1	Urban soil microbial community and microbial-related carbon storage are severely limited by sealing
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18	Abstract
19	Purpose Urbanisation causes changes in land use, from natural or rural to urban, leading to the sealing of soil
20	and the replacement of vegetation by buildings, roads and pavements. The sealing process impacts soil
21	properties and services and can lead to negative consequences for microbial attributes and processes in soil. At
22	present, information about the microbial community following soil sealing is limited. As such, we investigated
23	how changes in soil physical and chemical properties caused by sealing affect the soil microbial community and
24	soil ecosystem services.
25	Material and methods Soils were sampled beneath impervious pavements (sealed) and from adjacent pervious
26	greenspace areas (unsealed). Soil properties (total C, total N, C:N ratio and water content) and microbial
27	attributes (microbial biomass C, N-mineralisation and phospholipid fatty acids - PLFA) were measured and
28	correlated.
29	Results and discussion A reduction of total C, total N and water content were observed in sealed soil, while the
30	C:N ratio increased. Sealed soil also presented a reduction in microbial attributes, with low N-mineralisation
31	revealing suppressed microbial activity. PLFA data presented positive correlations with total C, total N and
32	water content, suggesting that the microbial community may be reduced in sealed soil as a response to soil
33	properties. Furthermore, fungal:bacterial and gram-positive:gram-negative bacterial ratios were lower in sealed
34	soil indicating degradation in C sequestration and a consequential effect on C storage.
35	Conclusions Sealing causes notable changes in soil properties leading to subsequent impacts upon the microbial
36	community and the reduction of microbial activity and soil C storage potential.
37	
38	Keywords
39	Urban soil, Soil sealing, Impervious surfaces, Microbial biomass, N-mineralization, PLFA, Soil carbon, Carbon

40 storage

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87 1 Introduction

88 Urbanisation causes considerable impacts on soil properties and services (Yan et al. 2015, 2016).

89 Changes in land use from natural and rural to urban are associated with the replacement of vegetation by

90 buildings, roads and pavements (Edmondson et al. 2012, 2015; Yan et al. 2016). The high degree of

91 impermeable surfaces in cities has many negative consequences for the environment and the services it

- 92 provides, particularly those provided by soil (Morgenroth et al. 2013; Wei et al. 2013; Piotrowska and
- 93 Charzynski 2015; Ziter and Turner 2018; Kelleher et al. 2020).

94 Carbon (C) storage is an important ecosystem service provide by soil in urban areas, with vegetation 95 biomass inputs and soil organic carbon (SOC) being key components of overall C storage (Edmondson et al. 96 2012; Ziter and Turner 2018). Soil sealing due to urbanisation leads to the removal of plants and topsoil during 97 the paving and construction process. This results not only in large losses of C stocks from urban soil (Wei et al. 98 2014), but also alters soil C dynamics, typically leading to a loss of SOC (Majidzadeh et al. 2018). Previous soil 99 C inventories suggested urban soil provides very little or no soil C storage (Bradley et al. 2005). However, more 100 recently, significant amounts of soil C have been reported in urban areas, in soils of greenspaces and beneath 101 sealed surfaces of pavements and houses (Edmondson et al 2012; Wei et al. 2014, Majidzadeh et al. 2017, Yan 102 et al. 2016, Hu et al. 2018, Vasenev et al. 2018). As such, urban soil C and the dynamics of C storage are

103 receiving increasing attention in research literature.

104 Many other key ecosystem services and soil properties are affected by soil sealing. Water infiltration is 105 prevented or reduced, changing surface runoff patterns and seasonal dynamics of soil water content (Majidzadeh et al. 2018; Hu et al. 2020; Kelleher et al. 2020). Paving materials can act as a reservoir for contaminants such 106 107 as heavy metals (Hu et al. 2018) and polycyclic aromatic hydrocarbons (Li et al. 2020); and soil temperatures 108 can be increased (Chen et al. 2016; 2017). Gas exchange between the soil and atmosphere is reduced which can 109 lead to higher CO₂ concentrations in sealed soil and increased CO₂ flux rate near pavement edges (Wu et al. 110 2016; Fini et al. 2017). Additionally, soil nutrient content can be altered, with sealed soils exhibiting increased calcium, potassium, sodium and phosphorous; and decreased aluminium, iron, magnesium and nitrogen (Zhao et 111 112 al. 2012; Morgenroth et al. 2013; Hu et al. 2018; Majidzadeh et al. 2018). The severe decrease in nitrogen (N) 113 can lead to very high CN ratios in sealed soils, despite the concurrent loss of soil C (Zhao et al. 2012, Hu et al. 114 2018).

These changes to the soil environment also affect soil microbes, which may impact the microbial processes and activities that underpin many important soil services (Zhao et al. 2012). Whilst sealed soils remain largely understudied, a small number of studies have observed that sealing can lead to a decrease in microbial biomass C, microbial biomass N, enzyme activities and respiration potential (Zhao et al. 2012, Wei et al. 2013, Piotrowska and Charzynski 2015), as well as a decrease in N-mineralisation potential (Zhao et al. 2012; Majidzadeh et al. 2018). Similarly, sealing has led to changes in bacterial communities, with a reduction in alpha diversity and a distinct community found in sealed soil when compared with unsealed soil (Hu et al.

122 2018, Yu et al. 2019). Research has shown that sealing has a negative effect on urban soil microbial attributes

123 and bacterial communities, although little is known about the dynamics of both bacterial and fungal

- 124 communities and their contribution to the soil microbial community in sealed soils. Furthermore, there is a gap
- 125 in knowledge into what these altered bacterial and fungal dynamics mean for important soil ecosystem services
- 126 such as nutrient cycling and C storage within sealed soils. Fungal:bacterial dominance is considered an

127 important factor in C sequestration (Strickland and Rousk 2010); and the ratio between gram positive:gram

- negative bacteria provides insight into the stability or recalcitrance of C in the soil (Fanin et al. 2019). At
- 129 present these dynamics have not been studied in sealed soil, and therefore the implications for soil C storage 130 across the urban landscape are currently unknown.

