

# Macroalgal meadow habitats support fish and fisheries in diverse tropical seascapes

**Short running title:** Tropical macroalgal fish and fisheries

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## 2 **ABSTRACT**

Canopy-forming macroalgae can construct extensive meadow habitats in tropical seascapes occupied by fishes that span a diversity of taxa, life history stages and ecological roles. Our synthesis assessed whether these tropical macroalgal habitats have unique fish assemblages, provide fish nurseries, and support local fisheries. We also applied a meta-analysis of independent surveys across 23 tropical reef locations in 11 countries to examine how shifts in macroalgal canopy condition are related to the abundance of macroalgal-associated fishes. Over 627 fish species were documented in tropical macroalgal meadows, with 218 of these taxa exhibiting higher local abundance within this habitat (cf. nearby coral reef) during at least one life history stage. Major overlap (40-43%) in local fish species richness among macroalgal and seagrass or coral reef habitats suggest macroalgal meadows may provide habitat refuges, particularly for macrophyte-associated taxa. Moreover, the prominence of juvenile fishes suggests macroalgal meadows facilitate the triphasic life cycle of many fishes occupying diverse tropical seascapes. Correlations between macroalgal canopy structure and juvenile abundance suggest changes in macroalgal habitat condition can influence the replenishment of tropical fish populations, including the majority of macroalgal-associated fishes that are targeted by commercial, subsistence, or recreational fisheries. While many macroalgal-associated fishery species are of minor commercial value, their local importance for food and livelihood security can be substantial (e.g., up to 60% of landings in Kenyan reef fisheries). Since macroalgal canopy condition can vary substantially with sea temperature, there is a high likelihood that climate change will impact macroalgal-associated fish and fisheries.

24 **Keywords:** coral reef, nursery, ontogenetic migration, recruitment, *Sargassum*, seagrass

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## 1. INTRODUCTION

26 Conservation and management of fish biodiversity requires an understanding of the habitats  
needed to support and replenish all of the species in a region of interest. While some species  
28 may be uniquely linked to a certain habitat type, many fish taxa follow a triphasic life cycle,  
where planktonic larvae settle into an initial habitat before migrating to different habitats as  
30 juveniles and/or adults. Moreover, adult fishes often move among habitats over daily or  
longer time-scales to fulfil foraging or reproductive activities. Characterisation of a fauna  
32 according to surveys within a single habitat type, therefore, can lead to a conclusion that a  
collection of species are dependent on that habitat type. A wider seascape perspective that  
34 tracks the abundance and activities of fishes across different patch habitat types is needed  
to reveal the full suite of connected habitats that sustain fish populations and communities  
36 (Brown et al., 2018; Olds et al. 2018; Sambrook et al., 2019).

38 Tropical seascapes often comprise a mosaic of patch habitats created by corals, seagrass,  
sponges, mangroves, and canopy-forming macroalgae, any of which may be utilised by  
40 fishes. Considerable effort has been devoted to understanding the fish-habitat functions  
performed by some of these patch types, particularly corals (e.g., Coker et al., 2014),  
42 seagrass (e.g., Gillanders, 2006), and mangroves (e.g., Faunce & Serafy, 2006). Despite the  
long-recognised importance of macroalgae for fish and fisheries in temperate waters  
44 (Bertocci et al., 2015), comparatively little attention has been directed to how tropical  
macroalgal habitats may influence patterns of fish diversity, replenishment and fisheries  
46 production (Fig. 1).

48 Macroalgal meadows can cover large areas of tropical seascapes (16-46% of shallow  
waters; Fulton et al., 2019) to provide food and shelter for shallow-water tropical fishes.  
50 Typically, these meadows are dominated by canopy-forming macroalgae (e.g., *Sargassum*,  
*Sargassopsis*, *Sirophysalis*, *Turbinaria*) with various understory genera (e.g., *Lobophora*,  
52 *Dictyota*, *Padina*) and an abundant epifaunal community, which provides a diverse prey base  
for higher-order consumers such as fish (Bittick et al., 2019; Tano et al., 2016; Fulton et al.,  
54 2019). Indeed, emerging evidence suggests these complex macroalgal meadows are  
occupied by diverse assemblages of tropical fishes at various life history stages (e.g., Ornellas  
56 & Coutinho, 1998; Rossier & Kulbicki, 2000; Eggertsen et al., 2019), some of which may  
support local fisheries (Hicks & McClanahan, 2012; Robinson et al., 2018). These tropical  
58 macroalgal meadows, however, are dynamic habitats that can vary dramatically in canopy  
structure across seasons and years (Fulton et al. 2019). Although individual studies have  
60 documented how changes in canopy condition can influence the abundance of certain  
macroalgae-associated fishes (e.g., Ornellas & Coutinho, 1998; Aburto-Oropeza et al., 2007;  
62 Wilson et al., 2017), large-scale assessments are needed to understand the generality and  
nature of these macroalgal habitat effects on patterns of tropical fish diversity, abundance  
64 and replenishment.

66 In this synthesis we assess whether: (i) tropical macroalgal meadows have unique fish  
assemblages based on overlap in species occurrence and relative abundance with adjacent  
68 coral and seagrass habitats (Section 2), (ii) macroalgal habitats provide fish nurseries in  
tropical seascapes (Section 3), and (iii) macroalgal-associated species support tropical  
70 fisheries (Section 4). We also used a meta-analysis to assess (iv) the relationship between  
macroalgae habitat condition and the abundance of macroalgal-associated tropical fishes

72 (Section 5). To do this we collated data from 23 independent studies in 11 countries (Fig. 2)  
that conducted underwater visual surveys across at least two adjacent habitat types within a  
74 tropical seascape (see Methods in Supporting Information for full details). When referring to  
habitats, we mean areas dominated by the biogenic components of canopy-forming  
76 macroalgae (macroalgal meadows), live hard corals (coral reef), or seagrass (seagrass beds).  
Due to data availability the majority of our analyses focused on the independent surveys  
78 that recorded the relative abundance of tropical fishes across adjacent areas of macroalgal  
meadow and coral reef (Table S1). In doing so, we identify how and why macroalgal habitats  
80 should be considered in the conservation and management of tropical fish and fisheries, and  
the emerging research fronts that are needed to bridge key knowledge gaps.

82

## **2. TROPICAL MACROALGAL FISHES: A DISTINCT ASSEMBLAGE?**

84 Studies exploring fish community structure in tropical macroalgal habitats have steadily  
increased over the past two decades, yet the majority of such studies have been directed  
86 towards the consequences of coral-algal regime shifts (Fig. 1). In some respects, this has  
skewed perspectives towards tropical macroalgae as a 'degraded' reef state for fishes in  
88 areas where a loss of live coral cover has led to substantial losses of biodiversity (e.g., Feary  
et al., 2007; Graham et al., 2006). However, studies that have documented tropical fish  
90 assemblages in macroalgal meadows co-occurring alongside patches of coral reef and  
seagrass provide a different seascape perspective. Drawing on results from 14 independent  
92 studies that deployed comparable levels of visual survey effort across macroalgal meadows  
and two other tropical habitats (coral and/or seagrass, Table S2), we found the average  
94 proportion of local fish species richness that was only found within tropical macroalgal

habitats was low (18%) relative to nearby coral reef (39%). However, we found the reverse  
96 for fish species found only in macroalgal meadows (40%) versus only in seagrass beds (20%).  
This suggests a sizeable portion of tropical fish biodiversity occupying macrophyte habitats  
98 are unique to macroalgal meadows, and that seagrass and macroalgae are not  
interchangeable habitats for the ecological connectivity of many macrophyte-associated  
100 fishes. Moreover, we found an average of 43% and 40% among-habitat overlap in local fish  
species among tropical macroalgae-coral and macroalgae-seagrass habitats, respectively  
102 (Fig. 3). In some locations this overlap was as high as 60-80% (e.g., Ningaloo, Seychelles;  
Table S2). Since over a third of fish species within a region can occupy both macroalgal and  
104 coral habitats, macroalgal meadows could provide stepping-stones or refuge habitats for  
fishes occupying a diverse tropical seascape subject to disturbance events. Depending on the  
106 trophic diversity of these macroalgal-associated fishes, such overlaps in habitat occupation  
could help stabilise ecosystem structure and function in the face of disturbances affecting a  
108 particular habitat type (e.g., mass-bleaching of corals).

110 Our compilation of fishes detected in tropical macroalgal meadows by 23 independent  
surveys (Table S1) found a broad range of taxonomic diversity, with 627 bony fish species  
112 from 75 families occurring as a juvenile and/or adult at some level of abundance (Table S3).  
At least some evidence suggests that a third of these macroalgal-associated fishes (218  
114 species) had most (more than half) of their local abundance within macroalgal habitats,  
either as juveniles (147 species) or adults (130 species; Table S3). Taking a smaller subset of  
116 species for which we had replicated surveys (at least  $n = 2$  for both life history stages) of  
relative abundance, we identified 44 fish species as being most strongly macroalgae-  
118 associated because one or both life history stages were predominantly abundant in

macroalgal meadows compared to nearby coral reef (Fig. 4). Focusing on the 35 species with  
120 more than half of their adults within macroalgae, which we call macroalgal residents, we find  
a diversity of trophic levels and groups, from herbivores (e.g., *Leptoscarus vaigiensis*, *Siganus*  
122 spp.) and omnivores (*Chrysiptera* spp.) to benthic invertivores (e.g., *Coris*, *Choerodon*,  
*Halichoeres* and *Stethojulis* species, *Lethrinus* spp., *Pseudojuloides elongatus*, *Pteragogus*  
124 *flagellifera*, *Xenojulis margaritaceus*). A relatively small component of this macroalgal  
resident fish fauna (9%) are higher trophic-level carnivores that are known to consume other  
126 fishes (e.g., *Cheilio inermis*, *Echidna nebulosa*, *Epinephelus rivulatus*, *Fistularia commersonii*).  
A larger group of fishes (78 species) also had a substantial proportion (between a quarter to  
128 a half) of their relative adult abundance within tropical macroalgal habitats. This lower  
relative abundance may reflect a more opportunistic or transient occupation of macroalgae,  
130 where these fishes could be exploiting a range of trophic resources, given they span  
herbivores (*Acanthurus*, *Chlorurus*, *Scarus*, and *Siganus* species) to omnivores (*Gerres*  
132 *oyena*), generalist carnivores (e.g., some *Lethrinus*, *Lutjanus*, *Thalassoma*, and Mullidae  
species), and some (8% of 77 species) known to consume other fishes (e.g., *Caranx ignobilis*,  
134 *Gymnothorax*, *Pterois*, *Saurida* and *Synodus* species; Fig. 4, Table S3). Notably, the mean  
trophic level of these resident (3.21, 35 species) and opportunistic (3.18, 78 species) fish  
136 groups is relatively similar. A typical adult fish found in tropical macroalgal habitat appears  
to be targeting invertebrate prey, such as the diverse and abundant epifauna found in  
138 macroalgal canopies (Martin-Smith, 1993; Tano et al., 2016; Wenger et al., 2018). While  
there is a relatively small component of tropical fish diversity that we may consider  
140 dependent on macroalgal habitat, our results point to a larger role of macroalgal-associated  
fishes in the functioning of marine ecosystems. Strong overlap in the occupation of  
142 macroalgal and other habitat types by species operating across several trophic levels

suggests these fishes are functioning as mobile links that consume productivity within  
144 macroalgal meadows and then disperse this across tropical seascapes during foraging and/or  
ontogenetic migrations (Berkström et al., 2013). Accordingly, macroalgal meadows should  
146 be considered one of several key habitats within a diverse and productive seascape that is  
needed to sustain healthy tropical fish populations and communities.

