

1 **Early historical forest clearance caused major degradation of water quality at Lake Væng,**
2 **Denmark**

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29

30 **ABSTRACT**

31 Although humans have impacted their environment over millennia, details of these impacts,
32 especially on aquatic systems, is still surprisingly scarce despite potential disturbance by early land
33 use. This study examined a high-resolution radiocarbon-dated Holocene record from the Danish
34 Lake Væng, using geochemical and biological proxies, and related the observed impacts to other
35 lake records with catchment disturbance. The results indicate a lengthy and varying history of
36 aquatic eutrophication linked to human activity. Modest impacts on the lake coincided with the first
37 signs of landscape disturbance during the Neolithic (c. 4500 cal. yrs BP). Observed impacts
38 intensified in the Late Bronze and Pre-Roman Iron Age. Viking Age/Medieval deforestation and

39 erosional inputs to the lake associated with new ploughing technology (1200 cal. yrs BP), however,
40 led to a major reorganisation of the aquatic ecosystem. Filamentous bloom-forming cyanobacteria,
41 common today in heavily culturally impacted lakes, reached a historical maxima. The lake
42 ecosystem subsequently recovered somewhat but remains eutrophic to date. The erosion record
43 from Lake Væng shows a striking similarity with other Danish lake records, especially the notable
44 increase in Medieval Period catchment inputs, which are observed in other European lacustrine
45 records. Numerous European lowland lakes may have shifted into a degraded ecological state
46 millennia ago, but degradation intensified during the onset of the Medieval Period. Hence,
47 assuming pre-industrial conditions as relatively pristine reference baselines for more recent cultural
48 eutrophication could be flawed in landscapes intensively used by humans for millennia.

49
50 *Keywords:* Catchment-lake processes; macrofossils; pigments; inorganic phosphorus; soil erosion;
51 Holocene

53 **1. Introduction**

54
55 Humans have altered their surrounding landscapes for sustenance since the Early Holocene
56 (Harris, 1996). As agriculture and the associated clearance of forests became more widespread,
57 surface soil erosion increased markedly. Globally, human-driven soil erosion became significant
58 around 4000 years ago (Jenny et al., 2019). While a wealth of palaeoecological evidence exists on
59 past climate, landscape disturbance and associated soil erosion, much less is known on early
60 human-induced changes in the structure and functioning of aquatic ecosystems. Lake ecological
61 status is, however, known to potentially be affected by increased soil erosion via the enhanced
62 external loading of sediment and nutrients, causing increased turbidity and eutrophication (Dubois
63 et al., 2018; Jenny et al., 2019 and references therein).

64 Proxy-based climate reconstructions for Northwest Europe invariably report a pattern of
65 increasing Early Holocene temperatures, a Mid-Holocene thermal maximum and decreasing
66 temperatures during the last 4000 years (Brown et al., 2012; Mauri et al., 2015). Seasonal
67 differences in humidity were more pronounced during the Mid-Holocene with slightly depressed
68 summer rain and marginally higher winter precipitation (Mauri et al., 2015). This humidity pattern
69 is corroborated by a raised bog record from Svanemose, in southern Jutland (Denmark), which
70 shows more dry conditions between 4000 and 2500 BP, increasing humidity towards a peak at

71 1400-800 BP followed by a drier period since then (Barber et al., 2004). Such climate change
72 patterns could potentially have resulted in variable nutrient loading to lake basins through effects on
73 surface erosion. In the central European area, however, 6000–7000 years of farming have resulted
74 in massive and increasing disturbance of land cover, hydrology, erosion regimes and even climate
75 itself. These disturbances have generally over-ridden the more subtle effects of climate change on
76 lake ecological status (Smith et al., 2016; Wang et al., 2017; Zanon et al., 2018).

77 The lowland lakes of northwest continental Europe located in such heavily impacted
78 cultural landscapes have been subject to increased anthropogenic disturbance since at least 6000 yr
79 BP (e.g., Kalis et al., 2003; Boyle et al., 2015). Presently, it is estimated that *c.* 60% of European
80 waterbodies are disturbed beyond good ecological status (EEA, 2018). Hence, a wealth of studies
81 has investigated eutrophication, and how lake system structure and functioning have responded to
82 intensive modern agricultural practices and urbanisation in recent centuries (e.g., Lotter, 1998;
83 Langdon et al., 2006; Bennion and Simpson, 2011; Anderson et al., 2014). Far fewer, however,
84 have focused on the impact of early land use on lakes over millennia (Fritz, 1989; Bradshaw et al.,
85 2005). Studies of early human impacts can help to resolve debates about how, where and whether
86 humans significantly changed their environment (Ruddiman, 2013; Lewis and Maslin, 2015;
87 ArchaeoGLOBE Projec; 2019). Multi-evidence (multi-proxy) palaeoecological investigations have
88 the unique potential to provide insights into past landscape and lake disturbance and help
89 understand the process of lake eutrophication.

90 Studies of Holocene environmental changes in Denmark over more than 150 years have
91 revealed how humans have cleared the forests of Denmark, starting in the Early Neolithic (*c.* 5900–
92 5200 cal. yr BP; Iversen, 1973). A number of Danish lakes have exceptionally high sedimentation
93 rates (e.g., Rasmussen and Anderson, 2005; Bradshaw et al., 2005), allowing for highly temporal
94 resolution analysis of lake system changes in the past. Lake Væng, the site of this study, has
95 accumulated almost 18 m of sediment during the Holocene.

96 A recent extensive review of first human impacts on aquatic systems stated that, while
97 human impacts at the landscape level are relatively easy to detect, the effects of human disturbance
98 on aquatic ecosystems need further investigation. While case studies exist, they are often based on
99 one aquatic proxy (e.g., diatom assemblages) and hence provide a limited view of lake ecosystem
100 change (Dubois et al., 2017 and references therein). This study investigates the entire Holocene
101 development of Lake Væng, to encompass variability in the pre-disturbance ‘natural’ state. We
102 reconstruct erosion intensity and lake response at a high temporal resolution, to detect and assess

103 changes in nutrient dynamics, and algae, cyanobacteria, invertebrate and fish communities. In doing
104 so, we aim to answer the following questions: 1) When was the lake first impacted by human
105 activity? 2) Which time periods witnessed the most intensive impacts, and what was the ecosystem
106 response? 3) How do these changed ecosystem states compare to modern lake conditions? 4) How
107 quickly did the lake respond to external sediment and nutrient loading and was the response
108 proportional to the external forcing? By exploring these questions, we gain new insights into the
109 impacts of early human-induced landscape changes on aquatic ecosystems.

110

111 **2. Study site**

112 Lake Væng is located in eastern Jutland, east of the most extended ice marginal position of
113 the Scandinavian Ice Sheet during the Last Glacial Maximum (Fig. 1). The region was deglaciated
114 about 18 000–19 000 years ago according to cosmogenic surface exposure dating of large erratic
115 boulders (Houmark-Nielsen et al., 2012). The lake has a surface area of 15.7 ha, a mean and
116 maximum water depth of 1.2 m and 1.9 m and a water table at 26.3 m above sea level. The shallow
117 lake is located in a northwest–southeast orientated valley with peat deposits in an area dominated by
118 glaciofluvial sandy deposits and Miocene sandy and coal-bearing fluvial and deltaic deposits
119 (Kidmose et al., 2013). The area close to the lake is mainly forested, but the forests are only some
120 decades old and mainly consist of coniferous plantations. Studies of historical maps show that
121 agricultural land dominated the area a hundred years ago (Fig. 1; <https://historiskatlas.dk/>). The
122 topographical catchment is 9.8 km² (Kidmose et al., 2013).

123 Three small inlets contribute water to the lake, but groundwater comprises 74% of the
124 water entering the lake (Kazmierczak et al., 2016). An outlet is also present at the south end of the
125 lake. The climate of the region is temperate, with a mean July air temperature of *c.* 15.4 °C and mean
126 annual precipitation of *c.* 780 mm during the period from 1961 to 1990. The lake's residence time is
127 relatively short, about three weeks (Kidmose et al., 2013).

128 Lake Væng is a eutrophic and turbid lake, where groundwater-controlled input of dissolved
129 inorganic phosphorus (DIP) is one of the factors maintaining eutrophic conditions (Kidmose et al.,
130 2013; Kazmierczak et al., 2020). Previous studies (Kidmose et al., 2013; Kazmierczak et al., 2020)
131 have indicated that the origin of DIP discharged via groundwater is geogenic. Total phosphorus
132 (TP) concentrations in the lake water varied between 36 and 196 µg L⁻¹ during the period of 1985–
133 2015 (Meijer et al., 1994; Søndergaard et al., 1990, 2017). The mean summer concentration of TP
134 in the water was reduced from about 120 µg L⁻¹ in 1986 to 60 µg L⁻¹ in 1989 and to 50 µg L⁻¹ in

135 2011 as a result of bio-manipulations involving removal of zooplanktivorous and benthivorous fish
136 (Søndergaard et al., 1990, 2017; Jeppesen et al., 2012). The secchi depth increased from *c.* 60 cm in
137 1986 to *c.* 170 cm in 2011 and submerged macrophytes spread over the lake bottom. These
138 observed improvements in water quality, however, lasted only a few years (Meijer et al., 1994;
139 Søndergaard et al., 1990, 2017). The geographical coordinates at the coring site are 56°2.17'N,
140 9°39.32'E and the water depth was 175 cm.

141

142 **3. Methods**

143 *3.1. Field work*

144 Coring was carried out from a Uwitec coring platform. We used a Kajak corer (Renberg,
145 1991) to sample the surface sediment (down to 40 cm), a Russian peat corer (Jowsey, 1966) to
146 sample the upper sediments and a Usinger piston corer (Mingram et al., 2007) to sample the deeper
147 parts of the sediment sequence in the lakes. The Russian corer had a chamber length of 1 m and a
148 diameter of 7.5 cm. We used 2 m long 8 cm diameter coring tubes for the Usinger corer. Two
149 parallel Usinger cores, labelled A and B were retrieved. In the deepest parts of the coring, the corer
150 was hammered down using a percussion hammer. The Kajak samples were extruded and
151 subsampled in the field whereas the Russian cores were transferred to half-tubes, packed and
152 brought to the laboratory. The Usinger cores were extruded, cut into 1 m long sections, transferred
153 to plastic tubes, packed and brought to the laboratory where the cores were split length wise.

154

155 *3.2. ITRAX scanning*

156 We analysed the cores using an ITRAX X-ray fluorescence (XRF) core scanner to measure
157 elemental variations (Croudace et al., 2006). A Rhodium (Rh) tube with a resolution of 5 mm
158 provided XRF scans. The voltage and current were set to 30 kV and 50 mA, respectively, with an
159 XRF count time of 60 seconds. X-ray line-scan images were obtained in 1 mm resolution using a
160 tube setting of 60 kV and 30 mA and 1000 ms exposure time. The ITRAX core scanner is equipped
161 with an optical camera for high resolution digital line-scan images, as well as a Bartington magnetic
162 susceptibility detector used in 4 mm resolution. The XRF data were used for core correlations, as a
163 record of soil erosion (titanium, Ti) and to trace sediment hypoxia/anoxia (iron/manganese ratio,
164 Fe/Mn).

165 Core correlations were performed using a Matlab program developed by Rasmussen
166 (2006). Stratigraphic markers were found in the records of Fe, Zirkonium (Zr), radiography grey

167 scale, magnetic susceptibility and the coherent/incoherent scatter ratio. The Usinger A and B cores
168 were correlated using 123 correlation points (Supplementary).

169

170 *3.3. Inorganic phosphorus and loss-on-ignition*

171 We determined inorganic phosphorus by extraction in 1 M HCl (16 h, 1:50 solid to
172 solution ratio) with detection by molybdate colorimetry (Murphy and Riley, 1962) following
173 neutralization with dilute NaOH. The procedure primarily extracts phosphate associated with
174 calcium (i.e. apatite or secondary carbonates), as well as soluble phosphate and phosphate sorbed to
175 secondary metal oxides.

176 Loss-on-ignition (LOI) was determined by quantifying mass loss following combustion at
177 550°C for 4 hours. We first dried samples at 105°C for 24 hours to remove water. Values for all
178 procedures are expressed on an oven-dry (105°C × 24 h) mass basis.