131 In this paper we investigate how changes to soil physical and chemical properties caused by sealing 132 affect the microbial community and microbial attributes. The city of Lancaster (UK) and surrounding urban areas were used as a study site. We measure soil properties (total C, total N, C:N ratio and water content) and 133 134 microbial attributes (microbial biomass C, phospholipid fatty acids and N-mineralisation) to make a comparison 135 across sealed and unsealed soils. To our knowledge, we present the first investigation into bacterial and fungal 136 dynamics in sealed soil using phospholipid fatty acid analysis, and consider their contributions to the soil 137 microbial community and consequences for important soil services. We hypothesise that, (i) sealing leads to 138 large changes in soil properties; and (ii) sealing leads to changes in microbial attributes, significantly altering 139 community composition and reducing microbial activity. Measurements of soil total C, total N, C:N ratio and 140 water content provided indicators of the impacts of sealing on soil properties (hypothesis 1). Microbial biomass 141 C, phospholipid fatty acids and N-mineralisation were used as indicators of changes in microbial attributes, with 142 biomass C and phospholipid fatty acids pointing to changes in community composition; and N-mineralisation to 143 changes in microbial activity (hypothesis 2).

144

145 2 Materials and methods

146 2.1 Study area

147The study area consisted of the medium-sized UK city of Lancaster and the surrounding urban areas148(Fig. 1). The National Soil Map for England and Wales, accessed on the Soilscapes viewer online (Cranfield,1492020), shows that across much of Lancaster city there are freely draining slightly acid loamy soils, while150sampling sites in the surrounding areas tended to be on slowly permeable seasonally wet acid loamy and clayey151soils.

152

153 2.2 Soil sampling

154 Sealed soils were collected from 25 roadworks sites where works had exposed the soil beneath pavements and roads. Sealing had occurred at different times in the past, and further research is still needed to 155 156 determine if the time since sealing has an impact on the measured variables. Soil was collected from the top 10 157 cm of soil below the sealed surface and human-made layers. To allow a comparison between soils, an unsealed sample was collected from the nearest available greenspace after each sealed soil was collected. Unsealed 158 159 samples were collected from the top 10 cm of soil, primarily from grass covered road verges, amenity 160 greenspaces and residential gardens, with a distance ranging from 0.5 to 15 m of the respective sealed site. 161 Approximately 500 g of both soils (50 samples) were collected with a trowel and were immediately returned to 162 the lab for refrigeration prior to fresh soil tests. 163 164 2.3 Soil preparation and analysis

165 Soil properties and CN analysis

166

Soil water content was determined gravimetrically by drying the samples at 105 °C for 24 hours. The

167 dried sample was ball milled to a powder and analysed for total C and total N using a dry combustion CN

168 analyzer (Vario Max CN).

169 Microbial biomass C and N-mineralisation

170 Microbial biomass C (MBC) was determined using the chloroform fumigation-extraction method 171 (Brookes et al. 1985; Vance et al. 1987). Two subsamples of 5 g of moisture adjusted soil were prepared for each sample, one fumigated with alcohol-free CHCl₃ for 24 hours; and one non-fumigated stored at 4 °C. After 172 173 removal of the CHCl₃, both subsamples were extracted with 25 mL of K₂SO₄ (0.5 M) for 30 minutes. The 174 filtrate was analyzed for extracted C using a TOC analyzer (Shimadzu TOC-L_{CPN} TN). 175 Soil potential N-mineralisation was measured before and after incubation. Subsamples were prepared for 176 water saturation to determine moisture adjustments for each sample. The subsamples were placed in a funnel 177 with Whatman nº 1 filter paper, wet with MilliQ water, and periodically re-wet over a 2-hour period. They were 178 then covered with cling film and drained for 2 hours, weighed, and oven dried at 105 °C for 24 hours. They 179 were re-weighed and moisture adjustments were calculated to 60 % for each sample. For extractions, 5 g of 180 moisture adjusted soil was put in an extraction bottle, covered with covered with polythene and incubated at 25 181 °C for 14 days. A second sample was extracted immediately. The incubated and non-incubated subsamples were

182 extracted using KCl (1 M), and the filtrate was analysed for inorganic N using an auto-analyzer (Elementar

183 Vario EL III). Potential N-mineralisation was calculated as the difference in inorganic N before and after

184 incubation.

185 Phospholipid fatty acid analysis

Phospholipid fatty acid (PLFA) analysis was used to determine the overall microbial community 186 187 composition and dominance. Soil subsamples were taken from soils previously stored at - 80 °C and extracted 188 for PLFA determination by gas chromatography (Vestle and White 1989; Willers et al. 2015). Microbial PLFA 189 markers were identified and measured as per the method by Frostegård et al. (2011) to estimate the total and 190 group-specific microbial marker biomass. The i15:0, a15:0, i16:0, a17:0 and i17:0 PLFA markers were used as 191 gram positive (GP) bacteria markers; and 16:1007, cy17:0, cis18:1007 and cy19:0 as gram negative (GN) bacteria 192 markers. Total bacteria were estimated from the sum of GP and GN bacteria, and 15:0 marker mass. Total fungi 193 were measured using $18:1\omega9$ and $18:2\omega6,9$ as markers. The $16:1\omega5$ was used as a proxy measurement for 194 arbuscular mycorrhizal (AM) fungi. Total PLFA expresses total microbial marker biomass and was estimated as 195 the sum of total bacteria, total fungi, AM fungi and 16:0, 16:107, br17:0, 17:108, 17:0 7-methyl, 18:0, br18:0, 196 18:1ω5 and 19:1 markers. The fungal:bacterial and GP:GN ratios were calculated by dividing the respective 197 biomarker masses.

