2004-18. *Lancet*. 398(10294), 53-63.
871 [https://doi.org/10.1016/S0140-6736\(21\)00798-4](https://doi.org/10.1016/S0140-6736(21)00798-4).

872 Wang, X., et al., 2025. Bundled measures for China's food system transformation reveal social and
873 environmental co-benefits. *Nature Food* 6(1), 72-84. <https://doi.org/10.1038/s43016-024-01100-z>.

874 Wang, X.L., et al., 2009. Stunting and 'overweight' in the WHO Child Growth Standards - malnutrition
875 among children in a poor area of China. *Public Health Nutrition*. 12(11), 1991-1998.
876 <https://doi.org/10.1017/S1368980009990796>.

877 Wang, X.X., et al., 2022. The triple benefits of slimming and greening the Chinese food system. *Nature*
878 *Food*. 3(9), 686-693. <https://doi.org/10.1038/s43016-022-00580-1>.

879 Wang, X.Z., et al., 2023. Global food nutrients analysis reveals alarming gaps and daunting challenges.
880 *Nature Food*. 4(11), 1007-1017. <https://doi.org/10.1038/s43016-023-00851-5>.

881 Wang, Y., et al., 2025. Exploring dietary transition and its environmental mitigation effects in China.
882 Sustainable Production and Consumption 61, 37-47. <https://doi.org/10.1016/j.spc.2025.10.006>.

883 Wang, Y.F., et al., 2021. Obesity in China 3 Health policy and public health implications of obesity in
884 China. *Lancet Diabetes Endocrinol.* 9(7), 446-461. [https://doi.org/10.1016/S2213-8587\(21\)00118-](https://doi.org/10.1016/S2213-8587(21)00118-2)
885 [2](https://doi.org/10.1016/S2213-8587(21)00118-2).

886 Willett, W., et al., 2019. Food in the Anthropocene: the EAT–Lancet Commission on healthy diets from
887 sustainable food systems. *Lancet.* 393(10170), 447-492. [https://doi.org/10.1016/s0140-](https://doi.org/10.1016/s0140-6736(18)31788-4)
888 [6736\(18\)31788-4](https://doi.org/10.1016/s0140-6736(18)31788-4).

889 Wu, G.Y., et al., 2014. Production and supply of high-quality food protein for human consumption:
890 sustainability, challenges, and innovations. *Ann. N.Y. Acad. Sci.* 1321(1), 1-19.
891 <https://doi.org/10.1111/nyas.12500>.

892 Wu, H.J., et al., 2022. A new dietary guideline balancing sustainability and nutrition for China's rural and
893 urban residents. *iScience.* 25(10). <https://doi.org/10.1016/j.isci.2022.105048>.

894 Xiong, X., et al., 2022. How urbanization and ecological conditions affect urban diet-linked GHG
895 emissions: New evidence from China. *Resources, Conservation and Recycling.* 176, 105903.
896 <https://doi.org/10.1016/j.resconrec.2021.105903>.

897 Xu, X.Y., et al., 2015. Evaluation of older Chinese people's macronutrient intake status: results from the
898 China Health and Nutrition Survey. *Br. J. Nutr.* 113(1), 159-171.
899 <https://doi.org/10.1017/S0007114514003444>.

900 Yang, Y., 2004. *China Food Composition (Book 2)*, 1st ed. Peking University Medical Press, Beijing.

901 Zeng, L., et al., 2003. Analysis on malnutrition of children under 3 years old in 40 poor counties in the
902 western areas of China(in Chinese). *Chinese Journal of Public Health.* 19(1), 55.

903 Zhang, H., et al., 2022. The greenhouse gas footprints of China's food production and consumption
904 (1987–2017). *Journal of Environmental Management.* 301, 113934.
905 <https://doi.org/10.1016/j.jenvman.2021.113934>.

906 Zhang, H.Q., et al., 2021. Seafood consumption patterns and affecting factors in urban China: A field
907 survey from six cities. *Aquaculture Reports.* 19, 100608.
908 <https://doi.org/10.1016/j.aqrep.2021.100608>.

909 Zhang, H.Z., et al., 2023. Pork Consumption Patterns among Rural Residents in China: A Regional and
910 Cultural Perspective (2000-2020). *Agriculture.* 13(10).
911 <https://doi.org/10.3390/agriculture13101888>.

912 Zhang, Y., Wang, J., Fan, S., 2025. Healthy and sustainable diets in China and their global implications.
913 *Agricultural Economics* 56(3), 349-359. <https://doi.org/10.1111/agec.70020>.

914
915