

1 **A case study on the effects of data temporal resolution on the simulation of water**
2 **flux extremes using a process-based model at the grassland field scale**

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14

15 **Abstract**

16 Projected changes to rainfall patterns may exacerbate existing risks posed by flooding.
17 Furthermore, increased surface runoff from agricultural land increases pollution through
18 nutrient losses. Agricultural systems are complex because they are managed in individual
19 fields, and it is impractical to provide resources to monitor their water fluxes. In this respect,
20 modelling provides an inexpensive tool for simulating fluxes. At the field-scale, a daily time-
21 step is used routinely, however, it was hypothesized that a finer time-step will provide more
22 accurate identification of peak fluxes. To investigate this, the process-based SPACSYS
23 model that simulates water fluxes, soil carbon and nitrogen cycling as well as plant growth
24 with a daily time-step was adapted to provide sub-daily simulations. As a case study, the
25 water flux simulations were checked against a 15-minute measured water flux dataset from
26 April 2013 to February 2016 from a pasture within a monitored grassland research farm,
27 where the data were up-scaled to hourly, 6-hourly and daily. Analyses were conducted with
28 respect to model performance for: (a) each of the four data resolutions, separately (15-minute
29 measured versus 15-minute simulated; hourly measured versus hourly simulated; etc.); and
30 (b) at the daily resolution only, where 15-minute, hourly and 6-hourly simulations were each
31 aggregated to the daily scale. Comparison between measured and simulated fluxes at the four
32 resolutions revealed that hourly simulations provided the smallest misclassification rate for
33 identifying water flux peaks. Conversely, aggregating to the daily scale using either 15-
34 minute or hourly simulations increased accuracy, both in prediction of general trends and
35 identification of peak fluxes. For the latter investigation, the improved identification of
36 extremes resulted in 9 out of 11 peak flow events being correctly identified with only 2 false
37 positives, compared with 5 peaks being identified with 4 false positives of the usual daily
38 simulations. Increased peak flow detection accuracy has the potential to provide clear field
39 management benefits in reducing nutrient losses to water.

40 **Key words:** SPACSYS; extreme flows; North Wyke Farm Platform; scale effects; grassland;

41

42 **1 Introduction**

43 Flooding in the UK puts more than 5 million people in 2.4 million properties at risk each year
44 (Environment Agency, 2009). Projected changes to rainfall patterns (Watts and Anderson,
45 2016) may exacerbate the existing risks posed by flooding. Flash flooding or surface water
46 flooding, defined as those flood events where the rise in water is either during or within a few
47 hours of the rainfall that produces the rise, is one of the most common types of flooding in
48 the UK. The utilised agricultural area, of which almost 60% is permanent grassland, covers
49 71% of the total land of the UK (Department for Environment, Food and Rural Affairs,
50 2019). Water fluxes or surface runoff generated from agricultural land can contribute
51 significantly to local floods and nutrient losses that cause water pollution. Flooding of
52 farmland is likely to become more frequent in some areas under projected climate change
53 (Brown et al., 2016), although intriguingly, studies have found increases in precipitation
54 extremes do not necessarily mean increases in flood magnitude, due to decreased soil
55 moisture at storm onset and reduced storm durations (Sharma et al., 2018; Wasko et al.,
56 2019). Further, soil erosion is accelerating due to more intense rainfall, leading to the loss of
57 valuable topsoil and the pollution of watercourses (Morison and Matthews, 2016).

58

59 Accurate forecasting of water runoff (or water fluxes) from agricultural land is, therefore, not
60 only a vital component of flood early-warning systems, but also for associated management
61 strategies for nutrient loss and water pollution. Water fluxes from the soil surface are
62 controlled by soil properties. Long-term hydrological studies have shown that sandy Alfisols
63 can generate higher runoff compared to clayey Vertisols (Pathak et al., 2013), and a greater

64 risk of flooding on clay soils has been reported (Charlton et al., 2010). The wetness of the soil
65 before a precipitation event (Merz and Plate, 1997) and soil compaction also affect water
66 fluxes. Farm machinery and livestock (Adimassu et al., 2019; Alaoui et al., 2018; Newell
67 Price et al., 2012) can cause serious compaction and so exacerbate flood risk. Natural events,
68 particularly long and intense precipitation events (Archer and Fowler, 2018), and land cover
69 variation (Dadson et al., 2017; Keesstra et al., 2018) also affect flux.

70

71 Agricultural systems are complex because they are generally managed at the field scale and
72 each field has its own unique set of soil conditions and topology. Monitoring water surface
73 fluxes in fields is costly both in time and financially. In this respect, modelling provides an
74 effective tool for simulating or forecasting water fluxes. The SPACSYS model (Wu et al.,
75 2007) is one such process-based model. It is a field scale and weather-driven dynamic
76 simulation model. Since it was first published in 2007, it has been developed to provide
77 added functionality (Bingham and Wu, 2011; Liu et al., 2013; Wu et al., 2019; Wu et al.,
78 2015). The model can simulate the interactions of soil carbon (C), nitrogen (N) and
79 phosphorus (P), plant growth and development, water re-distribution and heat transformation
80 in agricultural fields. The model has been used to investigate several issues including
81 resource use efficiency by crops (Wu et al., 2009), greenhouse gas (GHG) emissions (Abalos
82 et al., 2016; Perego et al., 2016), the responses of cropping/grassland systems to
83 environmental change (Wu et al., 2016) and the forecasting of crop yield and stocks of C and
84 nutrients (Zhang et al., 2016) under various climatic and soil conditions.

85

86 The SPACSYS model has been developed to investigate not only temporal dynamics, but
87 also within-field spatial variation in processes such as water runoff, using a linked, grid-based

88 approach (grid-to-grid) (Liu et al., 2018). As in all previous implementations of SPACSYS,
89 and common to many agriculture-focused models (Ahuja et al., 2002), a daily time-step was
90 used. However, model predictions of water flux did not increase in accuracy when
91 considering grid connectivity. We hypothesise, that a finer time-step might provide this
92 improvement instead; not only in the grid-to-grid model, but also in the (non-grid-to-grid)
93 standard model, as investigated here. Although not demonstrated within this study, increasing
94 the accuracy of water flux simulations should implicitly increase the accuracy of associated
95 SPACSYS simulations, such as those for nutrient loss that use predicted water flux in their
96 calculation.

97

98 For our case study, we used measured 15-minute water flux data from one field (or sub-
99 catchment) of the North Wyke Farm Platform (NWFP). The NWFP is a systems scale
100 research facility in the south-west of England for investigation of the sustainability of
101 lowland ruminant production systems (Orr et al., 2016). South-west England has a relatively
102 wet climate where the greatest rainfall is in winter and the driest times are between April to
103 July. August tends to show an increase in rainfall over July and starts the inexorable rise in
104 rainfall into autumn and early winter. More recently, the number of flood events has
105 increased (Stevens et al., 2016), mostly in the autumn and winter months; all as a likely
106 consequence of increased surface water runoff (Palmer and Smith, 2013).

107

108 For this study, the NWFP's 15-minute water flux data were up-scaled to hourly, 6-hourly and
109 daily data and the SPACSYS model was adapted to provide corresponding downscaled
110 simulations at 15-minute, hourly and 6-hourly resolutions (in addition to its usual daily
111 output). This provided four measured water flux datasets and four simulated water flux

112 datasets over a study period of 34 months (April 2013 to February 2016). Simulations were
113 generated using the same field management practices and parameter configurations. These
114 rich water flux datasets enabled investigation of the effects of temporal scale on model
115 performance not only in terms of extreme water runoff, which is the study focus and provides
116 it's novelty, but also in terms of general trends.

117

118 **2 Materials and Methods**

119 *2.1 Model description*

120 The SPACSYS model includes a plant growth and development component, an N cycling
121 component, a C cycling component, a P cycling component, plus a soil water component that
122 includes representation of water flow to field drains as well as downwards through the soil
123 layers, together with a heat transfer component. The equations to quantify such different
124 processes have been described elsewhere (Liu et al., 2013; Wu et al., 2019; Wu et al., 2007;
125 Wu et al., 2015). Here, only the processes influencing directly the soil water component are
126 presented.

127

128 For SPACSYS, the Richard's equation for water potential and Fourier's equation for
129 temperature are used to simulate water and heat fluxes, which are inherited from the SOIL
130 model (Jansson, 1998). If the water content in a layer rises above a specified value a
131 proportion is held in macropores such that rapid downward water movement takes place due
132 to gravitational forces alone. Water flow from the soil profile to a drainage pipe occurs when
133 the ground water table is above the bottom level of the pipe and the soil below the ground
134 water table is saturated. The Hooghoudt drainage flow equation with modification is adopted
135 for the subsurface drainage flow.

136

137 The main processes concerning plant growth in SPACSYS are plant development,
138 assimilation, respiration, root growth and development, water uptake, nutrient uptake,
139 biological N fixation for legume plants and partitioning of photosynthate and nutrients from
140 uptake estimated with various mechanisms implemented in the model. N cycling coupled
141 with C cycling covers the transformation processes for organic matter and inorganic N. The
142 main processes and transformations causing size changes to mineral N pools are
143 mineralization, nitrification, denitrification including N gaseous emission and plant N uptake.
144 P cycling is linked to other components such as the plant component, heat transformation and
145 the water cycle. Organic P is subdivided into certain sub-pools with different forms which are
146 connected with transformation rates.