148

An important limitation of our analyses is that by only considering species with data across  
150 multiple independent studies, we are likely to exclude endemic species whose limited range  
inhibits the number of studies of their habitat ecology. For example, recruitment of the  
152 leopard grouper, *Mycteroperca rosacea*, is reliably predicted by *Sargassum* cover (Aburto-  
Oropeza et al., 2007), but this species only occurs in the Eastern Central Pacific and there  
154 were insufficient independent empirical studies for this species to be highlighted in the  
trends discussed above. Similarly, cryptic species are not easily detected using underwater  
156 visual census (Ackerman & Bellwood, 2000), the prominent method for surveying fish  
(Murphy & Jenkins, 2010) and are not recorded frequently enough for inclusion in analyses.  
158 For example, abundance of the wrasse, *Xenojulis margaritaceus*, is known to respond to  
canopy cover and composition (Wenger et al., 2018), but small body size and cryptic  
160 colouration means this species is often not recorded in multi-taxa visual surveys.  
Accordingly, we see the above as a conservative estimate of the number of macroalgal-  
162 associated fish species across tropical reef locations, with a bias towards diurnally-active,  
conspicuous fish species of relatively large body size. Further research to identify small-  
164 bodied cryptic macroalgal-associated fishes is warranted, as these could be a considerable  
component of the overall tropical fish diversity with important implications for trophic flows  
166 of nutrients and energy (Depczynski et al., 2007; Brandl et al., 2019).

### 168 3. TROPICAL MACROALGAL MEADOWS AS FISH NURSERIES

170 While the presence of a high number of juveniles seems an obvious requirement for a  
nursery habitat, this abundance of juveniles is irrelevant to future adult breeding  
populations unless they grow and survive to reproductive age (i.e., recruitment success). As  
172 such, the identification of a fish nursery habitat requires various lines of evidence, including  
the relative density, growth and survival rates of juveniles (Beck et al., 2001; Gillanders et al.,  
174 2003; Dahlgren et al., 2006). This means the connectivity of habitats within a seascape is also  
of key importance for nurseries to be effective in replenishing adult fish populations (Beck et  
176 al., 2001; Berkström et al., 2012; Whitfield, 2017). Marine macrophyte habitats such as  
seagrass and mangroves have long been thought to provide this seascape nursery function  
178 (e.g., Beck et al., 2001; Gillanders et al., 2003; Whitfield, 2017). However, the comparative  
importance and ecological significance of tropical macroalgae as fish nursery habitats have  
180 yet to be fully assessed (Adams et al., 2006; Mellin et al., 2007). Here we synthesise the  
evidence for macroalgal habitats to work alongside other common marine subtidal habitats  
182 to support the life cycles of tropical fishes.

184 Our compilation of relative fish abundance across 23 tropical locations confirms that  
macroalgal habitat use by juveniles is globally widespread and includes a remarkably wide  
186 range of tropical fish taxa. Of the 627 fish species found within macroalgal habitats across  
these locations, 64% (399 species) were present as juveniles (Table S3). Over a third (147) of  
188 the 399 species present as juveniles in macroalgal habitats had their highest proportional  
abundance within macroalgal versus coral reef habitat (Table S3). Notably, several species

190 with a high proportion of their juveniles within macroalgal habitat had the majority of their  
adults occupying nearby coral-dominated reef (e.g., *Cephalopholis boenak*, *Lethrinus*  
192 *atkinsoni* and *L. nebulosus*, *Lutjanus carponotatus* and *L. fluviflamma*, *Stethojulis strigiventer*;  
Table S3). As such, macroalgal habitats seem to provide a key middle step in the triphasic life  
194 cycle of some tropical “coral reef” fishes (Mellin et al., 2007; Wilson et al., 2010; Sambrook  
et al., 2019). Importantly, these macroalgal-coral reef species play vital functional roles in  
196 tropical marine ecosystems (e.g., mesopredator *C. boenak*), and many are targeted by  
commercial and/or recreational fisheries (e.g., *Lethrinus* and *Lutjanus* species).

198

Juvenile survivorship rates are a key quantum for identifying a fish nursery habitat, with  
200 piscivory a major driver of early life history mortality (Beck et al., 2001). We found generally  
fewer piscivorous fish species within macroalgal meadows relative to nearby coral reef (e.g.,  
202 fishes of highest trophic level in Fig 4; Fulton et al., 2019). While this suggests juvenile fish  
are subject to fewer types of piscivores in macroalgal habitats, more information is required  
204 to determine if this translates to lower predation risk. Chief among these requirements is  
whether the local density of the relatively few resident piscivorous fish species is low relative  
206 to alternative habitats like coral reef or seagrass. There is the potential that some other  
resident fish species become facultative fish-feeders during seasonal periods of high juvenile  
208 abundance in macroalgal meadows (Holmes et al., 2012). Indeed, several species we  
identified as generalists/transients (Section 2) are piscivores that could periodically increase  
210 their abundance and foraging time within macroalgal meadows during periods of peak fish  
settlement. While it is possible that juvenile fish are subject to a relatively low diversity of  
212 piscivores relative to coral reef habitats, we have little evidence to conclude that macroalgal  
habitats confer higher rates of juvenile survival. We see this as a key research front that

214 requires investigation of relative rates of predator mortality imposed on juveniles occupying  
a range of subtidal habitats within tropical seascapes.

216

218 Within a given habitat type there is potential for a certain combination of optimum local  
conditions to create patches that are particularly effective fish nurseries that contribute to  
future adult populations (Dahlgren et al., 2006; Nagelkerken et al., 2015). Macroalgal  
220 habitats can vary considerably in canopy condition in ways that help explain differences in  
juvenile fish abundance over space and time. Percent macroalgal canopy cover, canopy  
222 height and/or density, as well as underlying attributes such as understory macroalgal cover,  
live coral and/or degree of underlying reef complexity have been linked to spatial and  
224 temporal variation in juvenile fish abundance (e.g., Eggertsen et al., 2019; van Lier et al.,  
2018; Wenger et al., 2018). Besides direct selection by juvenile fish for certain microhabitat  
226 shelters, these variations in canopy condition are likely to influence the availability of  
preferred prey (either the macroalgae or their epibionts; Lim et al., 2016; Wenger et al.,  
228 2018), with direct consequences for fish growth and survival. Indeed, a combination of  
macroalgal meadow condition and juvenile abundance were key predictors for the future  
230 abundance of *Lethrinus* sub-adults in the eastern Indian Ocean (Wilson et al., 2017); a  
finding that is consistent with an earlier study linking *Sargassum* condition to the  
232 recruitment success of the Pacific Ocean leopard grouper *Mycteroperca rosacea* in the  
eastern Pacific (Aburto-Oropeza et al., 2007). Changes in canopy complexity could also  
234 influence the success of piscivores to capture juvenile fish. However, the evidence for the  
latter remains equivocal, with studies in marine macrophyte habitats finding fish predator  
236 success was either affected greatly by differences in canopy structural complexity, or not at  
all (e.g., Horinouchi, 2007; Perez-Matus et al., 2016).

238

Seascape context could also play a key role in determining the nursery quality of macroalgal  
240 habitats (Nagelkerken et al., 2015). Structural connectivity, measured as the proximity of a  
meadow patch to other viable habitats within the seascape, is emerging as an important  
242 predictor of juvenile abundance and diversity in macroalgal habitats (e.g., Mellin et al., 2007;  
van Lier et al., 2018; Bradley et al., 2019). This is likely to be particularly important for  
244 ontogenetic migrations, such as the movement of key fishing target species (*Lethrinus* spp.)  
from macroalgal to adjacent coral reef habitat as they increase in body size from juveniles to  
246 sub-adults (Wilson et al., 2017). In such species, both the canopy quality (e.g., composition,  
percent cover, height, density) and proximity of macroalgal habitats are likely to shape the  
248 magnitude of fish recruitment across diverse tropical seascapes (Nagelkerken et al., 2015;  
Wilson et al., 2017). From a management perspective, identifying these high quality and  
250 connected macroalgal patches should be a priority to ensure protection of key sources of  
fish population replenishment.

252

Published and emerging evidence lends support to two criteria for tropical macroalgal  
254 meadows functioning as fish nurseries: (i) they are widely used by juveniles of tropical reef  
fishes, many of which have the majority of their juvenile abundance within macroalgal  
256 habitats but are later found on coral reefs as adults; and (ii) juvenile macroalgal habitat  
quality can influence the future abundance of sub-adult and adult populations (e.g., Aburto-  
258 Oropeza et al., 2007; Wilson et al., 2017). Evidence for the former is much stronger and  
widespread than for the latter. Our analyses do not provide evidence of the link between  
260 macroalgal habitat availability and future fish population sizes, only evidence of patterns of  
occupation and proportional abundance. Nonetheless, for species whose juveniles are

262 exclusively found within macroalgal habitats, it is likely this habitat type provides a nursery  
function, as long as all available habitats in the seascape have been adequately surveyed  
264 (Beck et al., 2001; Dahlgren et al., 2006). For future work, we suggest fish taxa with juveniles  
that utilise a range of purported nursery habitats (Fig. 3 – “both” category) could be prime  
266 targets for testing whether macroalgal habitats facilitate increased fish growth, survivorship  
and recruitment success (*sensu* Beck et al., 2001; Dahlgren et al., 2006).