198

199 2.4 Statistical analysis

Data were evaluated using R (version 4.0) on the software RStudio (version 1.1.463). Since only water content and total C in unsealed soil presented data with a normal distribution according to the Shapiro-Wilk test, the non-parametric Wilcoxon test was applied. Where microbial attributes presented values equal to zero they were considered null values (Table 1); while some soil samples did not present detectable amounts of PLFA during gas chromatography and so were excluded from the analysis. Boxplots were constructed using the *ggplot*

- 205 package and statistical significance was presented to compare sealed and unsealed soils. The correlations

206 between soil properties and microbial attributes were estimated using the Spearman's rank correlation

207 (ggcorrplot package).

208

209 3 Results

210 Sealed soils exhibited consistently lower values than unsealed soils across all measured soil properties 211 and microbial attributes, other than the C:N ratio (Table 1). Total C (p = 0.0026), total N (p < 0.001), and water content (p < 0.001) were all significantly lower in sealed soil than unsealed soil (Fig. 2A, B and D), while the 212 213 C:N ratio (p = 0.023) was higher in sealed soil (Fig. 2C). All microbial attributes exhibited significantly lower 214 values in sealed soil than unsealed soil: MBC, N-mineralisation, total PLFA, total fungi, AM fungi, total 215 bacteria, GP bacteria and GN bacteria presented p < 0.001; fungal:bacterial ratio presented p = 0.019; and 216 GP:GN bacterial ratio presented p = 0.0017 (Fig. 3 and Fig. 4). 217 Significant correlations were observed between soil properties and microbial PLFA attributes, however, 218 MBC and N-mineralisation potential showed no correlation with soil properties in this study (Table 2). In sealed 219 soil, total bacteria had a strong and positive correlation with total N (rho = 0.63, p = 0.038) and water content 220 (rho = 0.71, p = 0.015); GP bacteria a strong and positive correlation with total N (rho = 0.63, p = 0.038) and 221 water content (rho = 0.71, p = 0.015); and GN bacteria a strong and positive correlation with total C (rho = 0.64, p = 0.032), total N (rho = 0.71, p = 0.015) and water content (rho = 0.79, p = 0.004). In unsealed soil, total 222 223 PLFA, total fungi, total bacteria and GP bacteria presented moderate to strong positive correlations with total C 224 (rho = 0.58, p = 0.020; rho = 0.59, p = 0.019; rho = 0.56, p = 0.025 and rho = 0.52, p = 0.042, respectively);225 total N (rho = 0.62, p = 0.012; rho = 0.54, p = 0.034; rho = 0.68, p = 0.005; and rho = 0.69, p = 0.004, 226 respectively); and water content (rho = 0.75, p = 0.001; rho = 0.75, p = 0.001; rho = 0.68, p = 0.005; and rho = 227 0.66, p = 0.007, respectively). GN bacteria had a strong positive correlation with total N (rho = 0.61, p = 0.015) 228 and water content (rho = 0.65, p = 0.008); and the GP:GN bacterial ratio showed a moderate positive correlation 229 with total C (rho = 0.52, p = 0.040). 230

231 4 Discussion

232 In contrasting soil samples from sealed and unsealed areas, we observed that sealing affects soil 233 properties, reduces the microbial community and limits microbial processes; changes which may disrupt 234 important soil ecosystem services. Soil properties were notably altered in sealed areas, with a reduction of total 235 C, total N and water content, and a consequent increase in C:N ratio. Sealing had a negative impact on microbial 236 attributes, with a large reduction of the microbial community (MBC and PLFA biomarkers) and activity (N-237 mineralisation). Additionally, microbial attributes that correlated with soil properties in unsealed soil did not 238 show equivalent correlations in sealed soil, such as those between total PLFA and total fungi to total C, and total 239 N and water content. These results suggest that the microbial community in sealed soil may respond differently 240 to that in unsealed soil, indicating that sealing may disrupt the microbial response to changes in soil properties 241 and lead to negative impacts on microbial services. The PLFA data provides an indicator of the microbial 242 community in sealed soil, where low fungal:bacterial and gram-positive:gram-negative bacterial ratios indicate 243 degradation in microbial C sequestration and a consequential effect on soil C storage in sealed soil. 244

245 4.1 Soil sealing leads to depletion of C, N and water content

246 The sealed soils exhibited lower total C, total N and water content than unsealed soils (Table 2 and Fig.

247 2A). Soil sealing leads to a reduction of soil C due to topsoil removal during the construction process and the