147

148 2.2 *The North Wyke Farm Platform*

149 The study site is located in south-west England, at the NWFP, Rothamsted Research,
150 Okehampton, Devon (50°46'10''N, 30°54'05''W). For the period 1985-2015, the mean
151 annual temperature in North Wyke ranges between 6.8 and 13.4 °C, the mean annual rainfall
152 is 1033 mm and the climate is classed as cool temperate. The platform is a 63 ha systems-
153 based experimental facility divided into 15 hydrologically isolated sub-catchments across
154 three 21 ha farmlets with five sub-catchments in each. At the time of this study, all three
155 farmlets were used solely for grazing livestock research (sheep and cattle) where each farmlet
156 was operating under different sward management strategies: no re-seeding (permanent
157 pasture); re-seeded monoculture; and re-seeded legume mix. The platform monitors routinely
158 water runoff and water chemistry in each of the 15 sub-catchments, together with other
159 primary data collections (e.g. greenhouse gas emissions, livestock performance) so that each

160 farming system can be described, contrasted and compared according to its level of
161 sustainability (Orr et al., 2016). Datasets are freely available from
162 <https://www.rothamsted.ac.uk/north-wyke-farm-platform>, including those used in this study.

163

164 2.3 Model configuration

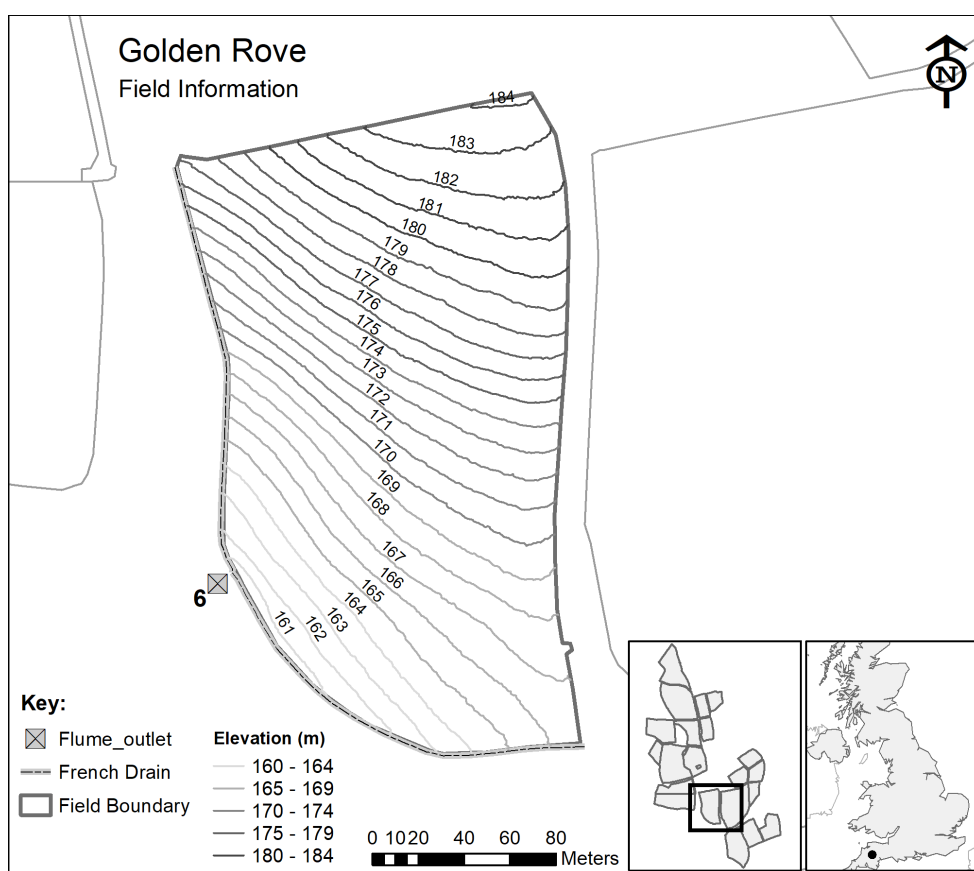
165 For this study, we focused on water fluxes for one sub-catchment in the permanent pasture
166 farmlet called ‘Golden Rove’; a single field that has been under permanent pasture since the
167 outset of the platform in 2010 (Fig. 1). The soil class for this field is primarily Halstow,
168 which comprises a slightly stony clay loam topsoil (approximately 36% clay) that overlies a
169 mottled stony clay (approximately 60% clay), derived from underlying Carboniferous Culm
170 rocks (Harrod and Hogan, 2008). The study field also has a smaller, but not insignificant area
171 of Denbigh-Cherubeer soil class. In the simulations, the soil type was ignored.

172

173 To mimic the grazing system, daily grass intake and excretion of sheep and cattle in the field
174 was estimated before running the simulations (Carswell et al., 2019; Wu et al., 2016). Soil
175 physical and chemical properties of the field were adopted from a previous study of the same
176 field (Wu et al., 2016). The temporal frequency for the measured water fluxes ($l\ s^{-1}$) from a
177 NWFP water flume has been 15-minutes since the outset of the platform’s setup in October
178 2012. However, meteorological measurements at the same 15-minute resolution were only
179 initiated from 30 April 2013. Thus, to ensure consistency in the frequency of the driving
180 variables and the water flux as an output variable, simulations also started from 30 April
181 2013. An end-date of 15 February 2016 was chosen to give an interrupted data collection
182 time period of 34 months. A longer time period would entail having significant periods of
183 missing data due to instrument failure (i.e. there were no measurements on water flux

184 between 15 February and 24 October 2016). A previous scale-focused study, analysing the
185 measured 15-minute water fluxes together with aggregations at 30-minute, hourly, 3-hourly,
186 6-hourly, 12-hourly and daily resolutions, indicated that 15-minute, hourly, 6-hourly and
187 daily resolutions is sufficient to communicate all key outcomes adequately (Curceac et al.,
188 2020). Thus, the same four temporal resolutions were adopted for the model simulations of
189 this study.

190



191

192 Figure 1. Details of the NWFP sub-catchment selected for this study (sub-catchment number
193 6 of 15, consisting of a single field called Golden Rove).

194

195 2.4 *Statistical analysis*

196 Two distinct sets of statistical analyses were conducted with respect to model performance
197 and data resolution: (a) model performance for each of the four data temporal resolutions,
198 separately (i.e. 15-minute measured versus 15-minute simulated; hourly measured versus
199 hourly simulated; 6-hourly measured versus 6-hourly simulated; daily measured versus daily
200 simulated); and (b) model performance conducted at the daily temporal resolution only,
201 where 15-minute, hourly and 6-hourly simulations were each aggregated to the daily scale.
202 The latter analyses provide valuable insights into the worth of using fine temporal resolution
203 data to increase the accuracy of daily simulations, especially with respect to the accurate
204 identification of extremes. This is important as many process-based models in the literature
205 simulate only at a daily time-step (e.g. Del Grosso et al., 2009).

206

207 2.4.1 Model performance graphics

208 Model performance graphics consist of time-series plots, density plots and scatterplots of
209 measured and simulated datasets. For the latter, the ideal 1:1 line, a linear regression fit, and a
210 non-linear regression fit (i.e., a Loess smoother fit; Cleveland, 1979) are given where the
211 estimated intercept and slope parameters from the linear fit should equal zero and one for
212 perfect model simulations, respectively. Results (p -values) from a linear hypothesis test are
213 reported comparing this ideal model with the estimated model using a finite sample F test
214 (see Fox, 2016). The non-linear regression provides added insight into where the simulated
215 values tend to over- or under-predict (e.g., at measured low or high values, respectively).
216 Time-series plots for the errors (i.e. measured minus simulated data) are also given.

217

218 2.4.2 Model performance indices

219 To further assess the accuracy of the simulations, six accuracy indices were calculated: the
 220 mean absolute error (MAE), the normalized root mean square error (NRMSE), the percentage
 221 bias (PBIAS), the Nash-Sutcliffe efficiency (NSE), the index of agreement (d) and the Kling-
 222 Gupta efficiency (KGE), as given in Table 1.

223

224 Table 1. Accuracy indices formulae, where \hat{z}_i are the simulated values, z_i are the measured
 225 values, \bar{z}_i is the mean of the measured values, r is the Pearson product-moment correlation
 226 coefficient (between measured and simulated) and σ is the standard deviation.

Index form	Index formula	Min.	Max.	Ideal
Error	$MAE = \frac{1}{N} \sum_{i=1}^N \hat{z}_i - z_i $	0	∞	0
Error	$NRMSE = 100 \frac{\sqrt{\frac{1}{N} \sum_{i=1}^N (\hat{z}_i - z_i)^2}}{z_{max} - z_{min}}$	0	∞	0
Error	$PBIAS = 100 \frac{\sum_{i=1}^N (\hat{z}_i - z_i)}{\sum_{i=1}^N z_i}$	0	∞	0
Agreement	$NSE = 1 - \frac{\sum_{i=1}^N (\hat{z}_i - z_i)^2}{\sum_{i=1}^N (z_i - \bar{z}_i)^2}$	$-\infty$	1	1
Agreement	$d = 1 - \frac{\sum_{i=1}^N (\hat{z}_i - z_i)^2}{\sum_{i=1}^N (\hat{z}_i - \bar{z}_i + z_i - \bar{z}_i)^2}$	0	1	1

Agreement	$KGE = 1 - \sqrt{(r - 1)^2 + \left(\frac{\sigma_z}{\sigma_z} - 1\right)^2 + \left(\frac{\bar{z}}{\bar{z}} - 1\right)^2}$	$-\infty$	1	1
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227

228 2.4.3 Simulation accuracy of measured peaks

229 To investigate model accuracy in simulating water flux peaks, a threshold at the 99th
 230 percentile of each measured water flux dataset was used to identify peak flows. Model
 231 simulations were then assessed to determine if they could similarly exceed this threshold
 232 coinciding with a measured exceedance. Incidences of correct peak flow simulations, false
 233 negatives (simulation does not exceed threshold when measured flow does), false positives
 234 (simulation exceeds threshold when measured flow does not) and corresponding Kappa
 235 values are reported. The Kappa statistic provides a measure of agreement beyond the level of
 236 agreement expected by chance alone. General guidelines for Kappa values are as follows: less
 237 than 0.2 slight agreement, 0.2 to 0.4 fair agreement, 0.4 to 0.6 moderate agreement, 0.6 to 0.8
 238 substantial agreement, greater than 0.8 almost perfect agreement, and equal to 1 perfect
 239 agreement.