268

#### 4. MACROALGAE-ASSOCIATED TROPICAL FISHERIES

270 Tropical macroalgal habitats likely play several roles in supporting local fisheries production.  
First, as with kelp forests (see Bertocci et al., 2015; Brown et al., 2018), macroalgal  
272 production may enhance fishable biomass through direct consumption by browsing  
herbivores (e.g., *Siganus* species, *L. vaigiensis* and *Calotomus spinidens*; Unsworth et al.,  
274 2007; Fox & Bellwood, 2008; Hoey et al., 2013; Table 1). These herbivorous taxa have some  
of the highest rates of secondary production (i.e., rapid somatic growth and short longevity)  
276 among targeted reef fishes, enabling them to withstand high fishing pressure (Hicks &  
McClanahan, 2012; Morais & Bellwood, 2018). Secondly, macroalgae-derived detrital  
278 subsidies may be substantial in regions with extensive *Sargassum* beds because of the  
annual canopy loss in these habitats (Fulton et al., 2019). Macroalgal detritus may then be  
280 consumed by a wide range of invertebrates and grazing fishes across macroalgal and other  
habitat types that receive biomass subsidies through the drift of algal rafts and wrack  
282 (Stimson, 2013; Wilson et al., 2003; Zubia et al., 2015). Thirdly, canopy-forming macroalgae  
promote the production of epifaunal invertebrates that are preyed upon by smaller  
284 carnivorous fishes, thereby facilitating multiple routes for higher-order production involving

resident and transient carnivorous fishes, in addition to predation on herbivorous fishes  
286 (Edgar & Aoki, 1993; Rossier & Kulbicki, 2000; Wenger et al., 2018). The importance of  
macroalgal habitats as foraging areas for some guilds of carnivorous fish may even exceed  
288 that of seagrass beds due to higher diversity, abundance and biomass of epifauna (Tano et  
al., 2016). Aside from these trophic pathways, structurally complex macroalgal communities  
290 may also serve as nursery habitats for the juveniles of targeted reef fish (Section 3). This  
means strong seasonal and interannual fluctuations in macroalgal habitat quality are likely to  
292 have direct implications for recruitment and future fishery yields (Lim et al., 2016; Wilson et  
al., 2017).

294

Over half of the 44 fish species most strongly associated with tropical macroalgal meadows  
296 (Section 2, Table S3) are targeted by commercial, subsistence or recreational fisheries (Table  
1). Targeted species include herbivores in the families Siganidae and Labridae (subfamily  
298 Scarinae), as well as larger-bodied (>30 cm maximum length) generalist carnivores  
(Lethrinidae, Serranidae) and invertivores (Labridae). While many of these species are of  
300 minor commercial importance for industrial-scale fishing, they collectively represent a major  
component of production in small-scale fisheries that are significant for local communities.  
302 Estimating the contribution of macroalgal habitats to tropical fisheries is, however,  
problematic due to the lack of species- or habitat-specific data in global fisheries statistics  
304 (FAO, 2018). Furthermore, global data are likely to under-represent or completely exclude  
small-scale fisheries that target macroalgae-associated species (McManus et al., 1992;  
306 McClanahan et al., 2008; Pauly & Zeller, 2016). Taking one family of herbivorous and  
detritivorous fishes as a model, the rabbitfishes (Siganidae; FAO, 2018), we may get some  
308 indication of the importance of tropical macroalgal habitats to food security on a global scale

(Table S4). Catch data for this family is mainly for herbivorous rabbitfish that school in large numbers and utilise macroalgal habitats, rather than the coral reef-associated pair-forming rabbitfish species (Campos et al., 1994; Hicks & McClanahan, 2012; Hoey et al., 2013; Robinson et al., 2018). Rabbitfishes are targeted in at least 23 countries where they typically account for a small proportion (median = 1.3%) of national aggregated marine fish landings, although their contribution can be more substantial in some countries (17% in Kenya; 32% in Bahrain; Table S5). Importantly, 77% of the reported global rabbitfish catch occurs in Indonesia and the Philippines, two low-income countries, with high human populations and extensive coral reefs threatened by overfishing (Burke et al., 2012), and where the importance of rabbitfishes as a food source is increasing (Fig. 5). In the Philippines, annual rabbitfish catch has grown gradually over the past half-century, surpassing snapper (Lutjanidae) landings within the last two decades. In contrast, the trend in Indonesia suggests a more recent and rapid increase, with rabbitfish approaching emperor (Lethrinidae) landings in the past decade (Fig. 5). Some of this growth in herbivorous fish catches may have been linked to the expansion of seaweed farming in these countries, which has increased the macroalgal habitat and trophic resources available to fishes in shallow reef habitats (Hehre & Meeuwig, 2016).

Case studies of small-scale fisheries operating in back reef and lagoon areas often dominated by macroalgal habitat suggest high yields of macroalgae-associated fishes at high levels of fishing effort (McManus et al., 1992; Unsworth & Cullen, 2010). Total fish yields of 12 to 16 metric tons  $\text{km}^2 \text{yr}^{-1}$  from back reefs have been reported in the Philippines and Kenya, respectively (McManus et al., 1992; McClanahan et al., 2008). These yields exceed mean reported annual fish yields from coral reefs in the Pacific and Indian Ocean (McClanahan,

2006) and are more than 2 to 3 times the global maximum sustainable yield from coral reefs  
334 (Newton et al., 2007). However, macroalgae-associated fishes that dominate yields may be  
at a high risk of overfishing in some localities. In Kenya, two herbivorous species (*Siganus*  
336 *sutor* and *Leptoscarus vaigiensis*), together making up >60% of the total catch of small-scale  
reef fisheries by weight, have shown symptoms of growth and recruitment overfishing (Hicks  
338 & McClanahan, 2012). Similarly, in the Philippines, stocks of *Siganus fuscescens* and *S. spinus*  
are at severe risk of recruitment overfishing because both the adults and very young  
340 juveniles are targeted (McManus et al., 1992; Soliman & Yamaoka, 2010). As a consequence,  
smaller size-at-maturity and lower fecundity has been observed where fishing pressure is  
342 high (Jumawan-Nanual & Metillo, 2008; Soliman & Yamaoka, 2010). Since these macroalgae-  
associated rabbitfish can form large transient spawning aggregations targeted by fishers  
344 (Bijoux et al., 2013; McManus et al., 1992; Robinson et al., 2011), they are particularly  
susceptible to overfishing. However, fast somatic growth, early maturation and high  
346 fecundity, along with variable catchability are thought to limit the vulnerability of this fishery  
to overexploitation (Robinson et al., 2017).

348

Tropical macroalgal habitats may appear to make a minor contribution to global fisheries  
350 production, but they can underpin and enhance the food security and livelihoods of coastal  
communities that have few other sources of income (Cabral & Geronimo, 2018). Even in  
352 wealthy countries, the indirect contribution of macroalgal habitats to recreational fisheries  
and local economies may be substantial. For example, at Ningaloo in Western Australia,  
354 recreational fishing is a major driver of tourism (Smallwood et al., 2013) and two commonly  
caught species, *Epinephelus rivulatus* and *Lethrinus nebulosus* (Ryan et al., 2017), utilise  
356 macroalgal habitats (Table 1). The potential for tropical macroalgal habitat to be fish

nurseries, trophic facilitators, and corridors for fish movement between habitats suggest  
358 they warrant the same recognition as mangroves, seagrass beds and coral reefs in  
ecosystem-based fisheries management. Spatial management and monitoring of tropical  
360 macroalgal habitats could then provide capacity for the adaptive management of habitat-  
based fluctuations in fishable biomass of coastal target species (Brown et al., 2018; Green et  
362 al., 2014). This will require more quantitative information on the key aspects of macroalgal  
habitat quality and connectivity that affect both standing fish biomass and recruitment, and  
364 how these habitat traits respond to climatic cycles and disturbance events (Aburto-Oropeza  
et al., 2007; Wilson et al., 2017, 2018).

366

Apart from overexploitation, the other major threat to the sustainability of tropical fisheries  
368 is habitat loss, especially the loss of hard corals due to climate change (Newton et al., 2007;  
Bell et al., 2013). One of the responses of coral reefs to live coral loss is a regime shift from a  
370 coral-dominated to a macroalgal-dominated state (Hughes, 1994; Graham et al., 2015). The  
implications of such a shift in habitat are anticipated to be detrimental to the yield of coral  
372 reef fisheries (Graham et al., 2007; Pratchett et al., 2008, 2014; MacNeil et al., 2010).

However, data to assess long-term implications of regime shifts on tropical fisheries are  
374 scant. An exception is in the Seychelles, where long-term assessments of the inshore trap  
fishery found that yield and CPUE have been maintained or even increased following  
376 widespread bleaching and a shift to macroalgal-dominated habitat on some reefs (Robinson  
et al., 2018). Variability of the catch, however, has increased, and on reefs that underwent a  
378 regime shift the trophic structure of the fish assemblage was altered to shortened food  
chains with increases in biomass of low-trophic level herbivores and mid-level carnivores,  
380 such as emperors (Hempson et al., 2018b). Therefore, while the catch has been maintained

with a shift towards macroalgae-associated herbivorous fish, the predictability of catch per  
382 fishing trip has become less certain (Robinson et al., 2018).

## 384 **5. VULNERABILITY OF MACROALGAL FISHES TO HABITAT LOSS**

Macroalgal meadows are sensitive to environmental fluctuations and local disturbances  
386 (Olsen et al., 2019). Annual productivity of canopy-forming macroalgae is closely related to  
seasonal shifts in water temperature, although the range and optimal temperature for  
388 growth varies among taxa and regions (Fulton et al., 2019). As such, both local seasonal and  
large-scale climatic oscillations in sea temperature, such as those associated with the El Niño  
390 Southern Oscillation, can alter the structure of tropical macroalgal habitats (Wilson et al.,  
2014, 2018; Fulton et al., 2019). Similarly, extreme climatic events like heatwaves and  
392 cyclones can cause extensive loss of macroalgal habitat, with reduced macroalgal canopy  
cover often persisting for several years after the event (McCourt, 1984; Rogers, 1997; but  
394 see Loffler & Hoey, 2018). Long-term shifts in climate are also expected to alter the  
distribution and abundance of macroalgal species, leading to changes in community  
396 composition and ecosystem function (Diaz-Pulido et al., 2007). Over finer spatial scales,  
experimental manipulation of nutrient concentrations (Schaffelke & Klumpp, 1998) and  
398 sediment loads (Umar et al., 1998) demonstrate that environmental conditions can have a  
direct effect on macroalgae canopy phenology. Nutrient pulses and increased sediment  
400 loads associated with natural disturbances such as flooding, or human activities such as  
dredging and coastal development, are therefore expected to have a local impact on canopy  
402 cover and structure within macroalgal habitats.