179

180 *3.4. Macrofossils and radiocarbon dating*

181 ITRAX scanning of the cores was followed by visual sediment descriptions and
182 subsampling of the cores. Contiguous samples were taken for macrofossil analyses, except for the
183 bottom part of the Lake Væng succession. We could not extrude the bottom part because the
184 sediments were too stiff. Hence, we had to dig out samples from the core tube. The samples were
185 wet sieved on 0.4 and 0.2 mm sieves and the residue left on the sieves transferred to a petri dish and
186 analysed using a Leica Wild M3C dissecting microscope. A total number of 414 samples was
187 analysed. Selected remains of terrestrial plants from 16 levels were submitted for accelerator mass
188 spectrometry radiocarbon dating at the Aarhus AMS ¹⁴C Centre (Table S1). The widespread
189 occurrence of carbonate-rich glacial deposits in Denmark results in large hard-water effects, and
190 hence selecting remains of terrestrial plants for dating is important. Calibration to calendar years is
191 according to the INTCAL20 data (Reimer et al., 2020). The age-depth model was generated using
192 OxCal version 4.3 (Bronk Ramsey, 2013).

193

194 *3.5. Chlorophyll and carotenoid pigments*

195 Pigments were quantitatively extracted in an acetone: methanol: water (80:15:5) mixture.
196 Extracts were left overnight at -10 °C, filtered with a PTFE 0.2 µm filter and dried under nitrogen
197 gas. A known quantity was re-dissolved into an injection solution of a 70:25:5 mixture of acetone:
198 ion-pairing reagent (IPR; 0.75 g of tetra butyl ammonium acetate and 7.7 g of ammonium acetate in
199 100 mL water): methanol and injected into the HPLC unit. Separations used an Agilent 1200 series

200 module with quaternary pump. The mobile phase comprised solvent A (80:20 methanol: 0.5 M
201 ammonium acetate), solvent B (9:1 acetonitrile: water) and solvent C (ethyl acetate) with the
202 stationary phase consisting of a Thermo Scientific ODS Hypersil column (205 × 4.6 mm; 5 μm
203 particle size). The separation conditions started with 100% solvent A and ramped to 100% solvent
204 B over 4 minutes, followed by a linear increase towards 75% solvent C for 34 minutes, isocratic
205 hold at 25% B and 75% C for 1 minute, and a return to initial conditions of 100% solvent A over 4
206 minutes (a modification of Chen et al., 2001). Eluted pigments passed through a photo-diode array
207 detector and ultraviolet-visible spectral characteristics were scanned between 350 and 750 nm.
208 Quantification was based on scanning peak areas at 435 nm and calibrating to a set of commercial
209 standards (DHI Denmark). Pigment concentrations are reported as molecular weights of pigments in
210 nanomoles per unit organic matter of sediments.

211

212 *3.5.1. Pigment affinities*

213 Produced by algae, phototrophic bacteria and higher plants, chlorophyll and carotenoid
214 pigments are useful to infer past phototrophic production and shifts in primary producer
215 communities, providing essential information on lake trophic state. Pigments detected in the
216 sediments from Lake Væng include fucoxanthin and diatoxanthin (from siliceous algae),
217 alloxanthin (cryptophytes), chlorophyll *b* and lutein (chlorophyte algae and also higher plants
218 including macrolimnophytes), aphanizophyll (filamentous cyanobacteria), zeaxanthin,
219 canthaxanthin, echinenone (all cyanobacteria), chlorophyll *a* and β-carotene (all algae and higher
220 plants), okenone (Chromatiales; the obligate anoxic microbes purple sulphur bacteria) and the
221 ultraviolet radiation (UVR)-absorbing compound identified in Leavitt et al. (1997), which is often
222 produced by cyanobacteria as a photo-protectant. An index of water clarity (the UVR index) was
223 created by determining the amount of UVR pigment produced relative to quantity of algal pigments.
224 Thus, the UVR index is estimated as UVR pigment divided by the sum of the most abundant
225 pigments (alloxanthin + diatoxanthin + lutein-zeaxanthin) and multiplied by 100 (see Leavitt et al.,
226 1997 for further details).

227

228 *3.6. Numerical analyses*

229 The continuous and semi-quantitative pigment concentration profiles were divided into
230 zones using constrained optimal sum of squares partitioning available in the software package Zone
231 v. 1.2 (Juggins 1991). We estimated the number of significant zones through comparison with a

232 broken-stick model as described by Bennett (1996). All proxies were subsequently presented and
233 discussed using this pigment-based division of the core data.

234 Pigment concentration assemblages were summarized using detrended
235 correspondence analysis (DCA by segments, no transformation) to reflect overall changes in the
236 major groups of algal primary producers. Such procedure helps to understand the ecological
237 trajectory over time. While pigment gradient length was only 1.03 standard deviations, we chose
238 DCA over PCA because it resulted in better separation in the sample plot. Remains of
239 macrolimnophytes were too rare in the Lake Væng record to allow numerical analysis. We
240 quantified ecological (pigment) change since pristine (pre-agricultural) conditions by first selecting
241 the samples from 7000 to 6000 cal. yrs BP and then calculating the mean pigment assemblage of
242 these 39 samples. Samples earlier than 7000 BP were left out in this calculation because catchment
243 ecosystem adjustments following the Early Holocene warming were very slow (e.g., Giesecke et al.,
244 2017), and such slow changes may have had cascading effects on lake status. In contrast, the period
245 from 7000 to 6000 BP has been identified as ecologically less variable in terms of regional
246 catchment vegetation (Odgaard, 1994). We compared all younger samples to this pre-agricultural
247 baseline assemblage using chord distance as a dissimilarity measure. To do this, we converted the
248 pigment data to percentage of all pigments (excluding the UV-index). Following linear detrending
249 against time, the correlation of chord distance values with detrended Ti counts (a proxy for
250 erosional input) was calculated for the period 6000-0 BP. All ordinations were performed using
251 Canoco 4.5. Stratigraphic diagrams were plotted using the program C2 (Juggins, 2005) and edited
252 in Adobe Illustrator.

253

254 **4. Results and interpretations of proxy records**

255 *4.1. Sediments, ages and sedimentation rates*

256 We recovered a 1780 cm long sediment succession from Lake Væng. The succession
257 comprised glaciofluvial sand, a thin layer of clay, peat, carbonate-rich gyttja and fine-grained detritus
258 gyttja (Fig. 2). The peat was ¹⁴C-dated to *c.* 10 930 cal. yrs BP (Table S1). Gyttja started to accumulate
259 about 9 400 cal. yrs BP, and the sedimentation rate during gyttja deposition was fairly constant at about
260 1.7 mm/year (Fig. 2). One sample, AAR-24965 (Table S1; Fig. 2) from a depth of 1078–1082 cm is
261 considered an outlier, because it is younger than expected. Given the depth of the Early Holocene
262 basin, and the lack of topographic gradient in the lake catchment, the lake likely formed as a result
263 of the melting of buried stagnant ice (kettle hole lake). The fairly constant sedimentation rates at the

264 central coring point may indicate sediment focusing in the centre of the lake during the earlier part
265 of the record (steeper slopes), and an overall increase in lake-wide sedimentation, as the area of
266 accumulation tends to increase over time in lake basins (Davis et al., 1984).

267

268 4.2. *The titanium record*

269 The titanium (Ti) record from Lake Væng is interpreted as an indicator for catchment
270 erosion, as Ti has been shown as a useful proxy of clastic input and resulting nutrient flows to lakes
271 (Kylander et al., 2011; Boyle et al., 2015; Davies et al., 2015). The record shows a low and fairly
272 constant level until c. 5000 cal. yrs BP. Somewhat elevated contents of Ti are observed between c.
273 5000 and 3000 cal. yrs BP. Within this period, several small peaks are identified. From c. 3000 to c.
274 1250 cal. yrs. BP the average Ti level is similar to the previous period, but variability is much
275 higher with pronounced peaks around c. 2700, 2300 and 1600 cal. yrs BP. (Figs 3-5). A distinct low
276 level is seen around c. 1250 cal. yrs BP, which is followed by an abrupt and large increase peaking
277 at c. 1000 cal. yrs BP, indicating a marked increase in catchment erosion. These high Ti contents
278 are maintained until c. 500 years ago, when sediment Ti began to decline towards present-day
279 values.

280

281 4.3. *Lake and catchment biota*

282 Figures 3-5 present detailed results of the macrofossil and pigment assemblage analyses,
283 alongside results of the inorganic phosphorus (P) and LOI analyses, and the sediment Fe/Mn ratio.
284 These proxy diagrams were divided into six zones (see 3.6. for methodology).

285

286 4.3.1. *Zone 1 (c. 11 000 – 7370 cal. yrs BP)*

287 *Catchment vegetation*

288 The five lowest peat samples are characterised by common wood fragments and remains of
289 *Betula sect. Albae* (tree birch), *Pinus sylvestris*, *Populus tremula*, *Carex* sp. and *Phragmites* (rare).
290 The peat is characterized by many wood fragments, suggesting that the site was probably
291 overgrown by open woodland vegetation. The presence of *Carex* and *Phragmites* shows that the
292 ground was waterlogged. *Betula sect. Albae* and *Pinus sylvestris* dominated the Early Holocene
293 woodlands in Denmark (Iversen, 1973). *Betula sect. Albae* spread at the onset of the Holocene,
294 whereas *Pinus* arrived somewhat later and in western Jutland it spread at about 10 500 cal. yrs BP
295 (Odgaard, 1994).

296 The three upper samples of peat consisted almost entirely of stems of the bryophyte
297 *Tomentypnum nitens*, which is well known as a peat forming species of alkaline mires (Dickson,
298 1973). In addition to *T. nitens*, remains of the bryophytes *Fontinalis antipyretica* and *Sphagnum* sp.
299 were also recorded, as well as remains of the trees *Betula* sect. *Albae*, *Pinus sylvestris* and *Populus*
300 *tremula*, which were common in the region in the Early Holocene (Odgaard, 1994). The moss peat
301 indicates that the site became wetter, perhaps due to melting of buried stagnant ice below the peat.
302 The sediment probably accumulated in a small moss-rich mire surrounded by open woodland.

303 Since core material is not available to show the transition from peat to lake gyttja, we do
304 not know if the transition from mire to lake was abrupt or gradual. The common presence of *Betula*
305 sect. *Albae* and *Pinus sylvestris* may indicate that the coring site was close to the palaeo-shore, and
306 that the lake at that time was relatively small. The first record of *Alnus glutinosa* is at 9200 cal. yrs
307 BP, which is in accordance with studies of other Danish locations (Aaby, 1993; Odgaard, 1994),
308 although some marine sites show older ages of macrofossils [Storebælt c. 9500 cal. yrs BP (Bennike
309 et al., 2004), Aarhus Bay c. 9800 cal. years BP (Rasmussen et al., 2020)]. *Sphagnum* sp. remains are
310 abundant between 9000 and 8500 cal. yrs BP, after which they nearly disappear from the record and
311 reappear at greater numbers again in Zones 5 and 6. These remains indicate that swampy conditions
312 were found along the shores of the lake.

313 *Pinus* disappears from the record around 8000 cal. yrs BP and *Betula* becomes rare. The
314 tree macrofossil record is thereafter characterised by *Alnus glutinosa*. The decline in *Pinus* and
315 *Betula* probably reflects that the forests around the lake were replaced by closed temperate
316 deciduous forests. Regional pollen diagrams show that the forests were dense and dominated by
317 broad-leaved deciduous trees such as *Corylus*, *Tilia*, *Quercus* and *Ulmus* (Iversen, 1973; Odgaard,
318 1994). The general lack of macrofossils of these species is a common feature of macrofossil studies
319 of lake sediments (Watts, 1978).

320

321 *Lake macrophyte, invertebrate and fish communities*

322 Remains of lake macrophytes are rare and largely confined to *Najas marina*, *Nymphaea*
323 *alba* and *Nuphar* sp. The scarcity of submerged macrophytes, particularly before 8000 cal. yrs BP,
324 could be due to brownification, limiting light availability at the lake floor. This interpretation is in
325 line with the low values of the pigment-based UVR-index (Fig. 5; low values indicate
326 turbid/coloured water, high values clear water) and inferred swampy conditions along the shores of
327 the lake.

328 The presence of the thermophilous plants *Najas marina* and *Cladium mariscus* in the early
329 part of the record indicates that summer temperatures around 9000 – 8000 cal. yrs BP were at least
330 similar to present temperatures. *Najas marina* is a submerged macrophyte that grows in brackish
331 water or carbonate-rich fresh water with high conductivity (Bennike et al., 2001), whereas *Cladium*
332 *mariscus* is a calciphilous reed plant that was common and widespread in eastern Denmark in the
333 Early Holocene. While *Najas marina* survived in many Danish lakes until recently, *Cladium*
334 *mariscus* declined due to leaching of the calcareous soils throughout the Holocene and due to
335 decreasing temperatures after the Holocene thermal maximum. Only a few studies have previously
336 found *Cladium mariscus* macrofossils from Holocene deposits in Jutland (Jonassen, 1950).