- reduction of C inputs from organic matter, plant root exudates and residue decomposition (Edmondson et al.
- 249 2012; Raciti et al. 2012; Wei et al. 2013, 2014; Piotrowska and Charzynski 2015; Yan et al. 2015; Majidzadeh
- et al. 2017, 2018). Indeed, sealed soils have been recorded as having significantly lower C stores when
- compared with unsealed or greenspace soils in urban areas (Wei et al. 2014; Piotrowska-Długosz and
- 252 Charzyński 2015; Majidzadeh et al. 2017). Additionally, if C decomposition continues within sealed soil, even
- at a low rate (Wei et al. 2014; Piotrowska and Charzynski 2015), and there are negligible C inputs (Majidzadeh
- et al. 2018), this will contribute to C losses. In this context, elevation of microbial C respiration in sealed soil
- has been linked to increases in water content (Piotrowska and Charzynski 2015; Majidzadeh et al. 2017, 2018).
- In sealed soil, water content is affected by the type and size of pavement or sealing surface (Morgenroth et al.
- 257 2013), and beneath impervious and semi-permeable pavements the water content is, in general, lower than in
- greenspace soils (Hu et al. 2018; Piotrowska and Charzynski 2015). In soil under semi-permeable surfaces,
- 259 water moving from adjacent greenspaces into sealed soil can promote C inputs beneath sealed surfaces
- 260 (Majidzadeh et al. 2018); however, this can also increase the microbial processes of C decomposition and lead
- to C losses (Majidzadeh et al. 2017, 2018). In soil under house crawl spaces of different ages, most C was lost in
- 262 the first 50 years after construction, but after 50 years, C sequestration became the dominant process
- 263 (Majidzadeh et al. 2018). Overall, it is not clear whether longer periods of sealing lead to an increase or decrease
- 264 in the C balance of sealed soils, and this is an area which requires further investigation.
- The notable depletion of total N, as seen in our results (Fig. 2B), is a commonly observed consequence of
 soil sealing, often being greater in magnitude than losses of total or organic C (Raciti et al. 2012; Zhao et al.
 2012; Wei et al. 2014; Majidzadeh et al. 2018; Hu et al. 2018). Our results indicate that in sealed soil total N
- 268 was reduced by over 60 % compared to unsealed soil (Fig. 2B); while total C was reduced by nearly 40 %
- 269 compared to unsealed soil (Fig. 2A), leading to a higher C:N ratio in sealed soil (Fig. 2C). Our results are
- 270 comparable to other observations of sealed soil where total C reduction was between 42 and 57 %; and N
- depletion was between 47 and 97 % (Majidzadeh et al. 2018; Piotrowska et al. 2015; Raciti et al. 2012; Zhao et
- al. 2012). The effect of sealing appears to be most notable and variable for N dynamics and processes, which
- 273 can be connected to the length of time sealed, organic C availability and water content; influencing the sealing
- impact on microbial processes (Zhao et al. 2012; Piotrowska et al. 2015; Majidzadeh et al. 2017, 2018) and N-
- 275 mineralisation potential (Fig. 3B, Zhao et al. 2012). Previous research has shown that sufficient water content
- 276 can promote microbial decomposition and N-mineralisation where there is available organic C (Zhao et al.
- 277 2012; Majidzadeh et al. 2018), leading to inorganic N production (Zhao et al. 2012; Majidzadeh et al. 2018),
- and potential leaching of NH_4^+ -N and NO_3^-N and accumulation in the sub-soil (Zhao et al. 2012). Where water
- 279 can infiltrate into sealed soils from adjacent unsealed areas (Majidzadeh et al. 2018), we speculate that
- 280 mineralization of remaining organic matter could be stimulated. Considering, the reduced levels of C and the
- absence of plant roots, N assimilation by microorganisms and plants is likely to be low, resulting in N losses
- 282 over time by leaching, subsoil accumulation and groundwater transport. Beyond that, these circumstances may
- lead to inorganic N pollution of urban groundwater and water courses (Zhao et al. 2012).
- 284

285 4.2 Sealing alters microbial attributes and community composition

286 Soil sealing leads to a drastic reduction in microbial attributes. Our results showed that sealed soil

exhibited a reduction in MBC (Fig. 3A), as consistently reported in previous studies (Wei et al. 2013;

Piotrowska and Charzynski 2015; Majidzadeh et al. 2017, 2018). Observations of low MBC in sealed soil have

commonly been associated with low C, N and water content (Wei et al. 2013; Piotrowska and Charzynski 2015;

290 Majidzadeh et al. 2017, 2018; Hu et al. 2018). Our PLFA data also demonstrated the negative impact of sealing

on the microbial community (Fig. 3), with sealed soil exhibiting significantly lower mass of total PLFA and
 microbial markers, consistent with reductions in MBC, total C, total N and water content. It has been observed

292 microbial markers, consistent with reductions in MBC, total C, total N and water content. It has been observed 293 that a reduction in the microbial community reflects low microbial activity (Zhao et al. 2012; Piotrowska and

Charzynski 2015), a pattern also observed in our results with the significantly reduced N-mineralisation
 potential in sealed soil.

In studies of urban soil, few have considered the relationship between soil properties and microbial
 attributes in both sealed and unsealed soil. Indeed, physical and chemical properties, in particular water content,

298 have been shown to have significant effects on microbial attributes in unsealed soils (Wei et al. 2014;

299 Piotrowska and Charzynski 2015); and have exhibited positive correlations with MBC, catalase activity and

300 β-glucosidase activity in unsealed soil, but not in sealed soil (Piotrowska and Charzynski 2015). Here, neither

301 MBC nor N-mineralisation potential had significant correlations with any soil properties across sealed or

302 unsealed soils. Conversely, the PLFA data does show significant responses of the microbial community to soil

303 properties (Table 2). In unsealed soil, increases in C, N and water content correlated with growth of the

304 microbial community (total PLFA, bacteria and fungi), which is typical for natural soils or those under

305 agricultural conservation management (Helgason et al. 2014; Bai et al. 2020). However, in sealed soil, only

306 bacteria correlated with soil properties, suggesting that sealing disrupts the relationships normally seen in

307 natural and agricultural soils between microbial attributes and soil properties. And the importance of total N and

308 water content could be highlighted from our data, once both affected positively total, GP and GN bacteria of

309 sealed and unsealed soil (Table 2), indicating that input of water and N promoted bacterial growth. Other studies

310 have found additional soil properties associated with sealing-driven microbial depletion, including potassium

311 and phosphorus availability, heavy metals and dissolved organic C (Hu et al. 2018; Yu et al. 2019). Low

312 respiration and metabolic quotient observed on sealed soil (Piotrowska and Charzynski 2015) can still suggest

313 organic matter of low quality. Thus, sealing results in alterations to soil properties and negative impacts on the

314 soil microbial community and processes.

315 Sealing also caused alterations to the microbial community composition, notably the fungal:bacterial

ratio and GP:GN ratio. The effect of sealing was seen more strongly in fungi, with sealed soils having ~ 93 %

317 less fungi than unsealed soils, and ~ 78 % less bacteria than unsealed soils. Consequently, the fungal:bacterial

318 ratio decreased in sealed soils indicating greater numbers of bacteria to fungi (Fig. 4G). Fungi have been shown

319 to be resistant to conditions of low total N, high C:N ratio and low water content (Six et al. 2006; Strickland and

320 Rousk 2010; Fang et al. 2020); conditions which are commonly observed in sealed soils. However, these

321 conditions did not lead to greater dominance of fungi in this study. Conversely, soils affected by degradation

322 processes such as tillage, deforestation, trampling and contamination usually present a greater impact on the

323 fungal community and show a proportional decrease on the fungal:bacterial ratio (Kaur et al. 2005; Malmivaara-

Lämsä et al. 2008; Simmons and Coleman 2008; Bischoff et al. 2016; Montiel-Rozas et el. 2018; Lopes and

Fernandes 2020). Thus, our results suggest that fungi in sealed soils may be more affected by aspects of soil

326 sealing not included in this study but that commonly arise due to the degradation processes of urbanization, such 327 as contamination and disturbance.