404 Acute and continuous stressors that affect the structure of tropical macroalgal habitats are  
also expected to have an impact on the associated fishes. Ecological theory suggests that  
406 species most severely affected by disturbances will be those that feed or shelter exclusively  
within macroalgal habitats (Vázquez & Simberloff, 2002). Such macroalgal specialists might  
408 be especially vulnerable during the early life history stages when juvenile fishes are highly  
susceptible to predation and take shelter within particular aspects of macroalgal habitat  
410 microstructure. Changes in availability of dietary resources within macroalgal meadows may  
also influence fish growth and abundance, given the role they play in supporting direct  
412 (herbivorous) and indirect (carnivorous) food sources (Fulton et al. 2019). However, studies  
that have explicitly assessed the influence of disturbances on tropical macroalgal meadows  
414 and their associated fish fauna have been limited to fine-scale macroalgal removal  
experiments (McClanahan et al., 1999, 2001), or long-term comparisons of reefs that have  
416 undergone regime shifts from coral to macroalgal-dominated states (Graham et al., 2015).

418 Our meta-analysis of the correlation between fish abundance and varying macroalgal cover  
across a range of tropical locations around the world (Table S1) identified fish species and  
420 life history stages that are likely to be macroalgal-dependent and most affected by  
macroalgal habitat loss. Using Pearson's correlation adjusted for survey area in a weighted z-  
422 score, averaged across a minimum of three independent surveys per life history stage of  
each species (see section 4 of Methods in Supporting Information), we found a spectrum of  
424 relationships between the abundance of macroalgal-associated fishes and percent  
macroalgal cover (Fig. 6). Some species had a significant positive relationship with  
426 macroalgal canopy cover (i.e., a positive mean z-score with confidence interval above zero,  
indicated by hashed bars above the centre line, Fig. 6), increasing in abundance when cover

428 was high and declining when it was low. For *Leptoscarus vaigiensis* and *Cheilio inermis*, this  
positive relationship was strong for both adults and juveniles, while for other species the  
430 general relationship differed in direction or significance among life history stages, suggesting  
ontogenetic shifts in macroalgal habitat dependence. For example, abundance of juvenile  
432 *Lethrinus atkinsoni* was strongly correlated with canopy cover, but this relationship was not  
apparent for larger conspecifics, which are typically found on nearby coral reef (Wilson et al.,  
434 2017). Conversely, the abundance of adult *Epinephelus rivulatus* positively correlated with  
macroalgal canopy cover, while the abundance of their juveniles did not, which suggests an  
436 increased dependence on macroalgal habitats as fish become older. There were also several  
species that consistently declined in abundance as macroalgal canopy cover increased  
438 (*Pomacentrus trilineatus*, *Canthigaster solandri*, *Plectrorhincus chaetodonoides* and *Fistularia*  
*commersonii*), implying that these macroalgal-associated fishes, which were predominantly  
440 found in macroalgal meadows rather than coral reef, may prefer macroalgal habitats with  
low cover.

442

Our meta-analysis suggests that the extent of macroalgal dependence in tropical fishes  
444 varies both interspecifically and ontogenically when considering only percent cover of  
canopy-forming macroalgae. The within-meadow canopy structure of macroalgal meadows  
446 is, however, a conglomerate of multiple factors and fishes may associate with specific  
architectural components of the three-dimensional macroalgal canopy habitat. Previous  
448 studies have, for example, identified that abundance of fishes can also correlate with canopy  
height (Evans et al., 2014; Lim et al., 2016; Eggertsen et al., 2019) or the density of canopy-  
450 forming macroalgae holdfasts (Wilson et al., 2014, 2017). Using six years of fish and habitat  
surveys at 19 sites spread across the Ningaloo lagoon in the eastern Indian Ocean, we

452 compared annual patterns of fish abundance with canopy cover, height, and density to  
identify which macroalgal structural elements consistently predict fish abundance (Fig. 7).  
454 Some species, like *Leptoscarus vaigiensis* and *Lethrinus nebulosus*, clearly have strong  
positive correlations with multiple facets of macroalgal canopy structure, whilst the  
456 abundance of others (e.g. juvenile *Stethojulis strigiventer*) primarily correlate with a single  
canopy feature such as height (Fig. 7b). Interestingly, the abundance of adults and juveniles  
458 of the same fish species often correlate with the same elements of macroalgal canopy  
structure, but juveniles typically have stronger canopy-abundance relationships (Figures 6 &  
460 7). This suggests the early life history stages of most species may be more habitat dependent  
and vulnerable to environmental disturbances acting on macroalgal meadows. Temporal  
462 mismatches that arise between peak macroalgal habitat availability and seasonal pulses of  
larval fish settlement may, therefore, directly affect the survival of juveniles with long-term  
464 consequences for the replenishment of adult populations. However, experimental  
manipulations of habitat structure and reciprocal removal of competitive fishes are required  
466 to understand the true extent of habitat limitation, competition and recruitment facilitation  
in macroalgal meadows. A key challenge in these experiments will be manipulating certain  
468 aspects of canopy condition while keeping others constant (e.g., reducing height while  
maintaining cover) in order to tease apart specific habitat effects.

470

Different levels of dependence upon particular canopy structural features may allow  
472 partitioning of resources and co-existence of species (Wilson et al., 2014, 2017), which could  
explain the spectrum of relationships in the diversity and abundance of macroalgal-  
474 associated fishes across 19 meadows of varying canopy structure at Ningaloo (see Figures  
S1-S4). Although herbivorous fishes on coral-dominated reefs can respond negatively to

476 increasing canopy density in experimental patches of macroalgal habitat (Hoey & Bellwood,  
2011), we found the overall abundance of macroalgae-associated herbivorous and  
478 carnivorous fishes generally increased with macroalgal canopy density, height and cover at  
Ningaloo (Figures S1 & S2). Species-specific relationships are less clear (Figures S3 & S4), and  
480 while we found some evidence for negative correlations between canopy density and the  
abundance of two common macroalgal-associated herbivorous fishes (*Scarus ghobban* and  
482 *Siganus fuscescens*; Fig. S3), the most abundant macroalgal meadow herbivore in our  
dataset, *Leptoscarus vaigiensis*, responded positively to increases in canopy cover, height  
484 and density (Fig. S3). We are prevented from making generalisations on the nature and  
drivers of these relationships by a lack of diversity in the types of macroalgal habitat data  
486 collected by studies – the majority to date have been focused on percent cover. Macroalgal  
percent cover is unlikely to be a good predictor for other measures of canopy structure,  
488 given the disparity between the two metrics (Wilson et al., 2014; Lim et al., 2016). Indeed,  
studies at various locations around the world report that neighbouring macroalgal meadows  
490 within a tropical seascape can vary considerably in canopy cover, height and/or density, and  
that taxonomic and trophic groups of macroalgal fishes respond in different ways to this  
492 canopy complexity over space and time (e.g., Eggertsen et al., 2017, 2019; Wilson et al.,  
2017). Based on the evidence to date, canopy height and cover could arise as some of the  
494 best habitat-based predictors for temporal trends in macroalgal fish-habitat relationships, as  
these aspects often vary the most over time, while canopy density is relatively stable within  
496 meadows but can vary considerably among sites (e.g., Lim et al., 2016; Umar et al., 1998;  
Wilson et al., 2014). To test these general hypotheses, we need more studies to include  
498 measures of canopy height and density (alongside percent cover) to resolve which aspects  
best predict fish community structure and function in tropical macroalgal meadows.

500

502 Limited evidence suggests fishes may also associate with certain macroalgal species or  
genera (e.g., *Sargassum*; Lim et al. 2016, Wenger et al. 2018). While we lack the information  
504 needed to assess macroalgal-specificity in a wider suite of tropical fishes, this should be an  
important line of research because this lack of ecological versatility can render fishes more  
506 susceptible to disturbances that affect specific macroalgae. Moreover, habitat features other  
than the composition and soft habitat structure afforded by the macroalgae could influence  
508 fish abundance. For instance, the structural complexity of the underlying hard substratum or  
availability of shelter holes may be a better predictor of abundance for some macroalgal fish  
510 species (Eggleston, 1995). The influence of hard substratum complexity (e.g., presence of  
live/dead corals, holes, crevices and/or general topographic variation) on fish communities  
512 has been well documented in coral-dominated systems (Graham & Nash, 2013), but has  
received minimal attention in macrophyte systems. Recent surveys in the western Indian  
514 Ocean found the abundance of juvenile fishes was greater in macroalgal meadows with  
more structurally complex hard substrate (Eggertsen et al., 2019). Hard habitat complexity  
516 has also emerged as an important predictor of macroalgae-associated adult fish  
communities at Ningaloo (van Lier et al., 2018). A more in-depth analysis of data from van  
518 Lier et al. (2018) revealed macroalgal meadows with high hard substratum complexity were  
often characterised by species typical of coral reefs (e.g. *Thalassoma lunare*; Fig. S5). Fishes  
520 exclusively found in macroalgal habitats, however, displayed a mixed relationship with hard  
substratum complexity (Fig. S5). For instance, abundance of the herbivorous browser  
522 *Leptoscarus vaigiensis* was not unduly related to hard substratum complexity, while a  
predatory serranid (*Epinephelus rivulatus*) tended to occupy meadows with greater hard

524 substratum complexity (Fig. S5). Before any generalities can be drawn we need more  
assessments of how fine-scale changes in both soft macroalgal and hard substratum habitat  
526 structures influence macroalgal fish communities across a greater range of tropical settings.

528 Spatial arrangement and proximity of macroalgal meadows to other tropical habitat-forming  
taxa may also have an important bearing on fish diversity and abundance (Berkström et al.,  
530 2012; Martin et al., 2018; Bradley et al., 2019). Macroalgal meadows are often located in the  
vicinity of other habitats, forming a mosaic of interlinked patches in diverse tropical  
532 seascapes. Functional linkages among habitat patches support the persistence of marine  
populations and communities and are pivotal for enabling ecosystems to persist and recover  
534 from disturbances (Cumming, 2011; Olds et al., 2018). Consequently, disturbances that  
disrupt connectivity among habitats are expected to have the greatest effect on those  
536 species that migrate among habitats within the seascape (Dahlgren & Eggleston, 2000), such  
as the many fish species that appear to migrate from macroalgal to other habitats during  
538 their ontogeny (Section 3). Adult fishes may also regularly move among different patches or  
habitats types to forage. For example, some of the strongest and most consistent  
540 relationships between canopy cover and abundance occur with adult *Thalassoma lunare* and  
*Lethrinus nebulosus* (Figures 6 & 7), highly mobile species that may move among meadows  
542 and other habitat types within diverse tropical seascapes. Clearly, more seascape studies  
that incorporate macroalgal habitats are required to improve our understanding of the  
544 consequences of habitat destruction and fragmentation within tropical seascapes.

