

25 **Abstract**

- 26 1. Unsustainable fishing is a major driver of change in marine ecosystems. The ways
27 that fishing gears target fishes with different ecological functions are unclear,
28 particularly in complex multi-species fisheries.
- 29 2. Here, we examine whether artisanal fishing gears compete for fishes with unique
30 combinations of ecological traits (diet, body size, depth, position in water column,
31 period of activity, schooling behaviour) in a coral reef ecosystem. We use coral reef
32 fish landing data from 25 sites along the Kenyan coast collected over a seven year
33 period.
- 34 3. All fishing gears targeted a wide diversity of traits, but with some differentiation
35 among gears. Fish assemblages captured by spearguns were significantly different
36 from the other gear types, specialising on diurnal species that feed on sessile
37 invertivores. Nets, including gillnets and beachseines, targeted the most functional
38 diversity. Escape slot traps targeted the least functionally-diverse assemblages. Basket
39 traps and escape-slot traps targeted the most functionally similar species of all two-
40 gear combinations.
- 41 4. There were 163 functional entities (unique combinations of traits) captured in the
42 fishery, however 50% of the catch by each gear was from only 2-6 functional entities.
43 Most of the differences in gear selectivity were due to unique and rarely targeted
44 functional entities, that made up only a small proportion of the catch.
- 45 5. Synthesis and applications. Coral reef fisheries target a breadth of functional entities,
46 but catches are heavily skewed towards relatively few functional entities. While
47 banning specific gears will benefit rare functional entities in the catch, effort
48 reductions will be necessary to alleviate pressure on commonly targeted functional
49 entities.

50 Key words: ecosystem-based management, functional diversity, gear-based management,
51 gear interactions, niche breadth, trait-based approach

52 **Introduction**

53 Fishing is a major driver of marine ecosystem change worldwide (Worm *et al.* 2006).
54 Consequently, a variety of tools have been proposed to evaluate the environmental effects of
55 fishing on marine ecosystems (McClanahan, Hicks & Darling 2008; Guillemot *et al.* 2014).
56 Most work aimed at understanding the effects of fishing on the marine environment tends to
57 focus on species abundances and catch composition (Jennings, Greenstreet & Reynolds 1999;
58 Hiddink *et al.* 2006). Yet, the growing interest in an ecosystem-based approach has stressed
59 maintaining and sustaining ecological functions (Sinclair *et al.* 2002; Tillin *et al.* 2006).
60 Trait-based ecology has much to offer to this objective. Indeed, choosing relevant traits that
61 represent the complementary roles of organisms has become a cornerstone of functional
62 ecology (McGill *et al.* 2006; Violle *et al.* 2007). By considering biological traits as proxies
63 for function, the trait-based approach (Mouillot *et al.* 2013; Villéger *et al.* 2017) may help
64 uncover ecosystem processes and functional implications of changes in fisheries
65 assemblages.

66 Trait-based approaches were initially applied in plant ecology (Cornwell, Schilck & Ackerly
67 2006) and are now widely used across other organisms, such as birds (Naeem, Duffy &
68 Zavaleta 2012), bats (Norberg 1994), corals (Darling *et al.* 2012), insects (Poff *et al.* 2006),
69 and fish (Albouy *et al.* 2011). The approach has proven to be exceptionally versatile, offering
70 functional insights into changes in assemblages through time (Friedman 2009; Villéger,
71 Novack-Gottshall & Mouillot 2011), the impacts of species invasions (Olden, Poff & Bestgen
72 2006; Corbin & D'Antonio 2010), and responses to environmental change (Laughlin *et al.*
73 2011; Graham *et al.* 2015).

74 Several studies have used traits to assess how fishing modifies aquatic ecosystems, however
75 most are based on in-situ observations in temperate countries, and none have assessed how
76 different fishing gears remove specific ecological traits from the ecosystem (Tillin *et al.*
77 2006; Guillemot *et al.* 2014; Koutsidi *et al.* 2016). In multi-species coral reef fisheries,
78 fishing gears are known to exhibit some degree of overlap in the species they capture
79 (McClanahan & Mangi 2001) and to reduce fish biodiversity (McClanahan 2015)
80 highlighting the need to understand how competitive interactions among gear types affect
81 outcomes (McClanahan & Kosgei 2018). However, the degree to which specific fishing gears
82 target different traits remains unclear.

83 Here, we employ a trait-based approach to assess the functional selectivity of seven fishing
84 gears, many of which are commonly used in small scale coral reef fisheries around the world.
85 Specifically, we ask the following questions; (i) do specific fishing gears target certain traits
86 and (ii) what overlaps are there in trait composition among gears? We use field data on
87 fisheries landings collected over a seven-year period from 25 coral reef and lagoon sites in
88 Kenya.

89 **Material and Methods**

90 ***Catch sampling***

91 We used catch data from 25 landing sites conducted monthly between 2010 and 2016 (Fig.
92 1). Observers identified landed catch to species level and recorded the number, size (total
93 length in cm), gear used, landing site name, and date. Although all sampling was conducted
94 during daylight hours, these include catches attributed to night-time fishing activities as
95 observers also intercepted fishers returning from their overnight fishing. At least 8 days of
96 data collection were achieved every month, translating into a total of 599 sampling days over
97 the survey period. We excluded 60 species that were represented by only one individual in

98 any gear to avoid potential misidentification. Our analysis is based on 19,401 fish
99 representing 245 species from 25 families, with a mean of 777 ± 546 fish per site. We
100 produced cumulative frequency curves to determine whether enough samples were collected
101 to reach asymptotes of observed functional entities. All curves reached saturation as
102 evidenced by the asymptote plateaus in the number of functional entities (unique
103 combinations of traits), suggesting that our sampling for each gear was adequate (Fig. 2).

104 We assessed functional selectivity of seven fishing gears: hook and line, speargun, gillnet,
105 beach seine, basket trap, and escape slot trap (modified basket traps that allows juveniles and
106 narrow-bodied species to escape through a gap). Apart from gillnet and beach seine, artisanal
107 fishers use a variety of other nets, such as ringnets, scoop nets, cast nets, and mosquito nets;
108 we therefore include a separate gear category of other nets. Gillnets and beach seines were
109 separated from 'other nets' because they are more frequently used.

110 *Associations of gears with traits*

111 Fish species were assigned to a set of categorical trait values relating to their diet, body-size,
112 mobility, time of activity, schooling behaviour, and position in the water column (Table S1).
113 These ecological traits are thought to be important for determining trophic role and have been
114 used in other studies examining functional diversity, vulnerability, and redundancy of fish
115 assemblages in tropical ecosystems (Micheli *et al.* 2014; Mouillot *et al.* 2014). While we use
116 ecological traits to calculate these widely used metrics, we acknowledge uncertainties
117 regarding specific links to ecosystem function and the need for more detailed ecological
118 studies to refine such approaches in the future (Bellwood *et al.* 2018). Each unique
119 combination of these six traits is considered a distinct functional entity (which may be
120 comprised of one or more species) (Mouillot *et al.* 2014). Of the 245 species sampled, we
121 derived 163 unique functional entities (FEs).