337 The most abundant remains of fresh water animals include ephippia of *Daphnia pulex* and
338 statoblasts of the bryozoan *Cristatella mucedo*. Remains of ostracods and gastropods are rare. The
339 high abundance of *Daphnia pulex* ephippia indicates that the predation pressure by fish was limited
340 before c. 9000 cal. yrs BP (Davidson et al., 2010). *Perca fluviatilis* (perch) remains are found
341 throughout the zone, but their abundance is low compared to following zones.

342

343 *Algal communities, nutrients and primary production*

344 The pigment data (β -carotene) indicate relatively low algal production throughout the first
345 c. 5000 years of Lake Væng. The loss-on-ignition (LOI) record representing the organic matter
346 content of the sediments agrees with this observation, as values are clearly lowest in this zone
347 (starting below 10% and stabilising at c. 20% around 8000 cal. yrs BP). The inorganic P profile
348 largely mirrors LOI, showing lowest values prior to Zone 4 at the bottom of the core (c. 3.6 mg/g).
349 As a large part of the HCl-extracted inorganic P is in the form of calcium phosphates (see methods),
350 the slowly increasing trend in Zone 1 (and into Zone 2) is likely due to natural leaching of Ca from
351 the catchment soils.

352 All primary producer groups found in the sediment record are represented in this zone (Fig.
353 5), apart from purple sulphur bacteria (okenone). Especially cyanobacteria, however, but also
354 siliceous algae occur at clearly lower concentrations compared to the following zones. The absence
355 of the UVR-screening pigment in Zone 1, the scarcity of macrolimnophytes, and evidence of
356 *Sphagnum*-rich vegetation surrounding the basin indicate that the water was stained brown by
357 humic substances. Such colouring may have limited primary production through light limitation, as
358 is common in dystrophic waters. Around c. 8000 cal. yrs BP, low levels of the UVR index are
359 observed. These low levels could indicate improving water clarity in the lake in concert with the

360 decline in swampy *Sphagnum*-rich vegetation, as algae produce sunscreen compounds when
361 exposed to UVR light (Leavitt et al., 1997). Water level fluctuations can also account for them,
362 when marginal habitats (which have higher UVR exposure) are colonized by cyanobacteria, which
363 produce UVR-protective compounds as protection from periodic UVR exposure (Cantonati et al.,
364 2014).

365

366 4.3.2. Zone 2 (7370–6050 cal. yrs BP)

367

368 *Catchment vegetation*

369 *Alnus glutinosa* and *Populus tremula* characterize the tree macrofossil record in this zone.
370 The decline in *Pinus* and *Betula* in the previous zone likely reflects the spread of closed temperate
371 deciduous forests (as documented by local and regional pollen diagrams, e.g. Odgaard 1994), which
372 then dominated during this zone.

373

374 *Lake macrophyte, invertebrate and fish communities*

375 Macrophyte findings are also in this zone largely confined to *Najas marina*, *Nymphaea*
376 *alba* and *Nuphar* sp. The submerged *Najas marina* shows a continuous presence from c. 6500 cal.
377 yrs BP until ca. 4500 cal. yrs BP and *Chara* sp., a genus consisting of likewise submerged water
378 plant species appears in the record at the same time. *Perca fluviatilis* remains are most abundant in
379 the whole sediment record during the Holocene thermal maximum (between 8000 and 5000 cal. yrs
380 BP), when the lake was at its clearest based on the UVR-index (Fig. 5).

381

382 *Algal communities, nutrients and primary production*

383 We observe increased concentrations of pigments belonging to siliceous algae (diatoms
384 and chrysophytes), cryptophytes, chlorophytes and cyanobacteria. These concentrations indicate
385 increased production in these groups as light availability increased, while the overall algal
386 production was still relatively low (β -carotene). Within the whole sediment record, the UVR index
387 is highest in this zone and in the beginning of Zone 3. The steady increase of inorganic P stabilises
388 around 7.5 mg/g in the end of this zone, until ca. 4500 cal. yrs BP.

389

390 4.3.3. Zone 3 (6050–2700 cal. yrs BP)

391 *Catchment vegetation*

392 The land plant record is characterised by the dominance of *Alnus glutinosa*, which
393 probably grew along the margin of the lake and in areas with wet soil. *Betula* sect. *Albae* remains
394 are rare but occur through the zone, whereas *Populus tremula* remains disappear around 3700 cal.
395 yrs BP. Other woody plants are represented by a few remains of *Sorbus aucuparia* and *Prunus*
396 *padus*. Remains of wetland plants are rare. Regional pollen diagrams show some disturbance,
397 especially by animal husbandry, meaning that the forests during this period became more open and
398 *Betula* and *Corylus* increased, whereas *Tilia*, *Fraxinus* and *Ulmus* decreased (Iversen, 1973;
399 Odgaard, 1994).

400

401 *Lake macrophyte, invertebrate and fish communities*

402 Water plants are only present in low numbers. Remains of *Najas marina* show a
403 continuous record and highest abundances during the first half of the zone, simultaneously with the
404 only near-continuous occurrence of *Chara* sp., before disappearing entirely from the sediment
405 record around 4500 cal. yrs BP. The most abundant water animal remains in this zone are
406 statoblasts of the bryozoan *Plumatella* sp. (mainly *Plumatella repens*, especially around 4000–3000
407 cal. yrs BP) and the ostracod *Cyclocypris laevis* (in the middle of the zone). Other ostracods are
408 rare. Freshwater snails (*Valvata* spp.) and freshwater bivalves (*Pisidium* sp.) disappear from the
409 sediment record by the middle of this zone (around 4000 cal. yrs BP). Egg cocoons of the fish leech
410 *Piscicola geometra* first appear in this zone and display a near-continuous record (albeit with low
411 abundances) to the top of the core. The fish fauna includes *Scardinius erythrophthalmus* (common
412 rudd), which first appears in this zone, and *Perca fluviatilis*, which shows a decline in the
413 abundance of remains compared to the previous zone.

414 The continuous presence of *Najas marina* in Lake Væng during the Mid-Holocene may
415 reflect peak summer temperatures during this time. It survived in many Danish lakes until recently,
416 however, and may have disappeared rather recently from Lake Væng because the water became
417 more eutrophic and turbid, as documented by Anderson and Odgaard (1994). The pigment data
418 (UVR-index) indicate greatest water clarity at the same time as the continuous *N. marina*
419 occurrence, which lends further support to the reasoning outlined above on its disappearance. While
420 favourable light conditions for submerged macrophytes may also occur during lower lake levels,
421 palaeohydrological data from Lake Bliden on Zealand, Denmark indicate that the Holocene thermal
422 optimum was wet, becoming drier towards its end around 5700 cal. yrs BP (Olsen et al., 2010).
423 Regional lake level reconstructions from southern Sweden indicate lower lake levels in the Early

424 Holocene and again around 5600–5400 cal. yrs BP (Gaillard and Digerfeldt, 1991). Thus, these
425 studies do not support lake level lowering during the presence of photo-protectant pigments and *N.*
426 *marina* in the Lake Væng record.

427 With infilling of the lake, the littoral zone expands and (submerged) macrophyte cover
428 would be expected to increase. In the Lake Væng record, however, we witness the opposite, i.e.
429 increasingly sparser macrophyte cover towards the present times, which would indicate increased
430 lake turbidity, in line with the decrease in the UVR-index (see below). The decline in *Perca*
431 *fluviatilis* and increase in *Scardinius erythrophthalmus* remains could indicate a change in the fish
432 community towards dominance of planktivorous cyprinid fish, commonly observed in eutrophied
433 lakes (Sandström and Karås, 2002). *Plumatella* species have likewise been found to benefit from
434 higher nutrient concentrations (Hartikainen et al., 2009).

435 Studies have found positive correlations between the frequency of fish leech (*Piscicola*)
436 and both submerged macrophyte and fish abundance (Odgaard and Rasmussen, 2001; Sayer et al.,
437 2016). Although rare throughout the core, macrophyte and fish remains are even rarer in the section
438 of the core (from *c.* 4500 cal. yrs BP to the top), when *Piscicola* is present. On the other hand,
439 *Scardinius* remains appear around the same time as *Piscicola*, possibly indicating an increase in
440 cyprinid fish.

441

442 *Algal communities, nutrients and primary production*

443 Several simultaneous changes are observable around 4500 cal. yrs BP. Concentrations of
444 β -carotene (reflecting total algal production) show a *c.* 4-fold increase compared to the very low
445 concentrations of the previous zones. Both LOI and inorganic P increase to *c.* 30% and *c.* 10 mg/g,
446 respectively. The UVR index decreases clearly, and we observe the initial change towards higher
447 Fe/Mn values. Concentrations of diatoxanthin, which is produced by diatoms, increase towards the
448 upper part of this zone. After *c.* 3500 cal. yrs BP, okenone from purple sulphur bacteria appears for
449 the first time.

450 The increases in pigment concentrations and LOI indicate that the lake became more
451 productive. The simultaneous increases in inorganic P and Ti (erosion) imply increased nutrient (P)
452 inputs from the catchment. The observed decrease in the UVR-index indicates lower water clarity,
453 which could be translated into increased turbidity due to higher algal production. By this time, the
454 lake had likely become shallower due to infilling (Fig. 2), but it was probably still stratified, at least
455 in the deeper parts of the lake. Infilling leads to expanded benthic and epiphytic areas for algal

456 colonisation, which could also have increased sedimentary pigment concentrations. The onset of the
457 presence of okenone in tandem with the first increase in the Fe/Mn ratio indicates development of
458 anoxic habitats in the lake, which may derive from the decomposition of aquatic vegetation or
459 planktonic seston. The decomposition leads to pockets of anoxic habitat in benthic areas, or (in
460 stratified lakes) hypolimnetic anoxia during lake stratification.

461

462 4.3.4. Zone 4 (2700–1270 cal. yrs BP)

463

464 *Catchment vegetation*

465 *Alnus glutinosa* continues to dominate the land plant record. *Betula* sect. *Albae* is rare and
466 nearly disappears at the transition between this zone and the uppermost two zones. Apart from one
467 finding of *Sorbus aucuparia*, no other woody plant remains are recorded and wetland plant remains
468 are very rare. Regional pollen studies from Jutland indicate that during the early part of this zone a
469 reduction in woodland caused by extensive domestic grazing was prevalent. In the late part of the
470 period (from about 1800 BP), a general relaxation of grazing pressure resulted in a reforestation
471 during which *Fagus* spread in the forests (Iversen, 1973; Odgaard, 1994; Rasmussen, 2005).

472

473 *Lake macrophyte, invertebrate and fish communities*

474 Water plant remains are present in very low numbers. *Stratiotes aloides* appears for the
475 first time in this zone. This species is commonly found at sheltered sites in eutrophic lakes (Hämet-
476 Ahti et al., 1998). Its presence is noteworthy, because this plant is rarely recorded in Holocene lake
477 sediments from Denmark (Bennike and Hoek, 1999), but its leaf spines are common in subsurface
478 sediments of recently eutrophied shallow lakes (Odgaard and Rasmussen, 2001). *Stratiotes aloides*
479 has been reported to show allelopathic suppression of phytoplankton growth in laboratory
480 experiments (Mulderij et al., 2005), but any such effect is not detectable in our data (see algal
481 discussion below). The bryozoan *Fredericella* sp. first appears around 2000 cal. yrs BP and its
482 remains are relatively abundant from here to the top of the core. The bryozoan *Plumatella* sp.
483 (mainly *Plumatella repens*) is near absent in the first half of this zone, but the abundance of its
484 remains increases clearly at the same time with *Fredericella* before declining to rare occurrences at
485 the end of the zone. *Perca fluviatilis* remains continue to decline towards the top of the core and are
486 near absent in this zone. One occurrence of *Scardinius erythrophthalmus* is recorded.

487 The water plants and animals observed in this zone each indicate eutrophic conditions.
488 Bryozoan abundance (*Fredericella* and *Plumatella*) is known to increase with higher nutrient
489 concentrations, especially with increases in phosphorus (Hartikainen et al., 2009).

490

491 *Algal communities, nutrients and primary production*

492 The pigment record becomes more variable with oscillations in concentrations of most
493 pigments and a pronounced assemblage readjustment where pigments from cyanobacteria
494 (including filamentous cyanobacteria) become more dominant relative to siliceous, cryptophyte, and
495 chlorophyte algal pigments (Fig. 5). LOI increases clearly in this zone, with highest values around
496 40% while, similar to the pigments, variability in inorganic P increases to display the highest values
497 in the whole sediment record (1.94 mg/g) at c. 2100 cal. yrs BP, around the same time with a
498 marked increase in Fe/Mn. Okenone increases abruptly at c. 1700 cal. yrs BP. The increase in
499 pigment concentrations and LOI, and the shift towards cyanobacteria, which are strongly associated
500 with nutrient enrichment (Taranu et al., 2015) each indicate that the lake became more productive
501 compared to the previous zone, while lake bottom anoxia also increased (indicated by okenone and
502 Fe/Mn). Increases in bottom water anoxia are commonly associated with eutrophication in stratified
503 lakes as the supply and subsequent decomposition of organic matter to the lake bottom waters leads
504 to oxygen drawdown (Lami et al., 1994; Maheaux et al., 2016). Shallow lake sediments may also be
505 anoxic, however, despite overlying well-aerated water. Their well-oxidized sediment surface does
506 not preclude the potential release of mobile P to the overlying water column (Tammeorg et al.,
507 2020). Based on the inorganic P profile of Lake Væng, it was mobilised from the sediment into the
508 water column, further increasing eutrophication.