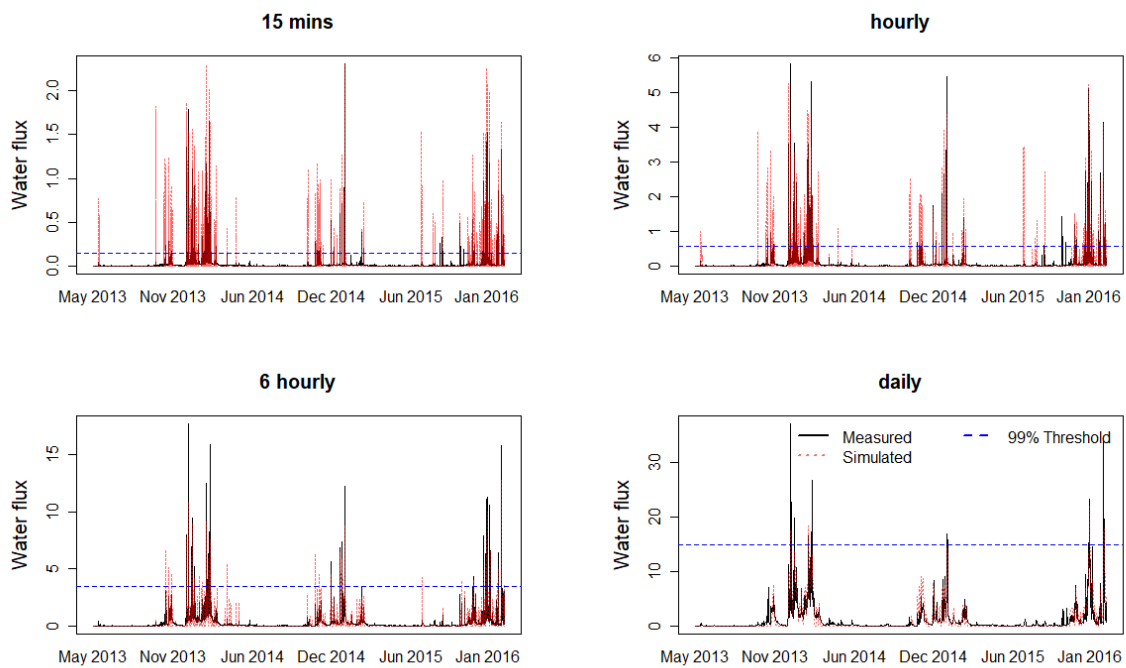
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241 3 Results

242 3.1 Model performance for each of the four data temporal resolutions, separately

243 Comparisons between the measured and simulated water flux rates at different temporal
 244 resolutions are shown in Fig. 2. Visually, it appears that simulations of daily and 6-hourly
 245 water fluxes tend to under-predict the measured data, often missing high peaks, while
 246 simulations of 15-minute and hourly data possibly tend to over-predict. However, the
 247 scatterplots of the measured and simulated data, together with the ideal 1:1 line, a linear

248 regression fit, and a Loess smoother fit (Fig. 3) present a more complete picture. Simulations
 249 for all four temporal resolutions clearly tend to over-predict, with the level of over-prediction
 250 increasing as the resolution increases. Over-prediction is shown with each linear regression
 251 fit lying below the 1:1 line; and increasingly so, as the resolution increases. All linear
 252 regression fits were found to be significantly different to the 1:1 line, each with F -test p -
 253 values < 0.0001 .



254

255

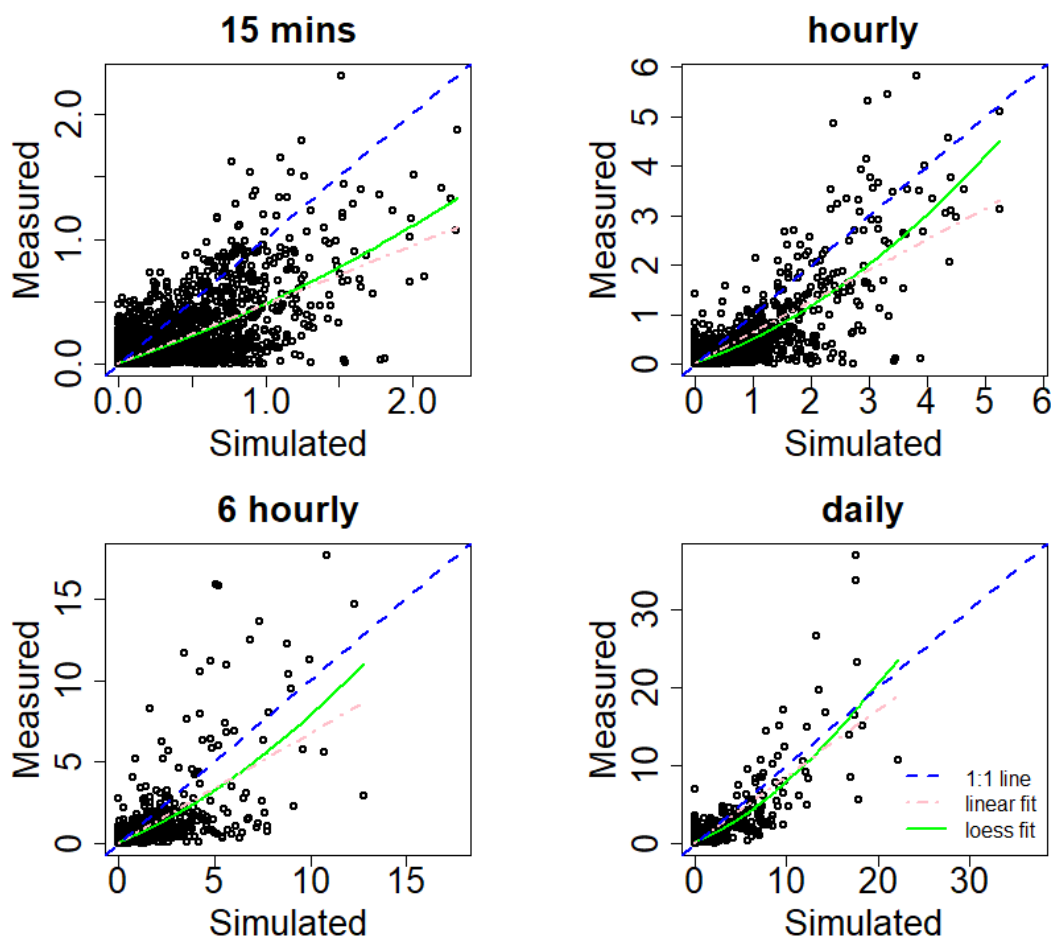
256 Figure 2. Time-series plots for measured and simulated water flux data (not aggregated) for
 257 15-minute, hourly, 6-hourly and daily data (in units of $\text{mm } 15\text{min}^{-1}$, mm h^{-1} , $\text{mm } 6\text{h}^{-1}$ and mm
 258 d^{-1} , respectively). All plots are shown with a threshold at the 99th percentile of measured data
 259 (at $0.138 \text{ mm } 15\text{min}^{-1}$, 0.553 mm h^{-1} , $3.45 \text{ mm } 6\text{h}^{-1}$ and 14.9 mm d^{-1} , respectively).

260

261 For all four temporal resolutions, the tendency to over-predict decreases at the largest
 262 measured water fluxes, as shown by the concave behaviour of the loess smoother fit (Fig. 3),

263 with daily simulations tending to under-predict at very large fluxes, thus, missing extreme
264 events that may cause flooding and associated nutrient and sediment losses. Clearly,
265 ‘smoothing bias’ increases as temporal resolution decreases. The 15-minute simulations
266 maintain the variation shown in the measured data (i.e. observations range from 0 to 2.306
267 mm 15min⁻¹ while simulations range from 0 to 2.310 mm 15min⁻¹), while the daily
268 simulations do not (i.e. observations range from 0 to 36.97 mm d⁻¹ while simulations only
269 range from 0 to 22.20 mm d⁻¹). As each ‘simulation-to-observation’ comparison is on a
270 different scale, it is not useful to present further model fit diagnostics, such as error and
271 agreement indices.

272



274 Figure 3. Scatterplots of the measured and simulated data (not aggregated) for 15-minute,
275 hourly, 6-hourly and daily data. Scatterplots are shown with the 1:1 line, a linear regression
276 fit and a loess smoother fit. Units are in $\text{mm } 15\text{min}^{-1}$, mm h^{-1} , $\text{mm } 6\text{h}^{-1}$ and mm d^{-1} ,
277 respectively.

