

1 **Repurposing distillation waste biomass and low-value mineral resources through biochar-**
2 **mineral-complex for sustainable production of high-value medicinal plants and soil quality**
3 **improvement**

4
5 B. B. Basak^{a¶}, Ajoy Saha^{a,b}, Binoy Sarkar^{c*}, B. Prem Kumar^d, N. A. Gajbhiye^a, Atanu Banerjee^e

6
7 ^a ICAR-Directorate of Medicinal and Aromatic Plants Research, Anand – 387310, India

8 ^b ICAR- Central Inland Fisheries Research Institute, Bangalore Research Centre, Bangalore –
9 560089, India

10 ^c Lancaster Environment Centre, Lancaster University, Lancaster, LA14YQ, United Kingdom

11 ^d Department Soil Science and Agricultural Chemistry, Anand Agricultural University, Anand –
12 388110, India

13 ^e Dr. K C Patel Research & Development Centre, Charotar University of Science and
14 Technology, Changa, Anand- 388421, India

15
16 *Corresponding author: Dr Binoy Sarkar; Lancaster University; e-mail: b.sarkar@lancaster.ac.uk

17 ¶e-mail: biraj.basak@icar.gov.in (B. B. Basak; ICAR-Directorate of Medicinal and Aromatic
18 Plants Research)

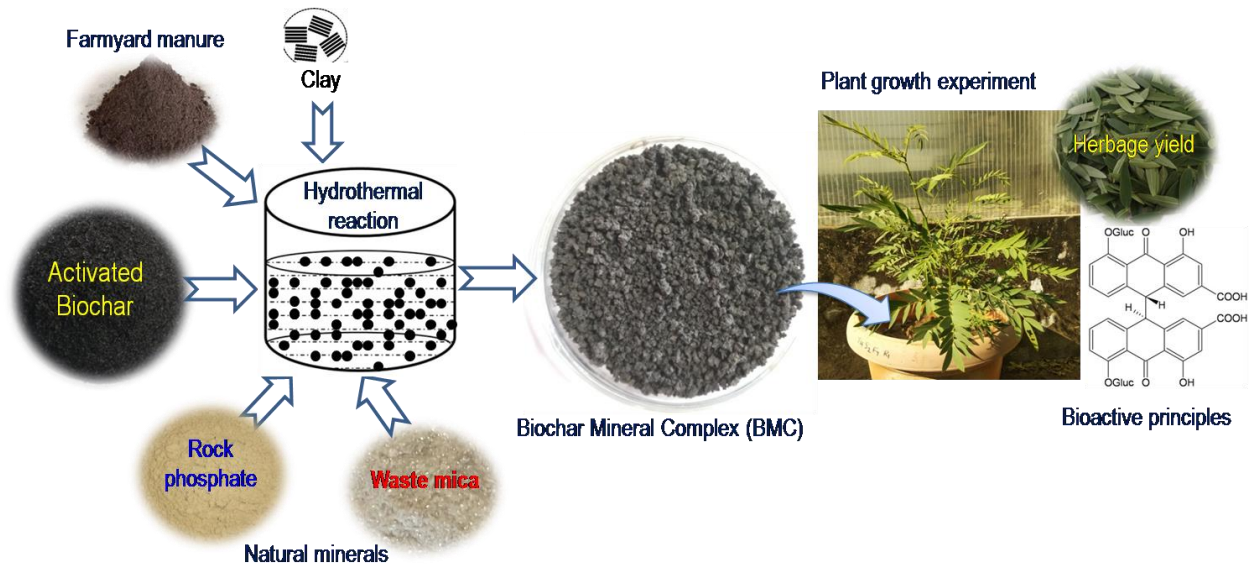
19 **Highlights**

- 20 ➤ Novel biochar-mineral-complex (BMC) was prepared as an effective soil amendment.
- 21 ➤ BMC had better surface properties and nutrient contents than pristine biochar.
- 22 ➤ BMC performed better than conventional organic and chemical fertilizers.
- 23 ➤ BMC improved herbage and bioactive compound yields of senna plants.
- 24 ➤ BMC improved soil OC, available nutrients and biological properties.

25

26 **Graphical abstract**

27



28

29

30 **Abstract**

31 High cost of synthetic fertilizers and their hazardous effects catapult the exploration of
32 alternative nutrient formulations and soil amendments. This study aimed to synthesize a novel
33 biochar-mineral-complex (BMC), and evaluate its nutrient supplying and soil improvement
34 performances. In a hydrothermal reaction, the BMC was prepared using a biochar derived from
35 distillation waste of Lemongrass (*Cymbopogon flexuosus*) and farmyard manure, for the first
36 time via fortification with low-grade rock phosphate and waste mica. The BMC showed
37 improved physico-chemical properties and nutrient availability than the pristine biochar. When
38 applied to a deeply weathered acidic soil, the BMC significantly ($p<0.05$) improved the herbage
39 and bioactive compound (sennoside) yields of a medicinal plant (senna; *Cassia angustifolia*
40 Vahl.) compared to the pristine biochar, farmyard manure, vermicompost, and chemical
41 fertilizers. The BMC also improved the soil quality by increasing nutrient and carbon contents,
42 and microbial activities. Soil quality improvement facilitated greater nutrient uptake in senna
43 plants under BMC compared to the pristine biochar, and conventional organic and chemical
44 fertilizer treatments. This study thus encourages the development of BMC formulations not only
45 to overcome the limitation of sole biochar application to soils, but also to phaseout chemical
46 fertilizers in agriculture. Moreover, BMC could bestow resilience and sustainability to crop
47 production via value-added recycling of waste biomass and low-grade mineral resources.

48

49 **Key words:** Alternative fertilizers; Distillation waste biomass; Low-grade minerals; Medicinal
50 plants; Resource recycling; Soil quality improvement

51

52 **1. Introduction**

53 A sharp increase in the price and demand of chemical fertilizers in global agriculture and the
54 huge carbon footprint to produce them emphasized the need to invent efficient nutrient
55 management strategies based on cost-effective and environment-friendly nutrient sources.
56 Recycling of waste biomass, crop residues, animal manures, and also naturally occurring low-
57 grade minerals has been regarded as an avenue to reduce the need of costly synthetic fertilizers
58 (Basak, 2017). Farmyard manure (FYM) and compost are frequently used as soil amendments
59 and important nutrient sources in organic agriculture (Basak et al., 2020). Recently, biochar is
60 also considered as a promising soil amendment to conserve carbon and nutrients present in
61 organic materials and biomass, and thereby address the environmental issues concerning with
62 sustainable agricultural nutrient management (Mandal et al., 2016; El-Naggar et al., 2019;
63 Purakayastha et al., 2019).

64 Extraction of essential oil through distillation of aromatic plants contributes 3 Mt of solid
65 biomass every year in India (Saha and Basak, 2019). The waste biomass generated after the
66 extraction of essential oil (0.5 – 1.0% of total fresh biomass) may contribute to emission of
67 greenhouse gases if the biomass is dumped openly or burnt (Saha et al., 2018). Recycling of the
68 aromatic plant waste biomass (APWB) into effective soil amendments (e.g., compost and
69 biochar) might not only mitigate the greenhouse gas issues but also increase the farmers' benefits
70 by supplying and retaining nutrients, at least as a partial replacement of chemical fertilizers.

71 Application of biochar and compost is also known to improve the physical, chemical, and
72 biological qualities of agricultural soils (Sanchez-Monedero, 2018; Saha et al., 2019; Basak et
73 al., 2020). However, nutrient contents in the biochar prepared from APWB is not sufficient to
74 support crop demand; particularly major nutrients such as N, P and K are a concern (Jha et al.,

75 [2010](#)). Fortification of biochar with rock phosphate (RP) and waste mica via biochar-mineral-
76 complex (BMC) preparation could be a promising approach to increase the nutrient contents in
77 biochar-based products because these low-grade mineral materials are an inexpensive source of P
78 (Basak et al., [2020](#)) and K (Basak, [2018](#)). The P and K release from these minerals could be
79 accelerated by thermal alteration (dos Santos Teixeira et al., [2015](#); Tumbure et al., [2020](#)) which
80 is an integral practice for biochar preparation. No report is currently available for the production
81 of BMC from APWB and RP/waste mica, and their nutrient release pattern during crop
82 cultivation. P and K are non-renewable nutrient sources, and many developing countries
83 including India are dependent on the import of commercial fertilizers (100% K fertilizers, and
84 >60% P fertilizers) from other countries (Basak, [2019](#)). An effective recycling of the locally
85 available minerals via BMC thus offers enormous potentials in crop production by reducing the
86 dependency on costly chemical fertilizers.

87 High specific surface area (SSA) and a range of reactive surface functional groups (e.g., -COOH,
88 -OH, and -NH₂) enable biochar to exhibit various chemical properties (e.g., hydrophilic vs
89 hydrophobic, acidic vs basic properties) (Wallace et al., [2019](#); Zhang et al., [2020](#)). The SSA of
90 biochar can significantly contribute to adsorption-desorption, acid-base, and redox reactions with
91 the soil matrix (Ding et al., [2016](#); Matin et al., [2020](#)). Recent findings revealed that biochar could
92 form organo-mineral complexes by interacting with surrounding organic matter, ash, clay, and
93 other minerals in the soil (Farrar et al., [2019](#); Zhao et al., [2019](#); Lu et al., [2020](#)).

94 BMC having properties similar to natural soil organo-mineral aggregates could be produced by
95 mixing organic matter, ash, clay, and other minerals, and subsequently heating at a moderate
96 temperature (up to 240°C) (Chia et al., [2014](#)). The organo-mineral reaction during torrefaction at
97 220-240°C could form a nutrient-rich BMC (Lin et al., [2013](#)). Li et al. ([2014](#)) reported an

98 increase of surface reactivity and cation exchange capacity (CEC) in BMC due to the formation
99 of oxygenated functional groups, and Lewis acid and base sites. The production of BMC could
100 conserve nutrient elements by incorporating them into the heterocyclic carbon structures, which
101 would provide a comparatively stable nutrient rich formulation with slow-release property.
102 Through improved physicochemical and biological activities, BMC application could facilitate
103 soil nutrient cycling and mobilization, and ultimately plant uptake (Ye et al., 2016). However,
104 nutrient availability from BMC and its role in plant growth and nutrition have been investigated
105 only sparsely.

106 Poor nutrient use efficiency from chemical fertilizers in highly weathered acidic tropical soils
107 covering nearly one-third of the continental surface is a significant global issue (Jien and Wang,
108 2013; Anda et al., 2015). Owing to high SSA, CEC and pH ameliorating ability, BMC could
109 improve nutrient retention in such deeply weathered soil, and support the production of crops
110 maximizing benefits from the degraded soils while simultaneously recycling biowaste and
111 mineral resources. It was hypothesized that BMC would be more effective than chemical
112 fertilizers and conventional organic nutrient sources (e.g., FYM and vermicompost) in sustaining
113 plant growth by improving the nutrient use efficiency in highly weathered soils. The specific
114 aims of this study are to (1) prepare BMC from APWB, RP and waste mica as an amendment of
115 a highly weathered (degraded) soil, (2) explore the physico-chemical characteristics of the BMC,
116 (3) examine BMC-induced soil property changes in comparison to chemical fertilizer, FYM and
117 vermicompost application, and (4) investigate the yield and quality of a medicinal plant (*Cassia*
118 *angustifolia* Vahl.) following soil application of the BMC.