546 Disturbances that alter the structure and composition of macroalgal communities can also  
result in ecologically novel situations for fish. Ecological novelty can be difficult to navigate

548 for animals that lack relevant evolutionary experience (Sih et al., 2011), and individuals that  
incorrectly assess habitat quality may fail to make the most of available habitat options (i.e.,  
550 fall into ecological traps; Hale & Swearer, 2016). As a result, fish populations could be  
affected disproportionately to the level of changes in overall habitat condition (Hale et al.,  
552 2015). Numerous ecological traps have been revealed in terrestrial habitats (Hale & Swearer,  
2016), and analogous cases are likely to arise in macroalgal systems. For example, via habitat  
554 fragmentation and subsequent failure of individuals to avoid risky patch edges (*sensu*  
Weldon & Haddad, 2005), or via fishes utilising non-native or range-expanding macrophytes  
556 that offer superficially similar habitat but lower quantity of resource provision throughout  
the relevant season (e.g., Rodewald et al., 2010). Small-bodied juveniles, as well cryptic and  
558 endemic fishes with limited capacity to move to alternate habitats may be the most  
vulnerable to changes in macroalgal habitat. Conversely, larger bodied species that have a  
560 generalist carnivore diet and generalist habitat associations (e.g., ‘transient’ fish taxa  
identified in Section 2) may have the capacity to readily move among patches and adapt to  
562 macroalgal habitat loss (Berkström et al., 2013). Fish species that overlap in their habitat  
occupation of macroalgal meadows and seagrass beds (Fig. 3) may be particularly resilient to  
564 habitat disturbances affecting one habitat-forming organism in a diverse tropical seascape.  
However, even among these species, our meta-analysis of macroalgal-associated fishes and  
566 published studies on seagrass fish faunas (e.g., Eggertsen et al., 2017; Tano et al., 2017)  
suggest strong preferences for certain types and quality of macrophyte habitat. The negative  
568 effects of changes in coral canopy condition have been well documented in reef fishes that  
prefer certain hard coral species and growth forms (e.g., Pratchett et al., 2004; Wilson et al.,  
570 2006). Accordingly, shifts in the availability of preferred macrophyte canopy habitats are

likely to have a detrimental impact on the abundance and/or condition of macrophyte  
572 specialist fishes.

## 574 **6. CONCLUSIONS**

Macroalgal habitats can be a substantial component of tropical seascapes around the world.  
576 Our synthesis has revealed that macroalgal meadows can house a unique component of  
tropical fish species richness, may be an important factor in shaping fish recruitment across  
578 diverse tropical seascapes, and provide a key habitat for productive species that support  
local fisheries. While over 600 species of bony fishes have been recorded in tropical  
580 macroalgal meadows around the world, at present there is only evidence to suggest a  
quarter to a third of those species have the majority of their juvenile and/or adult  
582 abundance within macroalgal habitat. Using the relatively few studies for which there are  
balanced visual surveys of fish species richness across different habitat types, we found  
584 considerable overlap between fish assemblages within macroalgal habitats and two other  
common subtidal habitats - seagrass and coral reefs - which suggests macroalgal meadows  
586 could also be important foraging habitats and/or stepping stones in the triphasic life cycles  
of fishes in diverse tropical seascapes.

588  
Juvenile fishes appear to be prominent in macroalgal habitats across the tropics, where the  
590 quality of macroalgal canopy (percent cover, height and/or density) can be positively  
correlated to the abundance of juveniles. Evidence from two studies (in the Eastern Pacific  
592 and Western Indian oceans) suggests these habitat effects during the juvenile phase could  
influence the future abundance of fish in larger/older size classes (Aburto-Oropeza et al.,

594 2007; Wilson et al., 2017). While this is suggestive that canopy-forming macroalgal habitats  
may serve as fish nurseries, we lack key lines of evidence needed to make that general  
596 conclusion. More information on rates of juvenile growth, survival and movement is needed  
before we can establish the potential nursery function of macroalgal meadows for tropical  
598 fish and fisheries. Similarly, we need greater resolution on the catch of macroalgal-  
associated fishes to fully understand the contribution of macroalgal habitats to tropical  
600 fisheries.

602 Compared to other prominent tropical habitats like coral reef, macroalgal meadows are  
spatially and seasonally dynamic in structure, which has major consequences for the  
604 abundance and distribution of many macroalgal-associated fishes. Environmental conditions  
such as sea temperature are important drivers of this dynamism, which suggests climate  
606 change may affect tropical canopy-forming macroalgae and the associated fish communities  
and fisheries. Significantly, not all tropical macroalgal patches respond in the same way to  
608 seasonal and inter-annual changes in climate, with some patches in the seascape retaining  
canopy and providing a key habitat refuge (Lim et al., 2016; Wilson et al., 2017). Accordingly,  
610 we need to identify what makes certain macroalgal patches resilient to disturbance, and at  
what scale this needs to be maintained in order to facilitate functional connectivity with  
612 other tropical habitats. This will require long-term monitoring of fish abundance across  
tropical seascapes, and assessments of how seasonal changes to macroalgal-canopy  
614 structure affect patterns of fish distribution and replenishment across a suite of habitat  
types within tropical seascapes. Indeed, a lack of information on macroalgal fish  
616 assemblages over prolonged periods has restricted our understanding of how these fish  
contribute to key ecological processes and ecosystem services. Filling this knowledge gap is

618 becoming increasingly important as the extent of macroalgal habitat may change according  
to climate-related coral mortality and regime shifts.

620

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### **DATA AVAILABILITY**

634 Metadata used in the analyses are provided in online Supporting Information.

### **REFERENCES**

- 636 Aburto-Oropeza, O., Sala, E., Paredes, G., Mendoza, A., & Ballesteros, E. (2007). Predictability  
of reef fish recruitment in a highly variable nursery habitat. *Ecology*, 88, 2220–2228.
- 638 Ackerman, J. L., Bellwood, D. R. (2000). Reef fish assemblages: a re-evaluation using enclosed  
rotenone stations. *Marine Ecology Progress Series*, 206, 227–237.

640 Adams, A.J., Dahlgren, C.P., Kellison, G.T., Kendall, M. S., Layman, C. A., Ley, J. A.,  
Nagelkerken, I., & Serafy, J. E. (2006). Nursery function of tropical back-reef systems.  
642 *Marine Ecology Progress Series*, 318, 287-301

Beck, M. W., Heck, K. L., Able, K. W., Childers, D. L., Eggleston, D. B., Gillanders, B. M., ... &  
644 Weinstein, M. P. (2001). The Identification, conservation, and management of estuarine  
and marine nurseries for fish and invertebrates. *BioScience*, 51, 633-641.

646 Bell, J. D., Ganachaud, A., Gehrke, P. C., Griffiths, S. P., Hobday, A. J., Hoegh-Guldberg, O., ...  
& Waycott, M. (2013). Mixed responses of tropical Pacific fisheries and aquaculture to  
648 climate change. *Nature Climate Change*, 3, 591–599.

Berkström, C., Gullström, M., Lindborg, R., Mwandya, A.W., Yahya, S.A.S., Kautsky, N., &  
650 Nyström, M. (2012). Exploring ‘knowns’ and ‘unknowns’ in tropical seascape  
connectivity with insights from East African coral reefs. *Estuarine, Coastal and Shelf  
652 Science*, 107, 1-21.

Berkström, C., Lindborg, R., Thyresson, M., & Gullström, M. (2013). Assessing connectivity in  
654 a tropical embayment: Fish migrations and seascape ecology. *Biological Conservation*,  
166, 43–53.

656 Bertocci, I., Araújo, R., Oliveira, P., & Sousa-Pinto, I. (2015). Potential effects of kelp species  
on local fisheries. *Journal of Applied Ecology*, 52, 1216–1226.

658 Bijoux, J.P., Dagorn, L., Berke, G., Cowley, P.D., Soria, M., Gaertner, J.-C., & Robinson, J.  
(2013). Temporal dynamics, residency and site fidelity of spawning aggregations of a  
660 herbivorous tropical reef fish *Siganus sutor*. *Marine Ecology Progress Series*, 475, 233-  
247.

- 662 Bittick, S. J., Clausing, R. J., Fong, C. R., Scoma, S. R., & Fong, P. (2019). A Rapidly Expanding  
Macroalga Acts as a Foundational Species Providing Trophic Support and Habitat in the  
664 South Pacific. *Ecosystems*, 22, 165–173.
- Bradley, M., Baker, R., Nagelkerken, I., & Sheaves, M. (2019). Context is more important  
666 than habitat type in determining use by juvenile fish. *Landscape Ecology*, 34, 427-442.  
doi:10.1007/s10980-019-00781-3
- 668 Brandl, S. J., Tornabene, L., Goatley, C.H.R., Casey, J.M., Morais, R.A., Côté, I.M., Baldwin,  
C.C., Parravicini, V., Schiettekatte, N.M.D., & Bellwood, D.R., (2019). Demographic  
670 dynamics of the smallest marine vertebrates fuel coral-reef ecosystem functioning.  
*Science*, 364, 1189–1192.
- 672 Brown, C.J., Broadley, A., Adame, F., Branch, T.A., Turschwell, M.P., & Connolly, R.M. (2018).  
The assessment of fishery status depends on the condition of fish habitats. *Fish and*  
674 *Fisheries*, 20, 1-14.
- Burke, L., Reyttar, K., Spalding, M., & Perry, A. (2012). *Reefs at Risk Revisited in the Coral*  
676 *Triangle*. Washington: World Resources Institute.
- Cabral, R.B., & Geronimo, R.C. (2018). How important are coral reefs to food security in the  
678 Philippines? Diving deeper than national aggregates and averages. *Marine Policy*, 91,  
136-141.
- 680 Campos, W.L., del Norte-Campos, C., & McManus, J.W. (1994). Yield estimates, catch, effort  
and fishery potential of the reef flat in Cape Bolinao, Philippines. *Journal of Applied*  
682 *Ichthyology*, 10, 82-95.
- Chaves, L.T.C., Pereira, P.H.C., & Feitosa, J.L.L. (2013). Coral reef fish association with  
684 macroalgal beds on a tropical reef system in North-eastern Brazil. *Marine and*  
*Freshwater Research*, 64, 1101-1111.