122 Associations between traits and fishing gears were examined using Principal Component
123 Analysis (PCA) based on the fourth root transformation of Wisconsin double standardized
124 abundance data (sum of all species). In this standardization, the fourth root of each element is
125 calculated after each element is divided by its column maximum and then divided by the row
126 total. This standardization is recommended for ordination of species data that exhibit
127 substantial differences in sample sizes across sampling units (Legendre & Gallagher 2001). A
128 permutational multivariate analysis of variance (PERMANOVA, Anderson 2001) on trait
129 categories (i.e., the category within each trait, such as size class) was performed to examine
130 the overall trait structure between the six traits. Differences in trait categories across gear
131 types for each of the six traits were determined using six separate PERMANOVA tests.
132 These analyses were based on Euclidean distances on fourth root transformation of
133 Wisconsin double standardized abundance data for each trait category (n = 999
134 permutations). Pairwise comparisons between trait categories were carried out to determine
135 differences in trait categories within the six traits. All pair-wise comparisons were carried out
136 at the site level based on unrestricted permutation of raw data to allow for a sufficient number
137 of unique permutations (>258) to be tested.

138 ***Functional structure***

139 A trait-based ordination analysis was used to describe variation in fish assemblage trait
140 structure among gear types. To build a multidimensional trait space (Mouillot *et al.* 2013), we
141 performed a Principal Coordinates Analysis (PCoA) using a FEs x traits matrix. FEs
142 coordinates on the first four principal axes of this PCoA were used to construct a synthetic
143 multidimensional ordination based on pairwise Gower's distances between functional entities
144 (Legendre & Legendre 2012) (Fig. 3). Gower's distances allows mixing of different types of
145 variables while giving them equal weight (Legendre & Legendre 1998). A square root
146 correction for negative eigenvalues was applied for Euclidean representation of distance

147 relationships among entities in order to avoid biased estimations of distances (Legendre &
148 Legendre 2012). Although there is no rule to choose *a priori* the number of dimensions,
149 spaces with higher dimensionality (i.e., with at least four dimensions) provide the best
150 assessment of diversity (Maire *et al.* 2015). We therefore selected *a posteriori* the first four
151 dimensions of the ordination, keeping a manageable number that reduced computing time and
152 allowed graphical representation. The third and fourth PCoA axes are presented as Fig. S1.

153 ***Associations of gears with assemblage functioning***

154 In describing how gears affect assemblage trait functioning, we utilised three indices:
155 functional volume (FV), functional redundancy (FR) and rarely targeted functional entities
156 (RFEs) (Mouillot *et al.* 2014). We define FV for each gear as the proportion of the functional
157 space the gear occupies relative to that of all fish caught (Villéger, Mason & Mouillot 2008).
158 Functional redundancy (FR) is the mean number of individuals per functional entity. Others
159 have used functional originality as a proxy of functional redundancy (Buisson *et al.* 2013;
160 Brandl *et al.* 2016). Here, we found a strong correlation between mean values of functional
161 originality and mean number of individuals per functional entity ($r = 0.95$, $n = 7$, $p < 0.001$)
162 (SI). The RFE index is expressed as the proportion of FEs that constitute less than 1% of
163 catch (total number of individuals) within a gear. With n_p being the relative abundance of a
164 functional entity in a gear, we express RFEs as the following ratio:

$$165 \quad n_p = \frac{n_i}{N_g}$$

166

$$167 \quad RFEs = \frac{FE - \sum_{i=1}^{FE} \min(n_p - 1, 1)}{FE}$$

168 N_g denotes the total number of individuals in a gear, FE the total number of functional
169 entities, and n_i the number of individuals in a functional entity i (Mouillot *et al.* 2014). We
170 also compute the number of unique FEs targeted by the different gear types.

171 In examining the distribution of catch by FEs, we found a very long tail i.e., although each
172 gear type caught dozens of FEs, the majority of the catch was typically comprised of only a
173 few functional entities. Consequently, we provide complementary analyses where we
174 examine, i) the entire catch, ii) the top 75% of the catch (i.e., the FEs representing the top
175 75% of catch abundance), and iii) the top 50% of the catch (i.e., only the most dominant
176 FEs).

177 ***Gear overlaps and exclusion analysis***

178 To show the levels of overlap in trait composition among gears, we quantified the number of
179 FEs and FV shared among gear types. Overlap in composition of FEs among gears was
180 visualised using Venn diagrams based on all catch, 75% and 50% of catch. Because of the
181 complexity in interpreting Venn diagrams with more than four elements, we only present
182 numbers and proportions of overlaps from the resultant groupings. To examine the potential
183 influence of gear restriction policies on functional diversity, we explored two sequential gear
184 exclusion case scenarios: i) one that minimises the number of FEs harvested by one or
185 combinations of gears regardless of overlaps across gears; and ii) one that minimises overlaps
186 notwithstanding the number of FEs targeted by a combination of two gears or more. All
187 values of FV are computed from the first four dimensions of the ordination.

188 ***Robustness analysis***

189 To determine whether our results are robust to the number and choice of traits, we reran all
190 analyses using all combinations of five traits out of six. We avoided reducing the number of
191 traits lower than five so as to retain important dimensions of the functional space defining

192 fish niches (Mouillot *et al.* 2013). As such, an over simplistic definition of FEs was avoided.
193 We further performed a crude categorization potentially inducing high functional redundancy
194 (many species in each FE) whereas a fine categorization would lead to few species in each FE
195 (Mouillot *et al.* 2014). For, example, instead of using all six categories on body size, we reran
196 the analysis testing the association between gears and three size categories (SI). In this
197 analysis, we reduced the number of categories for each trait and re-ran all analyses with 64
198 FEs (crude categorization) instead of 163 (fine categorization). To show the robustness of our
199 findings, we present the distribution of FEs contained in the entire catch, in addition to the
200 number of FEs and species in 50, 75, and 99% of the catch considering the reduced number
201 of FEs (163 – 64) for each gear (SI). We also show the proportion of RFEs and levels of FR
202 (SI).

203 ***Simulated random assignment models***

204 We tested whether the observed values of RFEs were significantly different from the null
205 hypothesis that individuals are randomly distributed into FEs. In each of the seven gears, we
206 simulated a random assignment of individuals to FEs while ensuring that each functional
207 entity had at least one individual. We simulated 999 random assemblages for each gear, and,
208 for each simulation, we computed rarely targeted FEs while the number of individuals and
209 the number of FEs were kept constant. All calculation of indices and statistical analyses were
210 performed with R statistical software (R Development Core Team, 2018) using packages
211 *cluster*, *ade4*, *stats*, and *vegan*. PERMANOVA tests were performed using the function
212 ‘*adonis*’ implemented in the R package *vegan*. Graphical representations were performed
213 with R and Sigma plot version 11.