278

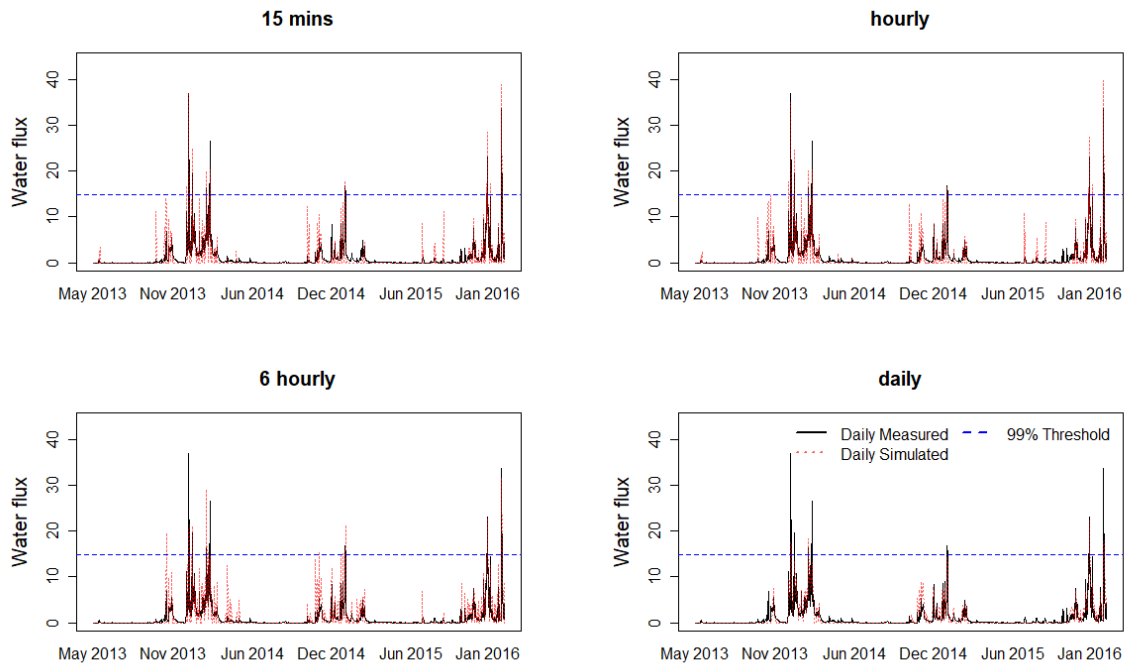
279 3.2 Model performance for simulations aggregated to the daily scale

280 Comparisons between the measured and simulated water flux rates aggregated to the same
281 daily scale are shown in Fig. 4. There are clear instances of both over- and under-prediction
282 for all four daily outputs. The scatterplots (Fig. 5) of daily measured and daily simulated data
283 from different aggregations, together with the 1:1 line, a linear fit, and a loess smoother fit,
284 again provide a clear visualisation of the relations in the time-series plots. Simulations for all
285 three aggregations to daily (15-minute, hourly and 6-hourly) again tend to over-predict (as
286 their linear fits lie below the 1:1 line), but this over-prediction is broadly similar across the
287 four datasets, and not as great as that found with the unaggregated data, above. The 6-hourly
288 aggregations appear to be the least accurate. Again, all linear regression fits were found to be
289 significantly different to the 1:1 line, each with F -test p -values < 0.0001 .

290

291 In this instance, ‘smoothing boas’ increases as aggregation resolution decreases, where
292 simulations for 15-minute and hourly aggregations both increase the variation shown in the
293 measured daily data (i.e. 0 to 36.97 mm d^{-1}); with 15-minute daily aggregations ranging from
294 0 to 38.94 mm d^{-1} and hourly daily aggregations ranging from 0 to 39.64 mm d^{-1} . Conversely,
295 the 6-hourly daily aggregations and the daily simulations reduce variation with the 6-hourly
296 daily aggregations ranging from 0 to 31.70 mm d^{-1} and the (unaggregated) daily simulations
297 ranging from 0 to 22.20 mm d^{-1} . In summary, daily simulations based on component 15-

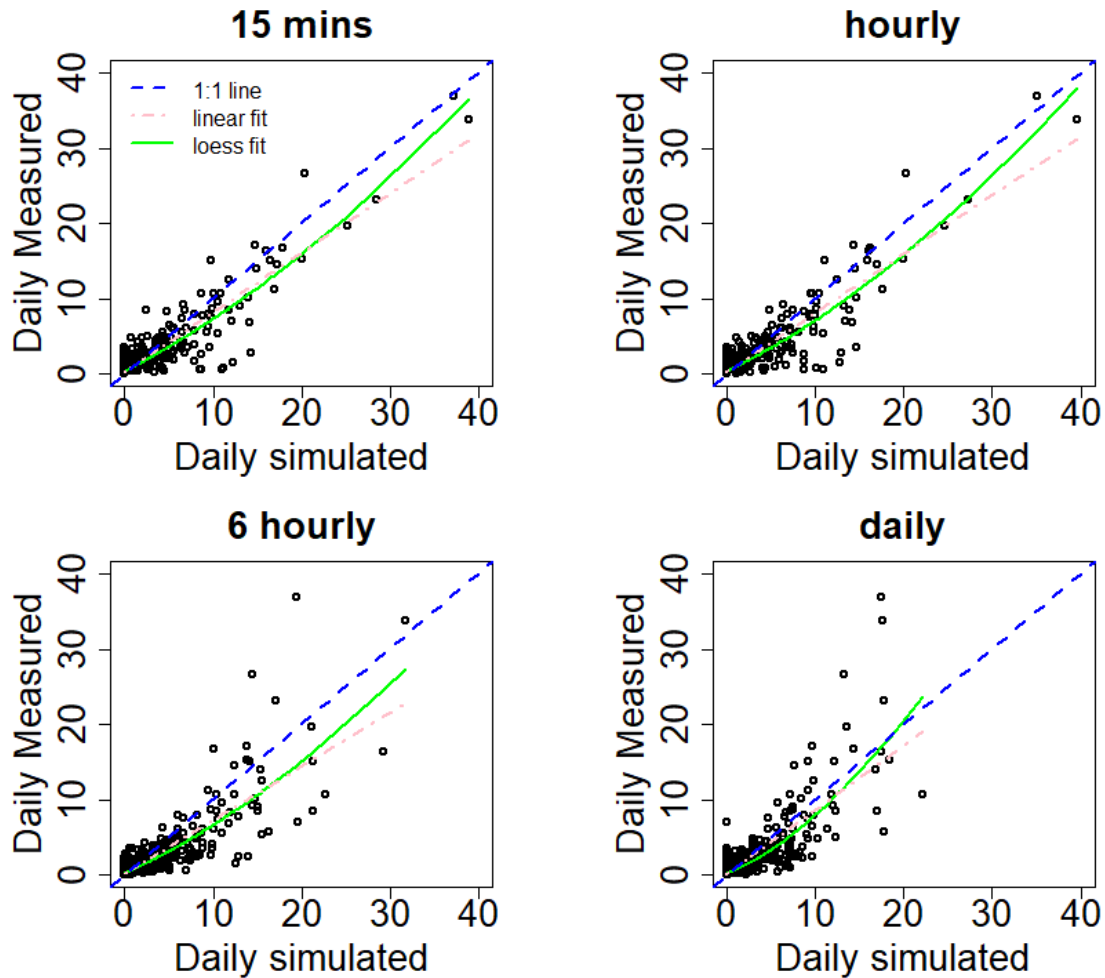
298 minute and hourly aggregations have the potential to identify peak water fluxes (and, thus,
299 flood events) and predict their magnitudes more accurately, relative to 6-hourly aggregations
300 and (unaggregated) daily simulations.



301

302 Figure 4. Time-series plots for daily measured and daily simulated water flux data (with the
303 first three plots having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to
304 daily). All units in mm d^{-1} . All plots are shown with a threshold at the 99th percentile of
305 measured data (14.90 mm d^{-1}).

306



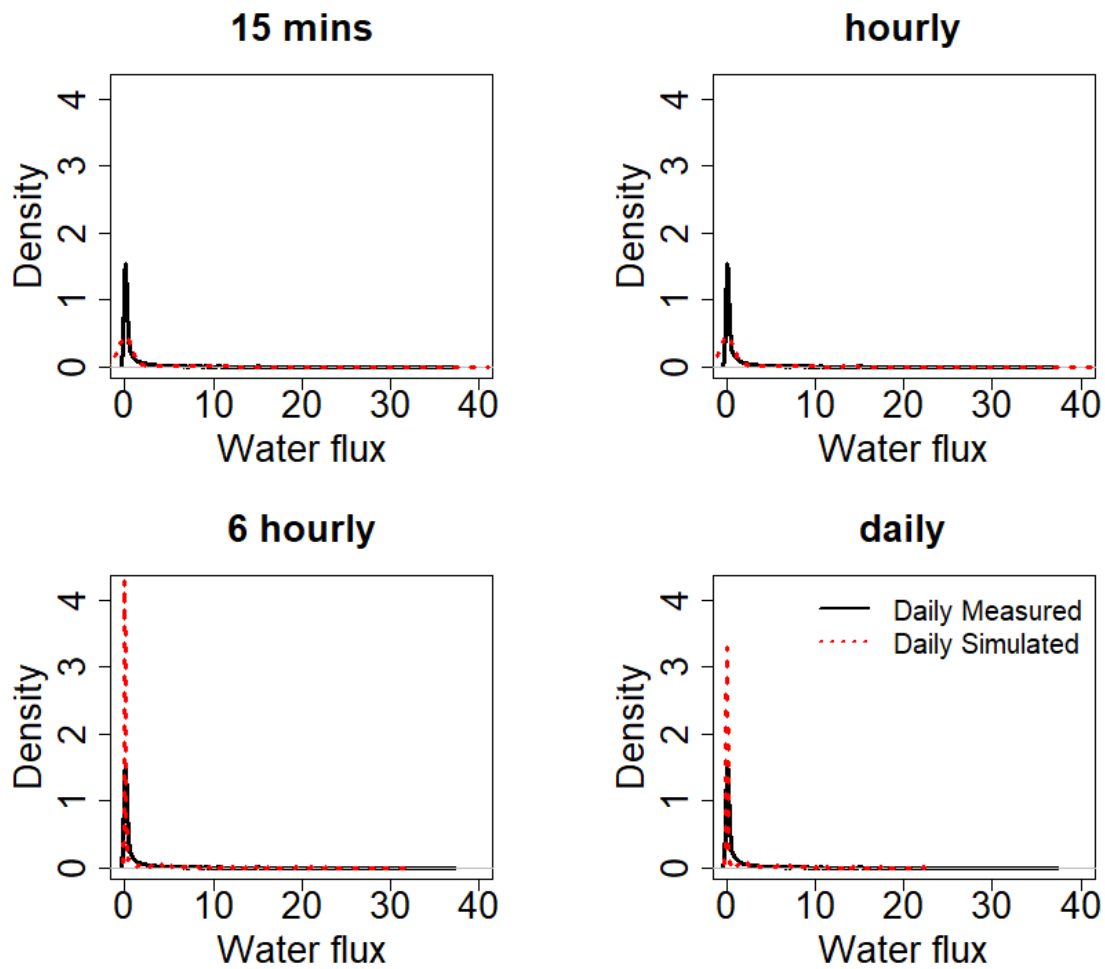
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308 Figure 5. Scatterplots of the daily measured and daily simulated water flux (with the first
 309 three plots having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to
 310 daily). Scatterplots are shown with the ideal 1:1 line, a linear regression fit and a loess
 311 smoother fit. All units in mm d^{-1} .