509

510 *4.3.5. Zone 5 (c. 1270–250 cal. yrs BP)*

511 *Catchment vegetation*

512 Land plant remains in this zone are dominated by *Sphagnum* sp. and *Juncus* sp. Remains of
513 woody plants are rare and confined to *Betula* sect. *Albae* and *Alnus glutinosa*, which occur in the
514 lower part of the zone. Regional pollen diagrams show that the landscape became progressively
515 more open due to expanding pastures and eventually also arable fields (Iversen, 1973; Odgaard,
516 1994; Rasmussen, 2005). The *Sphagnum* and *Juncus* remains indicate that boggy vegetation was
517 reappearing along the shores of the lake. Similar expansions of *Sphagnum* in the Late Holocene are

518 also seen at other sites in Jutland (Odgaard, 1994) and may indicate acidification of soils around the
519 lake.

520

521 *Lake macrophyte, invertebrate and fish communities*

522 Macrolimnophytes are very rare, represented by a few findings of the floating-leaved
523 *Nymphaea alba* and a single oospore of the submerged macrolimnophyte *Chara* sp. Water animals
524 are represented by cocoons of the leaches *Piscicola geometra* and *Erpobdella* sp., larval cases of the
525 caddis fly *Orthotrichia* sp. and statoblasts of the bryozoans *Cristatella mucedo*, *Plumatella* sp., and
526 *Fredericella* sp. The former are all rare, while *Fredericella* sp., which thrives in eutrophic
527 conditions, dominates this zone.

528

529 *Algal communities, nutrients and primary production*

530 Maximum pigment concentrations occur during this zone, with particularly elevated
531 concentrations of planktonic filamentous cyanobacteria pigments (despite the by now shallow lake
532 depth, see Fig. 2). LOI largely mirrors total algal production (β -carotene), showing highest (albeit
533 variable) values in the whole sediment record of up to *c.* 50%. Okenone concentrations are also
534 highest in the whole sediment record at the beginning of the zone, before declining and
535 disappearing entirely around 500 cal. yrs BP, largely mirrored by Fe/Mn. Maximum algal and
536 filamentous cyanobacterial abundance indicate very productive conditions. Inorganic P values
537 clearly decrease compared to the previous two zones, likely reflecting that a large proportion of the
538 inorganic P is mobilized into the water column (Anderson et al., 1993 and references therein).

539 Based on the proxy records, the time period from 1250 to *c.* 250 cal. yrs BP was the most
540 eutrophic (likely hypertrophic) period in the history of Lake Væng. Increases in okenone are
541 regularly associated with eutrophication in stratified lakes (Guilizzoni and Lami, 1992), but
542 okenone declined in Lake Væng during this highly productive phase. An increase in lake turbidity
543 from higher phytoplankton production might have reduced light penetration to the deeper waters,
544 eliminating the habitat for photosynthetic bacteria. Catchment deforestation, however, likely
545 increased wind fetch and lake mixing (Lotter, 2001; Romero et al., 2006) which, in combination
546 with progressive shallowing of the lake following sediment infilling, will have decreased stable
547 deep-water habitats for anaerobic phototrophs.

548

549 *4.3.6. Zone 6 (ca. 250 cal. yrs BP – present)*

550

551 *Catchment vegetation*

552 Apart from rare *Betula* sect. *Albae*, *Juncus* sp. and *Sphagnum* sp., which shows a steady
553 decline throughout the zone, no other land plant remains are found. At present, the area close to the
554 lake is forested, but according to historical records these are some decades old coniferous
555 plantations.

556

557 *Lake macrophyte, invertebrate and fish communities*

558 In this uppermost zone, the sediment record is devoid of remains of macrolimnophytes. All
559 water animal remains are rare, apart from *Fredericella* sp., and remains of both leaches and fish
560 disappear in the top of the core.

561

562 *Algal communities, nutrients and primary production*

563 During the last c. 250 years of our record, concentrations of β -carotene decrease markedly,
564 while concentrations of chlorophyll a show a slight increase. LOI follows the notable trend in β -
565 carotene and decreases from c. 50% to c. 25%. Among the different algal groups, cyanobacterial
566 pigments decline clearly, but siliceous algal, cryptophyte and chlorophyte pigments mostly increase
567 towards the top of the sediment record. At the same time the UVR index increases markedly. These
568 shifts in algal communities, decline in LOI, increase in the UVR-index, complete absence of
569 okenone and clear decrease in Fe/Mn indicate a moderate recovery towards a less-enriched lake
570 ecosystem and decreased turbidity of the shallow waters (light penetration to the lake bottom). This
571 is in line with the reduced catchment input represented by declining Ti-levels. However, there is no
572 indication of a recovery of submerged macrophyte vegetation.

573

574 *4.4. Numerical analyses*

575 The DCA of the pigment assemblages clearly shows a limited amount of change before c.
576 6000 cal. yrs BP, after which the differences among samples increased as the trajectory of
577 ecological change became more dynamic with increased catchment inputs (Fig. 6A). The
578 assemblages of Zone 5 deviate markedly from the rest of the core. Interestingly, the two modern
579 samples (last 50 years) show a clear trend back towards assemblages that are more similar to
580 pristine conditions.

581 The pigment chord distance results reflect a long period until about 2500 cal. yrs BP with
582 relatively little change compared to the baseline, albeit with some distinct spikes (Fig. 6B). This
583 initial change was followed by much more different assemblages from *c.* 1250 cal. yrs BP onwards
584 before finally returning to more similar biotic assemblages over the last century. A strong
585 correlation between Ti and pigment change (Fig. 6B, $R^2 = 0.37^{***}$ based on detrended values)
586 indicates that the changes in primary producers in the lake are directly correlated with inputs of
587 allochthonous minerals via processes such as soil erosion and shifts in land drainage.

588

589 **5. Discussion**

590

591 The lack of Late Glacial organic sediments and the presence of Early Holocene peat in the
592 bottom of the Lake Væng record show that the lake did not form until the Early Holocene, about
593 9400 cal. yrs BP. The lake probably formed by melting of a body of stagnant glacier ice and the
594 lack of Late Glacial and earliest Holocene sediments (11,700 to 9400 cal. yrs BP) indicates that
595 stagnant glacier ice persisted until several millennia into the Holocene. The presence of *Cladium*
596 *mariscus* in the record indicates summer temperatures similar to the present at *c.* 8500 cal. yrs BP.
597 The oldest dated Holocene finds of *Cladium* fruits from Denmark, however, gave an age of *c.* 11
598 000 cal. yrs BP (Bennike and Jensen, 2011), indicating rapid warming soon after the beginning of
599 the Holocene at 11 700 cal. yrs BP.

600 Iversen (1973) concluded that the mean July temperature was 2–3°C higher than at present
601 during the Holocene thermal maximum, in good accordance with more recent estimates based on
602 pollen data (Brown et al., 2011). The forests in the catchment of Lake Væng changed from pine
603 dominance until around 8000 cal. yrs BP to a closed temperate deciduous forests dominated by
604 broad-leaved species such as *Corylus*, *Tilia*, *Quercus* and *Ulmus* (Iversen, 1973; Odgaard, 1994),
605 which was gradually changed to secondary ecosystems by early agriculture during the Neolithic
606 (5900–3600 cal. yrs BP) (Iversen, 1973).

607

608 *The landscape*

609 Titanium is considered a reliable indicator of terrigenous catchment inputs (Kylander et al.,
610 2011; Boyle et al., 2015; Davies et al., 2015). Hence the Ti record from Lake Væng is interpreted as
611 a proxy of soil erosion, which was mainly controlled by the ratio of open to forested land.

612 Generally, lack of forest reduced the landscape resistance to erosion. In open landscapes, runoff

613 energy is also elevated. During prehistory, the degree of open land depended especially on the
614 intensity of grazing by domestic life stock and later by clearance for arable farming (Fig 7).
615 Terrestrial erosion tends to peak just after deforestations begin, and soils which were previously
616 protected by forests became exposed. In contrast, erosion relaxes to lower levels when the easily
617 eroded sediments gradually become depleted (Prosser and Williams, 1998). This means that
618 sedimentary Ti cannot be expected to be a direct proxy of landscape openness, but rather a
619 reflection of the availability of exposures of erodible catchment sediments. Erosion peaks can be
620 expected to reflect deforestation events while extended high erosion rates may represent
621 continuously declining forest cover or permanent crop cultivation. Lower erosion levels reflect
622 landscapes resistant to erosion, either because they are relatively densely forested, or because they
623 are mature in the sense that erodible sediments have already been flushed off.

624 The erosion record of Lake Væng is interpretable by reference to existing local pollen
625 records, such as from Ilsø (14 km east, Sørensen et al., 2017), Taastrup Sø (20 km northeast, Sørensen et al.,
626 2017) and Dallerup Sø (30 km southeast, Nielsen & Odgaard, 2010). When interpreting Ti as an
627 erosion proxy for land use history in Denmark (Fig. 7), it is clear that, before the introduction of
628 agriculture (the Mesolithic period until 5900 cal. yrs BP), erosion in the Lake Væng catchment was
629 low and fairly constant. This is in accordance with local pollen records, reflecting dense deciduous
630 woodlands in the area. The first clearly observable increase in erosion at Lake Væng is seen in the
631 Middle Neolithic B (Single Grave Culture, 4400–4800 cal. yrs BP), a period known from several
632 nearby pollen records to show extensive landscape opening in western and Central Jutland, caused
633 by animal husbandry (Odgaard, 1994; Sørensen et al., 2017). After a relaxation reflected in the pollen
634 records as reforestation, erosional inputs indicate that the intensity of landscape disturbance grew
635 again through the Late Neolithic and Early Bronze Age, a period of expansion of cultural grasslands
636 (Sørensen et al., 2017). During the Late Bronze Age and the Pre-Roman Iron Age, erosional inputs at
637 Lake Væng peak, which is in accordance with local palynological records showing a pronounced
638 transition to mostly very open pastoral landscapes during these periods (Odgaard and Rasmussen,
639 2000; Sørensen et al., 2017). Pollen records indicate that these open landscapes generally persisted in the
640 region through the Roman Iron Age (Nielsen and Odgaard, 2010), and the lower Ti counts here
641 probably reflect the maturity of the pastoral landscape. A conspicuous peak is observed at the
642 transition to the Germanic Iron Age, which otherwise shows low values in accordance with
643 palynological results indicating reforestation in many areas of Denmark as well as locally (Nielsen
644 and Odgaard, 2010; Sørensen et al., 2017). This peak is followed by an abrupt and very large increase in

645 catchment inputs at the transition to the Viking Age, which defined a period of strong deforestation
646 (Nielsen & Odgaard, 2010; Sørensen et al., 2017). While erosional inputs at Lake Væng decrease
647 somewhat in the Medieval Period, overall high levels of catchment disturbance persisted until
648 around 200 cal. yrs BP, reflecting a period with strong increases in arable farming well documented
649 by local pollen records. During the most recent century, partial afforestation of the area around
650 Lake Væng will have played a role in reducing landscape erosion. Overall, the very strong
651 correspondence between erosional peaks and local pollen records showing transitions to periods
652 with expansion of cultural grasslands or cropland evidences the strong importance of land use for
653 catchment erosion at Lake Væng. Since documented Mid-/Late Holocene climate changes in the
654 area are minor, possible direct effect of climatic variation on the erosion pattern of Lake Væng was
655 likely subtle in relation to the large impacts from land-use changes. We have not identified
656 exclusive climate signals in the erosional record.

657 Comparing the Lake Væng Ti-record with other records of catchment sediment yields from
658 Denmark reveals clear similarities (Fig. 7). While Lake Væng shows the earliest signs of catchment
659 erosion linked to human disturbance of the landscape during the Middle Neolithic, all records
660 indicate increased and prolonged deforestation events in the catchment during the Late Bronze Age
661 and/or Pre-Roman Iron Age (in some cases also evident during the Roman Iron Age). This increase
662 suggests that, overall, the early phase of agriculture in Denmark during the Early Neolithic had
663 relatively little effect on landscape openness and soil erosion, and more pronounced deforestation
664 only took place in the Bronze and Iron Ages. Areas with strong activity by the Single Grave
665 Culture, however, such as at Lake Væng, reflect elevated erosion already in the Middle Neolithic.
666 The marked minimum in erosion during the Germanic Iron Age is also a common feature, probably
667 related to reforestation caused by social unrest or economic depression and associated changes in
668 landscape management during the Migration Period.