214

215 **Results**

216 ***Trait composition in landed catch***

217 The majority of the fished assemblages were benthic species, species active during the day
218 and species actively mobile within reefs (Table S2; Fig. S2). Pelagics, species moving in
219 pairs, and those that feed on plants, animals, or plankton were rarely caught. Fish
220 assemblages with size ranges of 30-50 cm were disproportionately targeted whereas larger
221 species >80 cm were least caught (Fig. S2). Overall differences in trait structure between the
222 six traits were identified by the PERMANOVA ($F = 2.71$, $p < 0.001$, 978 permutations) (Table
223 S3). However, pairwise comparisons indicated differences in abundance between trait
224 categories for diet, activity, and mobility ($P \leq 0.05$; Table S4). No differences in abundance of
225 trait categories were detected for schooling behaviour and body size ($P > 0.05$).

226 *Associations of gears with traits*

227 Associations between gears and traits identified in the PCA were supported by the
228 PERMANOVA analyses with significant differences in trait composition between gears
229 ($P \leq 0.05$; Table S3; Fig. S3; Fig. 3a). Pairwise comparisons indicate that a vast majority of
230 categories within all six traits, except size, schooling behaviour, and mobility differed among
231 gear types (Table S5). Basket traps and escape slot traps predominately targeted benthic
232 herbivores/detritivores (e.g., rabbitfishes and parrotfishes) (Table S5; Fig. 3a). Sessile
233 invertivores (e.g., wrasses and porgies) and diurnal species were largely captured by
234 spearguns. A combination of other nets, including ring nets and cast nets, as well as hook
235 and lines, harvested species moving in large groups and feeding on plankton (e.g., mackerels
236 and jacks). Gillnet and beachseines were largely associated with pelagic species (e.g.,
237 mackerels and jacks).

238 Patterns of association between gears and traits did not change much when considering the
239 dominant 75% of the catch compared to those of all catch (Fig. S4a). For example, spearguns
240 and gillnets were consistently associated with diurnal species and fishes that moved between

241 reefs, respectively (Fig. S4a). Considering 50% of the catch, all types of nets (i.e., gillnets,
242 beachseines and other nets) exhibited substantial similarities in trait composition. Substantial
243 deviations in trait composition were observed between hook and line and spearguns (Fig.
244 S4b). Considering the dominant 75 or 50% of the catch, other nets were consistently
245 associated with pelagics (Fig. S4 a & b). There was no statistically significant difference in
246 traits composition among gear types when considering only 50% ($P = 0.18$) and 75% of catch
247 ($P = 0.77$).

248 *Functional diversity*

249 We detected substantial variability in the number of FEs (functional diversity) targeted by
250 gear types, ranging from 86 for spearguns to 57 for hook and line (Fig. 4). Distribution of
251 individuals among FEs was largely skewed with a few FEs containing a large number of
252 individuals, while the majority of FEs contained relatively few (Fig. 4). Having shown that
253 abundance was heavily packed in few FEs across all gear types, we decided to compare
254 functional diversity in the dominant catch, i.e., 75% and 50% of the total catch in each gear.
255 Here, we show that escape slot traps hosted the fewest FEs in 50% (2 FEs; six species) and
256 75% of catch (6 FEs; 31 species) (Fig. 5). The proportion of RFEs ranged between 11.9% for
257 gillnets and 25% for basket traps, and observed values were all significantly higher than
258 expected when abundance was randomly assigned to FEs (Fig. 4; Fig. S4b). Functional
259 redundancy, i.e. mean number of individuals per functional entity, ranged between 39.3 in
260 escape slot traps to 16.3 in gillnets (Fig. 4) with a species-functional entity gradient between
261 1.41 for escape slot traps to 0.87 for spearguns. This means that on average an escape slot
262 trap fisher had to catch about 40 fish individuals in order to remove one additional functional
263 entity from the ecosystem. In contrast, an additional functional entity was harvested when 16-
264 18 fish individuals were captured by fishers using beachseines, gillnets, or spearguns.

265 *Functional space and assemblage functioning*

266 The distribution of traits in the functional space showed that social grouping broadly changed
267 from left to right along the first axis of the PCoA, whereas fish body-size and mobility
268 increased from top to bottom along the second axis of the PCoA (Fig. 5). Herbivores,
269 detritivores and omnivores, typically associated closely with the benthos, were positioned
270 top-left in the functional space; sedentary, territorial and macroalgal herbivores were
271 positioned middle right; pairing invertivores targeting sessile invertebrates typically active
272 during the day were positioned to the top right; invertivores targeting mobile invertebrates
273 typically mobile within the reef in the top-right; planktivores in the middle-right; and larger
274 carnivores that were largely pelagic and typically mobile across reefs were located in the
275 bottom-left. The first four dimensions of the PCoA cumulatively explained 47.5% of the
276 projected inertia in the distribution of fish species traits (first two independent axes accounted
277 for 29.7% of the variance). Explained variances were not more than 17% per axis (Fig 5).

278 Nets in general (i.e. gillnets, beachseines, and other nets) filled more functional space (44.6-
279 50.7%), but did so targeting relatively fewer species (71-77 species) and FEs (59–64) than
280 many of the other gear types. For example, spearguns targeted 86 FEs comprised of 113
281 species, but only filled 34.3% of the FV. Similarly, hook and line targeted a slightly greater
282 number of species than nets (75 and 57, respectively), but occupied a much smaller FV
283 (28.2%). Trends in functional spaces occupied by the different gear types in the first two
284 dimensions of the ordination were mirrored in the third and fourth PCoA axes (Fig. 5; Fig.
285 S1).

286 RFEs in basket traps and spearguns affected a larger proportion of assemblage functioning
287 (48.8 – 55%). Only escape slot traps removed less than 20% of the total FV (assemblage
288 functioning) from the RFEs (Table 1). Spearguns had the highest number of unique FEs (12)

289 affecting 49.2% of the FV. Other gears caught between one (other nets) and eight (basket
290 traps) unique FEs, potentially affecting up to 38% of assemblage functioning. Unique FEs
291 associated with nets, such as gillnets and other nets, occupied the least FV (<4%) (Table 1).

292 ***Gear overlaps, exclusion and functional optimization analyses***

293 Gear overlap harvested 19 FEs, representing 12% of the functional diversity (i.e., the total
294 number of FEs) in this fishery (Fig. 3a). This shared component included 45 out of 245
295 species, representing an 18% overlap among gear types. This gear overlap represented 47.3%
296 of the total FV targeted by fishing. There was substantial similarity in traits captured by
297 fishing gears whose mode of operation and fishing grounds were related, e.g., basket traps
298 and escape slot traps (64% similarity); gillnet and other nets (53% similarity) (Fig. 3a).
299 Despite a reduction of two FEs (19 vs. 17) when considering 75% of the catch, the total FV
300 of the shared catch component remained the same (Fig. S4a). The 17 out of 29 FEs shared
301 across all gears represents a 59% overlap. Of the 63 species present in 75% of the catch, 42
302 were shared across all gear types, representing a 67% overlap. When considering 50% of the
303 catch, all gears together shared nine out of 11 FEs, representing 82% overlap (Fig. S4c). This
304 catch component included 29 out of 31 species and affected 30% of assemblage functioning.