- 686 Coker, D. J., Wilson, S. K., & Pratchett, M. S. (2014). Importance of live coral habitat for reef  
fishes. *Reviews in Fish Biology and Fisheries*, 24, 89–126.
- 688 Cumming, G. S. (2011). Spatial resilience: integrating landscape ecology, resilience, and  
sustainability. *Landscape Ecology*, 26, 899-909.
- 690 Dahlgren, C.P., & Eggleston, D.B. (2000). Ecological processes underlying ontogenetic habitat  
shifts in a coral reef fish. *Ecology*, 81, 2227–2240.
- 692 Dahlgren, C.P., Kellison, G.T., Adams, A.J., Gillanders, B.M., Kendall, M.S., Layman, C.A., Ley,  
J.A., Nagelkerken, I., & Serafy, J.E. (2006). Marine nurseries and effective juvenile  
694 habitats: concepts and applications. *Marine Ecology Progress Series*, 312, 291-295.  
<https://doi.org/10.3354/meps318303>
- 696 Depczynski, M., Fulton, C. J., Marnane, M. J., & Bellwood, D. R. (2007). Life history patterns  
shape energy allocation among fishes on coral reefs. *Oecologia*, 153, 111–20.
- 698 Diaz-Pulido, G., McCook, L.J., Larkum, A.W., Lotze, H.K., Raven, J.A., Schaffelke, B., Smith,  
J.E., & Steneck, R.S. (2007). Vulnerability of macroalgae of the Great Barrier Reef to  
700 climate change. In: Marshall, P. A., & Johnson, J. (Eds.) *Climate change and the Great  
Barrier Reef: A Vulnerability Assessment* (pp. 153-192). Townsville: Great Barrier Reef  
702 Marine Park Authority.
- Edgar, G.J., & Aoki, M. (1993). Resource limitation and fish predation: their importance to  
704 mobile epifauna associated with Japanese *Sargassum*. *Oecologia*, 95, 122-133.
- Eggertsen, L. (2019). *Identification and implications of fish nurseries in tropical and  
706 subtropical seascapes*. PhD thesis, Department of Ecology, Environment and Plant  
Sciences, Stockholm University, Stockholm.

708 Eggertsen, L., Ferreira, C.E.L., Fontoura, L., Kautsky, N., Gullström, M., & Berkström, C.  
(2017). Seaweed beds support more juvenile reef fish than seagrass beds in a south-  
710 western Atlantic tropical seascape. *Estuarine, Coastal and Shelf Science*, 196, 97–108.

Eggertsen, M., Chacin, D. H., C, Å., Halling, C., & Berkström, C. (2019). Contrasting  
712 distribution and foraging patterns of herbivorous and detritivorous fishes across  
multiple habitats in a tropical seascape. *Marine Biology*, 166, 51.

714 Eggleston, D. (1995). Recruitment in Nassau grouper *Epinephelus striatus*: post-settlement  
abundance, microhabitat features, and ontogenetic habitat shifts. *Marine Ecology  
716 Progress Series*, 124, 9–22.

Evans, R.D., Wilson, S.K., Field, S.N., & Moore, J. A. Y. (2014). Importance of macroalgal fields  
718 as coral reef fish nursery habitat in north-west Australia. *Marine Biology*, 161, 599–607.

Fairclough, D.V., & Nakazono, A. (2004). *Choerodon schoenleinii*. IUCN 2007. 2007 IUCN Red  
720 List of Threatened Species.

FAO (2018, September 8). *Fishery and Aquaculture Statistics. Global capture production  
722 1950-2016 (FishstatJ)*. Retrieved from:  
<http://www.fao.org/fishery/statistics/software/fishstatj/en>

724 Faunce, C. H., & Serafy, J. E. (2006). Mangroves as fish habitat: 50 years of field studies.  
*Marine Ecology Progress Series*, 318, 1–18.

726 Feary, D.A., Almany, G.R., McCormick, M.I., & Jones, G.P., (2007). Habitat choice,  
recruitment and the response of coral reef fishes to coral degradation. *Oecologia*, 153,  
728 727–737. <https://doi.org/10.1007/s00442-007-0773-4>

Fox, R.J., & Bellwood, D.R. (2008). Remote video bioassays reveal the potential feeding  
730 impact of the rabbitfish *Siganus canaliculatus* (f: Siganidae) on an inner-shelf reef of the  
Great Barrier Reef. *Coral Reefs*, 27, 605–615.

732 Froese, R. & Pauly, D. (2018, September 8). *FishBase* (version 06s/2018). Retrieved from  
http://www.fishbase.org.

734 Fulton, C.J., Abesamis, R.A., Berkström, C., Depczynski, M., Graham, N.A.J., Holmes, T.H.,  
Kulbicki, M., Noble, M.M., Radford, B.T., Tano, S., Tinkler, P., Wernberg, T., & Wilson,  
736 S.K. (2019). Form and function of tropical macroalgal reefs in the Anthropocene.  
*Functional Ecology*, 33, 989–999. <https://doi.org/10.1111/1365-2435.13282>

738 Gillanders, B. M. (2006). Seagrasses, fish, and fisheries. In: Larkum, A.W., Orth, R.J., &  
Duarte, C.M. (Eds.), *Seagrasses: Biology, ecology and conservation* (pp. 503–536). The  
740 Netherlands: Springer.

Gillanders, B. M., Able, K. W., Brown, J. A., Eggleston, D. B., & Sheridan, P. F. (2003). Evidence  
742 of connectivity between juvenile and adult habitats for mobile marine fauna: An  
important component of nurseries. *Marine Ecology Progress Series*, 247, 281–295.

744 Graham, N.A.J., Jennings, S., MacNeil, M.A., Mouillot, D., & Wilson, S.K. (2015). Predicting  
climate-driven regime shifts versus rebound potential in coral reefs. *Nature*, 518, 94–  
746 97.

Graham, N.A.J., & Nash, K.L. (2013). The importance of structural complexity in coral reef  
748 ecosystems. *Coral Reefs*, 32, 315-326.

Graham, N.A.J., Wilson, S.K., Jennings, S., Polunin, N.V.C., Robinson, J., Bijoux, J.P., & Daw,  
750 T.M. (2007). Lag effects in the impacts of mass coral bleaching on coral reef fish,  
fisheries and ecosystems. *Conservation Biology*, 21, 1291-1300.

752 Graham, N.A.J., Wilson, S.K., Jennings, S., Polunin, N.V., Bijoux, J.P., & Robinson, J. (2006).  
Dynamic fragility of oceanic coral reef ecosystems. *Proceedings of the National  
754 Academy of Sciences*, 103, 8425-8429.

Green, A.L., Fernandes, L., Almany, G., Abesamis, R., McLeod, E., Aliño, P.M., White, A.T.,  
756 Salm, R., Tanzer, J., & Pressey, R.L. (2014). Designing marine reserves for fisheries  
management, biodiversity conservation, and climate change adaptation. *Coastal*  
758 *Management*, 42, 143-159.

Hale, R., & Swearer, S.E. (2016). Ecological traps: current evidence and future directions.  
760 *Proceedings of the Royal Society London Series B*, 283, 494–499.

Hale, R., Treml, E.A., & Swearer, S.E. (2015). Evaluating the metapopulation consequences of  
762 ecological traps. *Proceedings of the Royal Society London Series B*, 282, 20142930.

Hehre, E. J., & Meeuwig, J. J. (2016). A global analysis of the relationship between farmed  
764 seaweed production and herbivorous fish catch. *PLoS One*, 11, e0148250.

Hempson, T.N., Graham, N.A., MacNeil, M.A., Bodin, N., & Wilson, S.K. (2018a). Regime shifts  
766 shorten food chains for mesopredators with potential sublethal effects. *Functional*  
*Ecology*, 32, 820–830.

768 Hempson, T.N., Graham, N.A.J., MacNeil, M.A., Williamson, D.H., Jones, G.P., & Almany, G.R.  
(2017). Coral reef mesopredators switch prey, shortening food chains, in response to  
770 habitat degradation. *Ecology & Evolution*, 7, 2626–2635.

Hempson, T.N., Graham, N.A.J., MacNeil, M.A., Hoey, A.S., & Wilson, S.K. (2018b). Ecosystem  
772 regime shifts disrupt trophic structure. *Ecological Applications*, 28, 191-200.

Hicks, C.C., & McClanahan, T.R. (2012). Assessing gear modifications needed to optimize  
774 yields in a heavily exploited, multi-species, seagrass and coral reef fishery. *PLoS One*, 7,  
e36022.

776 Hoey, A. S., & Bellwood, D. R. (2011). Suppression of herbivory by macroalgal density: a  
critical feedback on coral reefs? *Ecology Letters*, 14, 267–273.

- 778 Hoey, A.S., Brandl, S.J., & Bellwood, D.R. (2013). Diet and cross-shelf distribution of  
rabbitfishes (f. Siganidae) on the northern Great Barrier Reef: implications for  
780 ecosystem function. *Coral Reefs*, 32, 973-984.
- Holmes, T.H., Wilson, S.K., Vanderklift, M., Babcock, R., & Fraser, M., (2012). The role of  
782 *Thalassoma lunare* as a predator of juvenile fish on a sub-tropical coral reef. *Coral  
Reefs*, 31, 1113–1123. <https://doi.org/10.1007/s00338-012-0934-8>
- 784 Horinouchi, M. (2007). Review of the effects of within-patch scale structural complexity on  
seagrass fishes. *Journal of Experimental Marine Biology and Ecology*, 350, 111–129.
- 786 Hsu, T.-H., Adiputra, Y.T., Burrige, C.P., & Gwo, J.-C. (2011). Two spinefoot colour morphs:  
mottled spinefoot *Siganus fuscescens* and white-spotted spinefoot *Siganus*  
788 *canaliculatus* are synonyms. *Journal of Fish Biology*, 79, 1350-1355.
- Hughes, T. P. (1994). Catastrophes, phase shifts and large-scale degradation of a Caribbean  
790 coral reef. *Science*, 265, 1547–1551.
- Jumawan-Nanual, B., & Metillo, E. B. (2008). Population structure and reproductive biology  
792 of *Siganus fuscescens* Houttuyn 1782 (Perciformes, Siganidae) in Pujada Bay,  
Southeastern Mindanao, Philippines. *Philippine Scientist*, 45, 62-79.
- 794 Lim, I.E., Wilson, S.K., Holmes, T.H., Noble, M.M., & Fulton, C.J. (2016). Specialization within  
a shifting habitat mosaic underpins the seasonal abundance of a tropical fish.  
796 *Ecosphere*, 7, e01212.
- Loffler, Z., & Hoey, A.S. (2018). Canopy-forming macroalgal beds (*Sargassum*) on coral reefs  
798 are resilient to physical disturbance. *Journal of Ecology*, 106, 1156-1164.
- MacNeil, M.A., Graham, N.A.J., Cinner, J.E., Dulvy, N.K., Loring, P.A., Jennings, S., Polunin,  
800 N.V.C., Fisk, A.T., & McClanahan, T.R. (2010). Transitional states in marine fisheries:

adapting to predicted global change. *Philosophical Transactions of the Royal Society B*,  
802 365, 3753-3763.