669 The most conspicuous common feature in the records is the strong increase in soil erosion
670 at the beginning of and during the past millennium. This development reflects the large-scale
671 deforestation documented by pollen records and resulted in a landscape dominated by pastures and
672 agricultural fields (Odgaard and Rasmussen, 2000; Sørensen et al., 2017). These data further reveal
673 marked landscape disturbance continuing into modern times until a more recent decrease over the
674 last few hundred years (apart from Gudme Sø). A similar large increase in terrigenous inputs during
675 the Medieval Period has also been observed in other European lake records (Bork and Lang, 2003;
676 Boyle et al., 2015; Bajard et al., 2016). The timing of this marked landscape change coincides with

677 the introduction of the heavy mouldboard plough in Northern Europe. This plough could turn over
678 the heavy, fatty and moist clay soils of the region and plough manure into them, markedly
679 increasing yields and transforming European agriculture and economies (Andersen et al., 2016).
680 The decreasing trend in erosion of the most recent century may reflect two developments: 1)
681 increasing afforestation in Denmark resulting in a development in forest cover from 4% in AD 1805
682 (Fritzbøger, 1992) to 14.6 % in 2018 (Nord-Larsen et al., 2020) and 2) the reduction in livestock
683 during the same period (Dalgaard et al., 2009), which reduced grazing at lake shores and
684 accordingly led to denser marginal vegetation of streams and lakes with a more effective filtering of
685 surface overflow.

686 The exceptionally long continuous erosion record from Fuglsø includes early catchment
687 inputs that have a natural origin. It begins with a large Ti-peak related to the transition from
688 minerogenic lake sediments to gyttja. The following Early Holocene peaks are probably due to
689 short cooling events, such as the Erdalen Event 1 (Nesje, 2009) and the 8.2 event (Alley et al.,
690 1997) intermittently resulting in vegetation types less resistant to erosion.

691

692 *The lake*

693 The most striking change in the Lake Væng ecosystem is recorded by the chord distance
694 dissimilarity increase at about 1200 cal. yrs BP and also clearly evident from zones 5 and 6 samples
695 in the DCA plot (Fig. 6). This change occurs concurrent with the most marked increase in Ti counts
696 (Figs 6, 7). The lake most likely became hypertrophic, with high production of planktonic algae,
697 blooms of filamentous cyanobacteria, and sediment anoxia (Fig. 5). Blooms of filamentous
698 cyanobacteria are commonly associated with eutrophication, particularly increases in phosphorus,
699 which would have been supplied by both increased catchment nutrient inputs and internal loading
700 from anoxic sediments (Taranu et al., 2015; McGowan, 2016). This indicates that Viking
701 Age/Medieval land disturbances and farming practices led to major increases in nutrient loading to
702 the lake and increased the prevalence of these nuisance and potentially toxic taxa. The evidence
703 from this study signifies that the local introduction of the heavy mouldboard plough changed Lake
704 Væng far beyond previous ecosystem conditions, and also conditions prevailing in the 21st century.
705 At Dallund Sø, where one of the few existing detailed multiproxy studies on landscape-lake
706 linkages over the past millennia was conducted, diatom-inferred TP concentrations increased
707 markedly from *c.* 30 µg/L to > 200 µg/L at the beginning of the Medieval Period, resulting in large

708 changes in lake productivity in concert with enhanced catchment inputs (Fig. 7C; Bradshaw et al.,
709 2005).

710 Interestingly, Lake Væng began to recover after the period of intense soil erosion in the
711 Viking Age/Medieval Period towards prior pigment assemblages, and water clarity began to
712 increase (Fig. 5). This recovery indicates that the marked disturbance may have led to greater
713 degradation in water quality relative to more recent times, when Danish lakes are well-known for
714 having water quality problems (Jeppesen et al., 2000). Despite the increase in water clarity and
715 decreased catchment inputs, the submerged macrophyte vegetation has not recovered (Fig. 3).
716 Recent biomanipulation projects have had only temporary success in restoring submerged
717 vegetation (Søndergaard et al., 2017).

718 Notably, evidence also exists of prehistoric disturbances at Lake Væng as early as in the
719 Middle Neolithic *c.* 4500 years ago, which predates the oldest reported record of aquatic changes
720 caused by human activity in Denmark (Dubois et al., 2017). With the first increases in erosional
721 inputs to the lake, it became more turbid, as evidenced by the decline of the UVR-index, and
722 primary production increased from the background level of the previous almost 5000 years (first
723 clear increase in β -carotene). The increase in turbidity apparently affected the submerged
724 macrophyte vegetation adversely (seen as the decline of *Najas marina*), and possibly changed the
725 dominant fish towards cyprinid species (Figs 3, 4). Bryozoans indicating elevated nutrient levels
726 also become common from here on. From the beginning of the Late Bronze Age (*c.* 3000 cal. BP)
727 onwards, cyanobacterial production increased, including filamentous cyanobacteria, and the first
728 signs of anoxia appear (increase in okenone and Fe/Mn), which are likely associated with more
729 intense decomposition of organic matter (Maheux et al., 2016). In contrast, at Dallund Sø, clear
730 indications of ecosystem changes dated later towards the end of the Late Bronze Age (Bradshaw et
731 al., 2005).

732 In Denmark, intense landscape modifications inferred from sedimentary pollen started
733 during the mainly pastoral Single Grave culture (4800-4400 cal. yrs BP), but with a strong regional
734 bias towards landscapes to the west and north of the East Jutland stationary line of the last
735 glaciation. As an example, strong deforestations during this period caused by intense animal
736 husbandry are characteristic of northwestern areas and some parts of Central Jutland (Andersen,
737 1993; Odgaard, 1994). Landscape modification east of this line (this study; Bradshaw et al., 2005)
738 was apparently less intense in Denmark before the Late Bronze Age (Fig. 7), after which clearly
739 detectable influences on lake ecosystems are evident. With the extensive deforestation and new

740 agricultural practices in early historical time (such as the heavy mouldboard plow), these lakes
741 responded with a regime shift (Scheffer and Jeppesen, 2007). While lake-ecosystem response is
742 always dependent on regional and local factors including topography, soils, hydrology and climate,
743 it is plausible — based on existing evidence on modification of the European landscape since the
744 Neolithic (summarised in Dubois et al., 2017) – that a notable number of European lowland lakes
745 may have first shifted into a degraded ecological state already many centuries ago, especially about
746 1200 cal. yrs BP, and possibly followed a recovery pathway, before again responding to increasing
747 pressures that the modern society poses on lakes (Jenny et al., 2016; Haas et al., 2019). Hence,
748 considering pre-industrial conditions as relatively pristine (and hence suitable as reference levels for
749 recent eutrophication) is not necessarily a valid assumption in landscapes that have been intensively
750 used by humans over the past millennia. These findings corroborate the conclusion of Bradshaw et
751 al. (2006), who found high P concentrations in a number of Danish lakes as early as AD 1800.

752 Finally, it is noteworthy that cyanobacteria have been consistently present in Lake Væng
753 throughout the Holocene, albeit at mostly lower abundances. This may be connected with the lake's
754 supply of groundwater with high phosphorus concentrations. Such naturally elevated abundances of
755 cyanobacteria have been recorded from other lakes in moraine landscapes where groundwater
756 supply dominates (McGowan et al., 1999). Debate still exists about the relative roles of
757 groundwater supply and internal recycling of P in lakes (Kilinc and Moss, 2002; Nisbeth, 2019a,b).
758 In the case of Lake Væng, the amount of groundwater entering the lake and the groundwater-
759 delivered DIP have not changed significantly over recent time (Kazmierczak, 2020), and changes in
760 algal abundances are tightly coupled to erosional inputs (based on pigment chord distance analysis,
761 Fig. 6B). Hence the observed increases in cyanobacterial blooms and primary production are most
762 likely a result of catchment disturbance due to early human activity.

763

764 **6. Conclusions**

765 The results based on the multi-proxy study of the Lake Væng sediments show that the
766 record begins with peat, and the lake probably formed by melting of buried stagnant glacier ice in
767 the Early Holocene. Forests in the catchment of Lake Væng changed from pine and birch
768 dominance prior *c.* 8000 cal. yrs BP to closed temperate deciduous forests dominated by broad-
769 leaved species, which remained in the pristine landscape until the first human disturbance.

770 Following up the specific research questions in the Introduction (section 1.): 1) The first
771 signs of clearances of these primeval forests around the lake are seen in the Middle Neolithic at *c.*

772 4500 cal. yrs BP. Erosional inputs to the lake were relatively low, however, indicating that the
773 region was still dominated by dense deciduous forests. It is obvious from the archaeological record
774 that farming had started, but the local impact on the landscape was limited, presumably because the
775 human population was small (Gron and Rowley-Conwy, 2017). Despite the limited nature of the
776 impact, the lake still responded by becoming more turbid and nutrient rich, with clear negative
777 impacts on the submerged aquatic plants, possible changes in the dominance of fish species and
778 shifts towards more eutrophic invertebrate species. 2) Larger peaks in erosional inputs are seen in
779 the Late Bronze Age and Pre-Roman Iron Age, with clear lake responses including smaller blooms
780 of filamentous cyanobacteria and first signs of anoxia. A very strong erosional signal at about 1200
781 cal. yrs BP may be related to the introduction of the mouldboard plough. At this time the lake
782 responded with a regime shift. It became hypertrophic and very turbid, with high production of
783 potentially bloom-forming filamentous cyanobacteria, and bottom-water anoxia. 3) The human
784 impact in the Medieval Period led to especially severe degradation in water quality and a more
785 degraded ecosystem state compared to recent times. 4) Lake response is near-contemporaneous with
786 catchment inputs and the response is proportional, i.e. with greater sediment and nutrient loading
787 we observe progressively more marked deterioration of the lake ecosystem.

788 As several Danish and European lake records on catchment disturbances share similar
789 features to the record from Lake Væng, it is possible that a number of lowland European lakes
790 experienced poor water quality already millennia ago, and certainly at the time of extensive forest
791 clearances and new agricultural practices about 1200 cal. yrs BP.

792

793 **Appendix A. Supplementary data**

794 Supplementary material related to this article can be found in the online version, at xxx

795

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1116

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1118

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1124

1125 Figure captions

1126

1127 **Fig. 1. A.** Map of Denmark showing the location of Lake Væng (VS) discussed in this study. AB:
1128 Aarhus Bay, DS: Dallund Sø, LB: Lillebælt, SB: Storebælt. **B.** Google Earth satellite image of Lake
1129 Væng. The dot shows the core position. **C.** Historical map, ca. 100 years old.

1130

1131 **Fig. 2. A.** Simplified log for the Lake Væng sediment record. **B.** Age–depth model for the
1132 succession from Lake Væng. The depth of the lowermost sample is uncertain. One sample is
1133 considered an outlier.

1134

1135 **Fig. 3.** Simplified macrofossil concentration diagram for Lake Væng (terrestrial, wetland and limnic
1136 plants). Radiocarbon ages are shown on the left of the diagram. Zone boundaries (based on pigment
1137 data) are shown with horizontal dashed lines and the zones are numbered on the right of the
1138 diagram.

1139

1140 **Fig. 4.** Simplified macrofossil concentration diagram for Lake Væng (invertebrate and vertebrate
1141 remains). Radiocarbon ages are shown on the left of the diagram. Zone boundaries (based on
1142 pigment data) are shown with horizontal dashed lines and the zones are numbered on the right of
1143 the diagram.