305 Functional diversity (number of FEs) in the landed catch decreased when one or
306 combinations of gears were excluded from the analysis (Fig. 3b). However, the levels of
307 overlap across gears increased when one or combinations of gears were excluded from the
308 analysis (Fig. 3c). For example, about 157 (± 2.3) FEs on average were captured when one
309 gear was excluded from the fishery, representing a 3.9% decrease in functions removed from
310 the ecosystem relative to all gear types (Fig. 3b). As more gears were excluded from the
311 fishery, the number of FEs conserved (i.e. not removed from the ecosystem) gradually
312 increased from 9.3 (2 gears removed) to 58 (6 gears removed)% (Fig. 3b). Conversely, the

313 number of shared FEs gradually increased from 12.6% to 37.4% when combinations of two
314 to five gears were excluded from the fishery (Fig. 3c). The number of FEs harvested by
315 various combinations of six gears (i.e., random removal of one gear), notwithstanding
316 overlaps, were lowest when spearguns were excluded (Fig. 3b). In order to sequentially
317 conserve the optimal number of fish functions by removing more than one gear, the other
318 gears to be excluded were basket traps, escape slot traps, hook and lines, and gillnets in that
319 order (Fig. 3b). When considering a sequence of gear exclusion that minimizes the number of
320 shared FEs (overlaps), a totally different trend emerged. Other nets were the first candidates,
321 followed by basket traps, beachseines, gillnets, and spearguns (Fig. 3c).

322 **Robustness analysis**

323 Using the crude functional categorisations, the level of RFEs was surprisingly close to that
324 observed with a much finer categorization (Fig. S5 & S6). The observed distribution of
325 abundance between FEs was more right skewed than expected, with a long tail with few
326 individuals (Fig. S6). Overall, our sensitivity analyses showed the consistent and uneven
327 distribution of some FEs, regardless of the number, identity, or categorization of traits. We
328 also reran all analyses with all combinations of five traits out of six (Fig. S5). Whatever the
329 combination, all patterns were still close to those observed with six traits (Fig. S5).

330

331 **Discussion**

332 Fished assemblages in this fishery were characterized by high capture of multiple functional
333 groups (e.g., rabbitfishes, parrotfishes, and emperors). However, a vast majority of the catch
334 was dominated by a few FEs overlapping across all gears to make up the largest proportion of
335 the catch. The breadth of functional diversity represented by unique and RFEs varied greatly
336 among gears. This indicates that differences in gear selectivity along with specific impacts of

337 fishing are driven by depauperate FEs represented by species that are opportunistically
338 caught. Findings highlight the need to manage fishing of rare taxa with potentially important
339 ecological functions because overfishing these species could alter ecosystem functioning.

340 On average, one gear has the potential to affect at least 43% of functional diversity in the
341 overall catch landed. Importantly, most of the catch for each gear is represented by very few
342 dominant FEs. Our study presents the first empirical evidence that shows that beyond the
343 number of species (Mouillot *et al.* 2014), abundance also packs into a few FEs. Gears that are
344 less selective, such as gillnets and beachseines, targeted the greatest breadth of functional
345 diversity. This was not surprising because nets in general target a larger portion of the water
346 column and habitat, thus increasing the likelihood of catching many species and high
347 functional diversity. Gillnets and beachseines also seem to catch FEs in more even
348 proportions. Besides differences in selectivity in terms of functional diversity, there were
349 differences in the actual level of catch among gear types. The very few FEs that dominated
350 the landings, however, occupy less functional volume - at least for some gears such as traps
351 and hook and lines. This means that gears that are more selective, such as escape slot traps
352 and hook and lines, rarely captured species that are functionally diverse. Our results indicate
353 that these more selective gear types target a narrow range of ecosystem functions (Murawski
354 2000).

355 Only two gears, spearguns and other nets, exert strong trait specialization in this fishery.
356 Spearguns specialised on diurnal and sedentary/territorial species that feed on plankton.
357 Spearguns disproportionately caught species that move in pairs (e.g., butterflyfishes) and
358 diurnal invertivores that feed on sessile invertebrates (Kruer & Causey 2005). Differences in
359 selectivity between gears show that other nets disproportionately targeted pelagic species
360 (e.g., bonitos) because nets (e.g., ring nets) are often used further offshore – habitats that are

361 typically associated with pelagic fishes (Fock, Pusch & Ehrich 2004). Basket traps targeted
362 more herbivorous detritivores than hook and lines because trap fishers in Kenya use a mixture
363 of plants and animals (seagrass, algae, crushed sea urchins, brittle stars and molluscs) as bait
364 (Mbaru & McClanahan 2013). In sum, physical and operational differences, in addition to
365 behavioural activity among species can provide clear insights into the differences in gear
366 selectivity on coral reef fishes.

367 The identification of unique and rarely targeted traits or their combinations can have
368 significant fisheries management implications. Unique or rarely targeted species in fished
369 assemblages occupied a higher proportion of functional volume, even for some selective
370 gears such as hook and lines and spearguns. In heavily exploited areas, depauperate
371 functional entities can be represented by few opportunistic and tolerant fish species in the
372 community (Mouillot *et al.* 2014). Previous ecological investigations on in-situ fish data
373 show that rare species tend to represent a high level of functional diversity within
374 communities (Micheli *et al.* 2014). These findings suggest that functional volume removed
375 by unique functional entities and rarely targeted functional entities can significantly amplify
376 the effect of specific fishing gears on ecosystem functioning. Future management priorities
377 should include monitoring of gears such as hook and lines that are traditionally assumed to be
378 selective, especially on the type and size of hooks used.

379 Previous studies show that species diversity, functional diversity, and assemblage functioning
380 are intricately linked (Villéger, Mason & Mouillot 2008; Mouillot *et al.* 2013). Here, we
381 show that these relationships in fished assemblages are not linear. For example, we show that
382 beachseines captured slightly fewer species than other nets, yet had more FEs that occupied
383 slightly less FV. Fish assemblages in escape slot traps had the third highest number of
384 species, yet had the most redundant FEs and occupied the lowest FV compared to all gear

385 types. This implies that capturing a high species diversity for some gears may not strongly
386 contribute to loss of ecosystem functioning (Wackernagel & Rees 1998; Murawski 2000;
387 Cumming, Cumming & Redman 2006). The number of species, FEs and FV was lower in
388 escape slots compared to basket traps. The inclusion of escape slots in basket traps therefore
389 has the potential to increase selectivity in traps.

390 The overrepresentation of functional groups associated with benthic-attached species is
391 consistent with the role played by tropical coral reefs in supporting demersal fish diversity
392 (Micheli & Halpern 2005). This is not surprising because a majority of the gears analysed
393 here are associated with the bottom habitat (McClanahan & Mangi 2004). Because a vast
394 majority of demersal species do not often share common traits (Elleouet *et al.* 2014), this
395 perhaps explains why the landings are characterized by a high capture of multiple functional
396 groups. Our catch data contained few pelagics, macroalgae feeders, and large body sized fish
397 moving in pairs. Given the high fishing intensity in Kenya (Hicks & McClanahan 2012;
398 McClanahan 2018), low prevalence of functional redundancy in trait combinations is likely a
399 result of low abundance and diversity, and hence availability for capture. Specifically, it is
400 highly likely that the low prevalence of herbivorous fishes and large body sized fish is due to
401 chronic overfishing. However, the low prevalence of ecological traits associated with the
402 pelagic environment is likely because only a small proportion of fishers use fishing
403 techniques that target pelagic stocks in Kenya (Mbaru 2012).