328 The decrease in the GP:GN bacterial ratio in sealed soil (Fig. 4H) suggests that GN bacteria are more 329 adapted to sealing than GP bacteria. GN bacteria presented a positive correlation with total C, while GP bacteria 330 had no correlation with total C (Table 2). As GN bacteria are more dependent on simple sugars (Kramer and 331 Gleixner 2008; Fanin et al. 2019), the organic C that is promoting GN bacterial growth is likely to be labile and 332 soluble C transported by water from adjacent greenspaces, a process which has been suggested as a source of organic C in soils beneath house crawl spaces (Majidzadeh et al. 2018). Additionally, GN and GP bacteria had 333 334 positive correlations with total N and water content, suggesting there may also be transport of soluble N by 335 water from adjacent greenspaces, and that this may be an important source of nutrients for bacteria in sealed 336 soil.

In contrast to GN bacteria, GP bacteria are linked to more complex SOC (Kramer and Gleixner 2008;
Fanin et al. 2019). Therefore, the low biomass of GP bacteria can be related to low levels of complex SOC

remaining in sealed soil as a consequence of topsoil removal and microbial degradation over time.

340

341 4.3 Sealing limits the microbial community and affects the C storage service

Litter degradation plays an important role in C inputs into soil. Organic and inorganic compounds released during decomposition and the remaining complex organic compounds are essential components of soil organic matter synthesis (Jastrow et al. 2007). In sealed soil, the sealed surface acts as a barrier preventing this source of organic C from reaching the soil, such that low or no organic C or nutrients from litter can enter the soil (Zhao et al. 2012; Majidzadeh et al. 2017, 2018), which in turn, affects soil biological and nutrient processes.

Plants and roots also contribute greatly to soil C stores. The lack of plants growing on sealed surfaces
usually leads to a reduced root colonization, limiting the C inputs from plant exudates and dead roots.

350 Consequently, microbial processes that take place in the soil-root zone and depend on plant exudates are limited

351 beneath sealed surfaces. Many of these processes are related to N inputs and nutrient availability, highlighting N

biological fixation, N oxidation reactions and phosphate solubility (Sylvia et al. 2005; Paul 2007). Many fungal

353 species establish a mutualistic association with plant roots to obtain organic molecules and, as payment, they

colonize soil space to assimilate and transport nutrients directly back to the plant roots (Smith and Read 2008).

355 By enhancing the soil microbial community, roots enable microbial processes connected with organic matter

356 formation, such as the microbial release of biomolecules and dead biomass (Jastrow et al. 2007; Clemmensen et

al. 2013). Thus, it is likely that the lack of plant and root growth, litter inputs and microbial activity in the soil-

358 root zone all contribute to the lower C stores in sealed soil.

359 Fungal biomass in soil is, in general, suggested to contribute to high soil C storage (Strickland and Rousk

360 2010). Fungi exhibit low nutrient requirements and high C use efficiency which results in more C being

361 allocated to their biomass, per unit of substrate used, compared to bacteria, which have lower C use efficiency

362 (Six et al. 2006). Fungi have the ability to grow under a high C:N ratio, permitting their mycelial growth to

363 explore wider areas and translocate nutrients across the soil (Strickland and Rousk 2010). In addition, fungal

364 biomass is more complex and resistant to decomposition than bacterial biomass, introducing a more stable form

- of organic C in the soil (Jastrow et al. 2007; Clemmensen et al. 2013). While studies have presented different
- 366 insights into the functional implications of the fungal:bacterial ratio (Strickland and Rousk 2010; Soares and
- 367 Rousk 2019), in general, a higher fungal:bacterial ratio is assumed to promote an increase in soil organic matter
- 368 (Jastrow et al. 2007; Strickland and Rousk 2010). Therefore, the observed reduction in fungi and consequent
 369 bacterial dominance in sealed soil is likely to lead to notable limitations to C storage.
- The lower GP:GN bacteria ratio in sealed soil illustrates that there is more GN bacteria to GP. This indicates that there is less recalcitrant C in the sealed soil (Kramer and Gleixner 2008; Fanin et al. 2019), which suggests the reduced ability of sealed soils not only to store C, but to store it as stable C that may be more
- 373 protected from decomposition (Lal 2004, Marschner et al. 2008), highlighting the wider impacts of soil sealing
- on the ecosystem service of soil C storage.
- 375

376 **5** Conclusion

377 Soil properties were notably affected in sealed soil, with a large significant reduction in total C, total N and water content in sealed soils. Microbial biomass C, N-mineralisation potential and microbial PLFA markers 378 379 were also significantly reduced in sealed soils. Our results show that changes to soil properties, caused by 380 sealing, led to a drastic decrease in the microbial community and important microbial processes. The increase of 381 the C:N ratio and decrease of the F:B and GP:GN ratios suggest that sealed soils are degraded due to the loss of 382 C, which limits fungal and bacterial growth. In addition, the reduced inputs of C from litter degradation and 383 plant exudates, associated with the reduction of fungal dominance, indicate a limitation on the C storage potential of sealed soil. Furthermore, the correlation of bacteria with C, N and water suggests there may 384 385 transport of soluble C and N by water into sealed soils from adjacent greenspaces. This may be an important source of nutrients for microbes in sealed soil, and the investigation of this process would be beneficial to 386 387 further understand sealed soil nutrient cycling and implications for C and N fluxes. In this context, further work, 388 such chronosequence studies, would elucidate how urbanisation and soil sealing impact the dynamics of C and N and microbial processes over time, and as a consequence, the ecosystem services of sealed soil. 389