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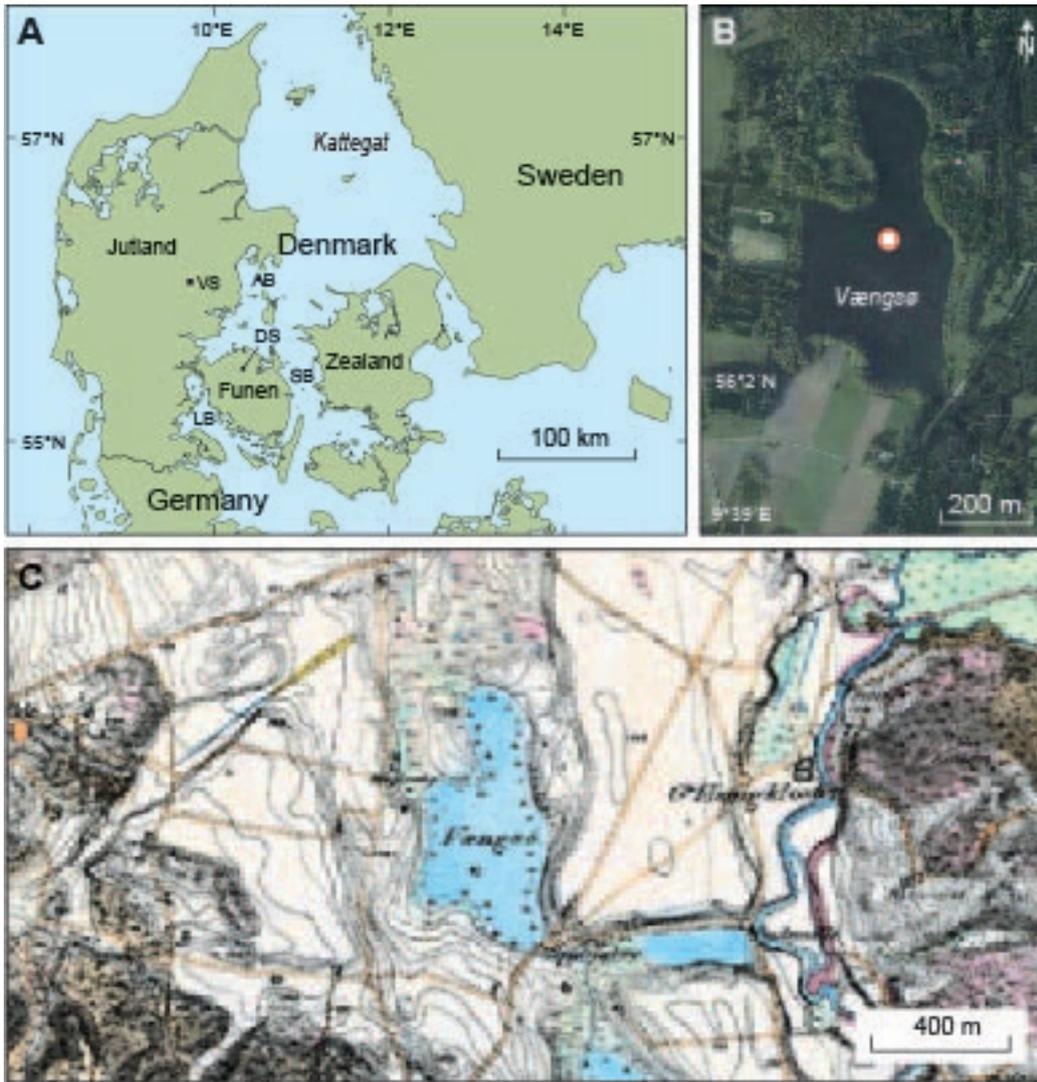
1145 **Fig. 5.** Sedimentary pigment stratigraphy, LOI, inorganic (HCl-extracted) phosphorus, Fe/Mn ratio
1146 and titanium records from the Lake Væng sediment core. Affinities of pigments with particular
1147 algal groups are indicated in the headers and calibrated radiocarbon ages are given on the left of the
1148 diagram. The UVR index indicates water clarity and is calculated as indicated in the methods
1149 section. Zone boundaries are shown with horizontal lines and zone numbers are shown on the right
1150 of the diagram.

1151

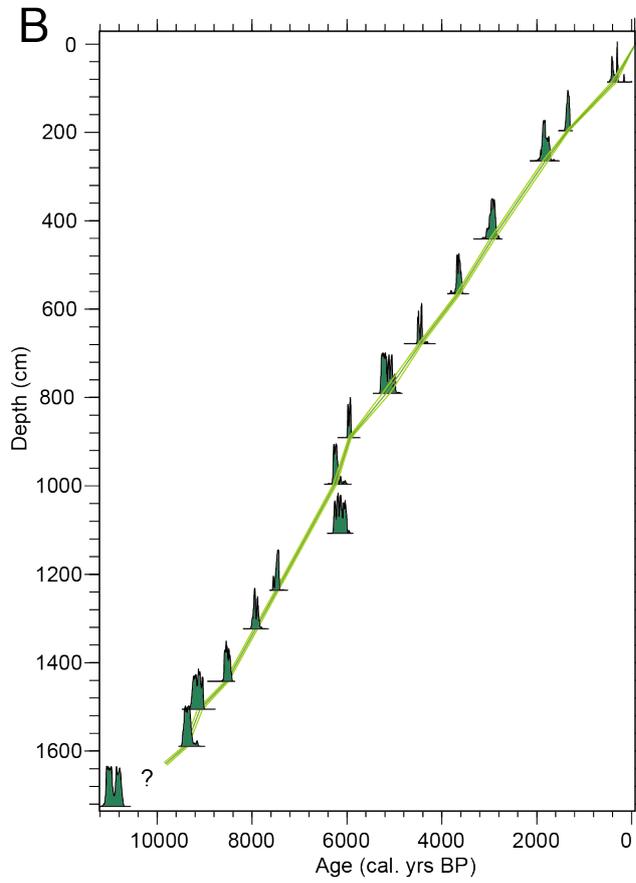
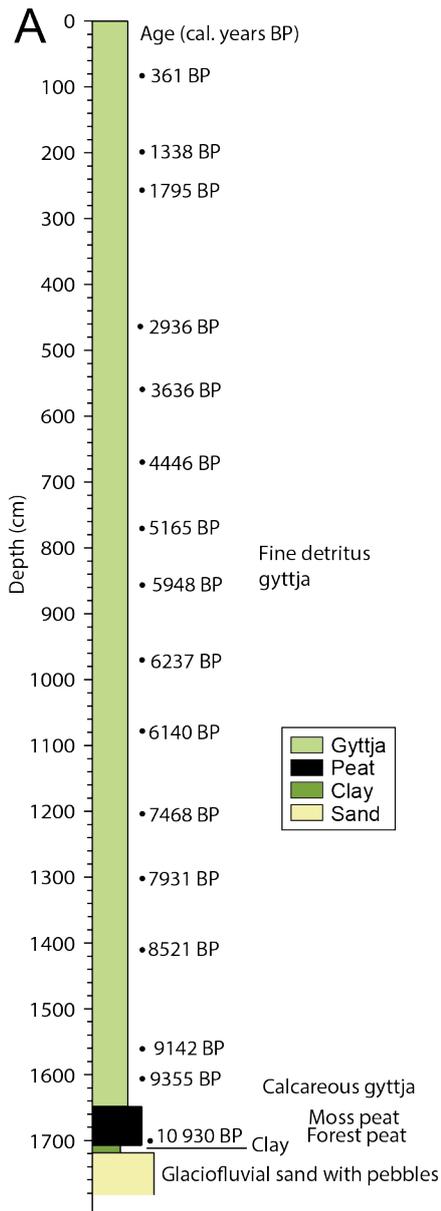
1152 **Fig. 6. A.** Detrended correspondence analysis (DCA) of pigment data with colour coding according
1153 to the zonation used in the paper. The ages of the two youngest samples are shown in years
1154 (Common Era). **B.** Ti counts (erosion proxy) and pigment chord distance values (perturbation
1155 proxy) for the last 6000 years plotted against age.

1156

1157 **Fig. 7.** Records of soil erosion from Lake Væng compared with other erosion records from Denmark.
1158 **A.** Lake Væng, this study; raw titanium counts per second), **B.** Fuglsø in eastern Jutland (Odgaard et
1159 al., 2017; normalized titanium counts), **C.** Dallund Sø on northern Funen (Rasmussen and Bradshaw,
1160 2005; minerogenic matter), **D.** Gudme Sø on south-eastern Funen (Rasmussen and Olsen, 2009;
1161 sediment accumulation rate of minerogenic matter), **E.** Ilsø in eastern Jutland (Søe et al., 2017;
1162 titanium counts per second). The inset map shows the location of the sites.
1163