404 The extent of overlap suggests that all gears together can affect >47% of the total breadth of
405 functional diversity targeted. This is not surprising because the mode of operation of the
406 fishing gears analysed and fishing grounds are related to traits of the catches, like type of
407 movement, predation, or habitat (McClanahan & Cinner 2008). There is hardly any change in
408 FV among shared FEs in 75% of the catch (47 vs 48.4%). Proportions of shared FEs increase

409 substantially when considering the dominant 50% of the catch suggesting that >80% of FEs
410 can potentially be harvested by just one gear. This shared component however affects only
411 27% of the total FV – emphasizing the point that the differences between fishing gears is
412 primarily from the non-dominant part of the catch.

413 To minimize the number of overlapping FEs based on all catch, we developed a sequential
414 gear exclusion process. Other nets were the first candidate gear to exclude, followed by
415 basket traps, beachseines, gillnets, and spearguns. However, given the extent of overlap in
416 functional diversity, we show that even just using two gears (i.e., excluding five gears) still
417 targets >60% of total FV. These findings indicate that the risk of targeting certain fish
418 functions is not necessarily or substantially reduced when one or more gears are excluded
419 from the fishery. Thus, initiatives to reduce overall effort should reduce pressure on
420 important trait combinations.

421 **Conclusion**

422 Multi-species, multi-gear small-scale fisheries are common in coral reef ecosystems around
423 the globe (McClanahan & Mangi 2004; Cinner *et al.* 2009). This study is the first attempt to
424 connect traits to catches of fishing gears in multispecies coral reef fisheries and provide
425 insights on how fishing can modify ecosystem biodiversity in tropical regions. Key findings
426 unveiled here could shape future assessments of the ecological implications of fishing in
427 multi-gear and multi-species fisheries. Across all gears, the distribution of individuals across
428 FEs is largely skewed, with the majority of the catch comprising relatively few FEs.
429 Although there was over-redundancy in the few functional entities in the dominant portion of
430 the catch, the low redundancy in the long-tail of FEs removed by fishing can potentially
431 affect ecosystem functioning. We show that gear selectivity in coral reef fisheries is primarily
432 due to unique, and rarely targeted species. Considering fished assemblages as one monolithic

433 catch component can be a weak indicator of differences in gear selectivity. The addition of a
434 traits approach to standard analyses could provide a concrete foundation for the formulation
435 of the Ecosystem Approach to Fisheries Management in tropical multispecies fisheries.

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443 We have no conflict of interest to declare.

444 **Author Contributions**

445 E.K.M, N.A.G, J.E.C conceived the idea and designed the methodology. T.R.M provided the
446 data. E.K.M analysed the data and led the writing of the manuscript. All authors contributed
447 critically to the drafts and gave final approval for the publication.

448 **Data Accessibility**

449 All data from the manuscript will be made publicly available on the Knowledge Network for
450 Biocomplexity.

451

452

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Tables

Table 1. Number of functional entities, species (in parenthesis), and functional volume (FV) in unique and rarely targeted functional entities (RFEs) for each gear. FV is expressed as a percentage of volume occupied by all fish caught.

Fishing gear	Unique FEs		Rarely targeted FEs	
	No. FEs(species)	% Functional volume	No. FEs(species)	% Functional volume
Escape slot trap	6(6)	16	6(6)	19.9
Basket trap	8(9)	37.9	13(13)	48.8
Hook and line	7(7)	21.5	8(8)	39.2
Speargun	12(12)	49.2	11(12)	55
Gillnet	3(4)	3.6	7(7)	22.2
Beachseine	7(9)	20.1	10(10)	53.2
Other nets	1(1)	-	7(7)	32.5