119

120 **2. Materials and methods**

121 2.1. Preparation of biochar-mineral-complex (BMC)

122 Residual biomass obtained after the distillation of Lemongrass (*Cymbopogon flexuosus*) in a
123 hydro-distillation unit of a local farmer in Anand, Gujarat, India, was collected for the
124 preparation of biochar. The solid residual biomass was air-dried, and processed into powder (≤ 2
125 mm). The powdered biomass was pyrolyzed at 350°C (heating rate: 5°C min⁻¹) for 2 h in a muffle
126 furnace equipped with N₂ purge (flow rate: 2 mL min⁻¹) to make a limited oxygen environment
127 (Saha et al., 2019). The BMC was prepared following a reported method (Lin et al., 2013; Chia
128 et al., 2014; Ye et al., 2016) with some modification. The BMC was synthesized with a
129 proportionate mixture of the above biochar (32 g), minerals (i.e., 30 g kaolinite clay, 12.5 g RP,
130 and 5.5 g waste mica) and 20 g organic matter (FYM) (Table S1; Supplementary Information).
131 The RP used in this study contained 8.6% and 0.003% total and water-soluble P (Basak, 2009a).
132 The waste mica contained 9.72% total K (K₂O), and 0.011% water soluble K (Basak, 2019b).
133 The biochar was pre-treated with 10% phosphoric acid (Chia et al., 2014) before mixing it with
134 other raw materials. The phosphoric acid treatment was employed to increase the porous
135 structure, and abundance and stability of surface functional groups such as carbonyl, phenolic,
136 alcoholic, hydroxyl, and ether groups in the biochar (Zhao et al., 2017; Chu et al., 2018). This
137 treatment also would lead to loss of hydrogen from the surface, and increase the aromaticity of
138 the biochar by increasing the proportional content of C (Lin et al., 2012). The biochar was
139 washed repeatedly with distilled water to remove the excess phosphoric acid until the solution
140 pH became neutral. Then, the proportionate amounts of other raw materials (Table S1) were
141 mixed thoroughly with the acid treated-biochar. Boiling double distilled water was poured into
142 the mixture, and stirred at 80°C for 2 h to let the materials coagulate, and finally a homogeneous
143 mixture was obtained. The mixture was transferred to a batch hydrothermal reactor (20 cm × 20

144 cm × 30 cm), and raised to 220°C within 1 h, and maintained at this temperature for 3 h to get the
145 BMC (Fig. S1). The moderate torrefaction temperature and duration were adopted to enhance the
146 dissolved organic C content in the BMC (Lin et al., 2013). The final product (BMC) was cooled
147 under atmospheric condition, and stored in an airtight container until further use.

148

149 2.2. *Physico-chemical properties of BMC*

150 The biochar and BMC were finely ground, and passed through ≤ 2-mm sieve for further analysis.
151 The pH and electrical conductivity (EC) were measured by using a digital pH and conductivity
152 meter (Aquamax KF, GR Scientific, UK), at a solid: deionized water ratio of 1:5 (w/v). The CEC
153 was measured by extracting the samples with 1N sodium acetate solution (pH = 8.2) (Sumner
154 and Miller, 1996). Acid neutralizing capacity (ANC) was determined by using the potentiometric
155 titration method (Johnson, 1990). The SSA was measured by the Brunauer-Emmett-Teller (BET)
156 adsorption method. Total C and N were determined by a CHN analyzer (PE-2400, Perkin Elmer,
157 USA). Other nutrient contents (e.g., P, K, Ca, Mg, Zn, Cu, Fe, and Mn) in the biochar and BMC
158 were analyzed by Inductively Coupled Plasma Optical Emission Spectrometer (ICP – OES)
159 (Optima 3300 RL, Perkin Elmer, USA) after digestion with HNO₃-HCl mixture [3:1(v/v)] in a
160 microwave digestion system (Discover[®] SPD, CEM Corporation, USA). The crystal structure,
161 morphology, and surface chemistry of BMC were studied through X-ray diffraction (XRD),
162 scanning electron microscopy (SEM), and Fourier transform infrared (FTIR)
163 spectroscopy(Supplementary Information).

164

165 2.3. *Plant growth experiment*

166 A pot culture study was conducted under natural conditions in a net house. A deeply weathered
167 acidic soil was selected for the experiment, and the bulk sample was collected from an
168 experimental field (0-15 cm depth) of the Agriculture Research Station at Kadapa, Andhra
169 Pradesh, India. The soil sample was air-dried in the laboratory, grinded, and sieved through a 2-
170 mm sieve for the analysis of initial properties. The experimental soil was light in texture (sandy
171 loam) with pH=6.9, EC=0.18 dS m⁻¹, and organic C (OC) = 3.89 g kg⁻¹. The soil belongs to the
172 order Alfisol, and subgroup Typic Haplustalf (Soil Survey Staff, 2010). Fertility status of the soil
173 was low with available N, P and K contents of 62.8, 12.7 and 72.4 mg kg⁻¹, respectively.
174 A medicinal herb, senna (*C. angustifolia*), was grown in the pot experiment. Senna is widely
175 used in traditional and modern medicines. India is the main producer of senna leaves and pods
176 (economic parts of the plant for medicinal use), and holds a monopoly in the international market
177 with an annual turnover of ₹300 million (Basak and Gajbhiye, 2018). Cultivation of aromatic and
178 medicinal plants in degraded marginal soils, as chosen in this study, could fetch both economic
179 and environmental benefits, and would be easily adopted by farmers (Basak and Gajbhiye,
180 2018). Medicinal and aromatic plants need less but a steady supply of nutrients for providing the
181 best yield of high-quality products (Basak and Gajbhiye, 2018), which is likely to be met by
182 BMC. FYM and vermicompost were obtained from the Anand Taluka Cooperative, and
183 Livestock Research Station, Anand Agricultural University, respectively. The chemical
184 fertilizers (CF) were applied according to the recommended dose of nutrients for senna crop: 60
185 kg ha⁻¹ N (urea), 40 kg ha⁻¹ P (single super phosphate), and 20 kg ha⁻¹ K (muriate of potash).
186 The six selected treatments comprised of: T₁- Control, T₂ – FYM (5 t ha⁻¹); T₃- Vermicompost (5
187 t ha⁻¹); T₄- Biochar (5 t ha⁻¹); T₅ - BMC (5 t ha⁻¹), and T₆ – CF.

188 Sieved (< 2-mm) 10 kg of soil was taken into earthen pots (32 cm inner diameter and 26 cm
189 height), and mixed thoroughly with different treatment combinations. Senna seeds were sown in
190 four replicated pots of each treatment, and placed in a randomized block design. A single plant
191 per pot was maintained to grow for the period of 120 days. The pots were irrigated at regular
192 intervals to maintain the soil moisture content at field capacity.

193

194 2.4. *Growth and yield parameters of senna plant*

195 The above-ground senna plant biomass was harvested after 120 days of sowing (DAS) which
196 coincided with the flowering (pod formation) stage of the plant. The plant height, number of
197 branches, and fresh and dry herbage (leaf + pod) yields were recorded at harvest. The fresh pods
198 and leaves yield per plant was recorded immediately after harvesting. The dry weight of pods
199 and leaves were recorded when the samples reached a constant weight under shade drying. The
200 shed-dried plant samples were stored for further analysis.

201

202 2.5. *Analysis of plant and soil*

203 Senna leaves and pods are known for their laxative properties due to the presence of sennoside
204 ([Reddy et al., 2015](#)). The sennoside content in pulverized (<0.5 mm) leaves and pods was
205 analyzed on a High-Performance Liquid Chromatography (HPLC) equipment (LC-20AD,
206 Shimadzu Corporation, Kyoto, Japan) following extraction in 70% methanol (Supplementary
207 Information).

208 The major nutrient (N, P and K) contents in the dried (65°C) plant biomass were analyzed by
209 standard methods (Supplementary Information). Soil sample (0-15 cm) from each pot was
210 collected after harvesting of senna plants. Immediately after collection, a set of soil samples was

211 stored separately in a refrigerator (4°C) for the analysis of soil biological parameters. Another set
212 of samples was air dried, grinded, and passed through a 2-mm sieve, and then analyzed for pH,
213 EC, OC content, and available nutrients (e.g., N, P and K) (Supplementary Information).
214 Samples were allowed to reach the room temperature after taking out from refrigerator, and used
215 for determination of inorganic N, microbial biomass carbon (MBC), soil respiration, and enzyme
216 activities (e.g., dehydrogenase and alkaline phosphatase) (Supplementary Information). The
217 moisture content of original soil samples was determined, and all results were expressed on oven
218 dry-basis.

219

220 2.6. *Data analysis*

221 The data generated from the laboratory and pot experiments were expressed as the mean of four
222 replicates. An analysis of variance (one-way ANOVA) was conducted according to the
223 experimental design followed (Completely Randomized Design). Statistical significance among
224 treatments was worked out by estimating the critical difference ($P < 0.05$) in SPSS-20 (SPSS
225 Inc., Chicago, USA) software package. Microsoft Excel (Microsoft Corporation, USA) was also
226 used for data calculation, tabulation and graphical representation.

227

228 **3. Results and discussion**

229 3.1. *Physico-chemical properties of amendments*

230 Physico-chemical properties of BMC in comparison with FYM, vermicompost and the pristine
231 biochar are given in Table 1, and Table S2. Among the amendments, the lowest total C was
232 recorded in BMC, but it contained the highest amount of water soluble, and total P and K (Table
233 1). BMC was found slightly acidic (pH=6.85±0.05) in nature as opposed to the alkaline pristine

234 biochar ($\text{pH}=8.12 \pm 0.06$), and it contained the highest CEC among all the amendments (Table
235 1). Low pH value in BMC was resulted likely from the release of organic acids (e.g., humic acid)
236 from FYM during BMC preparation (Sekhar et al., 2001). The CEC value [$66.1 \pm 2.3 \text{ cmol}(\text{p}^+)$
237 kg^{-1}] exhibited by BMC (Table 1) was greater than an enriched biochar [$40.67 \text{ cmol}(\text{p}^+) \text{ kg}^{-1}$]
238 reported earlier (Chia et al., 2014). Compared to the pristine biochar, total C content in the BMC
239 was low (Table S2) because the mineral addition created a dilution effect for the C content. In
240 addition, reaction of the biochar with phosphoric acid might have reduced the C content too by
241 an individual or combined action of (i) biochar surface oxidation, (ii) carbonate C elimination,
242 and (iii) inorganic P precipitation (dilution effect) (Chia et al., 2014). The hydrothermal reaction
243 of biochar with RP and waste mica most likely increased the available, and total P and K
244 contents in BMC (dos Santos Teixeira et al., 2015; Dissanayake et al., 2018). The SSA of the
245 BMC was $26.7 \text{ (m}^2 \text{ g}^{-1})$ with a total pore volume of $2.407 \text{ (mL g}^{-1})$ (Table S2). The SSA of BMC
246 was comparatively lower than a bamboo-based biochar-montmorillonite composite ($156 \text{ m}^2 \text{ g}^{-1}$)
247 (Viglasova et al., 2018), but higher than a municipal solid waste biochar-montmorillonite
248 composite ($6.51 \text{ m}^2 \text{ g}^{-1}$) (Ashiq et al., 2019). Riddle et al. (2019) reported that the reduction of
249 SSA in biochar-mineral composite compared to pristine biochar could be due to the coating of
250 biochar with metal hydr(oxide)s (e.g., magnesium hydroxide).