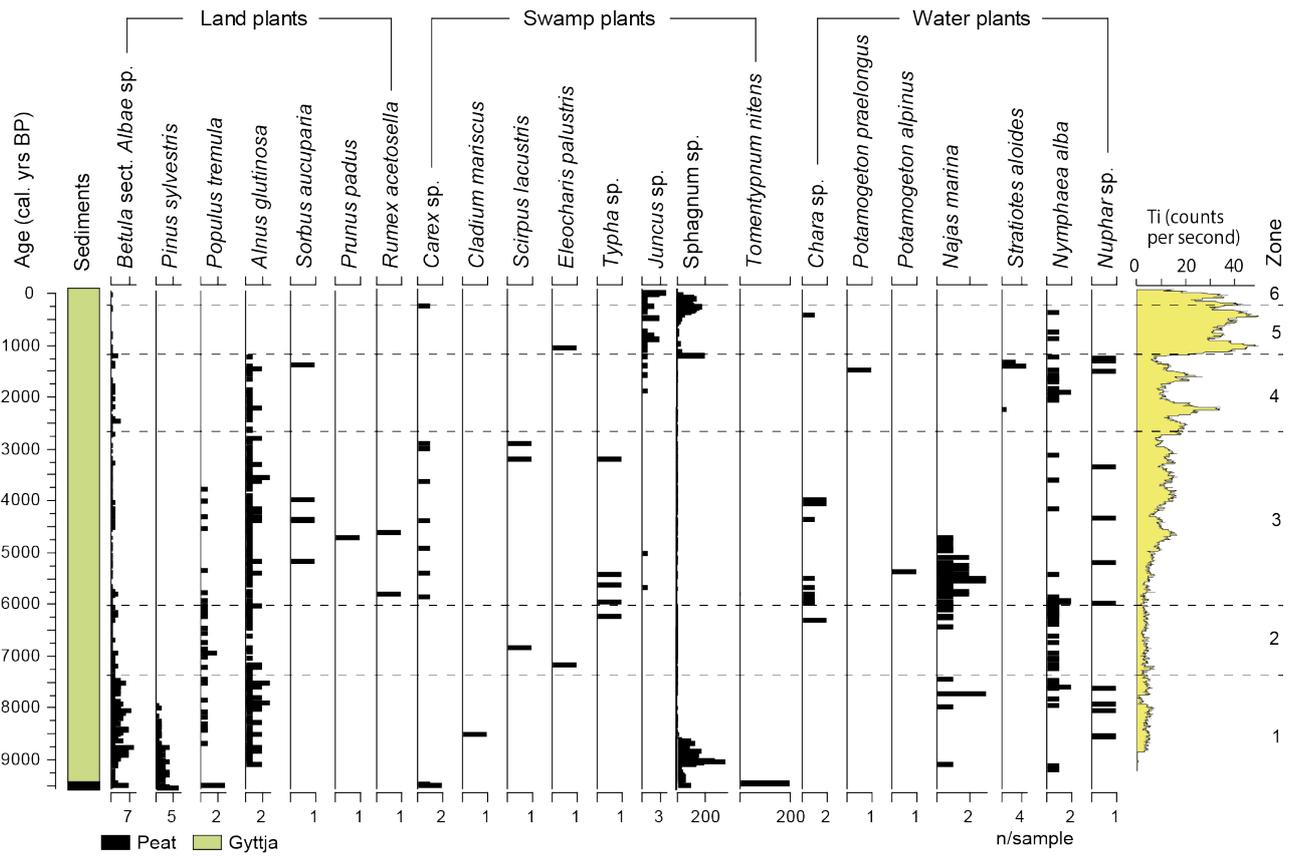
1164
1165 Fig. 1



1166

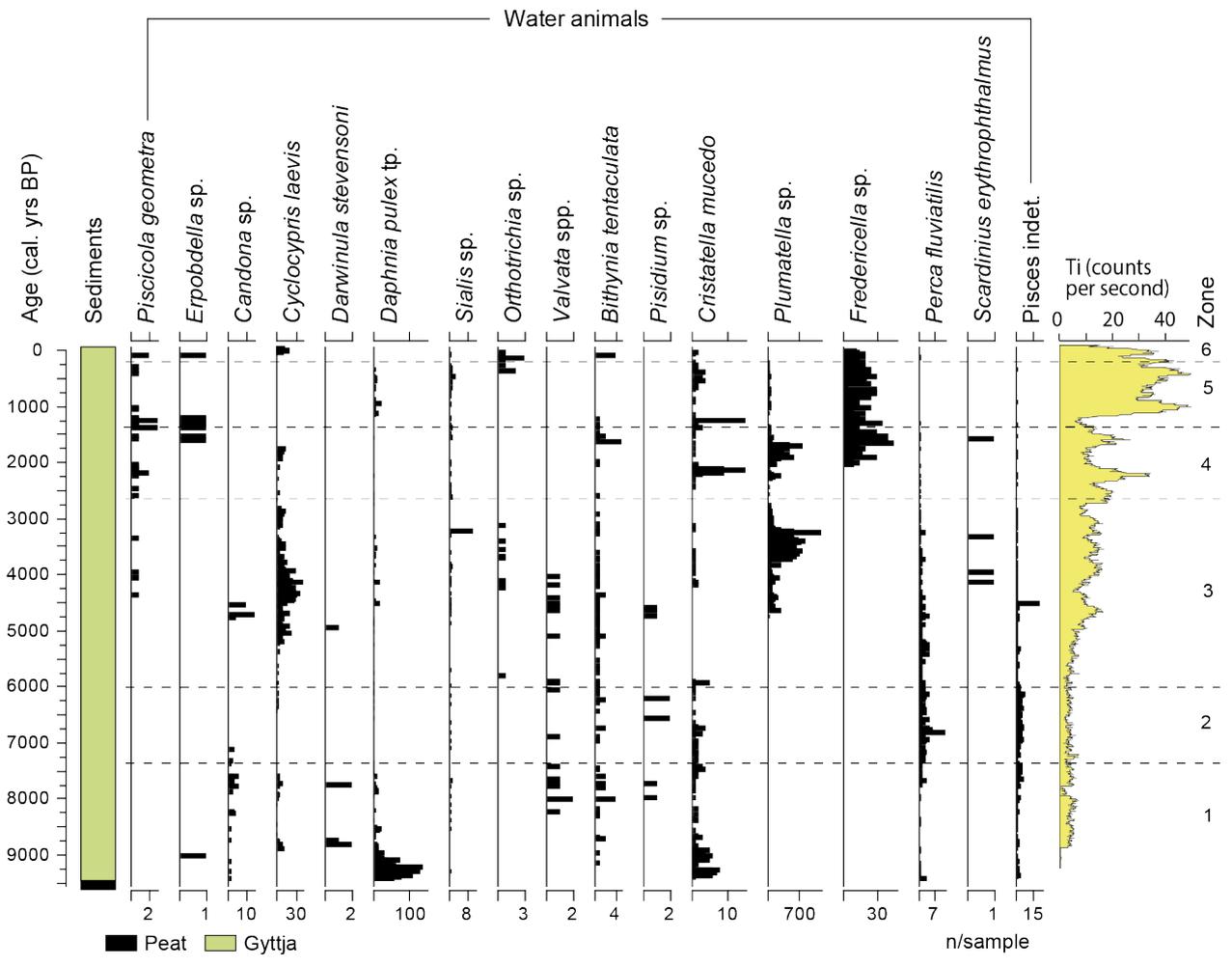
1167 Fig . 2

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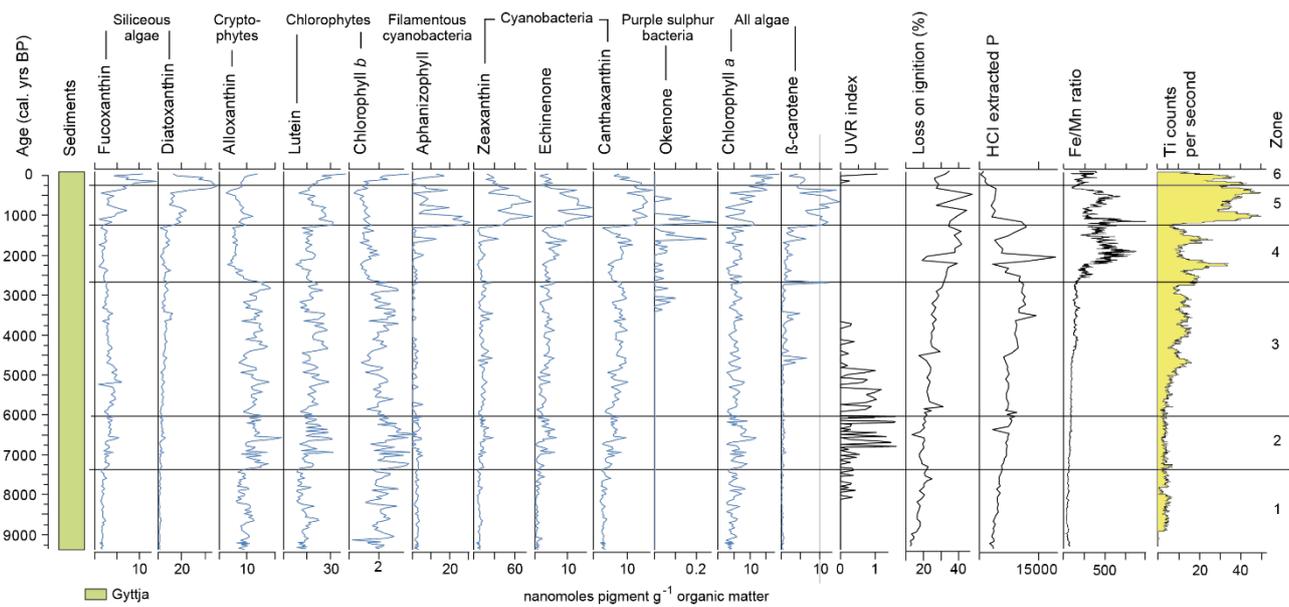
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1170 Fig. 3



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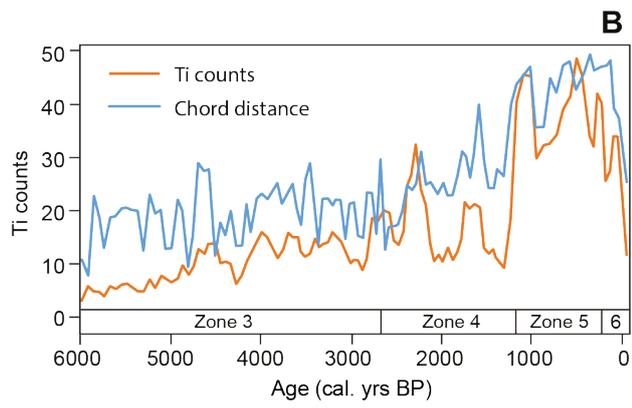
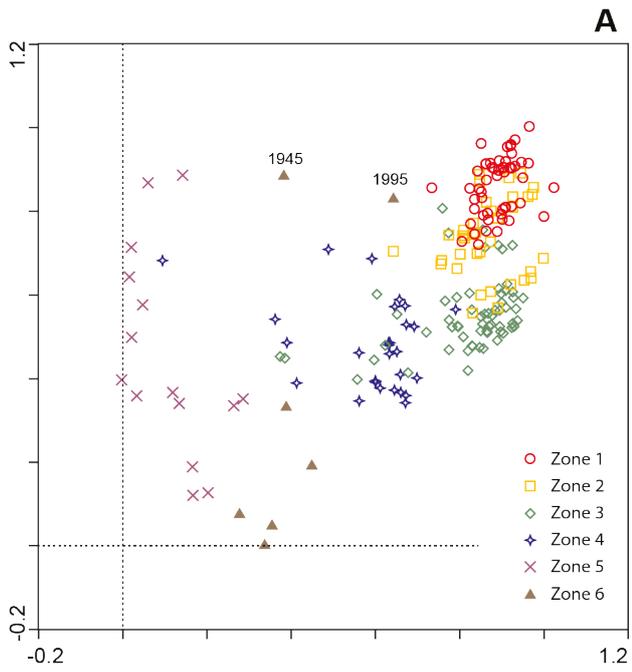
1172 Fig. 4



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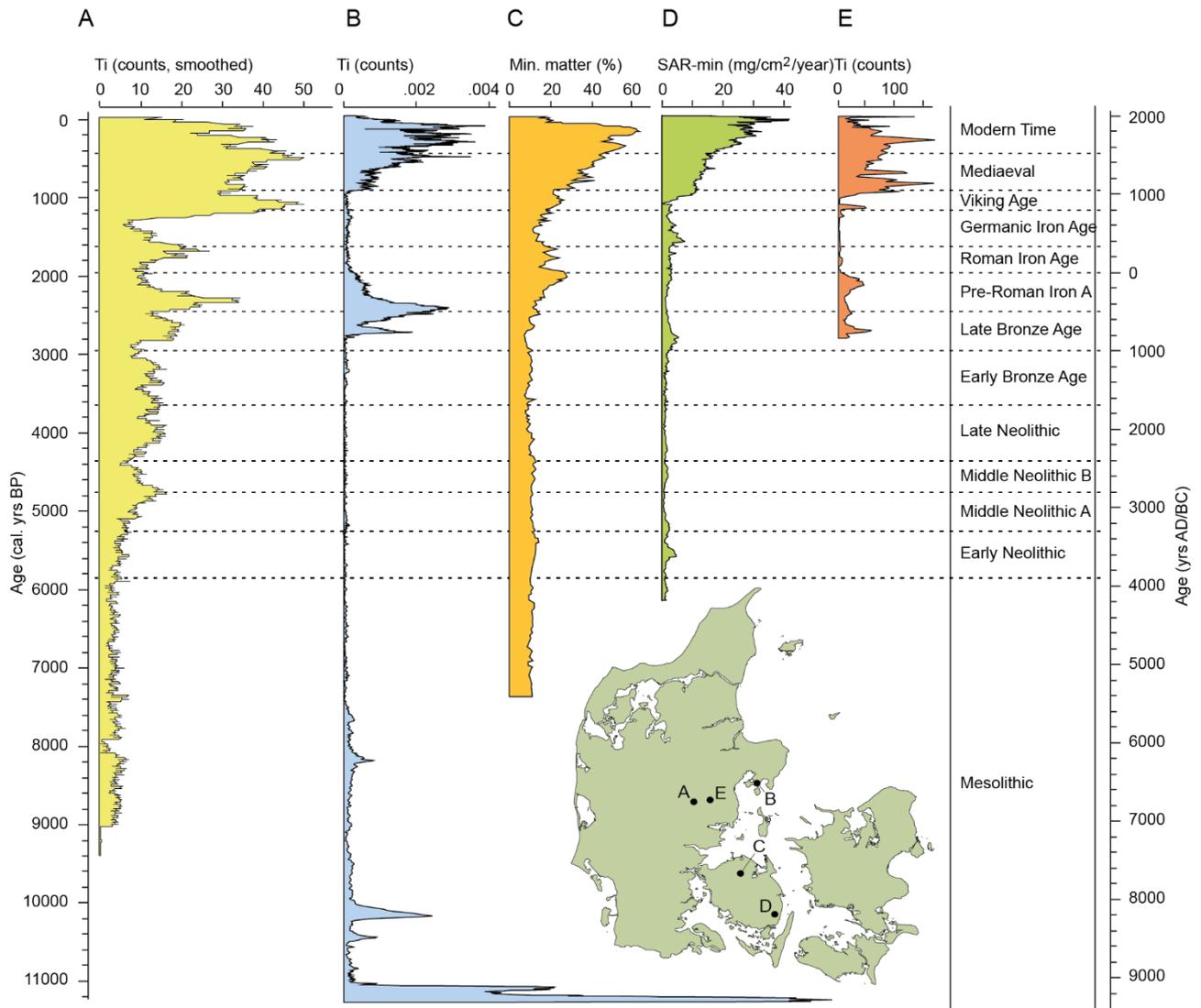
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Fig. 5



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1176 Fig. 6



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1178

1179

Fig. 7

Supplementary

Early historical forest clearance caused major degradation of water quality at Lake Væng, Denmark

Table S1

AMS radiocarbon ages from Lake Væng, Denmark

Laboratory number	Depth (cm)	Material ^a	¹⁴ C age ^b (yrs BP)	Cal. age ^c (cal. yrs BP)	Cal. age ^d (cal. yrs BP)
AAR-24956	82–86	<i>Nymphaea alba</i>	275 ± 30	154–443	361
AAR-24957	194–198	Leaf fragments	1456 ± 29	1301–1381	1338
AAR-24958	262–266	<i>Betula</i> sect. <i>Albae</i>	1886 ± 45	1707–1924	1795
AAR-24959	442–446	<i>Betula</i> sect. <i>Albae</i>	2830 ± 40	2805–3069	2936
AAR-24960	558–562	<i>Alnus glutinosa</i> , BA	3400 ± 30	3564–3815	3636
AAR-24961	670–674	Leaf fragments	3972 ± 28	4300–4522	4446
AAR-24962	766–770	Leaf fragments	4477 ± 32	4978–5290	5165
AAR-24963	866–870	Bark fragment	5189 ± 29	5905–5996	5948
AAR-24964	966–970	Ag, BA, Na	5430 ± 46	6014–6307	6237
AAR-24965	1078–1082	Leaf fragments	5359 ± 50	6000–6280	6140
AAR-24966	1206–1210	<i>Betula</i> sect. <i>Albae</i>	6562 ± 38	7422–7566	7468
AAR-24967	1306–1310	Ag, BA	7101 ± 46	7838–8013	7931
AAR-24968	1410–1414	AG, <i>Pinus sylvestris</i>	7750 ± 37	8430–8593	8521
AAR-24969	1522–1526	Ps, BA, Ag	8196 ± 38	9022–9277	9142
AAR-24970	1606–1610	BA, Ps	8335 ± 44	9143–9474	9355
AAR-24971	c. 1662	Twig	9562 ± 34	10 725–11 090	10 930

^a Full names are BA: *Betula* sect. *Albae* sp., Ag: *Alnus glutinosa*, Na: *Nymphaea alba*, Ps: *Pinus sylvestris*. The leaf fragments originate from deciduous woody plants.

