

1 **The need to integrate legacy nitrogen storage dynamics and time lags into policy and practice**

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## 26 Highlights

- 27 • Nitrogen (N) pollution from agriculture has negative environmental impacts
- 28 • Environmental benefits of initiatives to reduce N loads not always detectable
- 29 • N storage dynamics and time lag invalidate steady state models often used in policy
- 30 • Researchers should advocate for integrating N stores and time lags into policy
- 31 • Quantifying N storage aligns with phosphorus and carbon cycling research

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## 33 Abstract

34  
35 Increased fluxes of reactive nitrogen ( $N_r$ ), often associated with N fertilizer use in agriculture, have  
36 resulted in negative environmental consequences, including eutrophication, which cost billions of  
37 dollars per year globally. To address this, best management practices (BMPs) to reduce  $N_r$  loading to  
38 the environment have been introduced in many locations globally. However, improvements in water  
39 quality associated with BMP implementation have not always been realised over expected timescales.  
40 There is now a significant body of scientific evidence showing that the dynamics of legacy  $N_r$  storage  
41 and associated time lags invalidate the assumptions of many models used by policymakers for decision  
42 making regarding  $N_r$  BMPs. Building on this evidence, we believe that the concepts of legacy  $N_r$  storage  
43 dynamics and time lags need to be included in these models. We believe the biogeochemical research  
44 community could play a more proactive role in advocating for this change through both awareness  
45 raising and direct collaboration with policymakers to develop improved datasets and models. We  
46 anticipate that this will result in more realistic expectations of timescales for water quality  
47 improvements associated with BMPs. Given the need for multi-nutrient policy responses to tackle  
48 challenges such as eutrophication, integration of N stores will have the further benefit of aligning both  
49 researchers and policymakers in the N community with the phosphorus and carbon communities,  
50 where estimation of stores is more widespread. Ultimately, we anticipate that integrating legacy  $N_r$

51 storage dynamics and time lags into policy frameworks will better meet the needs of human and  
52 environmental health.

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## 54 [Keywords](#)

55 Nitrogen, legacy pollution, water pollution, time lag

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## 57 [1 Introduction](#)

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59 Nitrogen (N) is an essential macronutrient, fundamental for growth in both plants and animals  
60 (Schlesinger, 2005). Agricultural intensification and associated N fertilizer use has underpinned the  
61 world's growing population, resulting in a doubling of reactive N ( $N_r$ ) fluxes in the environment  
62 (Vitousek et al., 1997). Increased  $N_r$  fluxes have generated negative consequences for both human  
63 and environmental health, leading to costs associated with eutrophication and drinking water  
64 treatment alone in the billions of dollars per year (Dodds et al., 2009; House of Commons  
65 Environmental Audit Committee, 2018; Pretty et al., 2000).

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67 In response to the ecological impacts of increased  $N_r$  fluxes, best management practices (BMPs) have  
68 been implemented to reduce  $N_r$  fluxes in catchments. Some studies have shown BMPs to reduce  
69 nutrient export at the field to plot scale (Liu et al., 2017). However, at the catchment to basin scale,  
70 in many cases, the anticipated benefits of work to reduce  $N_r$  fluxes have not been realised (Hamilton,  
71 2012; Van Meter et al., 2018). For example, despite millions of dollars spent on implementation of  
72 best management practices (BMPs) to reduce  $N_r$  loadings from agricultural sources, the Gulf of Mexico  
73 hypoxic zone was the largest ever recorded in 2017, with the target date to reduce the size of the dead  
74 zone delayed to 2035. These observations at the catchment scale emphasise the need for the scientific  
75 community to address the apparent disconnect between action and environmental benefit in the case  
76 of  $N_r$ .

## 77 2 Disconnect between action and benefit at the catchment scale: 78 evidence for legacy $N_r$ storage dynamics and time lags 79

80 What is causing the apparent disconnect between actions and catchment scale benefits in the case of  
81  $N_r$ , despite some observations of benefits at the local scale? There is now a compelling body of  
82 scientific evidence from both field and modelling research that demonstrates legacy  $N_r$  storage in  
83 different compartments of the environment. Entry and subsequent release of  $N_r$  from these stores  
84 can result in significant time lags in the environmental benefits of actions designed to reduce new  $N_r$   
85 loads to the environment. The dynamics of legacy nitrogen storage and impacts of  $N_r$  release from  
86 stores on water quality have been shown to be significant in Europe (Ascott et al., 2016; Bell et al.,  
87 2021; Durand et al., 2011; Howden et al., 2011; Vero et al., 2018; Wang et al., 2016; Worrall et al.,  
88 2015), Asia (Jia et al., 2018; Turkeltaub et al., 2020; Wu et al., 2020; Wu et al., 2019), North America  
89 (Ator et al., 2020; Martin et al., 2021; Sprague et al., 2011; Tesoriero et al., 2013; Van Meter et al.,  
90 2016; Van Meter et al., 2018) and globally (Ascott et al., 2017; Chen et al., 2018; McCrackin et al.,  
91 2017; Xin et al., 2019). In the past delays in meeting water quality objectives due to time lags and  
92 legacy storage dynamics have been dismissed as a generic excuse (Schaure and Naus, 2010). More  
93 recently, however, policymakers are increasingly aware of the role of legacy storage in controlling the  
94 efficacy of BMPs at the catchment scale (e.g House of Commons Environmental Audit Committee  
95 (2018); Meals et al. (2010); Stuart et al. (2016)).