251

252 3.2. Characterization of BMC

253 3.2.1. Scanning electron microscopy

254 RP, waste mica, and kaolinite were mixed with the biochar, and thus the minerals were
255 incorporated into the biochar structure in BMC. Surface morphology of the pristine biochar was
256 random, disordered and porous in nature (Fig. 1a). The SEM image of BMC at 1000X

257 magnification showed a thin film-like structure covering the BMC surface (Fig. 1b). At 5000X
258 magnification, the thin film-like structure appeared to be a further prominent layered coating on
259 the BMC surface (Fig. 1c). This type of layered coating is common for the surface morphology
260 of clays, as reported in the literature (Ashiq et al., 2019a; 2019b). SEM images of BMC (Fig. 1b
261 and c) also showed partial filling of biochar pores with mineral particles that entered into the
262 biochar pores (Premarathna et al., 2019). The coarse mineral particulates coated the biochar
263 surface giving it a rough texture with irregularly shaped agglomerates and flaky structures, which
264 is again a confirmation of clays being incorporated in the BMC (Fig. 1b and 1c). Nevertheless,
265 the entire biochar surface was not covered by the mineral particles, and some pores were still
266 accessible. Pores originated from the raw biomass are important as they provide habitats for
267 beneficial microorganisms, and also retain and recycle nutrients to improve the soil fertility
268 (Viglasova et al., 2018). Partial covering of biochar pores in BMC was rather advantageous
269 because an excessive clay coating could have made the pores inaccessible to microorganisms
270 (Premarathna et al., 2019), hampering the nutrient cycling. The partial coating obtained in this
271 study thus could be ideal for the purpose of soil fertility improvement.

272

273 3.2.2. *Fourier transform infrared spectroscopy*

274 Considerable difference was observed between the FTIR spectra of pristine biochar and BMC
275 (Fig. 2). Compared to the pristine biochar, the presence of a sharp band at around 1034 cm^{-1} in
276 the BMC spectrum indicated the incorporation of Si-O functional groups onto the BMC, possibly
277 coming from the silicate minerals. Bands around $1030\text{-}1040\text{ cm}^{-1}$ confirmed an out-of-plane
278 deformation of the -CH groups, symmetric stretching vibration of C-O groups of cellulose,
279 hemicellulose and lignin, and O-P-O bending vibration due to the incorporation of RP. A small

280 band at around 915 cm^{-1} represented the Al-OH deformation of kaolinite clay mineral (Saikia et
281 al., 2016), which confirmed the incorporation of the mineral components on the BMC surface
282 (Darby et al., 2016; Premarathna et al., 2019). Sharp and intense bands for BMC at 3693 and
283 3619 cm^{-1} , which were absent in the pristine biochar, could also be attributed to the inclusion of
284 kaolinitic clay onto the BMC surface (Darby et al., 2016), and a broader band at around 3422
285 cm^{-1} indicated the stretching vibration of O-H groups of bonded water (Darby et al., 2016).
286 Bands at 794 cm^{-1} indicated the presence of silica (Darby et al., 2016). Representative bands for
287 aliphatic carbon, a shoulder like peak at about 1627 cm^{-1} might be associated with the C=O
288 stretching of ketone and carboxylate derivatives (Joseph et al., 2013). As compared to the
289 pristine biochar, aliphatic C-H stretching bands were missing at around 2925 cm^{-1} in the
290 spectrum of BMC, which was a characteristic feature of BMC. Li et al. (2018) reported that the
291 incorporation of major minerals (e.g., kaolinite, metakaolin, and quartz) could hinder the
292 expression of some functional groups from carbon fractions in the spectra of biochar-composite
293 materials. The intense bands at 600 cm^{-1} to 400 cm^{-1} might be associated with the stretching
294 vibrations of metal-O, halogen stretching, and phosphate bands (Li et al., 2017), which possibly
295 originated from the mineral materials added into the BMC.

296

297 3.2.3. X-ray diffraction

298 The XRD pattern of the pristine biochar revealed a reflection at $2\theta = 28^\circ$, which was the only
299 observable reflection of a crystallographic structural phase. Compared to the pristine biochar,
300 several new and intense reflections were observed in the XRD pattern of BMC (Fig. 3),
301 indicating that biochar formed a complex with minerals. The presence of kaolinite in the BMC
302 was confirmed by the sharp reflection at $2\theta = 12^\circ$ and 25° (Fig. 3), which was in line with the

303 SEM and FTIR results establishing the complex formation (Liu et al., 2019). Additionally, the
304 reflections at $2\theta = 27^\circ$ indicated the presence of quartz (SiO_2), and reflection at $2\theta = 32\text{-}34^\circ$
305 corresponded to calcite (CaCO_3) (Zhao and Narthey, 2014). Thus, a number of reflections in the
306 XRD pattern re-confirmed the formation of BMC. The presence of diffuse and broad reflections
307 at around $2\theta = 5$ to 70° were indicative of short-range order carbon structures in the BMC.
308 Moderate temperature pyrolysis (350°C) in the present study would have resulted in biochar with
309 a limited extent of order in the resultant carbon structure (Khan et al., 2015).

310

311 3.3. Plant growth and senna yield

312 Senna plant height and number of branches were significantly ($P < 0.05$) increased by
313 conventional organic treatments (FYM and vermicompost) over the control (Table 2). However,
314 the highest plant height (74.6 cm), and number of branches (16.7) were observed with the
315 application of BMC followed by chemical fertilizers (CF). Application of biochar and
316 vermicompost significantly improved the plant growth over the control, and FYM. However,
317 plant growth parameters recorded in the case of biochar and vermicompost treatments were
318 significantly lower than BMC. Treatment receiving CF was found effective in improving plant
319 growth parameters better than the control and FYM, but at par with biochar and vermicompost.
320 The growth and vigor shown by the treatments, particularly biochar, vermicompost and BMC,
321 were more effective in improving the total biological yield, which was also reflected in the
322 herbage yield (leaf + pod) of senna. Application of FYM, vermicompost, biochar and BMC
323 significantly increased the fresh as well as dry herbage yields over the control (Table 2).
324 Treatment receiving BMC (5 t ha^{-1}) recorded the highest fresh ($176.1 \text{ g plant}^{-1}$), and dry (57 g
325 plant^{-1}) herbage yields followed by CF. Application of biochar and vermicompost were found

326 equally effective to CF in improving the herbage yield. However, the BMC treatment recorded
327 18.1% higher total herbage yield than same rate of biochar application (5 t ha⁻¹).
328 Herbage (leaf +pod) is the main economic part of senna, and BMC had a more profound
329 influence on plant growth and herbage yield than biochar and vermicompost. The enriched
330 nutrient composition of BMC prepared from natural minerals (Table 1; Table S2) might have
331 immediate effect on plant growth and herbage yield. During BMC preparation, some physico-
332 chemical properties of biochar were improved (Table 1) due to the formation of organo-mineral
333 complexes, which likely contributed to high CEC and available nutrients (Chia et al., 2014). Due
334 to the improvement of CEC, BMC would retain plant nutrients more efficiently in soil than
335 biochar (Lin et al., 2013; Archanjo et al., 2017). Moreover, it could have a significant positive
336 influence on soil biological activity (Ye et al., 2016). So, the contribution of BMC in terms soil
337 quality improvement might have attributed to the highest plant growth and yield. The CF
338 treatment could not perform well (statistically at par with BMC) due to low use efficiency of
339 nutrient from chemical fertilizers in the highly weathered tropical soil (Basak, 2019b). The BMC
340 in this case would have behaved as a slow-release nutrient source as opposed to chemical
341 fertilizers (Farrar et al., 2019; Tumbure et al. 2020).

342

343 3.4. Bioactive compound production

344 Different treatments significantly influenced the leaf and pod sennoside contents (Fig. 4).
345 However, biochar and BMC were found more effective in improving the sennoside content in
346 leaf and pod than FYM and vermicompost. The treatment receiving BMC recorded the highest
347 sennoside content, both in leaf (2.73%) and pod (3.15%), which were significantly ($P < 0.05$)
348 higher than CF treatment. Treatments comprised of control, FYM, vermicompost, and CF did not

349 show any difference in sennoside contents. However, significant variation in sennoside yield was
350 found particularly in leaf, and total sennoside yield in individual treatment. Application of
351 vermicompost, biochar, BMC and CF significantly improved the sennoside yield over the
352 control. The sennoside yield ($1.58 \text{ g plant}^{-1}$) in the case of BMC was 15.9 and 12.8% higher than
353 CF and pristine biochar applications, respectively.

354 The sennoside yield results indicated that nutrient availability was indispensable for the
355 production of sennoside by senna plants irrespective of the nutrient sources (biochar, fertilizer, or
356 other organic amendments). The synergistic effects of nutrient availability and soil environment
357 improvement due to BMC application probably stimulated the plant growth and physiological
358 processes, which might have attributed to the highest sennoside content with BMC. BMC likely
359 improved the transformation, retention, and use efficiency of nutrients (Lin et al., 2013; Chia et
360 al., 2014), and also played a synergistic role in soil microbial colonization (Ye et al., 2016),
361 which most possibly influenced the primary plant metabolism, and consequently diversion to
362 secondary metabolites. Results of the present investigation were also supported by previous
363 reports, where sennoside content in senna (Basak and Gajbhiye, 2018), withanolide content in
364 ashwagandha (*Withania somnifera*) (Pratibha et al., 2013), and andrographolide content in
365 kalmegh (*Andrographis paniculata*) (Saha et al., 2019) were enhanced respectively by enriched
366 compost, bio-augmented vermicompost, and biochar applications.

367

368 3.5. Nutrient contents in plant tissues

369 Application of vermicompost, biochar, BMC, and CF significantly improved the plant nutrition
370 (N, P and K contents) over the control (Fig. 5). The highest N (1.23%), P (0.196%), and K
371 (0.691%) contents were recorded with the treatment receiving BMC. However, no significant (P

372 < 0.05) difference in N content was observed due to application of vermicompost, biochar, BMC,
373 and CF. Similar trend was also observed as in the case of plant nutrient uptake (Fig. 5).
374 Application of BMC recorded much higher P and K uptake over the control, FYM, and biochar,
375 but at par with CF treatment.
376 Results of the present study thus suggested that BMC was a superior nutrient source than FYM,
377 vermicompost, and biochar. Significantly higher nutrient contents in senna plant was recorded in
378 the treatment receiving BMC, which might be due to better availability of the nutrients in the soil
379 under this treatment, and also due to the effective supply of nutrients on plant demands. This
380 might be explained by the higher nutrient contents (Table 1; Table S2) in BMC than other
381 organic treatments. These results agree with a previous study (Ye et al., 2016) where a BMC
382 (made of wood biochar, chicken manure, and kaolinitic clay) performed better than chicken
383 manure compost in improving nutrient uptake by pakchoi (*Brassica rapa* L., ssp. *chinensis*).
384 Similarly, Basak and Gajbhiye (2018) reported enhanced senna plant growth and P uptake under
385 RP-enriched compost application. Therefore, BMC might have triggered the senna plant growth
386 and nutrition, which ultimately reflected in the total plant nutrient uptake.

387

388 3.6. Soil pH, EC and organic C

389 Notable changes in soil pH, EC, and OC were found due to the application of BMC, organic
390 amendments and CF after the harvest of senna (Table 3). Application of FYM, vermicompost
391 and biochar significantly improved soil pH, EC, and OC than the control and CF. However,
392 treatment receiving biochar and BMC showed distinctly higher soil pH, EC and OC values as
393 compared to other treatments. The highest soil pH (7.2), EC (0.34 dS m⁻¹), and OC (4.15 g kg⁻¹)
394 was recorded with the treatment receiving biochar, which was statistically at par with BMC.