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395 **REFERENCES**

- 396 Bai Y, Zha X, Chen S. (2020) Effects of the vegetation restoration years on soil microbial community
- 397 composition and biomass in degraded lands in Changting County, China. J For Res 28:1-14
- 398 Bischoff N, Mikutta R, Shibistova O, Puzanov A, Reichert E, Silanteva M, Grebennikova A, Schaarschmidt F,
- 399 Heinicke S, Guggenberger G (2016) Land-use change under different climatic conditions: Consequences
- 400 for organic matter and microbial communities in Siberian steppe soils. Agric Ecosyst Environ 235:253-264
- 401 Bradley RI, Milne R, Bell J, Lilly A, Jordan C, Higgins A (2005) A soil carbon and land use database for the
- 402 United Kingdom. Soil Use Manage 21(4):363-369

- Brookes PC, Landman A, Pruden G, Jenkinson DS (1985) Chloroform fumigation and the release of soil
 nitrogen: a rapid direct extraction method to measure microbial biomass nitrogen in soil. Soil Biol
 Biochem 17(6):837-842
- Chen Y, Wang X, Jiang B, Yang N, Li L (2016) Pavement induced soil warming accelerates leaf budburst of
 ash trees. Urban For Urban Green 16:36-42
- Chen Y, Wang X, Jiang B, Wen Z, Yang N, Li L (2017) Tree survival and growth are impacted by increased
 surface temperature on paved land. Landsc Urban Plan 162:68-79
- Clemmensen KE, Bahr A, Ovaskainen O, Dahlberg A, Ekblad A, Wallander H, Stenlid J, Finlay RD, Wardle
 DA, Lindahl BD (2013) Roots and associated fungi drive long-term carbon sequestration in boreal forest.
 Science 339(6127):1615-1618
- Edmondson JL, Davies ZG, McHugh N, Gaston KJ, Leake JR (2012) Organic carbon hidden in urban
 ecosystems. Sci Rep 2:963
- Fang X, Zhou G, Qu C, Huang W, Zhang D, Li Y, Yi Z, Liu J (2020) Translocating subtropical forest soils to a
 warmer region alters microbial communities and increases the decomposition of mineral-associated
 organic carbon. Soil Biol Biochem 142:107707
- Fanin N, Kardol P, Farrell M, Nilsson MC, Gundale MJ, Wardle DA (2019) The ratio of Gram-positive to
 Gram-negative bacterial PLFA markers as an indicator of carbon availability in organic soils. Soil Biol
 Biochem 128:111-114
- Fini A, Frangi P, Mori J, Donzelli D, Ferrini F (2017) Nature based solutions to mitigate soil sealing in urban
 areas: Results from a 4-year study comparing permeable, porous, and impermeable pavements. Environ
 Res 156:443-454
- Frostegård Å, Tunlid A, Bååth E (2011) Use and misuse of PLFA measurements in soils. Soil Biol Biochem
 43(8):1621-1625
- Helgason BL, Konschuh HJ, Bedard-Haughn A, VandenBygaart AJ (2014) Microbial distribution in an eroded
 landscape: Buried A horizons support abundant and unique communities. Agric Ecosyst Environ 196:94 102
- Hu Y, Dou X, Li J, Li F (2018) Impervious surfaces alter soil bacterial communities in urban areas: a case study
 in Beijing, China. Front Microbiol 9:226
- Hu S, Fan Y, Zhang T (2020) Assessing the effect of land use change on surface runoff in a rapidly urbanized
 city: A case study of the central area of Beijing. Land 9(1):17
- Jastrow JD, Amonette JE, Bailey VL (2007) Mechanisms controlling soil carbon turnover and their potential
 application for enhancing carbon sequestration. Clim Change 80(1-2):5-23
- Kaur A, Chaudhary A, Kaur A, Choudhary R, Kaushik R (2005) Phospholipid fatty acid–a bioindicator of
 environment monitoring and assessment in soil ecosystem. Curr Sci 10:1103-1112
- Kelleher C, Golden HE, Burkholder S, Shuster W (2020) Urban vacant lands impart hydrological benefits
 across city landscapes. Nat Commun 11(1):1-11
- Kramer C, Gleixner G (2008) Soil organic matter in soil depth profiles: distinct carbon preferences of microbial
 groups during carbon transformation. Soil Biol Biochem 40(2):425-433
- 441 Lal R (2004) Soil carbon sequestration to mitigate climate change. Geoderma 123:1–22