Martin, T.S.H., Olds, A.D., Olalde, A.B.H., Berkström, C., Gilby, B.L., Schlacher, T.A., Butler,  
804 I.R., Yabsley, N.A., Zann, M., & Connolly, R.M. (2018). Habitat proximity exerts opposing  
effects on key ecological functions. *Landscape Ecology*, 33, 1273-1286.

806 Martin-Smith, K. M. (1993). Abundance of mobile epifauna: the role of habitat complexity  
and predation by fishes. *Journal of Experimental Marine Biology and Ecology*, 174, 243-  
808 60.

McClanahan, T.R., Hendrick, V., Rodrigues, M.J., & Polunin, N.V.C. (1999). Varying responses  
810 of herbivorous and invertebrate-feeding fishes to macroalgal reduction on a coral reef.  
*Coral Reefs*, 18, 195–203.

812 McClanahan, T., McField, M., Huitric, M., Bergman, K., Sala, E., Nyström, M., Nordemar, I.,  
Elfwing, T., & Muthiga, N. (2001). Responses of algae, corals and fish to the reduction of  
814 macroalgae in fished and unfished patch reefs of Glovers Reef Atoll, Belize. *Coral Reefs*,  
19, 367–379.

816 McClanahan, T. R. (2006). Challenges and accomplishments towards sustainable reef  
fisheries. In: Côté, I.M., & Reynolds, J.D. (eds) *Coral Reef Conservation* (pp. 147-182).  
818 Cambridge: Cambridge University Press.

McClanahan, T.R., Hicks, C.C., & Darling, E.S. (2008). Malthusian overfishing and efforts to  
820 overcome it on Kenyan coral reefs. *Ecological Applications*, 18, 1516-1529.

McCourt, R. M. (1984). Seasonal patterns of abundance, distributions, and phenology in  
822 relation to growth strategies of three *Sargassum* species. *Journal of Experimental  
Marine Biology & Ecology*, 74, 141–156.

- 824 McManus, J.W., Nañola, C.L. Jr., Reyes, R.B. Jr., & Kesner, K.N. (1992). Resource ecology of  
the Bolinao coral reef system. *ICLARM Studies Review*, 22, 117.
- 826 Mellin, C., Kulbicki, M., & Ponton, D. (2007). Seasonal and ontogenetic patterns of habitat  
use in coral reef fish juveniles. *Estuarine, Coastal and Shelf Science*, 75, 481-491.
- 828 Morais, R. A., & Bellwood, D. R. (2018). Global drivers of reef fish growth. *Fish and Fisheries*,  
19, 874-889.
- 830 Murphy, H.M., & Jenkins, G.P. (2010). Observational methods used in marine spatial  
monitoring of fishes and associated habitats: a review. *Marine and Freshwater*  
832 *Research*, 61, 236-252.
- Nagelkerken, I., Sheaves, M., Baker, R., & Connolly, R.M. (2015). The seascape nursery: a  
834 novel spatial approach to identify and manage nurseries for coastal marine fauna. *Fish*  
*and Fisheries*, 16, 362–371.
- 836 Newton, K., Côté, I.M., Pilling, G.M., Jennings, S., & Dulvy, N.K. (2007). Current and future  
sustainability of island coral reef fisheries. *Current Biology*, 17, 655-658.
- 838 Olds, A. D., Nagelkerken, I., Huijbers, C. M., Gilby, B. L., Pittman, S. J., & Schlacher, T. A.  
(2018). Connectivity in coastal seascapes. In: Pittman, S. J. (Ed.) *Seascape Ecology* (pp.  
840 261-292). London: John Wiley & Sons.
- Olds, A.D., Albert, S., Maxwell, P.S., Pitt, K.A., & Connolly, R.M. (2013). Mangrove-reef  
842 connectivity promotes the effectiveness of marine reserves across the western Pacific.  
*Global Ecology and Biogeography*, 22, 1040-1049.
- 844 Olsen, Y.S., Mattio, L., Zavala Perez, A., Babcock, R.C., Thompson, D., Haywood, M.D.,  
Keesing, J., & Kendrick, G.A. (2019). Drivers of species richness and abundance of  
846 marine macrophytes on shallow tropical reefs of north-western Australia. *Journal of*  
*Biogeography*, 46, 170-84.

848 Ornellas, A. B., & Coutinho, R. (1998) Spatial and temporal patterns of distribution and  
abundance of a tropical fish assemblage in a seasonal *Sargassum* bed, Cabo Frio Island,  
850 Brazil. *Journal of Fish Biology*, 53 (Supplement A), 198-208.

Pauly, D., & Zeller, D. (2016). Catch reconstructions reveal that global marine fisheries  
852 catches are higher than reported and declining. *Nature Communications*, 7, 10244.

Pérez-Matus, A., Sánchez, F., González-But, J., & Lamb, R. (2016). Understory algae  
854 associations and predation risk influence broad-scale kelp habitat use in a temperate  
reef fish. *Marine Ecology Progress Series*, 559, 147–158.

856 Pratchett, M.S., Wilson, S.K., Berumen, M.L., & McCormick, M.I. (2004). Sublethal effects of  
coral bleaching on an obligate coral feeding butterflyfish? *Coral Reefs*, 23, 352–356.

858 Pratchett, M.S., Munday, P.L., Wilson, S.K., Graham, N.A.J., Cinner, J.E., Bellwood, D.R.,  
Jones, G.P., Polunin, N.V.C., & McClanahan, T.R. (2008). Effects of climate-induced coral  
860 bleaching on coral-reef fishes – ecological and economic consequences. *Oceanography  
and Marine Biology: An Annual Review*, 46, 251-296.

862 Pratchett, M.S., Hoey, A.S., & Wilson, S.K. (2014). Reef degradation and the loss of critical  
ecosystem goods and services provided by coral reef fishes. *Current Opinion in  
864 Environmental Sustainability*, 7, 37-43.

Robinson, J., Samoily, M.A., Grandcourt, E., Julie, D., Cedras, M., & Gerry, C. (2011). The  
866 importance of targeted spawning aggregation fishing to the management of Seychelle’s  
trap fishery. *Fisheries Research*, 112, 96-103.

868 Robinson, J., Graham, N.A.J., Grüss, A., Gerry, C., & Bijoux, J. (2017). Fishery benefits from  
exploiting spawning aggregations not solely dependent on enhanced fish density.  
870 *African Journal of Marine Science*, 39, 269-278.

- Robinson, J. P. W., Wilson, S. K., Robinson, J., Gerry, C., Lucas, J., Assan, C., ... & Graham, N. A. J. (2018). Productive instability of coral reef fisheries after climate-driven regime shifts. *Nature Ecology and Evolution*, 3, 183-190.
- 872
- 874 Rodewald, A.D., Shustack, D.P. & Hitchcock, L.E. (2010). Exotic shrubs as ephemeral ecological traps for nesting birds. *Biological Invasions*, 12, 33.
- 876 Rogers, R.W. (1997). Brown algae on Heron reef flat, Great Barrier Reef, Australia: Spatial, seasonal and secular variation in cover. *Botanica Marina*, 40, 113–118.
- 878 Rossier, O., & Kulbicki, M. (2000). A comparison of fish assemblages from two types of algal beds and coral reefs in the south-west lagoon of New Caledonia. *Cybium*, 24, 3–26.
- 880 Ryan, K.L., Hall, N.G., Lai, E.K., Smallwood, C.B., Taylor, S.M., & Wise, B.S. (2017). *Statewide survey of boat-based recreational fishing in Western Australia 2015/16*. Perth: Government of Western Australia Fisheries Research Division.
- 882
- Sambrook, K., Hoey, A.S., Andréfouët, S., Cumming, G.S., Duce, S., & Bonin, M.C., (2019). Beyond the reef: The widespread use of non-reef habitats by coral reef fishes. *Fish and Fisheries*, doi: 10.1111/faf.12383.
- 884
- 886 Schaffelke, B., & Klumpp, D. W. (1998). Nutrient-limited growth of the coral reef macroalga *Sargassum baccularia* and experimental growth enhancement by nutrient addition in continuous flow culture. *Marine Ecology Progress Series*, 164, 199–211.
- 888
- Sih, A., Ferrari, M.C.O., & Harris, D.J. (2011). Evolution and behavioural responses to human-induced rapid environmental change. *Evolutionary Applications*, 4, 367–387.
- 890
- Smallwood, C.B., Beckley, L.E., & Moore, S.A. (2013). Effects of adjacent land tenure on visitor use of Ningaloo Marine Park, Western Australia. *Australasian Journal of Environmental Management*, 20, 130-146.
- 892

- 894 Soliman, V. S., & Yamaoka, K. (2010). Assessment of the fishery of siganid juveniles caught by  
bagnet in Lagonoy Gulf, Southeastern Luzon, Philippines. *Journal of Applied Ichthyology*,  
896 26, 561–567.
- Stimson, J. (2013). Consumption by herbivorous fishes of macroalgae exported from coral  
898 reef flat refuges to the reef slope. *Marine Ecology Progress Series*, 472, 87–99.
- Tano, S., Eggertsen, M., Wikström, S.A., Berkström, C., Buriyo, A.S., & Halling, C. (2016).  
900 Tropical seaweed beds are important habitats for mobile invertebrate epifauna.  
*Estuarine, Coastal and Shelf Science*, 183, 1–12.
- 902 Tano, S., Eggertsen, M., Wikström, S.A., Berkström, C., Buriyo, A.S., & Halling, C. (2017).  
Tropical seaweed beds as important habitats for juvenile fish. *Marine and Freshwater*  
904 *Research*, 68, 1921–1934.
- Umar, M.J., McCook, L.J., & Price, I.R. (1998). Effects of sediment deposition on the seaweed  
906 *Sargassum* on a fringing coral reef. *Coral Reefs*, 17, 169–177.
- Unsworth, R.K.F., Taylor, J.D., Powell, A., Bell, J.J., & Smith, D.J. (2007). The contribution of  
908 scarid herbivory to seagrass ecosystem dynamics in the Indo-Pacific. *Estuarine, Coastal*  
*and Shelf Science*, 74, 53-62.
- 910 Unsworth, R. K., & Cullen, L. C. (2010). Recognising the necessity for Indo-Pacific seagrass  
conservation. *Conservation Letters*, 3, 63-73.
- 912 van Lier, J.R., Wilson, S.K., Depczynski, M., Wenger, L.N., & Fulton, C.J. (2018). Habitat  
connectivity and complexity underpin fish community structure across a seascape of  
914 tropical macroalgae meadows. *Landscape Ecology*, 33, 1287-1300.
- Vázquez, D. P., & Simberloff, D. (2002). Ecological specialization and susceptibility to  
916 disturbance: conjectures and refutations. *American Naturalist*, 159, 606–623.