395 Increase in EC and OC of soil was expected due to biochar and BMC because the materials
396 themselves had substantially higher EC and total C content (Table 1; Table S2) than that of the
397 experimental soil. The alkaline nature and acid neutralizing capacity of biochar and BMC might
398 have a role in increasing the soil pH (Lin et al., 2013; Saha et al., 2019). Previous studies too
399 indicated that biochar and BMC application increased soil EC and OC (Li et al., 2014). During
400 pyrolysis of biomass residues (high temperature), the biochar products would be composed of
401 oxides of Ca, Mg and K, which might have increased the EC of highly weathered soil (Jien and
402 Wang, 2013). The soil OC (SOC) content significantly ($P < 0.05$) improved with organic
403 amendments, biochar and BMC application than the control and CF. SOC contributes
404 significantly in improving soil physical and chemical properties which help in maintaining the
405 soil fertility (Rao et al., 2017). Improvement of SOC after biochar and BMC application
406 indicated their determining roles in increasing SOC which likely boosted the soil properties and
407 plant growth in this study (Ye et al., 2016; Arif et al., 2017).

408

409 3.7. Inorganic N, and available P and K in soil

410 Similar to SOC, available nutrient status was also significantly improved with the application of
411 vermicompost, biochar and BMC as compared to the control (Table 3). The highest inorganic N
412 (47.83 mg kg^{-1}) was recorded with BMC application, which was 9.7 and 14.2% higher than
413 biochar and CF applications. Available P content in soil was found distinctly higher in BMC
414 application, which was 46.2 and 17.1% higher than biochar and CF. The highest available K
415 (84.91 mg kg^{-1}) was recorded in the treatment receiving BMC, which was significantly ($P <$
416 0.05) higher than the biochar (80.23 mg kg^{-1}), and CF (79.37 mg kg^{-1}). As expected, available
417 nutrients would improve significantly in soil with the addition of different organic amendments

418 and chemical fertilizer treatments. An additional increase in nutrient content was observed due to
419 the application of BMC. This phenomenon could be explained by the nutrient composition of
420 BMC (Table 1) which was prepared from natural sources of P and K. However, application of
421 biochar also improved the soil nutrient availability, which indicated that apart from nutrient
422 composition, the physico-chemical properties of biochar and BMC also played an important role.
423 This hypothesis was supported by previous studies (Jien and Wang, 2013) where biochar notably
424 improved the physico-chemical properties of highly weathered soils, maintained the SOC level,
425 and increased nutrient use efficiency. Here, the increase in nutrient efficiency was expected from
426 biochar and BMC applications due to higher nutrient retention and availability in the soil.
427 Simultaneously, nutrient supply might have improved due to the prevention of nutrient loss
428 occurring through leaching and volatilization in highly weathered soils of tropical regions (Major
429 et al., 2010). The BMC was a nutrient rich formulation with improved physico-chemical
430 properties derived from the combinational effects of biochar and minerals through organo-
431 mineral complex formation (Lin et al., 2013). Therefore, BMC had some additional advantages
432 over biochar in terms of nutrient composition, which was reflected in the present study. The
433 outcomes of this study also corroborated with the findings a previous study (Ye et al., 2016)
434 where BMC was found a superior nutrient source than other organic amendments. In the present
435 study, BMC was produced from natural minerals, such as RP and waste mica, which were found
436 effective in improving P (Basak, 2019a) and K (Basak,2019b) use efficiency, respectively, in
437 highly weathered soil. These minerals might have contributed to the overall performance of
438 BMC as a nutrient rich soil amendment.

439

440 3.8. Soil biological properties

441 Application of organic amendments, biochar, BMC, and CF had significant impact on soil
442 biological properties (Fig. 6). A significant boost of MBC was observed in the soils treated with
443 FYM, vermicompost, biochar, BMC, and CF. However, the highest MBC (153.1 mg kg^{-1}) was
444 observed in the treatment receiving BMC, which was significantly ($P < 0.05$) higher than
445 vermicompost, biochar, and CF (Fig. 6a). Generally, FYM and biochar are known for rich source
446 of C, but having low nutrient content. On the other hand, chemical fertilizers are devoid of C
447 source, but rich in mineral nutrients. Here, BMC acted as a balanced source of C and nutrients
448 for soil microbes. The same was reflected in the results showing higher MBC in the soil treated
449 with BMC due to a faster growth and proliferation of soil microbes. Soil respiration (SR) rate
450 also followed a similar trend as MBC, where the treatment receiving vermicompost, biochar and
451 BMC showed significantly ($P < 0.05$) higher SR than the rest of the treatments (Fig. 6b). The
452 highest SR ($2.74 \text{ mg CO}_2\text{-C kg}^{-1} \text{ day}^{-1}$) was recorded in the treatment receiving BMC, however,
453 it was at par with vermicompost and biochar treatments (Fig. 6b). The SR represents the
454 metabolically active microbial population in the soil, which was stimulated by the supply of
455 substrates from vermicompost, biochar, and BMC. This was reflected in this study by higher SR
456 in the treatments receiving vermicompost, biochar, and BMC. In case of soil enzyme activities,
457 both DHA and ALP significantly ($P < 0.05$) increased in the treatment receiving vermicompost,
458 biochar, and BMC as compared to FYM and CF. The application of BMC recorded the
459 maximum DHA ($22.4 \text{ } \mu\text{g TPF g}^{-1} \text{ h}^{-1}$) and ALP ($49.2 \text{ } \mu\text{g PNP g}^{-1}\text{h}^{-1}$) in the soil, which was
460 significantly ($P < 0.05$) higher than vermicompost and biochar (Fig. 6c and 6d). DHA is a
461 metabolic enzyme which might have improved due to the balanced supply of substrates and
462 mineral nutrients from BMC. ALP plays a vital role in the cycling of soil P, and is governed by
463 the external P sources. Here, the application of P through RP in BMC had a much higher effect

464 in improving ALP activity, which was accounted for 38.9% higher than soluble P fertilizer
465 application (Fig. 6d).

466 The results of soil biological activities indicated that BMC was much more effective nutrient
467 source than other organic amendments and chemical fertilizers in the highly weathered soil.
468 MBC is considered as the most sensitive part of SOC, and is significantly influenced by the input
469 supply (Nielsen et al., 2014). Here, application of FYM, vermicompost, and biochar acted as
470 source substrates (C) for soil microbes, which was reflected in high soil MBC (Saha et al., 2019).
471 The high soil MBC was found in the treatments receiving biochar and BMC that stimulated the
472 microbial proliferation due to the supply of substrates, soluble nutrients, and congenial habitat.
473 This result corroborated the finding of other workers where application of enriched compost
474 (Basak, 2017), and enhanced biochar (Nielsen et al., 2014) improved soil MBC. The SR and
475 DHA represent the microbial population, specifically the physiologically active microbes
476 thriving in the soil (Nannipieri et al., 1990). The high SR and DHA with the addition of biochar
477 and BMC could be explained by relatively high substrate availability, which could stimulate soil
478 microbial population. Alkaline phosphatase is a predominant enzyme in soil, and is mainly
479 governed by the soil microbial activities with low P availability (Spohn and Kuzyakov, 2013). In
480 the present study, soluble chemical fertilizers showed a negative impact on ALP because of the
481 abundance of readily available P. On the contrary, the supply of OC as well as insoluble P source
482 (RP) from BMC stimulated the microbes for ALP activity. It was earlier observed that APL was
483 not only influenced by the P source but also by the nature of C source (Spohn et al., 2015).
484 Results of the current work had close conformity with earlier findings, where soil treated with P
485 enriched compost significantly improved ALP both in soil incubation (Basak, 2017), and pot
486 culture (Basak and Gajbhiye, 2018) experiments.

487

488 **4. Conclusions**

489 This study demonstrates successful utilization of waste biomass and low-grade mineral powders
490 for the production of an inexpensive BMC which can be a precursor for environmental
491 sustainability with the potential of phasing out chemical fertilizers in crop production. Blending
492 of low-grade rock phosphate and waste mica minerals with biochar significantly improved the
493 surface properties and nutrient contents of the BMC product. The enriched BMC had a
494 remarkable positive impact on senna plant growth and nutrition, and was even more effective
495 than chemical fertilizers in a highly weathered soil. The BMC also improved the soil quality by
496 enhancing available nutrients and biological activities. Thus, BMC served as an excellent starter
497 fertilizer, and could be a potential alternative of chemical fertilizers in medicinal plant
498 cultivation. However, intensive field experiment is needed to evaluate the full potential of BMC
499 for phasing out chemical fertilizers in order to sustain agricultural production. Nevertheless, this
500 study provides a new advent of BMC as a sustainable and alternative nutrient source for boosting
501 up crop production in highly weathered soil through effective recycling of biomass waste and
502 mineral resources.

503

504 **Acknowledgements**

505 The authors acknowledge ICAR-Directorate of Medicinal and Aromatic Plants Research, Anand,
506 India, for extending facilities required for this study. The authors thank Mr. Chintan Makwana
507 from Soil Chemistry Laboratory for his technical support in various laboratory experiments.

508

509 **Declaration of interest**

510 The authors declare no competing financial interests for this study.

511
512
513
514
515
516
517
518
519
520
521
522
523
524
525
526
527
528
529
530
531
532
533

References

Anda, M., Shamshuddin, J., Fauziah, C.I., 2015. Improving chemical properties of a highly weathered soil using finely ground basalt rocks. *Catena*, 124, 147-161.

Archanjo, B.S., Mendoza, M.E., Albu, M., Mitchell, D.R., Hagemann, N., Mayrhofer, C., Mai, T.L.A., Weng, Z., Kappler, A., Behrens, S., Munroe, P., 2017. Nanoscale analyses of the surface structure and composition of biochars extracted from field trials or after co-composting using advanced analytical electron microscopy. *Geoderma*, 294, 70-79.

Arif, M., Ilyas, M., Riaz, M., Ali, K., Shah, K., Haq, I.U., Fahad, S., 2017. Biochar improves phosphorus use efficiency of organic-inorganic fertilizers, maize-wheat productivity and soil quality in a low fertility alkaline soil. *Field Crops Res.* 214, 25-37.

Ashiq, A., Sarkar, B., Adassooriya, N., Walpita, J., Rajapaksha, A.U., Ok, Y.S., Vithanage, M., 2019a. Sorption process of municipal solid waste biochar-montmorillonite composite for ciprofloxacin removal in aqueous media. *Chemosphere*, 236, 124384.

Ashiq, A., Adassooriya, N.M., Sarkar, B., Rajapaksha, A.U., Ok, Y.S., Vithanage, M., 2019b. Municipal solid waste biochar-bentonite composite for the removal of antibiotic ciprofloxacin from aqueous media. *J. Environ. Manage.* 236, 428-435.

Basak, B.B., 2017. Phosphorus supplying capacity of value added compost prepared from low-grade Indian rock phosphates and crop residue. *Waste Biomass Valor.* 8, 2653-2662.

Basak, B.B., 2018. Recycling of waste biomass and mineral powder for preparation of potassium-enriched compost. *J. Mater. Cycles Waste Manage.* 20, 1409-1415.

Basak, B.B., 2019a. Evaluation of Indian rock phosphates for predicting agronomic potential through chemical and biological methods. *Arch. Agron. Soil Sci.* 65, 1599-1609.