Figures

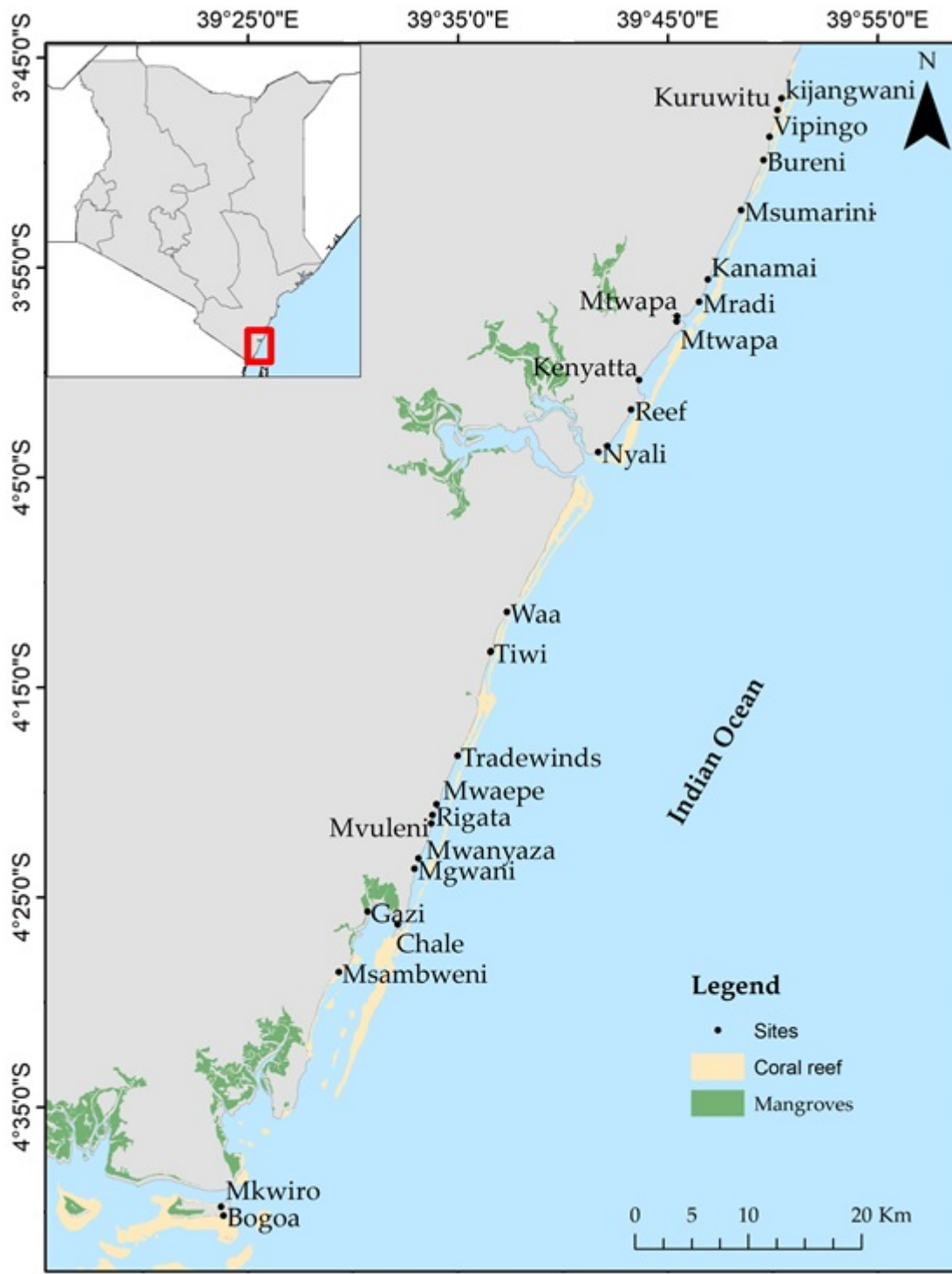


Fig. 1. Map of sampling sites

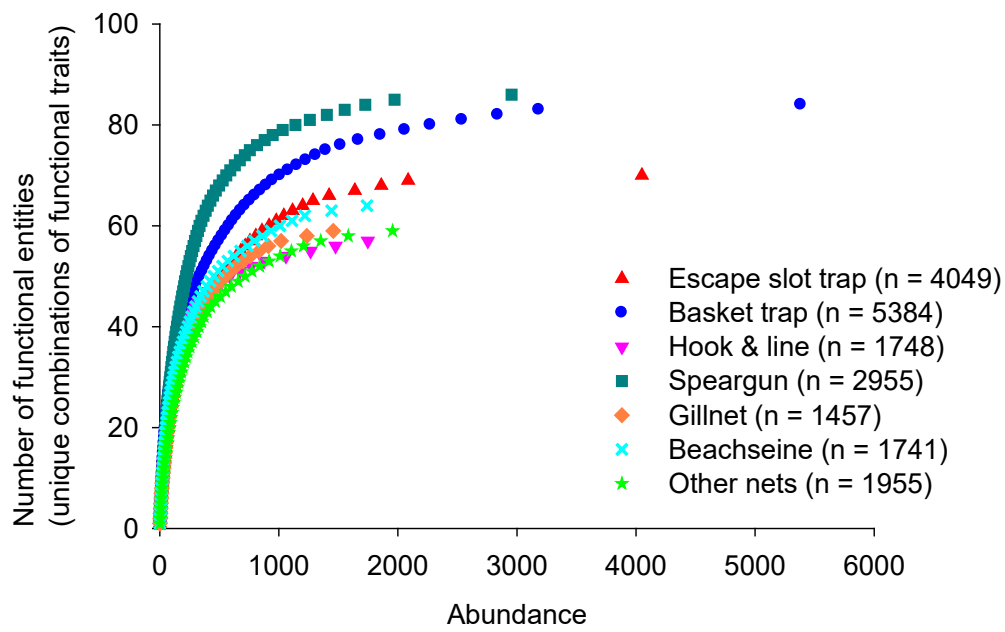
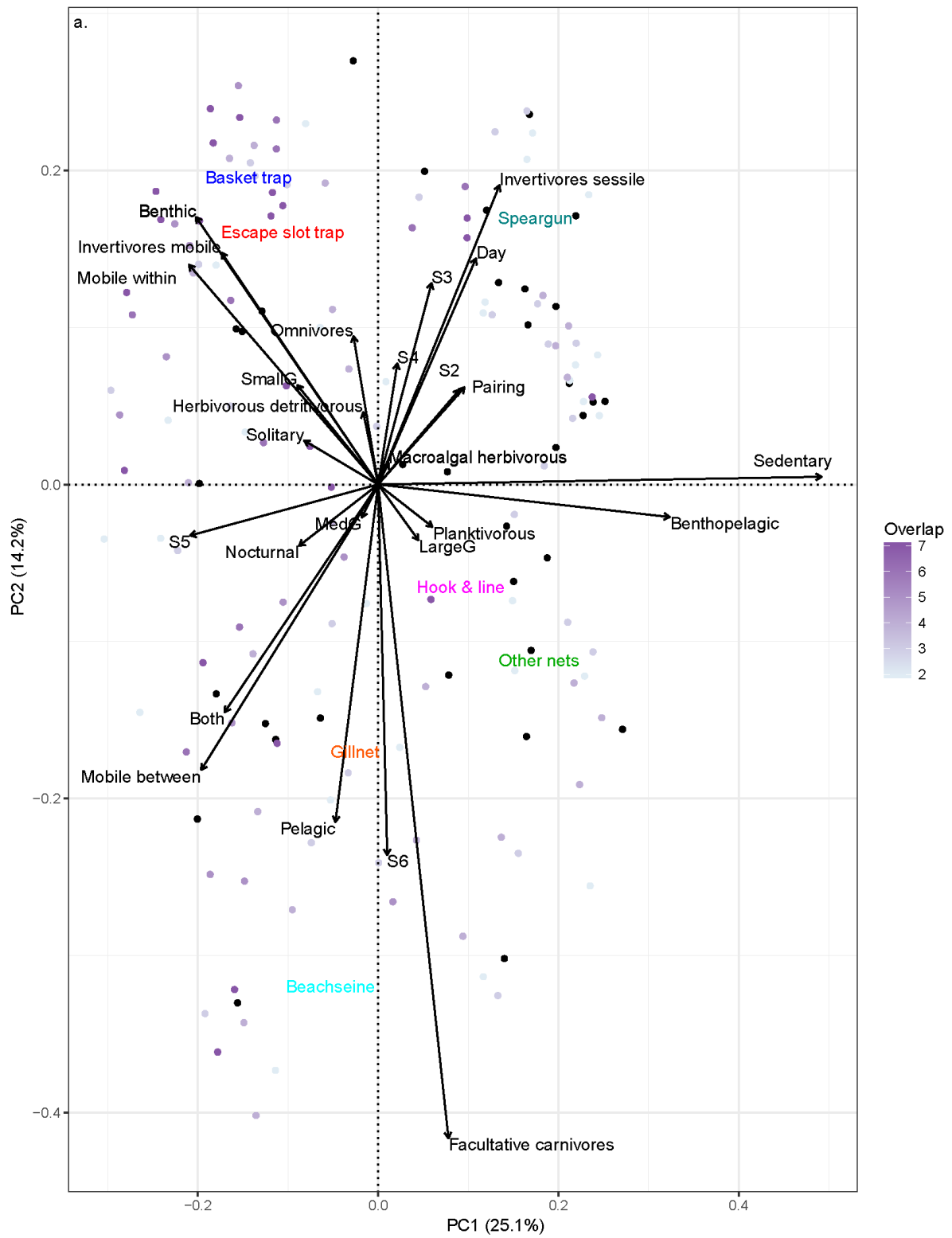


Fig. 2. Cumulative frequency curves of the number of functional entities (i.e., unique combinations of traits) present in sampled fish assemblages per gear. Sample sizes are displayed in parenthesis.