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97 Whilst there is now strong evidence for legacy  $N_r$  storage dynamics and increasing awareness of this  
98 amongst policymakers, a major challenge remains in how nutrient legacies are represented in models  
99 and budgets used in practice for decision making. A number of conventional modelling tools that  
100 inform policy and practice that underpins N management at the catchment scale invoke the steady  
101 state assumption (e.g. SPARROW, PoIFLOW, SAGIS, SEPARATE, NEAP-N, see Chen et al. (2018) for a  
102 recent summary of approaches). These models have been used to make decisions regarding control  
103 of  $N_r$  sources in the environment in order to reduce the risk of environmental damage, alongside

104 predicting the trajectory for recovery of the environment where impact has already occurred.  
105 Interventions made on the basis of these tools have not always been successful over predicted  
106 timescales, with time lags associated with legacy storage dynamics invalidating the steady state  
107 assumption over short (<50 year) timescales. There are also discrepancies between research and  
108 practice regarding the definition of the term 'store', with some practitioner studies (United States  
109 Environmental Protection Agency, 2011) reporting a store as flux, whilst the academic research  
110 community often deals with stores in terms of mass (Chen et al., 2018; Van Meter et al., 2016).

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### 112 3 The need for policy advocacy by the biogeochemical research 113 community

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115 Based on the body of scientific evidence highlighted above, we argue that the biogeochemical  
116 research community could play a more proactive role in advocating for integration of legacy storage  
117 dynamics and time lags into N<sub>r</sub> management strategies in policy and practice (Figure 1). We envisage  
118 that this would consist of both awareness raising and direct collaboration to develop the next  
119 generation of datasets and models to support decision making regarding BMPs.

#### 120 3.1 Awareness raising

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122 Whilst there is now some understanding in the policymaking community about the importance of  
123 legacy storage dynamics, we believe that researchers should continue to raise awareness of the issue,  
124 particularly amongst practitioners working in areas where implementation of BMPs is relatively recent  
125 and rapid improvements may be desired. We envisage that researchers could have direct engagement  
126 and discussions with policymakers, contributions to government enquiries, committees (e.g. Ascott  
127 and Ward (2018)) and evidence syntheses. Engagement at the local and regional level with key  
128 stakeholders (e.g. farmers, agri-environmental community groups) may also be beneficial.

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### 134 3.2 Data and model development

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136 Beyond awareness raising, we believe researchers should collaborate directly with policymakers to

137 develop the next generation of datasets and models to support BMP decision making. Initial

138 requirements for such collaboration would be to ensure a consistent terminology across both research

139 and practice regarding stores (e.g. as a mass in kg N), and sharing of existing models and datasets used

140 in N biogeochemical research with practitioners. Historic monitoring networks have often been poorly

141 set up to address legacy storage dynamics and associated time lags (England et al., 2008; Hamilton,

142 2012), and reviews of impacts of BMPs at the meso-scale have highlighted the need for long term

143 monitoring to assess water quality changes (Melland et al., 2018). Development of co-designed

144 monitoring networks that quantify long term fluxes to and from  $N_r$  stores and their magnitude would

145 be beneficial. For example, this could consist of porewater profiles in the unsaturated zone and soil

146  $N$  storage measurements, repeated every 5-10 years. Such monitoring would quantify reductions in

147 the magnitude of these  $N_r$  stores and provide the initial evidence that changes in management

148 practices designed to control  $N_r$  fluxes are having the desired effect. This would provide a sentinel

149 indicator of potential future changes in downstream components of the terrestrial water cycle.

150 Comparing the magnitude of different  $N_r$  stores could indicate the relative impacts of anthropogenic

151 activities on different components of the terrestrial environment such as soils, the unsaturated zone,

152 groundwater and riparian sediments. For example, large  $N_r$  storage in the unsaturated zone suggests

153 that future  $N_r$  concentration changes in linked receptors (i.e. groundwater and surface water) will

154 continue to be significantly affected by release of  $N_r$  from this store, before any impacts from changes

155 in soil  $N_r$  leaching associated with recent changes in management practices are detected in the

156 ultimate receptor. By combining consistent terminology, sharing of existing models, and improved

157 monitoring networks, we believe that researchers can support the development of new modelling

158 frameworks used in policy to provide better predictions of catchment nutrient trajectories and

159 timescales.