^b Radiocarbon ages are reported in conventional radiocarbon years BP (Stuiver and Polach, 1977). The ¹⁴C ages have been corrected for isotopic fractionation to a $\delta^{13}\text{C}$ value of -25‰ .

^c Calibration to calendar years is according to the INTCAL20 data (2 sigma).

^d Mean probability ages.

Core correlations

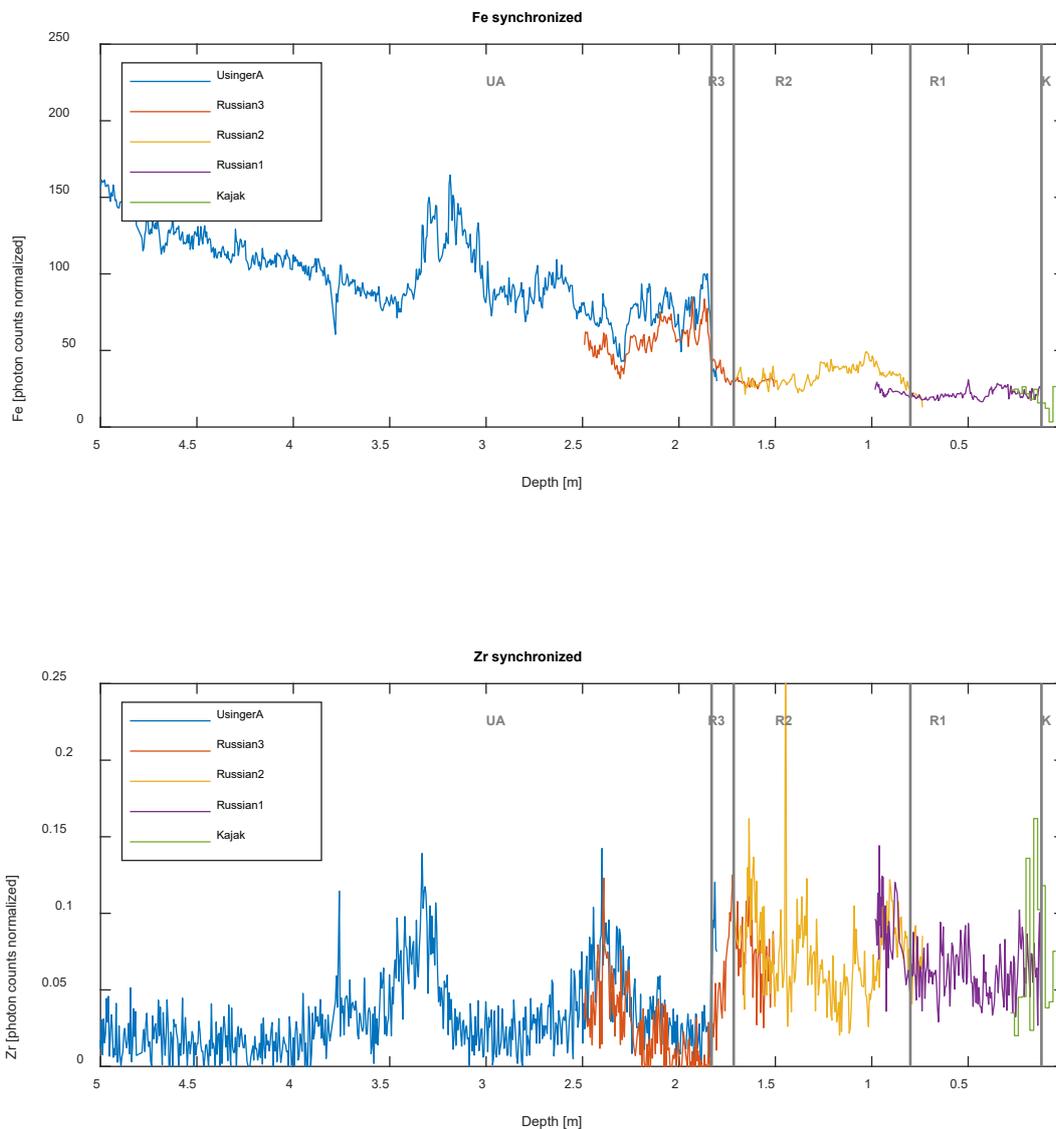
In the field the depths of the core sections were noted. The Kajak discrete samples, the Russian cores and the Usinger A and B cores overlaps in depth so they all together cover a continuous series of 16.30 m from the top of the sediment. The Kajak core was subsampled into 14 samples of 2 cm covering the top 28 cm of the sediment. The three Russian cores were each 1 m long overlapping in

1223 depth 18-20 cm, and XRF logged to respectively 10 to 110 cm, 90 to 190 cm, and 170 to 270 cm.
1224 The Usinger cores were 2 m long. The first Usinger A cores was retrieved from a logged depth of
1225 180 to 380 cm, and the following six Usinger A cores were retrieved from continuing depths. The
1226 Usinger B cores were shifted 50 cm in depth relative to the A cores with the first Usinger B cores
1227 retrieved from a logged depth of 230 to 430 cm, and the following six Usinger B cores from
1228 continuing depths.
1229

1230 In the XRF records the top of the sediment in each 2 m section of the Usinger A cores was assigned
1231 the depths from the top of the sediment, which was 175 cm below the water surface, of respectively
1232 1800, 3800, 5800, 7800, 9800, 11800 and 13800 mm. The Usinger B cores were correlated to the A
1233 cores using variations in the records of density proxy, Fe, Zr, radiography grey scale and magnetic
1234 susceptibility. The density proxy come from the ratio between the photon count numbers of the
1235 Rayleigh and the Compton scatter peaks in the XRF spectra, while the radiography grey scale come
1236 from the mean pixel value across the radiography image. The grey scale varies from core to core
1237 due to different thickness of the core half used for XRF scanning; however, variations in the grey
1238 scale on a depth scale below 1 m are useful for core correlations.
1239

1240 Table S2 shows the correlation points between all the cores. The depths listed in the table are
1241 measured from the logging depth at the top of the sediment in each core segment. The core
1242 segments of the Usinger A and B cores are indicated with colours.
1243

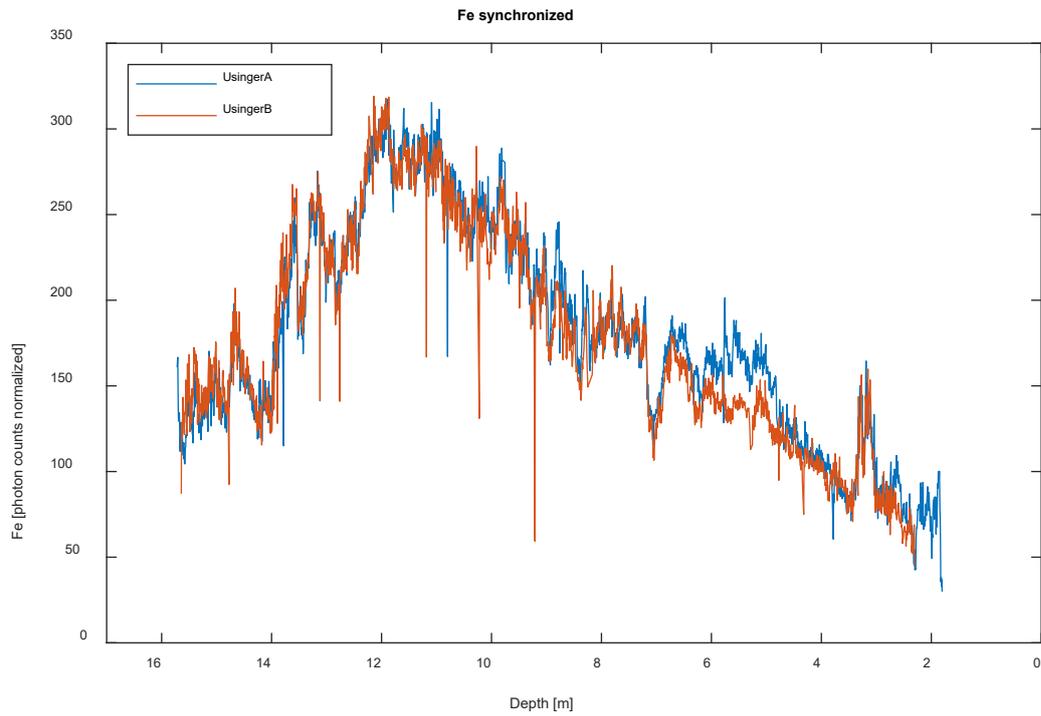
Usinger A extended to the top of the sediment



1245 Figure S1: Extension of Vængsø Usinger A core to the top of the sediment with the Russian1,2, 3 cores and discrete Kajak samples
1246 at the top. The top figure shows the upper 5 m of the iron record while the bottom figure shows the upper 5 m of the zirconium
1247 record. The vertical lines indicate the sections where cores labeled K: Kajak, R1,R2,R3: Russian1,2,3 and UA: Usinger A was used in
1248 the extension.

1249

UsingerA and UsingerB on the same depth scale



1251 *Figure S2: Iron records of Lake Væng Usinger A and Usinger B, on the Usinger A depth scale. The Usinger B was correlated to Usinger*
1252 *A using linear interpolation between 123 correlation points.*

1253
1254
1255
1256
1257

Kajak	Russian 1	Russian 2	Russian 3	Usinger A	Usinger B
0					
120	100				
	780	965			
	830	1080			
	945	1135			
		1695	1720		
		1745	1765		
		1785	1810		
		1860	1860		
		1880	1910		
			2030	1831	
			2131	1905	
			2189	1955	
			2222	2018	
			2346	2144	
			2376	2177	
			2414	2208	
			2460	2283	
				2341	2342
			2595	2394	
				2401	2418
				2438	2459
				2519	2516
				2546	2542
				2572	2563
				2639	2597
				2674	2648
				2714	2665
				2746	2716
				2774	2745
				2836	2840
				2867	2878
				2885	2894
				2913	2927
				2990	3015
				3029	3066
				3051	3096
				3074	3110
				3135	3133
				3172	3180
				3207	3214
				3237	3237

				3261	3269
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Table S2: Correlation points for synchronizing the Lake Væng cores to the Lake Væng Usinger A.

Usinger A	Usinger B		Usinger A	Usinger B		Usinger A	Usinger B
3332	3342		8472	8455		13471	13656
3386	3382		8530	8530		13530	13719
3412	3425		8605	8606		13634	13811
3541	3556		8646	8658		13713	13898
3652	3667		8677	8699		13995	14481
3703	3704		8737	8756		14685	15153
3768	3752		8767	8791		14748	15251
3916	3895		8828	8830		14822	15330
3996	3997		8997	9005		15028	15553
4062	4046		9041	9066		15148	15690
4116	4107		9106	9184		15376	15953
4229	4171		9180	9263		15485	16079
4383	4341		9243	9298		15548	16170
4482	4448		9391	9438			
4626	4560		9538	9579			
4707	4654		9579	9622			
4765	4718		9661	9726			
4944	4962		9855	9929			
5021	5043		10009	10045			
5254	5394		10146	10163			
5512	5504		10277	10328			
5618	5596		10349	10361			
5745	5755		10400	10418			
5915	5908		10542	10512			
6009	6018		10629	10570			
6160	6187		10729	10682			
6426	6354		10831	10812			
6725	6567		10902	10975			
6811	6827		11384	11473			
6920	6930		11469	11577			
7054	7075		11604	11708			
7165	7178		11907	11967			
7291	7316		12299	12377			
7449	7474		12381	12477			
7616	7598		12451	12562			
7729	7739		12536	12661			
7830	7855		12639	12746			
7901	7929		12713	12835			
7967	7985		12843	12965			
8027	8062		13016	13175			

8071	8109		13175	13353		
8250	8311		13323	13486		
8342	8391		13427	13589		

1259 *Table S3: Table S2 continued.*

1260

1261 Reference

1262 Stuiver, M., Polach, H.A., 1977. Discussion of reporting ¹⁴C data. Radiocarbon 19, 355–363.

1263 <https://doi.org/10.1017/s0033822200003672>

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