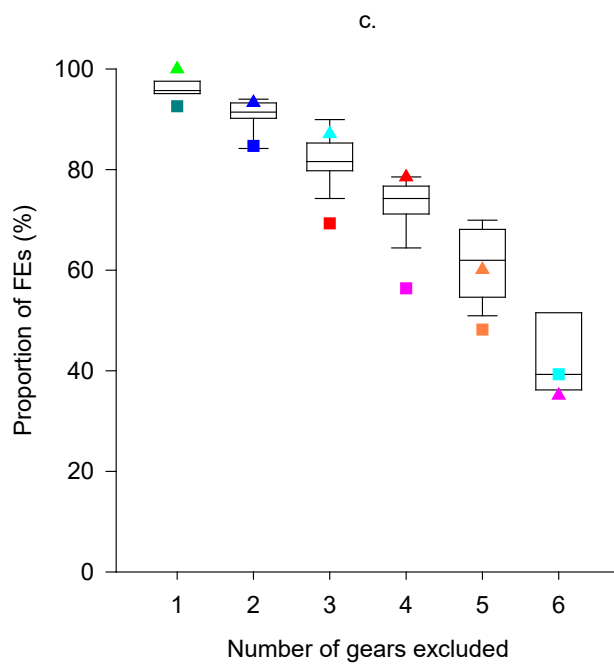
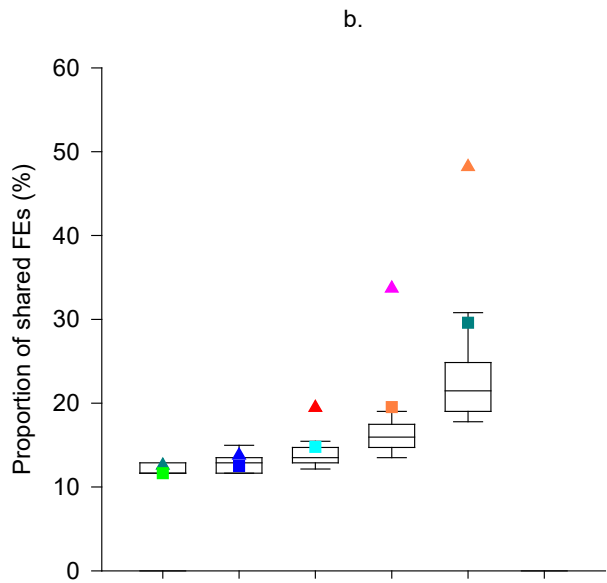
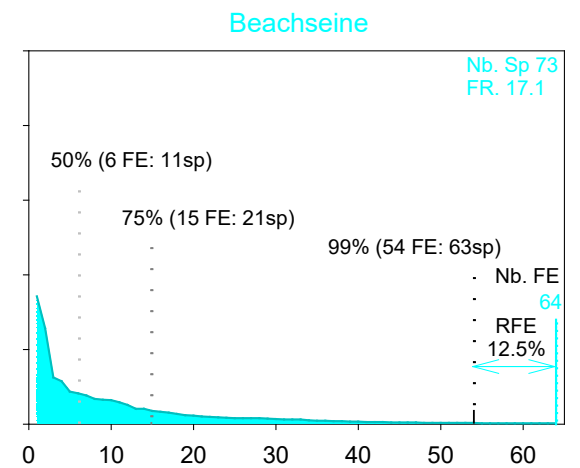
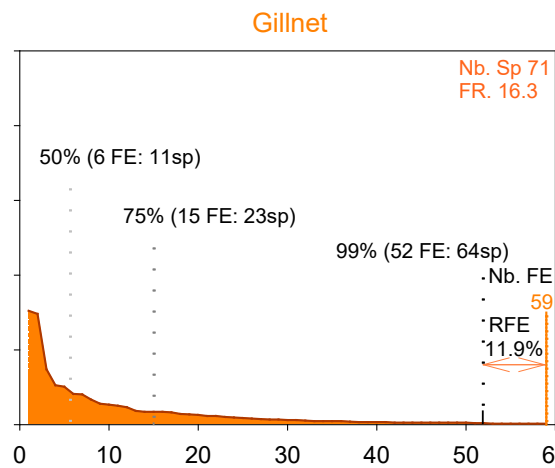
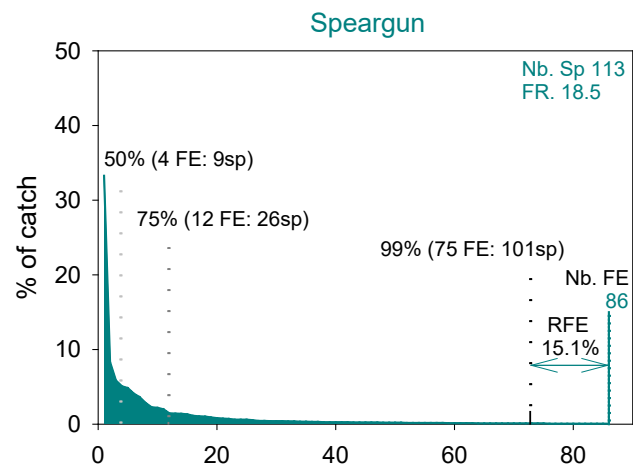
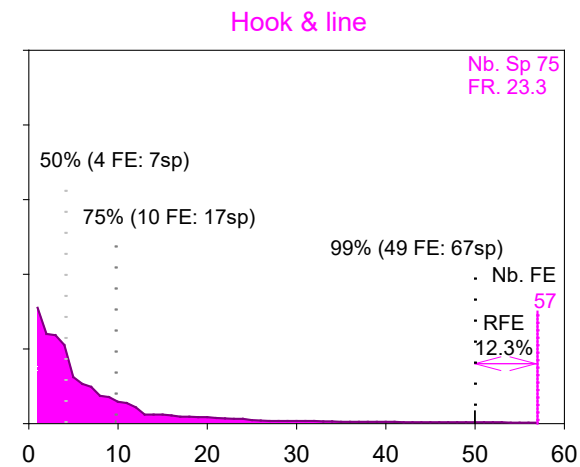
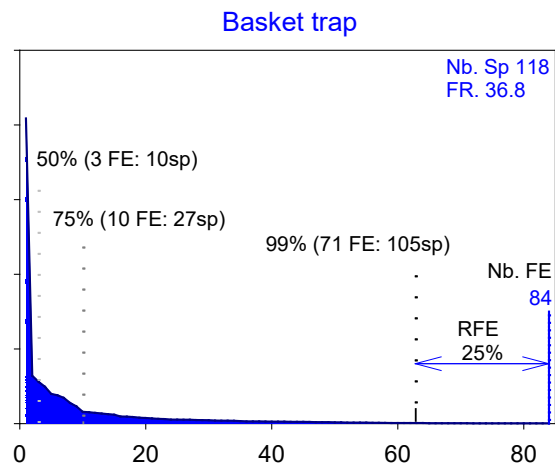
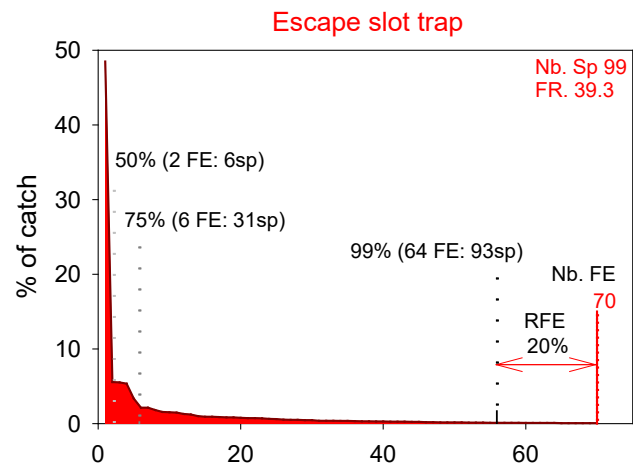