- Li Y, Liu M, Li R, Sun P, Xia H, He T (2020) Polycyclic aromatic hydrocarbons in the soils of the Yangtze
 River Delta Urban Agglomeration, China: Influence of land cover types and urbanization. Sci Total
 Environ 715:137011
- Lopes LD, Fernandes MF (2020) Changes in microbial community structure and physiological profile in a
 kaolinitic tropical soil under different conservation agricultural practices. Appl Soil Ecol 152:103545
- 447 Marschner B, Brodowski S, Dreves A et al (2008) How relevant is recalcitrance for the stabilization of organic
 448 matter in soils? J Plant Nutr Soil Sci 171(1):91-110
- Majidzadeh H, Lockaby BG, Governo R (2017) Effect of home construction on soil carbon storage-A
 chronosequence case study. Environ Pollut 226:317-323
- 451 Majidzadeh H, Lockaby BG, Price R, Governo R (2018) Soil Carbon and Nitrogen Dynamics beneath
 452 Impervious Surfaces. Soil Sci Soc Am J 82(3):663-670
- Malmivaara-Lämsä M, Hamberg L, Haapamäki E, Liski J, Kotze DJ, Lehvävirta S, Fritze H (2008) Edge effects
 and trampling in boreal urban forest fragments–impacts on the soil microbial community. Soil Biol
 Biochem 40(7):1612-1621
- Montiel-Rozas MM, Domínguez MT, Madejón E, Madejón P, Pastorelli R, Renella G (2018) Long-term effects
 of organic amendments on bacterial and fungal communities in a degraded Mediterranean soil. Geoderma
 332:20-28
- Morgenroth J, Buchan G, Scharenbroch BC (2013) Belowground effects of porous pavements—Soil moisture
 and chemical properties. Ecol Eng 51:221-228
- 461 Paul EA (2007) Soil microbiology, ecology, and biochemistry in perspective. Academic Press, Oxford
- 462 Piotrowska-Długosz A, Charzyński P (2015) The impact of the soil sealing degree on microbial biomass,
- 463 enzymatic activity, and physicochemical properties in the Ekranic Technosols of Toruń (Poland). J Soils
 464 Sediments 15(1):47-59
- 465 Raciti SM, Hutyra LR, Finzi AC (2012) Depleted soil carbon and nitrogen pools beneath impervious surfaces.
 466 Environ Pollut 164:248-251
- Simmons BL, Coleman DC (2008) Microbial community response to transition from conventional to
 conservation tillage in cotton fields. Appl Soil Ecol 40(3):518-528
- Six J, Frey SD, Thiet RK, Batten KM (2006) Bacterial and fungal contributions to carbon sequestration in
 agroecosystems. Soil Sci Soc Am J 70(2):555-569
- 471 Smith SE, Read DJ (2008) Mycorrhizal symbiosis. Academic Press, New York
- 472 Soares M, Rousk J (2019) Microbial growth and carbon use efficiency in soil: links to fungal-bacterial
 473 dominance, SOC-quality and stoichiometry. Soil Biol Biochem 131:195-205
- 474 Strickland MS, Rousk J (2010) Considering fungal: bacterial dominance in soils–methods, controls, and
 475 ecosystem implications. Soil Biol Biochem 42(9):1385-1395
- 476 Sylvia DM, Fuhrmann JJ, Hartel PG, Zuberer DA (2005) Principles and applications of soil microbiology.
 477 Pearson Prentice Hall, New Jersey
- 478 Vance ED, Brookes PC, Jenkinson DS (1987) An extraction method for measuring soil microbial biomass C.
 479 Soil Biol Biochem 19(6):703-707
- 480 Vasenev VI, Stoorvogel JJ, Leemans R, Valentini R, Hajiaghayeva RA (2018) Projection of urban expansion
- 481 and related changes in soil carbon stocks in the Moscow Region. J Clean Prod 170: 902-914

- Vestle J, White D (1989) Lipid analysis in microbial ecology–quantitative approaches to the study of microbial
 communities. Bioscience 39:535-541
- Wei Z, Wu S, Zhou S, Lin C (2013) Installation of impervious surface in urban areas affects microbial biomass,
 activity (potential C mineralisation), and functional diversity of the fine earth. Soil Res 51(1):59-67
- Wei ZQ, Wu SH, Zhou SL, Li JT, Zhao QG (2014) Soil organic carbon transformation and related properties in
 urban soil under impervious surfaces. Pedosphere 24(1):56-64
- Willers C, Jansen van Rensburg PJ, Claassens S (2015) Phospholipid fatty acid profiling of microbial
 communities–a review of interpretations and recent applications. J Appl Microbiol 119(5):1207-1218
- Wu X, Hu D, Ma S, Zhang X, Guo Z, Gaston KJ (2016) Elevated soil CO₂ efflux at the boundaries between
 impervious surfaces and urban greenspaces. Atmos Environ 141:375-378
- Yan Y, Kuang W, Zhang C, Chen C (2015) Impacts of impervious surface expansion on soil organic carbon–a
 spatially explicit study. Sci Rep 5:17905
- Yan Y, Zhang C, Hu Y, Kuang W (2016) Urban land-cover change and its impact on the ecosystem carbon
 storage in a dryland city. Remote Sens 8(1):6
- Yu W, Hu Y, Cui B, Chen Y, Wang X (2019) The effects of pavement types on soil bacterial communities
 across different depths. Int J Environ Res Public Health 16(10):1805
- Zhao D, Li F, Wang R, Yang Q, Ni H (2012) Effect of soil sealing on the microbial biomass, N transformation
 and related enzyme activities at various depths of soils in urban area of Beijing, China. J Soils Sediments
 12(4):519-530
- Ziter C, Turner MG (2018) Current and historical land use influence soil-based ecosystem services in an urban
 landscape. Ecol Appl 28(3):643-654

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Table 1 Descriptive statistics of soil properties and microbial attributes in sealed and unsealed soils