534 Basak, B.B., 2019b. Waste mica as alternative source of plant-available potassium: evaluation of
535 agronomic potential through chemical and biological methods. *Nat. Resour. Res.* 28(3),
536 953-965.

537 Basak, B.B., Gajbhiye, N.A., 2018. Phosphorus enriched organic fertilizer, an effective P source
538 for improving yield and bioactive principle of Senna (*Cassia angustifolia* Vhal.). *Ind.*
539 *Crops Prod.* 115, 208-213.

540 Basak, B.B., Maity, A., Biswas, D.R., 2020. Cycling of natural sources of phosphorus and
541 potassium for environmental sustainability. In: Dontsova, K., Balogh-Brunstad, Z., Le
542 Roux, G. (eds). *Biogeochemical Cycles: Ecological Drivers and Environmental Impact*,
543 American Geophysical Union, USA, pp. 285-299.

544 Chia, C.H., Singh, B.P., Joseph, S., Graber, E.R., Munroe, P., 2014. Characterization of an
545 enriched biochar. *J. Anal. Appl. Pyrol.* 108, 26-34.

546 Chu, G., Zhao, J., Huang, Y., Zhou, D., Liu, Y., Wu, M., Peng, H., Zhao, Q., Pan, B., Steinberg,
547 C.E., 2018. Phosphoric acid pretreatment enhances the specific surface areas of biochars
548 by generation of micropores. *Environ. Pollut.* 240, 1-9.

549 Darby, I., Xu, C.Y., Wallace, H.M., Joseph, S., Pace, B., Bai, S.H., 2016. Short-term dynamics
550 of carbon and nitrogen using compost, compost-biochar mixture and organo-mineral
551 biochar. *Environ. Sci. Pollut. Res.* 23, 11267-11278.

552 Ding, Y., Liu, Y., Liu, S., Li, Z., Tan, X., Huang, X., Zeng, G., Zhou, L., Zheng, B., 2016.
553 Biochar to improve soil fertility. A review. *Agron. Sustain. Dev.* 36, 36.

554 Dissanayake, D.K.R.P.L., Dharmakeerthi, R.S., Karunarathna, A.K., Dandeniya, W.S., 2018.
555 Changes in structural and chemical properties of rice husk biochar co-pyrolysed with
556 Eppawala rock phosphate under different temperatures. *Trop. Agric. Res.* 30, 19-31.

557 dos Santos Teixeira, A.M., dos Santos Garrido, F.M., Medeiros, M.E., Sampaio, J.A., 2015.
558 Effect of thermal treatments on the potassium and sodium availability in phonolite rock
559 powder. *Int. J. Miner. Process*, 145, 57-65.

560 El-Naggar, A., El-Naggar, A.H., Shaheen, S.M., Sarkar, B., Chang, S.X., Tsang, D.C.W.,
561 Rinklebe, J., Ok, Y.S., 2019. Biochar composition-dependent impacts on soil nutrient
562 release, carbon mineralization, and potential environmental risk: A review. *J. Environ.*
563 *Manage.* 241, 458-467.

564 Farrar, M.B., Wallace, H.M., Xu, C.Y., Nguyen, T.T.N., Tavakkoli, E., Joseph, S., Bai, S.H.,
565 2019. Short-term effects of organo-mineral enriched biochar fertiliser on ginger yield and
566 nutrient cycling. *J. Soil Sediments*, 19, 668-682.

567

568 Jha, P., Biswas, A.K., Lakaria, B.L., Rao, A.S., 2010. Biochar in agriculture—prospects and
569 related implications. *Curr. Sci.* 99, 1218-1225.

570 Jien, S.H., Wang, C.S., 2013. Effects of biochar on soil properties and erosion potential in a
571 highly weathered soil. *Catena*, 110, 225-233.

572 Johnson, J.J., 1990. Agricultural liming material. In: Helrich, K. (ed). *Official Method of*
573 *Analysis of the Association of Official Analytical Chemists Arlington, Virginia: AOAC*
574 *Inc*, 1–8.

575 Joseph, S., Graber, E.R., Chia, C., Munroe, P., Donne, S., Thomas, T., Nielsen, S., Marjo, C.,
576 Rutledge, H., Pan, G. X., Li, L., Taylor, P., Rawal, A., Hook, J., 2013. Shifting
577 paradigms: development of high-efficiency biochar fertilizers based on nano-structures
578 and soluble components. *Carbon Manage.* 4, 323–343.

579

580 Khan, A., Rashid, A., Younas, R., 2015. Adsorption of reactive black-5 by pine needles biochar
581 produced via catalytic and non-catalytic pyrolysis. *Arab. J. Sci. Eng.* 40, 1269-1278.

582 Li, F., Cao, X., Zhao, L., Wang, J., Ding, Z., 2014. Effects of mineral additives on biochar
583 formation: carbon retention, stability, and properties. *Environ. Sci. Technol.* 48, 11211-
584 11217.

585 Li, M., Tang, Y., Ren, N., Zhang, Z., Cao, Y. 2018. Effect of mineral constituents on
586 temperature-dependent structural characterization of carbon fractions in sewage sludge-
587 derived biochar. *J. Clean. Prod.* 172, 3342-3350.

588 Lin, Y., Munroe, P., Joseph, S., Henderson, R., 2012. Migration of dissolved organic carbon in
589 biochars and biochar-mineral complexes. *Pesqui. Agropecu. Bras.* 47, 677-686.

590 Lin, Y., Munroe, P., Joseph, S., Ziolkowski, A., Van Zwieten, L., Kimber, S., Rust, J., 2013.
591 Chemical and structural analysis of enhanced biochars: thermally treated mixtures of
592 biochar, chicken litter, clay and minerals. *Chemosphere*, 91, 35-40.

593 Liu, Y., Xu, Z., Hu, X., Zhang, N., Chen, T., Ding, Z., 2019. Sorption of Pb (II) and Cu (II) on
594 the colloid of black soil, red soil and fine powder kaolinite: effects of pH, ionic strength
595 and organic matter. *Environ. Pollut. Bioavail.* 31, 85-93.

596 Lu, J., Yang, Y., Liu, P., Li, Y., Huang, F., Zeng, L., Liang, Y., Li, S., Hou, B., 2020. Iron-
597 montmorillonite treated corn straw biochar: Interfacial chemical behavior and
598 stability. *Sci. Total Environ.* 708, 134773.

599 Mandal, S., Sarkar, B., Bolan, N., Novak, J., Ok, Y.S., Van Zwieten, L., Singh, B.P., Kirkham,
600 M.B., Choppala, G., Spokas, K., Naidu, R., 2016. Designing advanced biochar products
601 for maximizing greenhouse gas mitigation potential. *Crit. Rev. Environ. Sci. Technol.* 46,
602 1367-1401.

603 Major, J., Rondon, M., Molina, D., Riha, S.J., Lehmann, J., 2010. Maize yield and nutrition
604 during 4 years after biochar application to a Colombian savanna oxisol. *Plant Soil*, 333,
605 117-128.

606 Matin, N.H., Jalali, M., Antoniadis, V., Shaheen, S.M., Wang, J., Zhang, T., Wang, H., Rinklebe,
607 J., 2020. Almond and walnut shell-derived biochars affect sorption-desorption,
608 fractionation, and release of phosphorus in two different soils. *Chemosphere*, 241,
609 124888.

610 Nannipieri, P., Grego, S., Ceccanti, B., 1990. Ecological significance of biological activity. In:
611 Bollag, J.M., Stotzky, G. (eds). *Soil Biochemistry* Dekker, New York, pp. 293-355.
612

613 Nielsen, S., Minchin, T., Kimber, S., van Zwieten, L., Gilbert, J., Munroe, P., Joseph, S.,
614 Thomas, T., 2014. Comparative analysis of the microbial communities in agricultural soil
615 amended with enhanced biochars or traditional fertilisers. *Agr. Ecosyst. Environ.* 191, 73-
616 82.

617 Pratibha, G., Korwar, G.R., Venkateswarlu, B., Desai, S., Chary, R., Rao, S., Srinivas, K., Rao,
618 S., Ch, S.R., Amalraj, L.D., Choudhary, D.K., 2013. Utilization of composted bixa shell
619 with different bioinoculants as soil amendment for ashwagandha and bixa growth. *Ecol.*
620 *Eng.* 61, 235-244.

621 Premarathna, K.S.D., Rajapaksha, A.U., Adassoriya, N., Sarkar, B., Sirimuthu, N.M., Cooray,
622 A., Ok, Y.S., Vithanage, M., 2019. Clay-biochar composites for sorptive removal of
623 tetracycline antibiotic in aqueous media. *J. Environ. Manage.* 238, 315-322.

624 Purakayastha, T.J., Bera, T., Bhaduri, D., Sarkar, B., Mandal, S., Wade, P., Kumari, S., Biswas,
625 S., Menon, M., Pathak, H., Tsang, D.C.W., 2019. A review on biochar modulated soil

626 condition improvements and nutrient dynamics concerning crop yields: Pathways to
627 climate change mitigation and global food security. *Chemosphere*, 227, 345-365.

628 Rao, C.S., Indoria, A.K., Sharma, K.L., 2017. Effective management practices for improving soil
629 organic matter for increasing crop productivity in rainfed agroecology of India. *Curr. Sci.*
630 112, 1497-1504.

631 Reddy, N.R.R., Mehta, R.H., Soni, P.H., Makasana, J., Gajbhiye, N.A., Ponnuchamy, M. Kumar,
632 J., 2015. Next generation sequencing and transcriptome analysis predicts biosynthetic
633 pathway of sennosides from Senna (*Cassia angustifolia* Vahl.), a non-model plant with
634 potent laxative properties. *PloS One*, 10, 0129422.

635 Riddle, M., Bergstrom, L., Schmieder, F., Lundberg, D., Condrón, L., Cederlund, H. 2019.
636 Impact of biochar coated with magnesium (hydr)oxide on phosphorus leaching from
637 organic and mineral soils. *J. Soil Sediments*, 19, 1875-1889.

638 Saha, A., Basak, B.B., Gajbhiye, N.A., Kalariya, K.A., Manivel, P., 2019. Sustainable
639 fertilization through co-application of biochar and chemical fertilizers improves yield,
640 quality of *Andrographis paniculata* and soil health. *Ind. Crops Prod.* 140, 111607.

641 Saha, A., Tripathy, V, Basak B.B., Kumar, J., 2018. Entrapment of distilled palmarosa
642 (*Cymbopogon martinii*) wastes in alginate beads for adsorptive removal of methylene
643 blue from aqueous solution. *Environ. Prog. Sustain. Energy*, 37, 1942-53.

644 Saikia, B.J., Parthasarathy, G., Borah, R.R., Borthakur, R. 2016. Raman and FTIR spectroscopic
645 evaluation of clay minerals and estimation of metal contaminations in natural deposition
646 of surface sediments from Brahmaputra river. *Int. J. Geosci.* 7(07), 873.

647 Sanchez-Monedero, M.A., Cayuela, M.L., Roig, A., Jindo, K., Mondini, C., Bolan, N., 2018.
648 Role of biochar as an additive in organic waste composting. *Biores. Technol.* 247, 1155-
649 1164.

650 Sekhar, D.M.R., Aery, N.C., 2001. Phosphate rock with farmyard manure as P fertilizer in
651 neutral and weakly alkaline soils. *Curr. Sci.* 80(9), 1113-1115.

652 Soil Survey Staff, 2010. *Keys to Soil Taxonomy*. 11th Edition. United States Department of
653 Agriculture, Natural Resources Conservation Service, Washington, DC.