- Weldon, A.J. & Haddad, N.M. (2005). The effects of patch shape on indigo buntings:  
918 evidence for an ecological trap. *Ecology*, 86, 1422-1431.
- Wenger, L. N., Van Lier, J. R., & Fulton, C. J. (2018). Microhabitat selectivity shapes the  
920 seascape ecology of a carnivorous macroalgae-associated tropical fish. *Marine Ecology  
Progress Series*, 590, 187–200.
- 922 Whitfield, A. K. (2017). The role of seagrass meadows, mangrove forests, salt marshes and  
reed beds as nursery areas and food sources for fishes in estuaries. *Reviews in Fish  
924 Biology and Fisheries*, 27, 75-110.
- Wilson, S.K., Bellwood, D.R., Choat, J.H., & Furnas, M.J. (2003). Detritus in the epilithic algal  
926 matrix and its use by coral reef fishes. *Oceanography and Marine Biology: Annual  
Review*, 41, 279-310.
- 928 Wilson, S.K., Depczynski, M., Fisher, R., Holmes, T.H., O’Leary, R.A., & Tinkler, P. (2010).  
Habitat associations of juvenile fish at Ningaloo reef, Western Australia: the importance  
930 of coral and algae. *PLoS One*, 5, e15185.
- Wilson, S. K., Depczynski, M., Holmes, T. H., Noble, M. M., Radford, B. T., Tinkler, P., &  
932 Fulton, C. J. (2017). Climatic conditions and nursery habitat quality provide indicators of  
reef fish recruitment strength. *Limnology and Oceanography*, 62, 1868–1880.
- 934 Wilson, S. K., Depczynski, M., Fisher, R., Holmes, T. H., Noble, M. M., Radford, B. T., ... &  
Fulton, C. J. (2018). Climatic forcing and larval dispersal capabilities shape the  
936 replenishment of fishes and their habitat-forming biota on a tropical coral reef. *Ecology  
and Evolution*, 8, 1918-1928.
- 938 Wilson, S.K., Fulton, C.J., Depczynski, M., Holmes, T.H., Noble, M.M., Radford, B., & Tinkler,  
P. (2014). Seasonal changes in habitat structure underpin shifts in macroalgae-

- 940 associated tropical fish communities. *Marine Biology*, 161, 2597–2607.  
<https://doi.org/10.1007/s00227-014-2531-6>
- 942 Wilson, S. K., Graham, N. A. J., Pratchett, M. S., Jones, G. P., & Polunin, N. V. C. (2006).  
Multiple disturbances and the global degradation of coral reefs: Are reef fishes at risk or  
944 resilient? *Global Change Biology*, 12, 2220-2234.
- Zubia, M., Andréfouët, S., & Payri, C. (2015). Distribution and biomass evaluation of drifting  
946 brown algae from Moorea lagoon (French Polynesia) for eco-friendly agricultural use.  
*Journal of Applied Phycology*, 27, 1277–1287.
- 948

**Table 1.** List of 25 macroalgae-associated species targeted for food fisheries. Max TL:

950 maximum total length. Range: IO – Indian Ocean, WP – West Pacific, EP – East Pacific; RS –  
 Red Sea, M – Mediterranean, AO – Atlantic Ocean. Trophic groups: GC – generalist carnivore,  
 952 I – invertivore, H – herbivore, O – omnivore. Type of fishery: Comm – commercial, Rec –  
 recreational, Subs – subsistence. MA-LH stage denotes which life history stage mostly found  
 954 in macroalgal habitat (Table S3): Juv – juveniles, Both - juveniles and adults. Data on  
 maximum size, distribution, trophic group and fisheries from Froese & Pauly (2018).

Family (Subfamily)	Species	Max TL (cm)	Range	Trophic group	Type of fishery	MA-LH stage
Lethrinidae	<i>Lethrinus atkinsoni</i>	50	WP	GC	Comm; Rec	Juv
	<i>L. genivittatus</i>	25	IO,WP	GC	Comm*	Both
	<i>L. nebulosus</i>	87	IO,WP,RS	GC	Comm; Rec	Juv
	<i>L. semicinctus</i>	35	IO,WP	GC	Comm*	Adult
	<i>L. variegatus</i>	20	IO,WP	GC	Comm*	Both
Siganidae	<i>Siganus fuscescens</i> <sup>1</sup>	40	WP	H	Comm	Adult
	<i>S. spinus</i>	28	IO,WP	H	Comm*	Both
	<i>S. sutor</i>	45	IO	H	Comm	Both
Labridae	<i>Cheilio inermis</i>	50	IO,WP,RS	I	Comm*	Both
	<i>Choerodon schoenleinii</i>	100	IO,WP	I	Comm <sup>2</sup> ; Rec	Both
	<i>Oxycheilinus bimaculatus</i>	15	IO,WP	I	Subs	Adult
Serranidae (Epinephelinae)	<i>Cephalopholis boenak</i>	30	IO,WP	GC	Subs	Juv
	<i>Epinephelus</i> <i>coeruleopunctatus</i>	76	IO,WP	GC	Comm*	Adult
	<i>Epinephelus rivulatus</i>	45	IO,WP	GC	Comm*; Rec	Both
Labridae (Scarinae)	<i>Calotomus spinidens</i>	30	IO,WP	H	Comm	Adult
	<i>Leptoscarus vaigiensis</i>	35	IO,WP	H	Comm; Subs	Both
Lutjanidae	<i>Lutjanus carponotatus</i>	40	IO,WP	GC	Comm*; Rec	Juv
	<i>L. fulviflamma</i>	35	IO,WP,RS	GC	Comm; Rec	Juv
Carangidae	<i>Gnathanodon speciosus</i>	120	IO,WP,EP	GC	Comm*; Rec	Adult
Fistularidae	<i>Fistularia commersonii</i>	160	IO,WP,RS, EP,M	GC	Comm*	Adult
Mullidae	<i>Upeneus tragula</i>	25	IO,WP	I	Comm	Juv

Muraenidae	<i>Echidna nebulosa</i>	100	IO,WP,RS, EP,AO	GC	Comm*	Adult
Nemipteridae	<i>Scolopsis ghanam</i>	30	IO	GC	Subs	Both
Pomacentridae	<i>Abudefduf vaigiensis</i>	20	IO,WP,RS	O	Subs	Juv
Tetraodontidae	<i>Arothron hispidus</i>	50	IO,WP,EP	O	Comm* <sup>3</sup>	Both

956 <sup>1</sup>May be synonymous with *Siganus canaliculatus* (Hsu et al., 2011); \*Minor commercial status; <sup>2</sup>Near  
958 threatened globally (Fairclough & Nakazono, 2004); <sup>3</sup>Poisonous but traded in some countries.

960 **FIGURE LEGENDS**

**Figure 1.** Cumulative number of published studies on fish community structure within  
962 tropical macroalgae meadows (dark bars), and within macroalgae-dominated habitat that  
arose from a coral–macroalgal regime shift (grey bars). Published research effort on tropical  
964 fish communities within two other major macrophyte habitats - seagrass beds (solid line)  
and mangrove forests (dotted line) - are provided for comparison. Results are for 1985 to  
966 2018 inclusive, based on a Scopus search conducted on 17 March 2019.

**Figure 2.** Location of the 24 independent studies (with some geographic overlap) included in  
968 different aspects of our data syntheses and meta-analysis of macroalgal-associated tropical  
fishes (see Table S1 in Supporting Information).

970 **Figure 3.** Percentage of conspicuous fish species (readily detected by visual surveys)  
occupying only tropical macroalgae, coral reef, or seagrass habitats, or occurring in both  
972 habitat types based on studies that explored pairwise (macroalgae vs coral/seagrass)  
occupation patterns. Boxplots indicate number of fish species that fall into each category,  
974 expressed as a percentage of the total fish species recorded in each of 12 (macroalgae-coral)  
and 7 (macroalgae-seagrass) independent studies encompassing 7 or more tropical locations  
976 (underlying data in Table S2).

**Figure 4.** Proportional abundance of juveniles and adults for 350 fish species within  
978 macroalgal habitat relative to nearby coral reef, classified according to their adult trophic  
level (2.00-2.99 = white circles; 3.00-3.99 = grey; 4.00 or greater = black). Outer quadrants  
980 indicate species with higher macroalgal dependency due to majority (over half) of their  
juveniles (blue, 9 species, median trophic level = 3.68), adults (yellow, 13 species, 3.50), or  
982 both stages (green, 22 species, 3.50) occupying macroalgal habitats (Table S3). An additional

13 species had an equal proportion (0.5) of one or both life history stages among habitats  
984 (i.e., fall on the boundary lines of the quadrants).

**Figure 5.** Landings of rabbitfish (Siganidae), snapper (Lutjanidae) and emperor (Lethrinidae)  
986 in the Philippines (a) and Indonesia (b – Pacific Ocean; c – Indian Ocean) from 1950 to 2016  
(FAO, 2018).

988 **Figure 6.** Relationship between the abundance of 23 fish species and macroalgal habitat  
cover across geographic locations encompassed by our meta-analysis. Mean weighted z  
990 scores indicate species with positive (increased abundance) or negative (decreased  
abundance) relationships with macroalgal cover (replication per species and life history  
992 stage are provided in Table S3). Hatching denote scores with 95% confidence limits that fail  
to intersect zero, which indicates a consistent response across locations. An asterisk  
994 indicates a species is targeted for fishing in one or more of the survey locations (Table 1).

**Figure 7.** Correlations between the abundance of 28 fish species and mean annual (n = 6  
996 years) macroalgal canopy habitat structure in terms of (a) percent cover, (b) height, and (c)  
density of holdfasts across 19 study sites at Ningaloo, Western Australia. Pearson's  
998 correlation indicate species with positive (increased abundance) or negative (decreased  
abundance) relationships with macroalgal canopy structure. Hatching denote scores with  
1000 95% confidence limits that fail to intersect zero, which indicates a consistent response  
among years.

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