312

313 Further clarity on bias is provided in the density plots for the measured and simulated data
 314 (Fig. 6). Here, daily simulations based on 15-minute and hourly aggregations have a lower
 315 density at small daily water fluxes than that found with the measured data, while the 6-hourly
 316 aggregations and (unaggregated) daily simulations have a higher density at small daily water
 317 fluxes. This is combined with a longer tail in the density curve for the 15-minute and hourly

318 aggregations, as each can simulate large daily water fluxes, while the 6-hourly aggregation
319 and (unaggregated) daily simulations do not have this property.



320

321

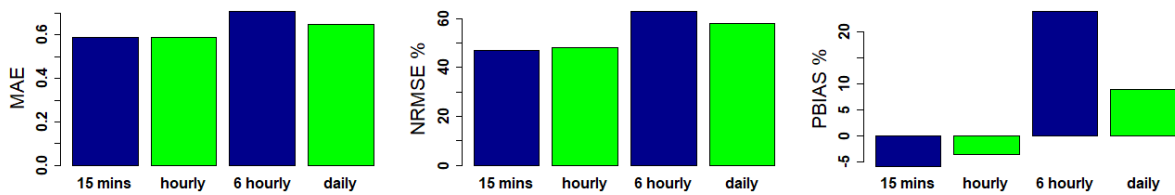
322 Figure 6. Density plots for daily measured and daily simulated data (with the first three plots
323 having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily). All units
324 in mm d^{-1} .

325

326 The error indices (MAE, NRMSE and PBIAS) are reported for each daily aggregation in Fig.
327 7, where the 15-minute and hourly aggregations clearly perform more accurately than the 6-
328 hourly aggregation and (unaggregated) daily simulations. Errors (i.e. residuals) are also

329 reported over the study time period in Fig. 8, where errors tend to be larger for the daily
 330 simulations based on the 6-hourly aggregation and the (unaggregated) daily simulations.
 331 Interestingly, the 6-hourly aggregation consistently is the least accurate, including being less
 332 accurate than the (unaggregated) daily simulations.

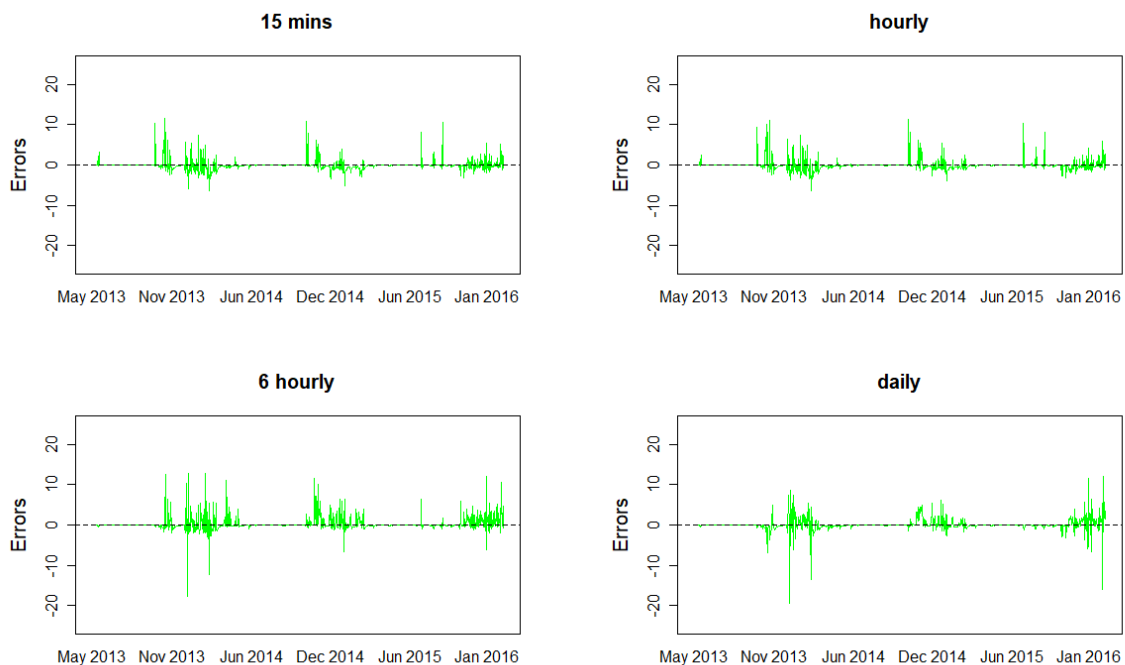
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335 Figure 7. Error (MAE, NRMSE, PBIAS) indices with respect to daily measured and daily
 336 simulated data (with 15-minute, hourly, 6-hourly data aggregated to daily).

337



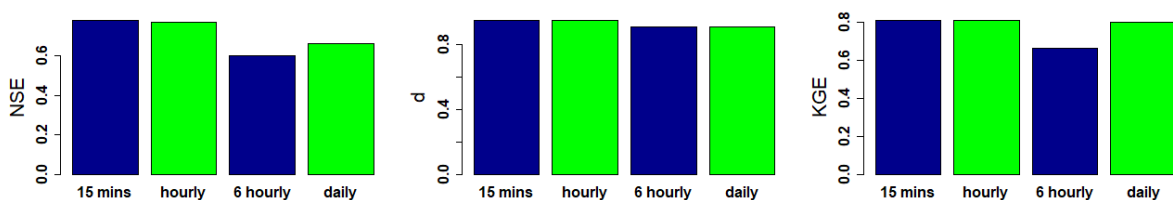
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339 Figure 8. Time-series of errors (simulated minus measured data) aggregated to the daily
340 temporal resolution. All units in mm d^{-1} . Positive errors represent over-prediction by the
341 model.

342

343 Agreement indices (NSE, d and KGE) are reported for each daily aggregation in Fig. 9,
344 where again the 15-minute and hourly aggregations perform more accurately than the 6-
345 hourly aggregation and (unaggregated) daily simulations (although daily simulations perform
346 relatively well according to the KGE index). From the given accuracy diagnostics, it is not
347 immediately apparent whether daily simulations based on 15-minute or hourly aggregations
348 are the most accurate, and as such, both appear to increase the accuracy relative to
349 SPACSYS' daily simulations. Again, the 6-hourly aggregation is the least accurate.

350



351

352 Figure 9. Agreement (NSE, d , KGE) indices with respect to daily measured and daily
353 simulated data (with 15-minute, hourly, 6-hourly simulations aggregated to daily).

354

355 3.3 *Simulation of measured peaks*

356 To investigate the ability of the model to simulate and identify water flux peaks, the 99th
357 percentile of each measured water flux dataset was used as a threshold to identify peak flows,
358 as highlighted in Figs. 2 and 4 (the dashed blue line). Incidences of correct peak flow

359 simulations, false negatives, false positives and the resultant Kappa values are given in Table
360 2. It appears that hourly simulations provide the largest correct classification rate (Kappa =
361 0.553) for the unaggregated approach, but with only moderate success in identifying
362 measured peak flows (as a promising 92% identification rate is tempered by a poor mis-
363 identification rate). Conversely, aggregating to the daily scale using either 15-minute or
364 hourly simulations was able to provide much greater agreement in identifying measured peak
365 flows at the daily scale, with each identifying 9 out of 11 peak flow events correctly, coupled
366 with only 2 false positives (Kappa = 0.816 in both cases). This level of agreement was far
367 greater than that found through directly simulating the daily data, which provided only
368 moderate success in identifying measured peak flows (Kappa = 0.495). Again, the 6-hourly
369 aggregation is the least accurate with a relatively high number of false positives (simulated
370 flow exceeds the threshold when measured flow does not).

371

372 Table 2. Accuracy at peak water fluxes according to simulation resolution. Peaks taken at 99th
373 percentile of measured data (see the dashed blue line in Figs. 2 and 4).

Simulation resolution	Sample size	Measured Peaks	Correctly Simulated	False Negative	False Positive	Kappa
Unaggregated						
15-minute	97920	980	759 (77)*	221	1224	0.506
hourly	24480	245	225 (92)	20	335	0.553
6-hourly	4080	41	32 (78)	9	52	0.503
daily	1020	11	5 (45)	6	4	0.495
Aggregated to daily						
15-minute	1020	11	9 (82)	2	2	0.816

hourly	1020	11	9 (82)	2	2	0.816
6-hourly	1020	11	6 (55)	5	9	0.455
daily	1020	11	5 (45)	6	4	0.495

374 * Value in brackets shows a percentage of correctly simulated peaks to measured peaks.

375 **4 Discussion**

376 *4.1 Model performance*

377 4.1.1 Unaggregated data

378 The statistical analyses for model performance suggested that the SPACSYS model simulates
379 the general trend of water fluxes at the four different temporal resolutions reasonably well
380 (Figs. 2 and 3). All simulations tended to over-predict water flux, however, and only
381 simulations at the finest resolutions maintained the variation in the measured data. The
382 accuracy of water flux peak simulations varied among the four resolutions (Table 2). Almost
383 92% of the measured peaks over the simulated period were modelled correctly at an hourly
384 resolution, the resolution with the smallest misclassification rate. However, this was tempered
385 by a high rate of predicting peaks that did not exist. A previous statistical analysis of peak
386 flows at different scales from a different NWFP sub-catchment (similar in size to the one
387 used here), modelled and simulated by a Generalized Pareto distribution, also showed the
388 greatest agreement at the hourly resolution (Curceac et al., 2020).