654 Spohn, M., Kuzyakov, Y., 2013. Phosphorus mineralization can be driven by microbial need for
655 carbon. *Soil Biol. Biochem.* 61, 69-75.

656 Spohn, M., Treichel, N.S., Cormann, M., Schloter, M., Fischer, D., 2015. Distribution of
657 phosphatase activity and various bacterial phyla in the rhizosphere of *Hordeum vulgare*
658 L. depending on P availability. *Soil Biol. Biochem.* 89, 44-51.

659 Tumbure, A., Bishop, P., Bretherton, M. and Hedley, M., 2020. Co-pyrolysis of maize stover and
660 igneous phosphate rock to produce potential biochar-based phosphate fertilizer with
661 improved carbon retention and liming value. *ACS Sustain. Chem. Eng.* 8, 4178-4184.

662 Viglasova, E., Galambos, M., Dankova, Z., Krivosudsky, L., Lengauer, C.L., Hood-Nowotny,
663 R., Soja, G., Rompel, A., Matík, M., Briancin, J. 2018. Production, characterization and
664 adsorption studies of bamboo-based biochar/montmorillonite composite for nitrate
665 removal. *Waste Manag.* 79, 385-394.

666 Wallace, C.A., Afzal, M.T., Saha, G.C., 2019. Effect of feedstock and microwave pyrolysis
667 temperature on physio-chemical and nano-scale mechanical properties of
668 biochar. *Bioresour. Bioprocess.* 6, 33.

669 Ye, J., Zhang, R., Nielsen, S., Joseph, S.D., Huang, D., Thomas, T., 2016. A combination of
670 biochar–mineral complexes and compost improves soil bacterial processes, soil quality,
671 and plant properties. *Front. Microbiol.* 7, 372.

672 Zhao, B., Nartey, O.D., 2014. Characterization and evaluation of biochars derived from
673 agricultural waste biomasses from Gansu, China. In of the World Congress on Advances
674 in Civil, Environmental, and Materials Research.

675 Zhao, Z., Zhou, W., 2019. Insight into interaction between biochar and soil minerals in changing
676 biochar properties and adsorption capacities for sulfamethoxazole. *Environ. Pollut.* 245,
677 208-217.

678 Zhang, X., Zhang, P., Yuan, X., Li, Y., Han, L., 2020. Effect of pyrolysis temperature and
679 correlation analysis on the yield and physicochemical properties of crop residue
680 biochar. *Bioresour. Technol.* 296, 122318.

681

682

683 **Title of tables**

684 Table 1. Comparative physico-chemical properties of farmyard manure (FYM), vermicompost
685 (VC), biochar (BC), and biochar-mineral-complex (BMC).

686 Table 2. Comparative effects of biochar mineral complex, organic amendments, and chemical
687 fertilizers on the growth and herbage yield of senna (*Cassia angustifolia*).

688 Table 3. Soil physio-chemical properties as influenced by the application of biochar-mineral-
689 complex, organic amendments, and chemical fertilizers.

690

691 **Figure legends**

692 Fig. 1. SEM images of (a) pristine biochar, and biochar-mineral-complex at (b) 1000X
693 magnification, and (c) 5000X magnification.

694 Fig. 2. FTIR spectra of pristine biochar, and biochar-mineral-complex.

695 Fig. 3. XRD patterns of pristine biochar, and biochar-mineral-complex.

696 Fig. 4. Comparative effects of biochar-mineral-complex, organic amendments, and chemical
697 fertilizers on bioactive molecule sennoside in *Cassia angustifolia* plant parts (Bars indicate
698 standard error of mean, n=3).

699 Fig. 5. Primary nutrient contents in *Cassia angustifolia* plant tissues as influenced by biochar-
700 mineral-complex, organic amendments, and chemical fertilizer applications (Bars indicate
701 standard error of mean, n=3).

702 Fig. 6. Comparative effect of biochar-mineral-complex, organic amendments, and chemical
703 fertilizers on soil biological properties under *Cassia angustifolia* growth (Bars indicate standard
704 error of mean, n=3).

705

706 **Tables**

707 Table 1. Comparative physico-chemical properties of farmyard manure (FYM), vermicompost (VC), biochar (BC), and biochar-
 708 mineral-complex (BMC).

	pH	CEC [cmol(p ⁺) kg ⁻¹]	Total C (%)	Mineral N (mg kg ⁻¹)	TKN (%)	WSP (mg kg ⁻¹)	TP (%)	WSK (mg kg ⁻¹)	TK (%)
FYM	6.87 ± 0.09	43.9 ± 1.3	31.9 ± 0.4	57.4 ± 2.7	0.53 ± 0.07	93.7 ± 2.3	0.21 ± 0.05	89.6 ± 1.1	0.47 ± 0.01
VC	7.51 ± 0.10	47.1 ± 2.1	35.8 ± 0.6	143.2 ± 4.8	1.12 ± 0.13	118.1 ± 2.6	0.66 ± 0.05	128.2 ± 1.8	0.72 ± 0.04
BC	8.12 ± 0.06	49.7 ± 1.7	47.6 ± 1.1	5.2 ± 0.3	0.72 ± 0.08	103.7 ± 1.9	0.19 ± 0.03	92.3 ± 1.4	0.68 ± 0.03
BMC	6.85 ± 0.05	66.1 ± 2.3	8.91 ± 0.8	47.5 ± 2.3	0.45 ± 0.07	209.4 ± 2.7	1.38 ± 0.07	143.5 ± 2.1	2.74 ± 0.06

709

710 [Mineral N: (NH₄⁺ + NO₃⁻); TKN: Total Kjeldhal nitrogen; WSP: water soluble P; TP: Total P; WSK: water soluble potassium; Total
 711 K]

712

713 Table 2. Comparative effects of biochar mineral complex, organic amendments, and chemical fertilizers on the growth and herbage
 714 yield of senna (*Cassia angustifolia*).

Treatments	Plant height (cm)	Number of branches plant ⁻¹	Fresh herbage yield (g plant ⁻¹)		Dry herbage yield (g plant ⁻¹)	
			Leaf	Pod	Leaf	Pod
T ₁ : Control	56.4	12.8	119.1	18.4	40.5	4.54
T ₂ : FYM	63.7	13.7	129.4	20.5	42.7	5.09
T ₃ : VC	71.7	14.6	140.5	22.8	45.3	5.83
T ₄ : BC	71.9	14.9	142.8	24.3	46.3	6.15
T ₅ : BMC	74.6	16.7	150.2	25.9	50.7	6.29
T ₆ : RFD	72.9	15.1	143.4	23.8	47.3	6.02
<i>CD (p = 0.005)</i>	2.86	1.72	6.14	1.94	3.12	0.21

715
 716 [FYM: Farmyard manure; VC: Vermicompost; BC: Biochar; BMC: Biochar mineral complex; RFD: Recommended dose of
 717 fertilizers; CD: Critical difference]

718

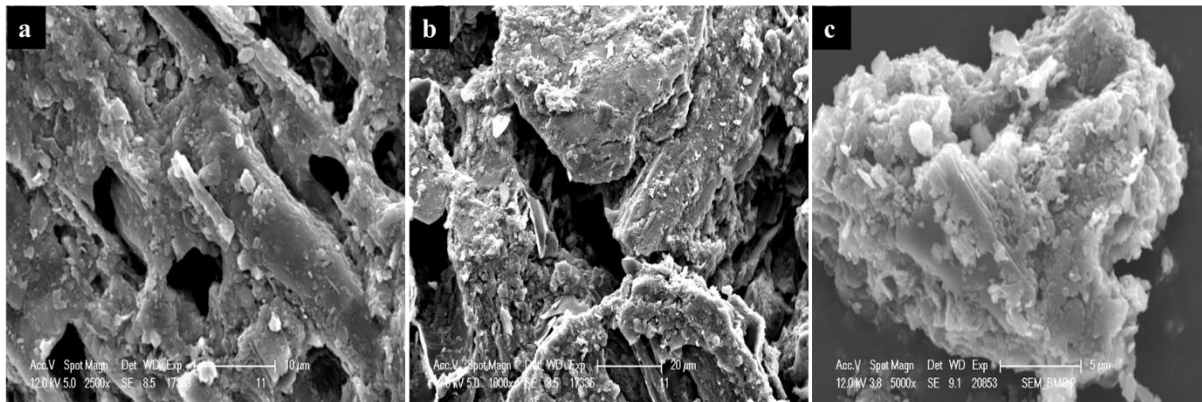
719 Table 3. Soil physio-chemical properties as influenced by the application of biochar-mineral-complex, organic amendments, and
 720 chemical fertilizers.

Treatments	pH	EC (dS m ⁻¹)	SOC (g kg ⁻¹)	Mineral N (mg kg ⁻¹)	Available P (mg kg ⁻¹)	Available K (mg kg ⁻¹)
T ₁ : Control	6.89	0.21	3.91	34.50	13.40	72.03
T ₂ : FYM	7.02	0.25	3.97	37.43	14.37	75.53
T ₃ : VC	7.06	0.26	4.00	42.57	17.20	76.00
T ₄ : BC	7.20	0.34	4.15	43.61	14.63	80.23
T ₅ : BMC	7.17	0.31	4.10	47.83	21.40	84.91
T ₆ : RFD	6.86	0.23	3.85	41.89	18.28	79.37
<i>CD (p = 0.005)</i>	<i>0.17</i>	<i>0.09</i>	<i>0.18</i>	<i>6.09</i>	<i>2.87</i>	<i>4.21</i>

721
 722 [FYM: Farmyard manure; VC: Vermicompost; BC: Biochar; BMC: Biochar mineral complex; RFD: Recommended dose of
 723 fertilizers; CD: Critical difference]