Variable	Variables	Pavement	n* Null*		Min – Max*	Mean ± SE*	CV* (%)
groups		types					
Soil	Total C g/Kg	Sealed	25	0	3.35 - 250.29	49.78 ± 10.67	107.12
properties		Unsealed	25	0	14.02 - 128.49	73.49 ± 5.33	36.27
	Total N g/Kg	Sealed	25	0	0.39 - 13.75	2.08 ± 0.587	141.03
		Unsealed	25	0	0.026 - 21.75	5.36 ± 0.79	73.24
	C:N ratio	Sealed	25	0	4.92 - 149.87	35.81 ± 7.13	99.57
		Unsealed	25	0	5.91 - 27.49	15.55 ± 1.11	35.62
	Water content g/g	Sealed	25	0	0.09 - 0.74	0.30 ± 0.03	54.09
		Unsealed	25	0	0.08 - 0.85	0.47 ± 0.03	32.83
Microbial	MBC g/Kg	Sealed	25	7	0 - 47.85	6.11 ± 2.17	177.67
attributes		Unsealed	25	0	1.99 - 58.59	19.79 ± 3.04	76.69
	Mineralization g/Kg	Sealed	25	12	0 - 2.87	0.42 ± 0.15	178.14
		Unsealed	25	1	0 - 21.22	5.61 ± 1.03	92.12
	Total PLFA mg/Kg	Sealed	11	0	0.007 - 2.176	0.311 ± 0.198	211.30
		Unsealed	16	0	0.338 - 2.996	1.101 ± 0.164	59.61
	Fungi mg/Kg	Sealed	11	3	0 - 0.239	0.036 ± 0.021	197.03
		Unsealed	16	0	0.118 - 0.867	0.357 ± 0.050	56.33
	AM fungi mg/Kg	Sealed	11	8	0 - 0.019	0.003 ± 0.002	230.69
		Unsealed	16	0	0.008 - 0.146	0.062 ± 0.009	60.98
	Bacteria mg/Kg	Sealed	11	5	0 - 0.832	0.094 ± 0.075	263.19
		Unsealed	16	0	0.075 - 0.821	0.304 ± 0.045	58.83
	GP bacteria mg/Kg	Sealed	11	5	0 - 0.364	0.043 ± 0.033	249.92
		Unsealed	16	0	0.044 - 0.572	0.187 ± 0.032	68.37
	GN bacteria mg/Kg	Sealed	11	6	0 - 0.468	0.050 ± 0.042	277.50
		Unsealed	16	0	0.031 - 0.236	0.113 ± 0.013	45.08
	Fungal:Bacterial	Sealed	10	4	0 - 2.470	0.663 ± 0.284	135.20
	ratio	Unsealed	16	0	0.717 - 1.585	1.206 ± 0.062	20.57
	GP:GN bacterial	Sealed	10	5	0 - 2.151	0.628 ± 0.237	119.30
	ratio	Unsealed	16	0	0.958 - 2.428	1.584 ± 0.104	26.16

507 *n: the number of values; null: the number of null values; min: the minimal value; max: the maximal value; SE:

508 the standard error of the mean; CV: the coefficient of variation.

Table 2 Spearman's rank correlation (rho) and p-values of correlations between microbial attributes and soil

properties in sealed and unsealed soils. Significant correlations with p-values < 0.05 are indicated in bold.										
Mianahial attuihuta	Soil status	Total C		Total N		C:N ratio		Water content		
witcrodial attribute		rho	p-value	rho	p-value	rho	p-value	rho	p-value	
MBC	Sealed	0.31	0.356	0.61	0.052	-0.27	0.418	0.47	0.146	
	Unsealed	0.50	0.051	0.20	0.450	0.35	0.188	0.41	0.114	
N-mineralisation	Sealed	-0.04	0.902	-0.21	0.534	0.02	0.951	-0.18	0.598	
potential	Unsealed	0.29	0.278	0.25	0.343	0.04	0.891	-0.04	0.891	
Total PLFA	Sealed	0.57	0.071	0.55	0.082	0.13	0.714	0.55	0.087	
	Unsealed	0.58	0.020	0.62	0.012	-0.08	0.771	0.75	0.001	
Total fungi	Sealed	0.46	0.156	0.5	0.113	0.03	0.936	0.5	0.121	
	Unsealed	0.59	0.019	0.54	0.034	-0.01	0.978	0.75	0.001	
Total bacteria	Sealed	0.56	0.072	0.63	0.038	-0.13	0.696	0.71	0.015	
	Unsealed	0.56	0.025	0.68	0.005	-0.13	0.633	0.68	0.005	
Fungal:Bacterial	Sealed	0.23	0.499	0.09	0.802	0.19	0.574	0.23	0.499	
ratio	Unsealed	0.29	0.283	-0.22	0.534 0.02 0.951 -0.18 0.343 0.04 0.891 -0.04 0.082 0.13 0.714 0.55 0.012 -0.08 0.771 0.75 0.113 0.03 0.936 0.5 0.034 -0.01 0.978 0.75 0.038 -0.13 0.696 0.71 0.005 -0.13 0.633 0.68 0.802 0.19 0.574 0.23 0.404 0.48 0.064 0.03 0.038 -0.13 0.696 0.71 0.004 -0.19 0.484 0.66 0.015 -0.19 0.569 0.79 0.015 -0.42 0.104 0.65 0.079 0.19 0.569 0.68	0.926				
GP bacteria	Sealed	0.56	0.072	0.63	0.038	-0.13	0.696	0.71	0.015	
	Unsealed	0.52	0.042	0.69	0.004	-0.19	0.484	0.66	0.007	
GN bacteria	Sealed	0.64	0.032	0.71	0.015	-0.19	0.569	0.79	0.004	
	Unsealed	0.21	0.443	0.61	0.015	-0.42	0.104	0.65	0.008	
GP:GN bacterial	Sealed	0.42	0.203	0.55	0.079	-0.19	0.569	0.68	0.022	
ratio	Unsealed	0.52	0.040	0.47	0.070	0.01	0.969	0.33	0.217	

FIGURE CAPTIONS

Fig. 1 Location of sampling sites, indicated on the map with black dots.

Fig. 2 Soil properties in sealed and unsealed soils. (A)Total C, (B) total N, (C) C:N ratio and (D) water content. A significant difference between sealed and unsealed soil was estimated by Kruskal-Wallis test, with "****",

"***", "**" and "*" indicating significance at p < 0.0001, p < 0.001, p < 0.001 and p < 0.05, respectively.

Fig. 3 Microbial biomass C (MBC) and N-mineralisation potential in sealed and unsealed soils. A significant difference between sealed and unsealed soil was estimated by Kruskal-Wallis test, with "****" indicating significance at p < 0.0001.

Fig. 4 Microbial community in sealed and unsealed soils. (A) total PLFA, (B) total fungi, (C) AM fungi, (D) total bacteria, (E) GP bacteria, (F) GN bacteria, (G) fungal:bacteria ratio and (H) GP:GN bacterial ratio. A significant difference between sealed and unsealed soil was estimated by Kruskal-Wallis test, with "****", "***" and "**" indicating significance at p < 0.0001, p < 0.001 and p < 0.01, respectively.