389

390 4.1.2 Aggregated to Daily

391 When simulations at a finer temporal resolution were aggregated to a daily rate, the
392 simulation results using both the 15-minute and hourly aggregations showed the greatest
393 accuracy broadly equally, both in the prediction of general trends (Figs. 4 to 9) and the

394 identification of peak flows (Table 2). This demonstrates clearly that the daily simulation of
395 water fluxes with the SPACSYS model, informed by finer temporal resolution data, can
396 increase simulation accuracy. This result is an important advance relative to previous
397 SPACSYS studies, which only used a daily time-step, and which similarly used sub-
398 catchments of the NWFP as the study site (Liu et al., 2018).

399 The simulation results using both the 15-minute and hourly aggregations showed the greatest
400 accuracy broadly equally. When the simulation time step is getting longer, average
401 precipitation intensity might be weaker, which causes simulated water fluxes smoother and
402 the deviations with monitored water fluxes larger. Given the complexity of the soil-water
403 processes that operate across the field, it is not surprising to see substantial variation around
404 the 1-1 line when predicting flow at 15-minute to 6-hour scales due to inherent variability.
405 We would expect to see a similar phenomenon with any similar process-based model. What
406 is important is that when these fine-scale predictions are aggregated we get substantial
407 improvements to daily predictions.

408 *4.2 Results in context and their generalisation*

409 Results are consistent with other studies that similarly showed that differences in the
410 (unaggregated or aggregated) time-step have the greatest impact on runoff simulation
411 accuracy relative to other factors, some of which, also investigated changes in spatial
412 resolution (i.e. aggregating over different spatial units) (Choi et al., 2018; Huang et al., 2019;
413 Jeong et al., 2010; Kavetski et al., 2011; Merz et al., 2009). Thus, the value of using
414 aggregated fine temporal resolution simulations to increase the accuracy of daily simulations
415 can be said to hold generally for other process-based models provided the hydrological
416 process is described appropriately. However, it does not follow that daily simulation accuracy
417 will continue to increase as the temporal resolution of the aggregated data becomes finer.
418 This study found aggregating hourly simulations to daily to be just as accurate as aggregating

419 15-minute simulations to daily, while aggregating 6-hourly simulations to daily performed
420 less well than the usual daily simulations.

421

422 The temporal resolution for process-based models should be chosen carefully to balance
423 between capturing all important processes, the study objectives and data availability. For our
424 study, with flooding as context, the identification of water flux extremes in a grassland field
425 (or small sub-catchment) with a heavy clayed soil, is viewed as *the* important process, more
426 so than capturing broad trends in water flux. It is well-known that running a model at a finer
427 resolution, then aggregate, will increase the prediction accuracy in a broad sense (see above).
428 What has received less attention in the literature is the effects of temporal resolution on a
429 model's ability to capture extremes (e.g. see Schaller et al., 2020, in the context of
430 streamflow). In this respect, our study has found daily peak flows to be more accurately
431 identified using aggregations of simulations at finer resolutions, than using coarse daily
432 simulations directly. Of course, measurement at a finer resolution comes at a cost and this
433 needs to be balanced with associated improvements in model accuracy. In this instance, this
434 interplay is simple to resolve since aggregating to the daily scale using both 15-minute and
435 hourly simulations were equally as accurate, meaning measuring at an hourly interval would
436 be sufficient for the case study site.

437

438 The appropriate temporal resolution to simulate water fluxes using a hydrological model
439 depends on hydro-climatological and geophysical characteristics, and the scale of the process.
440 It has been suggested that an appropriate temporal resolution could be between 12 hours for
441 middle-sized upstream areas and 48 hours for a complete river basin (Booij and Tran, 2005).
442 As the size of the field for this study is < 4 ha, the indicated hourly resolution appears

443 reasonable. Observed and projected changes in the UK's climate suggest an increase in heavy
444 rain events and wetter winters (Committee on Climate Change, 2017), where some UK
445 regions will be more affected than others. This will inevitably change agricultural
446 management practice and land use across the UK. Taking as an example the grazed pasture of
447 this study, introducing a deep-rooting grass suited to its heavy clay soils (Macleod et al.,
448 2013) and/or the mechanical loosening of topsoil (Newell-Price et al., 2011) would reduce
449 water runoff, whereas conversion to an arable crop (e.g. wheat) would provide its own set of
450 water runoff influences. Such changes would alter the characteristics of the water fluxes
451 generated, as the field's soil properties will change, meaning the determination of an
452 appropriate resolution to simulate water fluxes may also change from the hourly resolution
453 suggested here. This is analogous to other hydrological studies where, for example, different
454 overflow designs in roof drainage structures have markedly variable responses to rainfall
455 intensity increases (Verstraten et al., 2019).

456

457 *4.3 Inputs that impact hydrological model performance*

458 Key model input variables such as precipitation can determine the impacts of simulation
459 time-steps on the performance of hydrological models; for example, the duration and
460 temporal variability of a precipitation event in relation to the rainfall–runoff lag time (Ficchi
461 et al., 2016). A multiple-day precipitation event is the main cause of continuous runoff events
462 and related peaks. For example, for this study, there was almost an unbroken measured
463 precipitation period from 14 December 2013 to 5 March 2014, which brought a total of 541
464 mm of precipitation, 78% of which was measured as surface runoff (i.e., measured water
465 flux). Study simulations showed 70, 70, 81 and 85% as water fluxes over the period at the 15-
466 minute, hourly, 6-hourly and daily resolutions, respectively. Previous studies showed that
467 wetter soils had less capacity to store water, resulting in greater runoff volumes (Huang et al.,

468 2017; Kibet et al., 2014; Zehe et al., 2010). Both observations and simulations in this study
469 confirmed this finding. Conversely, for a single day event, a measured 92% of 40.2 mm daily
470 precipitation was discharged on 23 December 2013. The simulations generated 99, 99, 90 and
471 44% water losses at the 15-minute, hourly, 6-hourly and daily resolutions, respectively. Thus,
472 almost all of the precipitation contributed to the water loss on this day, where only the daily-
473 scale simulation did not capture this. However, although heavy rainfall is necessary to
474 generate water fluxes, it is not a sufficient condition for a higher surface runoff rate to occur
475 (Ledingham et al., 2019). For example, there was about 25 mm precipitation on 14 May 2013
476 and on 13 August 2015, but both the simulations and the observations (at all four resolutions)
477 did not show apparent water fluxes. Further, a daily precipitation of 4 mm on 27 February
478 2015 generated a measured 120% water loss, together with simulated values of 48, 148, 125
479 and 75% water loss at the 15-minute, hourly, 6-hourly and daily resolutions, respectively.

480

481 The generation of water fluxes not only depends on the intensity of precipitation, but also
482 surface coverage, topology and soil physical properties of the field. In hydrology, lag time,
483 defined as the time difference between the peak runoff and mass centre of rainfall excess
484 (Hall, 1984), is usually used to determine a runoff rate. Although the SPACSYS model does
485 not use this parameter to estimate the surface runoff rate, it uses the Richard's equation to
486 calculate soil water redistribution where soil hydraulic conductivity, saturated water content
487 and plant uptake play critical roles in water infiltration and consequently surface runoff. A
488 trial study on the spatial variation of soil hydraulic conductivity (unpublished data) in a
489 NWFP field, nearby to the study field, highlighted clear within-field variation, partially
490 because of compaction caused by grazed animal movement. However, the soil physical
491 properties used in the simulations were estimated based on soil texture, and at the field-scale
492 only. The error and uncertainty introduced by this approach are likely to be transferred to the

493 errors in simulating infiltration and surface runoff rates. To improve model simulation
494 accuracy, soil physical properties as core information should be provided wherever possible.

495

496 *4.4 Further considerations of scale*

497 The processes controlling water fluxes operate across a range of spatial and temporal scales,
498 and the time-series that are recorded represent an aggregation of these effects. For example,
499 effects of evapotranspiration will dominate at annual scales whereas more local impacts of
500 precipitation will manifest at finer scales (Rust et al., 2014). As noted above, the
501 characteristics of the study area, (e.g. size, soil condition and topography) will impact the
502 dominant scales of variation and hence the frequency at which it is most appropriate to model
503 or measure water fluxes. Rust et al. (2014) presented an analysis which aimed to determine
504 whether the process-based model they studied captured the scale-dependent variation
505 measured in catchment runoff. They analysed model residuals using wavelet-based signal
506 processing methods and found that although their model captured broadly the scale-
507 dependent variation in the data, fine scale variation was always under-predicted. Our results
508 shed light on their observation as we confirm that if the scale of the model prediction is not
509 sufficiently fine then model-damping will result in an under-prediction of extreme events.

510

511 **5 Conclusions**

512 For the grassland study site, the adapted process-based model (SPACSYS) could adequately
513 simulate the trends in measured water fluxes and identify their extremes. At a daily time-step,
514 model accuracy increased when simulations were run at finer temporal resolutions,
515 specifically 15-minute and hourly, and then aggregated to daily (a coarse output resolution

516 commonly used in field-scale agricultural settings). Aggregating using 6-hourly simulations
517 was less accurate. For the study site, which constitutes a field of a grassland research farm
518 platform (NWFP), simulation of water fluxes at an hourly resolution is likely optimal since
519 use of the 15-minute resolution did not increase prediction accuracy or the ability to identify
520 extremes in flow further. Therefore, for modelling purposes, monitoring frequency could be
521 reduced safely to hourly from the current 15-minute resolution.