724

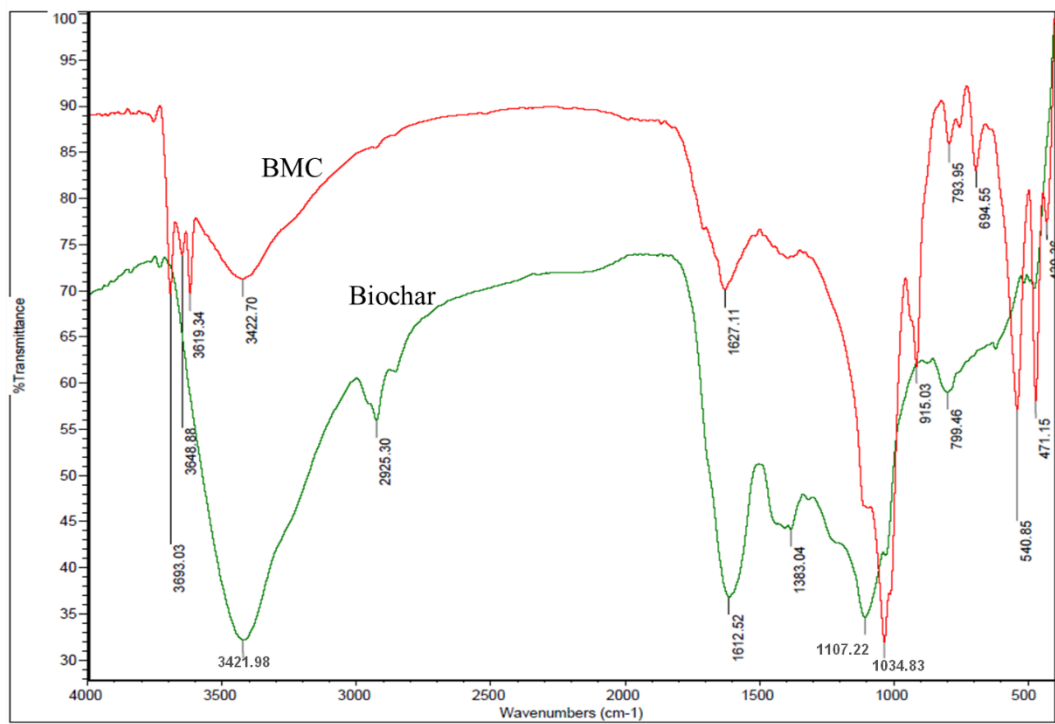
725 **Figures**



726

727 **Fig. 1.**

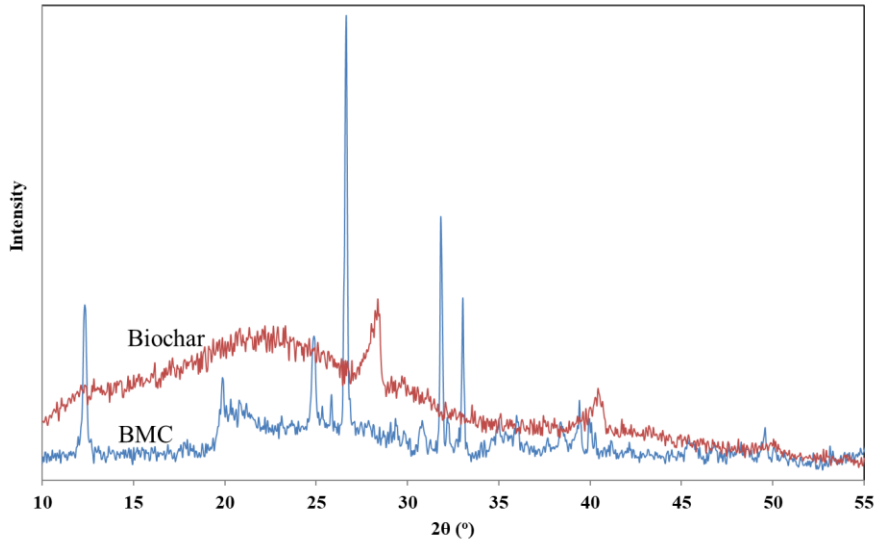
728



729

730 **Fig. 2.**

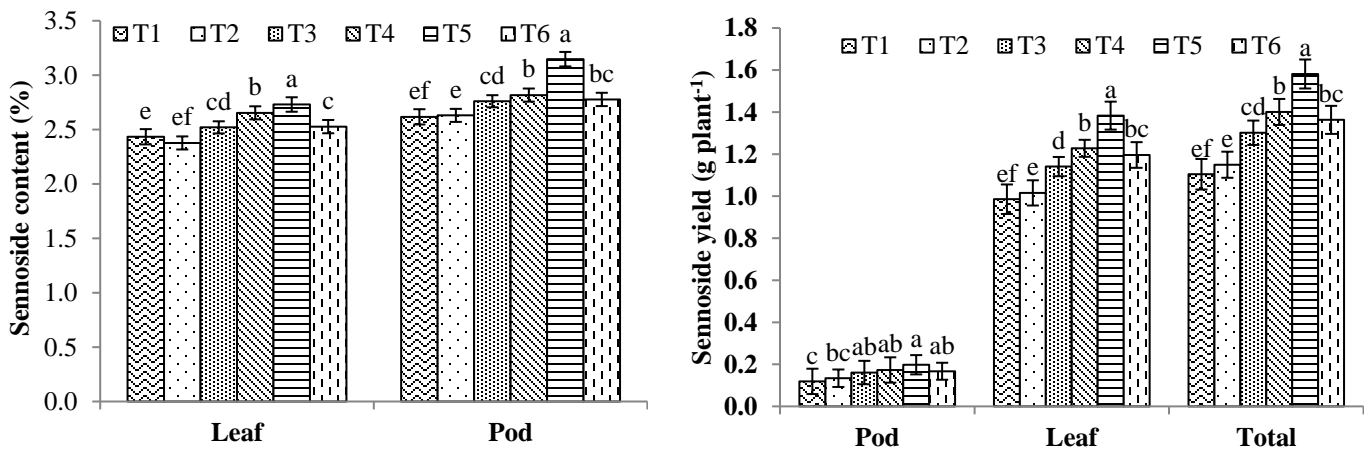
731



732

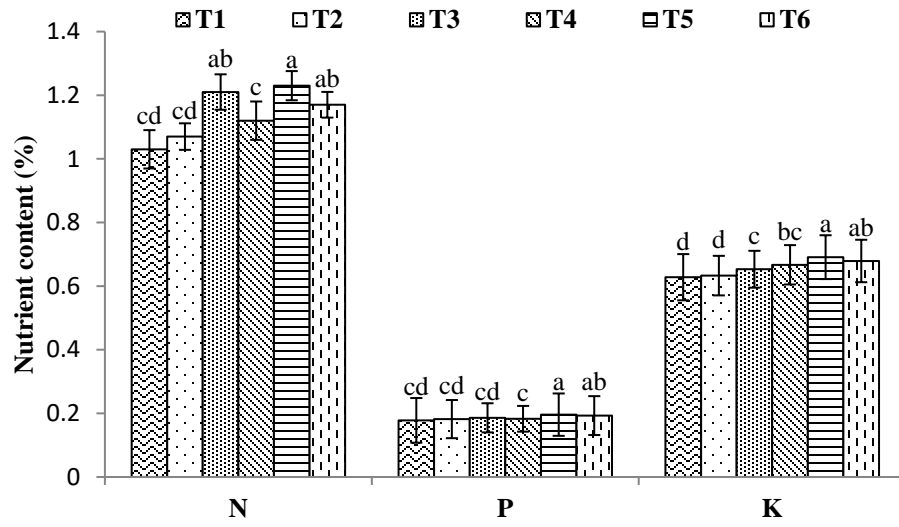
733 Fig. 3.

734



735 Fig. 4.

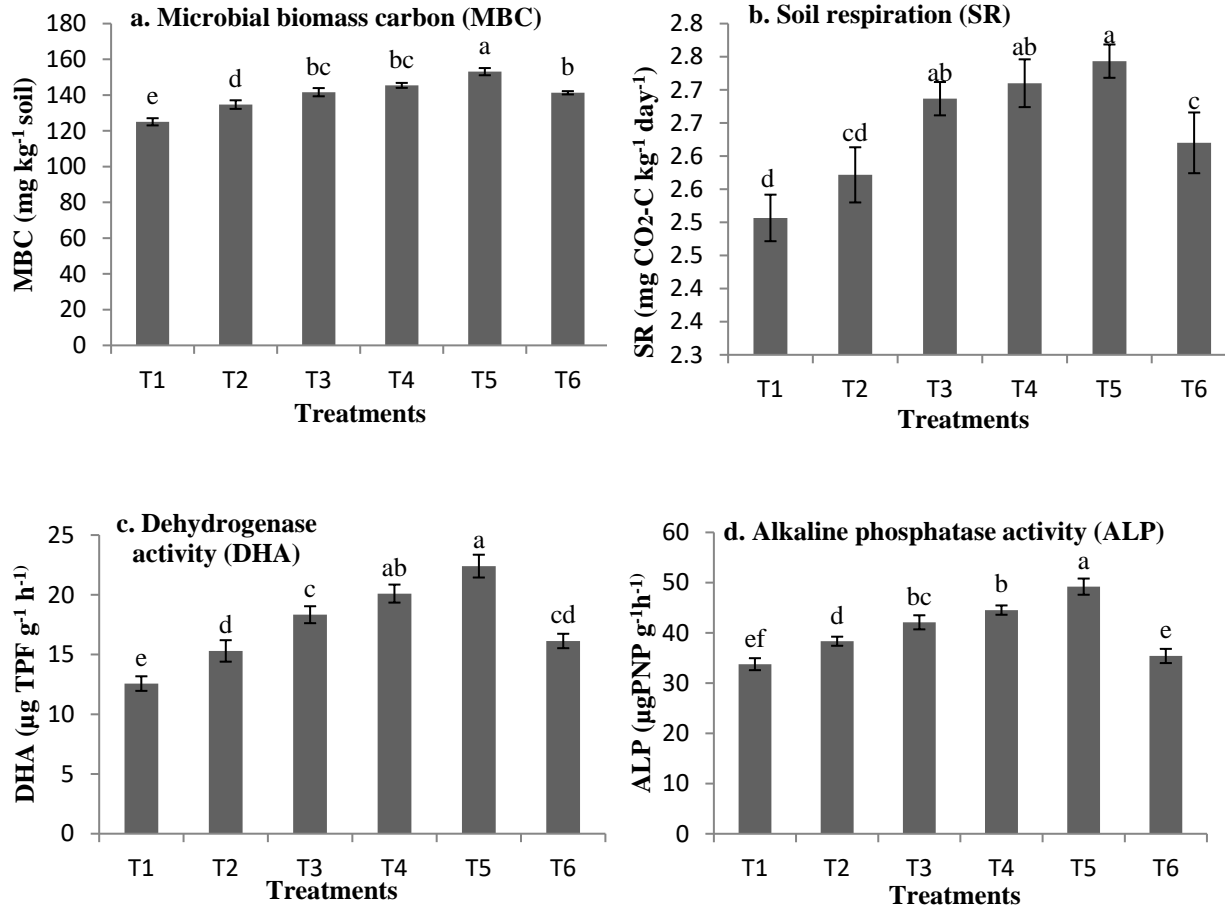
736



737

738 Fig. 5.

739



740 Fig. 6.

741

742 Supplementary Information:

743 **Repurposing distillation waste biomass and low-value mineral resources through biochar-**
744 **mineral-complex for sustainable production of high-value medicinal plants and soil quality**
745 **improvement**

746

747 B. B. Basak^{a¶}, Ajoy Saha^{a,b}, Binoy Sarkar^{c*}, B. Prem Kumar^d, N. A. Gajbhiye^a, Atanu Banerjee^e

748

749 ^a ICAR-Directorate of Medicinal and Aromatic Plants Research, Anand – 387310, India

750 ^b ICAR- Central Inland Fisheries Research Institute, Bangalore Research Centre, Bangalore –
751 560089, India

752 ^c Lancaster Environment Centre, Lancaster University, Lancaster, LA1 4YQ, United Kingdom

753 ^d Department Soil Science and Agricultural Chemistry, Anand Agricultural University, Anand –
754 388110, India

755 ^e Dr. K C Patel Research & Development Centre, Charotar University of Science and
756 Technology, Changa, Anand- 388421, India

757

758 *Corresponding author: Dr Binoy Sarkar; Lancaster University; e-mail: b.sarkar@lancaster.ac.uk

759 ¶e-mail: biraj.basak@icar.gov.in (B. B. Basak; ICAR-Directorate of Medicinal and Aromatic
760 Plants Research)

761

762 ***Structural characterization and surface morphology of BMC***

763 Fourier transform infrared (FTIR) spectra of the pristine biochar and BMC were taken on a
764 Shimadzu IR-Prestige-21[®] spectrometer (Shimadzu, Japan). Potassium bromide (KBr) pellets
765 were prepared by mixing powdered samples, and spectra were recorded in 4000-400 cm⁻¹ scan
766 range.

767 Morphology of the biochar and BMC surface was imaged on a scanning electron microscope
768 (SEM) (Philips XL-30[®], SEMTech Solutions, MA, USA) at different magnifications in order to
769 confirm the complex formation. For this purpose, powdered sample was smeared on specimen
770 stubs, and then gold coated (10 nm) under vacuum before being observed under SEM.