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### 3.3 An example from England (UK)

What would a proactive advocacy role for the biogeochemical research community look like in practice? Approaches to the integration of legacy N<sub>r</sub> storage dynamics and time lags into policy would need to be informed by dialogue between researchers and practitioners to identify discrepancies between the state of the science and models and tools used in policy within a particular setting. To illustrate the potential opportunities, here we provide an example of both awareness raising and data and model development from England (UK). In England researchers have raised awareness of the significance of legacy N storage dynamics to policymakers, national government and parliamentarians (Ascott and Ward, 2018). The methodology used to designate agricultural land in which N application may be restricted (known as Nitrate Vulnerable Zones (European Union, 1991)) is reviewed every four years in England. In the latest review in 2020, time lags between nitrate leaching from the base of the soil zone and changes in nitrate concentrations in groundwater are being considered in the methodology using outputs of previous modelling of unsaturated zone travel times by Wang et al. (2012) (Hart and Kieboom, pers. comm.).

## 4 Synergy across macronutrient cycles

Better integration of time lags and legacy N<sub>r</sub> stores would also align researchers and policymakers in the N community with those in the phosphorus (P) and carbon (C) communities. Successfully addressing challenges such as eutrophication requires policy responses that are coordinated across multiple nutrient elements (Conley et al., 2009; Harpole et al., 2011). However, different conceptual frameworks currently pervade across N, P and C communities. For example, P and C communities often more explicitly quantify the magnitude of stores compared to the N community. For P this is primarily due to issues of resource availability associated with finite resources of mineral phosphate rocks (Elser et al., 2014) and soil stores for agriculture (Haygarth et al., 2014; Sattari et al., 2012).

187 Consequently large-scale P budgets have been developed using substance flow analysis (SFA) methods  
188 and the principles of mass balance to calculate the absolute magnitude of a number of P stores (Chen  
189 and Graedel, 2016; Yuan et al., 2018). For C the quantification of the magnitude of stores is associated  
190 with climate change, with global scale budgets synthesizing fluxes and stores from a range of both  
191 observational and modelled data sources (Le Quéré et al., 2014). Whilst  $N_r$  is drawn from a large and  
192 renewable resource of atmospheric  $N_2$  (Erisman et al., 2008), the evidence for legacy  $N_r$  in the  
193 environment highlights the need to quantify  $N_r$  stores in the terrestrial environment. Whilst fluxes  
194 from agricultural systems are the primary source of  $N_r$  to freshwater systems (Fowler et al., 2013), the  
195 same principles of time lag and stores apply to other sources (e.g. contaminated land, sewer leakage  
196 (Wakida and Lerner, 2005), mains leakage (Ascott et al., 2018)).

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## 198 5 Concluding remarks

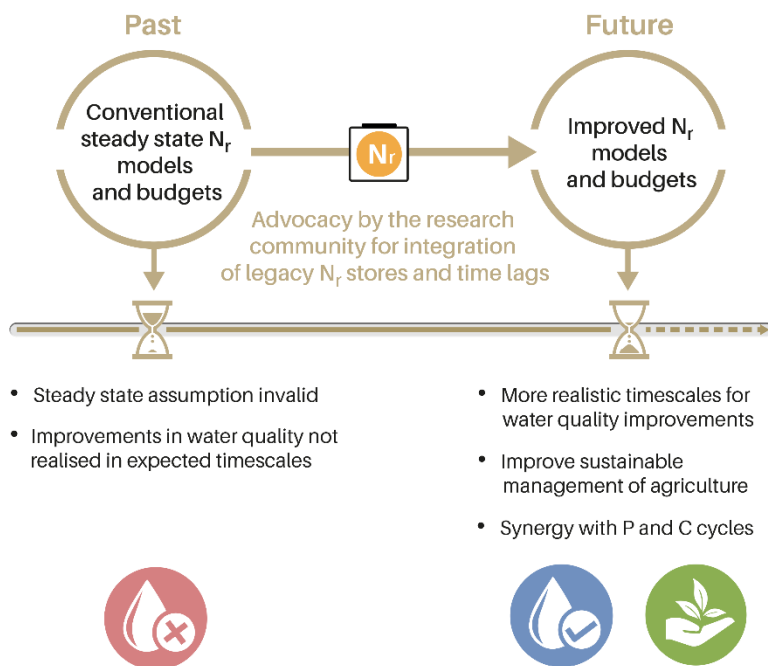
199

200 Despite a strong body of scientific evidence and increasing awareness amongst stakeholders, models  
201 and budgets used by policymakers in BMP planning often do not adequately represent legacy  $N_r$   
202 dynamics and associated time lags. Here we argue that the biogeochemical research community  
203 needs to proactively advocate for integration of time lags into future  $N_r$  management strategies  
204 through awareness raising and data and model development. This would support more realistic  
205 estimates of the trajectories of change following measures to reduce  $N_r$  loads, managing the  
206 expectations of stakeholders and supporting long term sustainable agriculture. Incorporating  $N_r$   
207 stores and time lags into improved models and budgets used in policy and regulatory frameworks for  
208 the sustainable management of agriculture can better meet the needs of human health and the  
209 environment.

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213 *Figure 1 Past and potential future approaches to management of legacy  $N_r$ , including the role of the research community to*

214 *advocate for integration of legacy  $N_r$  stores and time lags into policy and practice*

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