522

523 Results provide information not only for the NWFP experiment, but also and indirectly, the
524 UK grassland farming regions that its outputs upscale to (Pulley and Collins, 2019). Study
525 results are crucial in relation to meeting the increasing demand for reliable simulation-based
526 runoff forecasts at daily and sub-daily resolutions, where accurate knowledge of peak
527 discharge and stage are essential not only for flood protection, but also to help increase the
528 forecast accuracy of associated emissions such as nutrient or sediment loss, that each use
529 water flux as a component. Further research is called for in specifying the temporal resolution
530 amongst the wide range of field-scale hydrological/agricultural models currently applied.
531 This needs to be coupled with linked changes in climate and land use to increase model
532 forecast accuracy and to optimise data acquisition schemes on farms generally.

533

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References

- Abalos, D., Cardenas, L.M., Wu, L., 2016. Climate change and N₂O emissions from South West England grasslands: a modelling approach. *Atmos. Environ.*, 132: 249-257. DOI:<https://doi.org/10.1016/j.atmosenv.2016.03.007>
- Adimassu, Z., Alemu, G., Tamene, L., 2019. Effects of tillage and crop residue management on runoff, soil loss and crop yield in the Humid Highlands of Ethiopia. *Agric. Syst.*, 168: 11-18. DOI:<https://doi.org/10.1016/j.agsy.2018.10.007>
- Ahuja, L.R., Ma, L., Howell, T.A., 2002. *Agricultural System Models in Field Research and Technology Transfer*. CRC Press, Boca Raton.
- Alaoui, A., Rogger, M., Peth, S., Blöschl, G., 2018. Does soil compaction increase floods? A review. *J. Hydrol.*, 557: 631-642. DOI:<https://doi.org/10.1016/j.jhydrol.2017.12.052>
- Archer, D.R., Fowler, H.J., 2018. Characterising flash flood response to intense rainfall and impacts using historical information and gauged data in Britain. *J. Flood Risk Manag.*, 11: S121-S133. DOI:<https://doi.org/10.1111/jfr3.12187>
- Bingham, I.J., Wu, L., 2011. Simulation of wheat growth using the 3D root architecture model SPACSYS: validation and sensitivity analysis. *Eur. J. Agron.*, 34: 181-189. DOI:<https://doi.org/10.1016/j.eja.2011.01.003>
- Booij, M.J., Tran, H.M., 2005. Appropriate scales for hydro-climatological variables in the Red River basin. In: Wagener, T. et al. (Eds.), *Proceedings of symposium S6 held during the Seventh IAHS Scientific Assembly*. IAHS publication. IAHS Press, pp. 325-332.
- Brown, I., Thompson, D., Bardgett, R., Berry, P., Crute, I., Morison, J., Morecroft, M., Pinnegar, J., Reeder, T., Topp, K., 2016. *UK Climate Change Risk Assessment Evidence Report: Chapter 3, Natural Environment and Natural Assets*. Report Prepared for the Adaptation Sub-committee of the Committee on Climate Change, London.
- Carswell, A.M., Gongadze, K., Misselbrook, T.H., Wu, L., 2019. Impact of transition from permanent pasture to new swards on the nitrogen use efficiency, nitrogen and carbon budgets of beef and sheep production. *Agric. Ecosyst. Environ.*, 283: 106572. DOI:<https://doi.org/10.1016/j.agee.2019.106572>
- Charlton, M.B., Bailey, A., Arnell, N., 2010. *Water for Agriculture – Implications for Future Policy and Practice Report for the Royal Agricultural Society of England*.
- Choi, Y.S., Shin, M.-J., Kim, K.T., 2018. Preliminary study of computational time steps in a physically based distributed rainfall-runoff model. *Water*, 10: 1269. DOI:<https://doi.org/10.3390/w10091269>
- Cleveland, W.S., 1979. Robust locally weighted regression and smoothing scatterplots. *J. Am. Stat. Assoc.*, 74: 829-836. DOI:<https://doi.org/10.1080/01621459.1979.10481038>
- Committee on Climate Change, 2017. *UK Climate Change Risk Assessment 2017. Synthesis Report*, London.
- Curceac, S., Atkinson, P.M., Milne, A., Wu, L., Harris, P., 2020. An evaluation of automated GPD threshold selection methods for hydrological extremes across different scales. *J. Hydrol.*, 585: 124845. DOI:<https://doi.org/10.1016/j.jhydrol.2020.124845>
- Dadson, S.J., Hall, J.W., Murgatroyd, A., Acreman, M., Bates, P., Beven, K., Heathwaite, L., Holden, J., Holman, I.P., Lane, S.N., O'Connell, E., Penning-Rowsell, E., Reynard, N., Sear, D., Thorne, C., Wilby, R., 2017. *A restatement of the natural science evidence concerning catchment-based*

583 'natural' flood management in the UK. Proc. R. Soc. A, 473: 20160706.
584 DOI:<https://doi.org/10.1098/rspa.2016.0706>

585 Del Grosso, S.J., Ojima, D.S., Parton, W.J., Stehfest, E., Heistemann, M., DeAngelo, B., Rose, S., 2009.
586 Global scale DAYCENT model analysis of greenhouse gas emissions and mitigation strategies
587 for cropped soils. Glob. Planet. Change, 67: 44-50.
588 DOI:<https://doi.org/10.1016/j.gloplacha.2008.12.006>

589 Department for Environment, Food and Rural Affairs, 2019. Agriculture in the United Kingdom 2018,
590 London.

591 Environment Agency, 2009. Flooding in England: A National Assessment of Flood Risk, Bristol, UK.

592 Ficchi, A., Perrin, C., Andréassian, V., 2016. Impact of temporal resolution of inputs on hydrological
593 model performance: An analysis based on 2400 flood events. J. Hydrol., 538: 454-470.
594 DOI:<https://doi.org/10.1016/j.jhydrol.2016.04.016>

595 Fox, J., 2016. Applied Regression Analysis and Generalized Linear Models (Third Edition). SAGE
596 Publications Inc., California.

597 Hall, M.J., 1984. Urban Hydrology. Elsevier Applied Science Publishers, London.

598 Harrod, T.R., Hogan, D.V., 2008. The soils of North Wyke and Rowden, North Wyke Research, North
599 Wyke, Devon.

600 Huang, J., Kang, Q., Yang, J.X., Jin, P.W., 2017. Multifactor analysis and simulation of the surface runoff
601 and soil infiltration at different slope gradients. IOP Conf. Ser. Earth Environ. Sci., 82: 012019.
602 DOI:<https://doi.org/10.1088/1755-1315/82/1/012019>

603 Huang, Y., Bárdossy, A., Zhang, K., 2019. Sensitivity of hydrological models to temporal and spatial
604 resolutions of rainfall data. Hydrol. Earth Syst. Sc., 23: 2647-2663.
605 DOI:<https://doi.org/10.5194/hess-23-2647-2019>

606 Jansson, P.-E., 1998. Simulation model for soil water and heat conditions. Description of the SOIL model.
607 Division of Agricultural Hydrotechnics Communications 98:2, Swedish University of
608 Agricultural Sciences, Uppsala.

609 Jeong, J., Kannan, N., Arnold, J., Glick, R., Gosselink, L., Srinivasan, R., 2010. Development and
610 integration of sub-hourly rainfall-runoff modeling capability within a watershed model. Water
611 Resour. Manag., 24: 4505-4527. DOI:<https://doi.org/10.1007/s11269-010-9670-4>

612 Kavetski, D., Fenicia, F., Clark, M.P., 2011. Impact of temporal data resolution on parameter inference
613 and model identification in conceptual hydrological modeling: Insights from an experimental
614 catchment. Water Resour. Res., 47: W05501. DOI: <https://doi.org/10.1029/2010WR009525>

615 Keesstra, S., Nunes, J., Novara, A., Finger, D., Avelar, D., Kalantari, Z., Cerdà, A., 2018. The superior
616 effect of nature based solutions in land management for enhancing ecosystem services. Sci.
617 Total Environ., 610-611: 997-1009. DOI:<https://doi.org/10.1016/j.scitotenv.2017.08.077>

618 Kibet, L.C., Saporito, L.S., Allen, A.L., May, E.B., Kleinman, P.J.A., Hashem, F.M., Bryant, R.B., 2014. A
619 protocol for conducting rainfall simulation to study soil runoff. J. Vis. Exp.: e51664.
620 DOI:<https://doi.org/10.3791/51664>

621 Ledingham, J., Archer, D., Lewis, E., Fowler, H., Kilsby, C., 2019. Contrasting seasonality of storm rainfall
622 and flood runoff in the UK and some implications for rainfall-runoff methods of flood
623 estimation. Hydrol. Res., 50: 1309-1323. DOI:<https://doi.org/10.2166/nh.2019.040>

624 Liu, Y., Li, Y., Harris, P., Cardenas, L., Dunn, R.M., Sint, H., Murray, P., Lee, M., Wu, L., 2018. Modelling
625 field scale spatial variation in water run-off, soil moisture, N₂O emissions and herbage biomass
626 of a grazed pasture using the SPACSYS model. Geoderma, 315: 49-58.