771 The powdered biochar and BMC were also studied by X-ray diffraction (XRD) technique. XRD
772 pattern was acquired on a Philips PW 1710 diffractometer (Royal Philips, The Netherlands) with
773 monochromatic CuK α radiation ($\lambda=1.54 \text{ \AA}$) at generator voltage 40 mA, and tube current 40 kV,
774 with 2θ value ranging from 5 to 80°, where θ is the diffraction angle.

775

776 ***Analysis of bioactive compound in plant tissues***

777 Dried leaf and pod samples were pulverized to make powder (<0.5 mm). Pulverized samples
778 (100 mg) were extracted with aqueous methanol (70%, 20 mL) by sonication (10 min)
779 (Srivastava et al., 1983). Extracts were centrifuged and filtered by syringe filter (0.45 μm), and
780 analyzed on a High-Performance Liquid Chromatography (HPLC) equipment (LC-20AD,
781 Shimadzu Corporation, Kyoto, Japan) configured with SPD-20A UV-Vis detector at 270 nm, an
782 auto sampler (SIL-20AC HT), and GraceAlltima (100 \times 4.6 mm, 3 μm) column with mobile
783 phase consisting of methanol and 1.25% acetic acid in water.

784

785 *Analysis of plant nutrients*

786 For the analysis of N, P and K contents, plant biomass samples were dried at 65°C in an oven,
787 and crushed to powder in a Wiley mill (5-mm size). Samples were digested on an electric hot
788 plate with a di-acid mixture (HNO₃:HClO₄: 9:4) (Piper, 1967). The K content in the digest
789 solution was measured on a flame photometer (Model 128, Systronics, India). The P content in
790 the digest solution was measured on a spectrophotometer (Model 117 Systronics, India) after
791 developing a yellow color complex (vanadomolybdo-phosphate) (Jackson, 1973). For the
792 determination of total N, samples were digested using H₂SO₄ and a catalyst mixture (K₂SO₄:
793 CuSO₄: Se in the ratio of 200:10:1) at 400 ± 5°C for one hour in a micro-Kjeldahl digestion
794 system (Model KES20L, Pelican Equipment, India). The digested samples were distilled in the
795 micro-Kjeldahl system to measure the N content.

796

797 *Analysis of soil samples*

798 Soil pH and EC were analyzed on a digital pH and EC meter (Aquamax KF, GR Scientific, UK),
799 respectively, by suspending the soil in double distilled water (1:2.5(w/v) soil: water ratio)
800 (Jackson, 1973). Soil OC was measured by rapid titration following the Walkley-Black method
801 (Nelson and Sommers, 1996). Inorganic N (NH₄⁺ + NO₃⁻) was estimated by extracting the soil
802 sample with 2 M KCl solution (Keeney and Nelson, 1982) followed by micro-Kjeldahl
803 distillation. Available P (AP) was analyzed after extracting the soil with Mehlich I extractant
804 solution (Mehlich, 1953), and the P concentration in the extract was measured by the
805 spectrophotometer after developing blue color with ammonium molybdate (Watanabe and Olsen,
806 1965). Available K (AK) content in soil was analyzed by extracting with neutral 1N ammonium

807 acetate (NH₄OAc) solution (Hanway and Heidel, 1952) followed by estimation of K on the
 808 Flame Photometer.
 809 Chloroform fumigation-extraction method (Jenkinson and Powlson, 1976) was used for the MBC
 810 assay. Soil respiration (SR) was estimated by measuring CO₂-C released during incubation by
 811 alkali trap method (Anderson, 1982). Dehydrogenase activity (DHA) was assayed by measuring
 812 pink color intensity on the spectrophotometer due to the production of triphenyl formazan from
 813 triphenyl tetrazolium chloride (Klein et al., 1971). Alkaline phosphatase activity (ALP) in soil
 814 was estimated by measuring the yellow color intensity on the spectrophotometer due to the
 815 release of *p*-nitrophenol from *p*-nitrophenol phosphate substrate (Tabatabai and Bremner, 1969).
 816

817 ***Supplementary tables***

818 Table S1. Raw materials used for the preparation of enriched biochar mineral complex.

Raw material	Description	Content by weight (g)
Biochar	Lemon grass (<i>Cymbopogon flexuosus</i>) biochar, produced at 350°C for 2 h, and treated with phosphoric acid	32
Clay	Air dried kaolinite (Kaolin) procured from Molychem, Mumbai	30
Organic matter	Farm yard manure (FYM) obtained from local cattle shed at Anand	20
Mineral powder	Low-grade minerals such as rock phosphate and waste mica. - Rock phosphate from Rajasthan State Mines and Minerals Ltd., Udaipur, India	12.5

- Waste mica from mica mine of Koderma, Jharkhand, 5.5
India

Water Double distilled water 30

819

820

821 Table S2. Selected characteristics of biochar mineral complex (BMC) prepared from distillation

822 waste biomass of lemon grass and low-grade rock phosphate and waste mica.

Properties (dry weight basis)	BMC*
BET Surface area (m ² g ⁻¹)	26.7 (0.78)
Total pore volume (mL g ⁻¹)	2.41 (0.17)
pH	6.85 (0.03)
Electrical conductivity (d Sm ⁻¹)	0.62 (0.06)
Cation exchange capacity (c mol (p ⁺) kg ⁻¹)	66.1 (1.21)
ANC (% CaCO ₃ equivalent)	10.3 (0.21)
Total Carbon (%)	8.91 (0.3)
Total Nitrogen (g kg ⁻¹)	4.51 (0.08)
Total P (g kg ⁻¹)	13.8 (0.43)
Total K (g kg ⁻¹)	27.4 (0.47)
Ca (g kg ⁻¹)	30.8 (1.37)
Mg (g kg ⁻¹)	3.54 (0.83)
Zn (mg kg ⁻¹)	11.4 (0.23)
Cu (mg kg ⁻¹)	13.1 (0.41)
Fe (mg kg ⁻¹)	12212 (4.21)

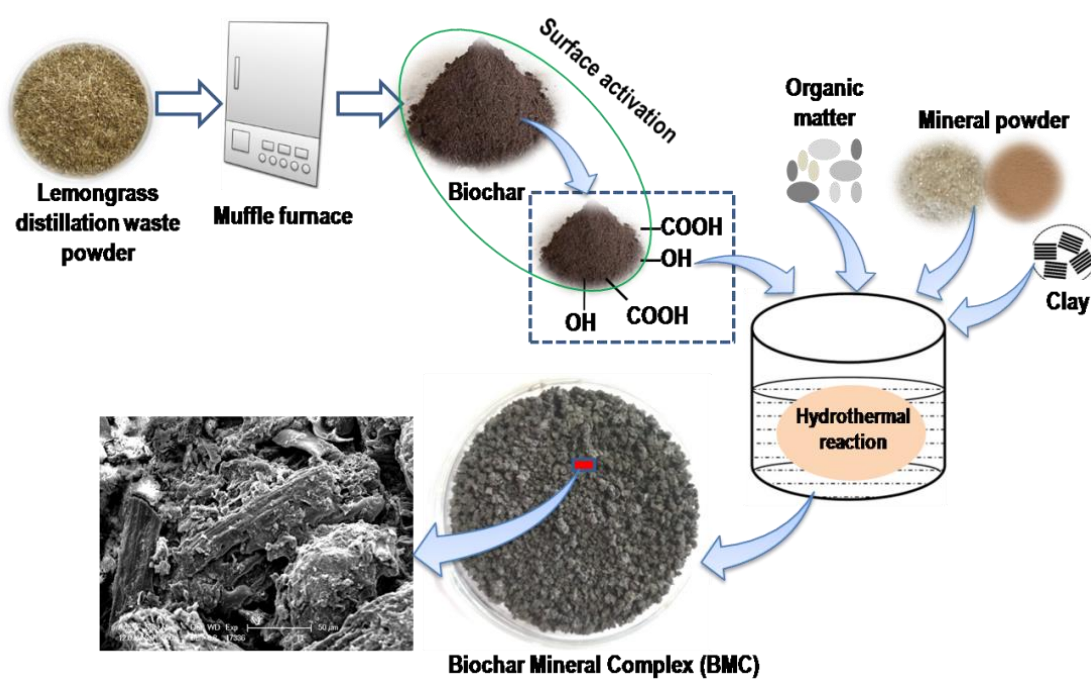
823

824 *Value in the parenthesis indicates (\pm SE).

825

826 *Supplementary figure*

827



828

829 Fig. S1. Schematic diagram showing the preparation steps of biochar mineral complex.

830

831 *References*

832 Anderson, J.P.E., 1982. Soil respiration. In: Page, A.L., Miller, R.H., Keeney, D.R. (eds).

833 Methods of Soil Analysis, Part 2: Chemical and Microbiological Properties, 2nd edition,

834 American Society of Agronomy and Soil Science Society of America, Madison,

835 Wisconsin, pp. 831-871.

836 Hanway, J.J., Heidel, H., 1952. Soil analysis methods as used in Iowa state college, Soil Testing
837 Laboratory. Iowa Agric., 57, 1–31.

838 Jackson, M.L., 1973. Soil Chemical Analysis. Prentice Hall India Pvt. Ltd., New Delhi.

839 Jenkinson, D.S., Powlson, D.S., 1976. The effects of biocidal treatment on metabolism in soil. I.
840 Fumigation with chloroform. Soil Biol. Biochem. 8, 167–177.

841 Keeney, D.R., Nelson, D.W., 1982. Nitrogen inorganic forms. In: Page, A.L., Miller, R.H.,
842 Keeney, D.R. (eds). Methods of Soil Analysis. Agronomy monograph 9 Part 2, 2nd
843 edition. American Society of Agronomy, Madison Wisconsin, pp. 643–698.

844 Klein, D.A., Loh, T.C., Goulding, R.L., 1971. A rapid procedure to evaluate dehydrogenase
845 activity in soil of low organic matter. Soil Biol. Biochem. 3, 385–387.

846 Mehlich, A., 1953. Determinations of P, Ca, Mg, K, Na, and NH₄⁺, North Carolina Soil Testing
847 Div. Mimeo. NC Dep. Agric., Raleigh.

848 Nelson, D.W., Sommers, L.E., 1996. Total carbon, organic carbon, and organic matter. In:
849 Sparks, D.L., Page, A.L., Helmke, P.A., Leppert, R.A., Soltanpour, P.N., Tabatabai,
850 M.A., Johnston, C.T., Sumner, M.E. (eds). Method of Soil Analysis, Part 3: Chemical
851 methods. Soil Science Society of America: Madison, WI, USA, pp. 995–996.

852 Piper, C.S., 1967. Soil and Plant Analysis. Asia Publishing House, Bombay, India.

853 Srivastava, V.K., Maheshwari M.L., Mandal, S., 1983. Investigation of chemical assay of
854 sennoside in Senna (*Cassia angustifolia* Vahl.). Int. J. Trop. Agric. 1, 231–238.

855 Tabatabai, M.A., Bremner, J.M., 1969. Use of p-nitrophenyl phosphate for assay of soil
856 phosphatase activity. Soil Biol. Biochem. 1, 301–307.

857 Watanabe, F.S., Olsen, S.R., 1965. Test of ascorbic acid method for determining phosphorus in
858 water and sodium bicarbonate extracts of soils. Soil Sci. Soc. Am. Proc. 29, 677–678.