- 627 DOI:<https://doi.org/10.1016/j.geoderma.2017.11.029>
- 628 Liu, Y., Wu, L., Watson, C.A., Baddeley, J.A., Pan, X., Zhang, L., 2013. Modeling biological dinitrogen
629 fixation of field pea with a process-based simulation model. *Agron. J.*, 105: 670-678.
630 DOI:<https://doi.org/10.2134/agronj2012.0412>
- 631 Macleod, C.J.A., Humphreys, M.W., Whalley, W.R., Turner, L., Binley, A., Watts, C.W., Skot, L., Joynes, A.,
632 Hawkins, S., King, I.P., O'Donovan, S., Haygarth, P.M., 2013. A novel grass hybrid to reduce flood
633 generation in temperate regions. *Sci. Rep.*, 3: 1683. DOI:<https://doi.org/10.1038/srep01683>
- 634 Merz, B., Plate, E.J., 1997. An analysis of the effects of spatial variability of soil and soil moisture on
635 runoff. *Water Resour. Res.*, 33: 2909-2922. DOI:<https://doi.org/10.1029/97WR02204>
- 636 Merz, R., Parajka, J., Blöschl, G., 2009. Scale effects in conceptual hydrological modeling. *Water Resour.*
637 *Res.*, 45: W09405. DOI: <https://doi.org/10.1029/2009WR007872>
- 638 Morison, J.I.L., Matthews, R.B., 2016. Agriculture and forestry climate change impacts summary report.
639 DOI:<https://nerc.ukri.org/research/partnerships/ride/lwec/report-cards/agriculture/>
- 640 Newell-Price, P., Chambers, B., Whittingham, M., 2011. The alleviation of grassland compaction by
641 mechanical soil loosening. Defra Project BD5001 Final Report.
- 642 Newell Price, B., Chambers, B., Whittingham, M., 2012. Characterisation of Soil Structural Degradation
643 Under Grassland and Development of Measures to Ameliorate its Impact on Biodiversity and
644 other Soil Functions. Defra Project BD5001 Final Report.
- 645 Orr, R., Murray, P., Eyles, C., Blackwell, M., Cardenas, L., Collins, A.L., Dungait, J., Goulding, K., Griffith,
646 B., Gurr, S., Harris, P., Hawkins, J., Misselbrook, T., Rawlings, C., Shepherd, A., Sint, H., Takahashi,
647 T., Tozer, K., Whitmore, A.P., Wu, L., Lee, M., 2016. The North Wyke Farm Platform: effect of
648 temperate grassland farming systems on soil moisture contents, runoff and associated water
649 quality dynamics. *Eur. J. Soil Sci.*, 67: 374-385. DOI:<https://doi.org/10.1111/ejss.12350>
- 650 Palmer, R.C., Smith, R.P., 2013. Soil structural degradation in SW England and its impact on surface-
651 water runoff generation. *Soil Use Manag.*, 29: 567-575.
652 DOI:<https://doi.org/10.1111/sum.12068>
- 653 Pathak, P., Sudi, R., Wani, S.P., Sahrawat, K.L., 2013. Hydrological behavior of Alfisols and Vertisols in
654 the semi-arid zone: Implications for soil and water management. *Agric. Water Manag.*, 118:
655 12-21. DOI:<https://doi.org/10.1016/j.agwat.2012.11.012>
- 656 Perego, A., Wu, L., Gerosa, G., Finco, A., Chiazzese, M., Amaducci, S., 2016. Field evaluation combined
657 with modelling analysis to study fertilizer and tillage as factors affecting N₂O emissions: a case
658 study in Po valley (Northern Italy) *Agric. Ecosyst. Environ.*, 225: 72-85.
659 DOI:<https://doi.org/10.1016/j.agee.2016.04.003>
- 660 Pulley, S., Collins, A.L., 2019. Field-based determination of controls on runoff and fine sediment
661 generation from lowland grazing livestock fields. *J. Environ. Manage.*, 249: 109365.
662 DOI:<https://doi.org/10.1016/j.jenvman.2019.109365>
- 663 Rust, W., Corstanje, R., Holman, I.P., Milne, A.E., 2014. Detecting land use and land management
664 influences on catchment hydrology by modelling and wavelets. *J. Hydrol.*, 517: 378-389.
665 DOI:<https://doi.org/10.1016/j.jhydrol.2014.05.052>
- 666 Schaller, N., Sillmann, J., Müller, M., Haarsma, R., Hazeleger, W., Hegdahl, T.J., Kelder, T., van den Oord,
667 G., Weerts, A., Whan, K., 2020. The role of spatial and temporal model resolution in a flood
668 event storyline approach in western Norway. *Weather and Climate Extremes*, 29: 100259.
669 DOI:<https://doi.org/10.1016/j.wace.2020.100259>
- 670 Sharma, A., Wasko, C., Lettenmaier, D.P., 2018. If precipitation extremes are increasing, why aren't
671 floods? *Water Resour. Res.*, 54: 8545-8551. DOI:<https://doi.org/10.1029/2018wr023749>

672 Stevens, A.J., Clarke, D., Nicholls, R.J., 2016. Trends in reported flooding in the UK: 1884–2013.
673 Hydrolog. Sci. J., 61: 50-63. DOI:<https://doi.org/10.1080/02626667.2014.950581>

674 Verstraten, L., Wasko, C., Ashford, G., Sharma, A., 2019. Sensitivity of Australian roof drainage
675 structures to design rainfall variability and climatic change. Build. Environ., 161: 106230.
676 DOI:<https://doi.org/10.1016/j.buildenv.2019.106230>

677 Wasko, C., Sharma, A., Lettenmaier, D.P., 2019. Increases in temperature do not translate to increased
678 flooding. Nat. Commun., 10: 5676. DOI:<https://doi.org/10.1038/s41467-019-13612-5>

679 Watts, G., Anderson, M., 2016. Water climate change impacts report card 2016.
680 DOI:<https://nerc.ukri.org/research/partnerships/ride/lwec/report-cards/water/>

681 Wu, L., Bingham, I.J., Baddeley, J.A., Watson, C.A., 2009. Modeling plant nitrogen uptake using three-
682 dimensional and one-dimensional root architecture. In: Ma, L., Ahuja, L.R., Bruulsema, T.W.
683 (Eds.), Quantifying and understanding plant nitrogen uptake systems modeling. CRC Press,
684 Boca Raton, FL, pp. 197-218.

685 Wu, L., Blackwell, M., Dunham, S., Hernández-Allica, J., McGrath, S.P., 2019. Simulation of phosphorus
686 chemistry, uptake and utilisation by winter wheat. Plants, 8: 404.
687 DOI:<https://doi.org/10.3390/plants8100404>

688 Wu, L., McGechan, M.B., McRoberts, N., Baddeley, J.A., Watson, C.A., 2007. SPACSYS: integration of a
689 3D root architecture component to carbon, nitrogen and water cycling - model description.
690 Ecol. Model., 200: 343-359. DOI:<https://doi.org/10.1016/j.ecolmodel.2006.08.010>

691 Wu, L., Rees, R.M., Tarsitano, D., Zhang, X., Jones, S.K., Whitmore, A.P., 2015. Simulation of nitrous
692 oxide emissions at field scale using the SPACSYS model. Sci. Total Environ., 530–531: 76-86.
693 DOI:<http://dx.doi.org/10.1016/j.scitotenv.2015.05.064>

694 Wu, L., Zhang, X., Griffith, B.A., Misselbrook, T., 2016. Sustainable grassland systems: A modelling
695 perspective based on the North Wyke Farm Platform. Eur. J. Soil Sci., 67: 397-408.
696 DOI:<https://doi.org/10.1111/ejss.12304>

697 Zehe, E., Graeff, T., Morgner, M., Bauer, A., Bronstert, A., 2010. Plot and field scale soil moisture
698 dynamics and subsurface wetness control on runoff generation in a headwater in the Ore
699 Mountains. Hydrol. Earth Syst. Sci., 14: 873-889. DOI:[https://doi.org/10.5194/hess-14-873-
700 2010](https://doi.org/10.5194/hess-14-873-2010)

701 Zhang, X., Sun, N., Wu, L., Xu, M., Bingham, I.J., Li, Z., 2016. Effects of enhancing soil organic carbon
702 sequestration in the topsoil by fertilization on crop productivity and stability: evidence from
703 long-term experiments with wheat-maize cropping systems in China. Sci. Total Environ., 562:
704 247-259. DOI:<https://doi.org/10.1016/j.scitotenv.2016.03.193>

705

706 **Figure captions**

707 Figure 1. Details of the NWFP catchment selected for this study (catchment number 6 of 15,
708 consisting of a single field called Golden Rove).

709 Figure 2. Time-series plots for measured and simulated water flux data (not aggregated) for
710 15-minute, hourly, 6-hourly and daily data (in units of $\text{mm } 15\text{min}^{-1}$, mm h^{-1} , $\text{mm } 6\text{h}^{-1}$
711 and mm d^{-1} , respectively). All plots are shown with a threshold at the 99th percentile
712 of measured data (at $0.138 \text{ mm } 15\text{min}^{-1}$, 0.553 mm h^{-1} , $3.45 \text{ mm } 6\text{h}^{-1}$ and 14.9 mm d^{-1} ,
713 respectively).

714 Figure 3. Scatterplots of the measured and simulated data (not aggregated) for 15-minute,
715 hourly, 6-hourly and daily data. Scatterplots are shown with the 1:1 line, a linear
716 regression fit and a loess smoother fit. Units are in $\text{mm } 15\text{min}^{-1}$, mm h^{-1} , $\text{mm } 6\text{h}^{-1}$ and
717 mm d^{-1} , respectively.

718 Figure 4. Time-series plots for daily measured and daily simulated water flux data (with the
719 first three plots having data aggregated from: 15 minutes to daily; hourly to daily; 6
720 hourly to daily). All units in mm d^{-1} . All plots are shown with a threshold at the 99th
721 percentile of measured data (14.90 mm d^{-1}).

722 Figure 5. Scatterplots of the daily measured and daily simulated data (with the first three plots
723 having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily).

724 Scatterplots are shown with the ideal 1:1 line, a linear regression fit and a loess
725 smoother fit. All units in mm d^{-1} .

726 Figure 6. Density plots for daily measured and daily simulated data (with the first three plots
727 having data aggregated from: 15 minutes to daily; hourly to daily; 6 hourly to daily).
728 All units in mm d^{-1} .

729 Figure 7. Error (MAE, NRMSE, PBIAS) indices with respect to daily measured and daily
730 simulated data (with 15-minute, hourly, 6-hourly data aggregated to daily).

731 Figure 8. Time-series of errors (simulated minus measured data) aggregated to the daily
732 temporal resolution. All units in mm d^{-1} . Positive errors represent over-prediction by
733 the model.

734 Figure 9. Agreement (NSE, d , KGE) indices with respect to daily measured and daily
735 simulated data (with 15-minute, hourly, 6-hourly simulations aggregated to daily).
736