Fig. 3. (a). Principal component analysis of functional entities (FEs) contained in the entire catch (n=163). Coloured dots represent FEs captured by the seven gear types analysed. Colour gradient represents FEs shared across a range of gear combinations. Black dots represent unique FEs targeted by a single gear. LargeG (>50 individuals), MedG (20-50 individuals), and SmallG (3-20 individuals) indicate

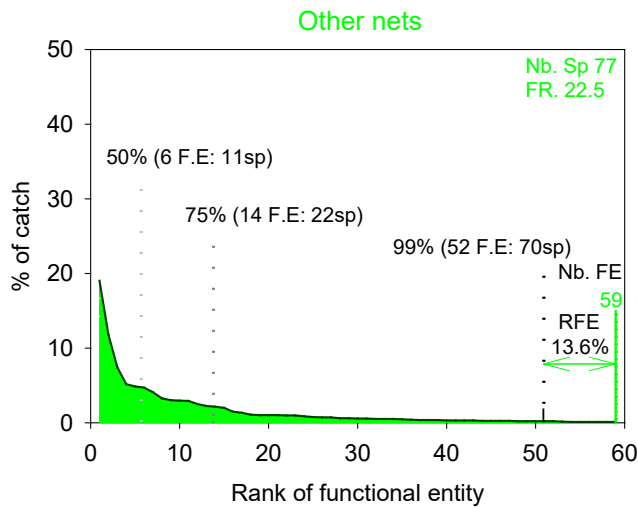
schooling behaviour. Fish size is coded using six categories: 7.1-15 cm (S2), 15.1-30 cm (S3), 30.1-50 cm (S4), 50.1-80 cm (S5), and >80 cm (S6). 'Both' denotes species active during the day and night. (b). Proportion of FEs targeted when one or a combination of gears are excluded from the analysis (box plot). Coloured squares denote the minimum number of FEs targeted by one or a combination of gears. Coloured triangles denote the number of FEs targeted by one or a combination of gears when the number of shared FEs (overlaps) is minimized. (c). Box plot shows the proportional distribution of shared FEs for each number of gear removal. Coloured squares show the lowest proportions of shared FEs when the number of overlapping FEs is minimized on the gear exclusion process. Coloured triangles depict the lowest proportions of shared FEs when the number of FEs targeted is minimized. Minimum values of targeted and shared FEs were based on all possible permutations of in the gear exclusion process. Colour codes represent gears as in

Fig

2.

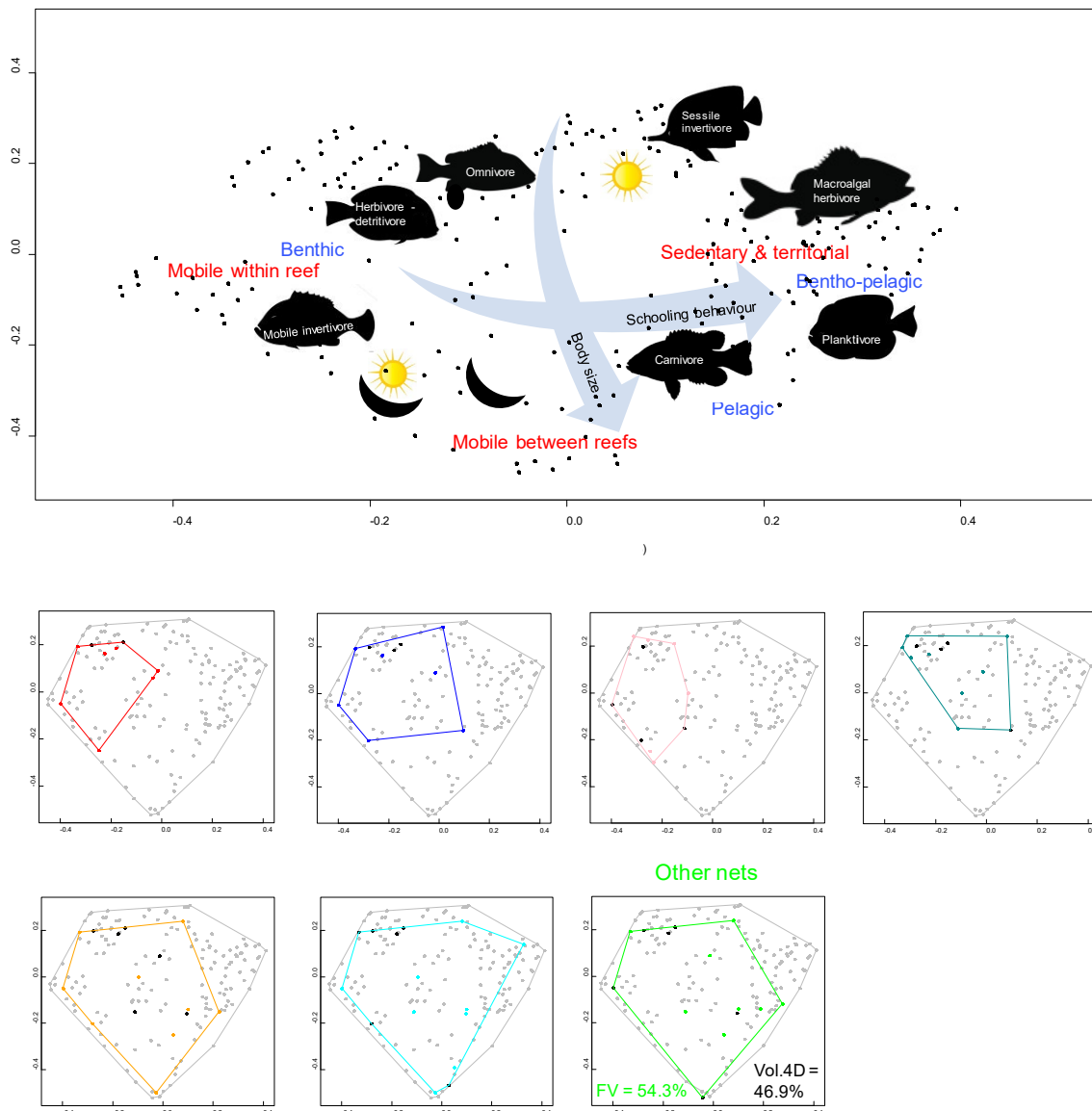


Rank of functional entity



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2 Fig. 4. The distribution of fish individuals into functional entities (FEs) is displayed for
 3 each gear type. The number of functional entities (“Nb FE.”) present in each gear is
 4 shown at the bottom right of the distribution. Functional redundancy (FR) (i.e., the
 5 mean number of individuals per functional entity) and number of species are
 6 displayed at far right top corner. The light grey dashed line illustrates number of FEs
 7 contained in 50% of catch. The grey dashed line illustrates number of FEs contained
 8 in 75% of catch while the black dashed line illustrates number of FEs contained in
 9 99% of catch. Rarely targeted functional entities (RFE) i.e., functional entities
 10 contained in 1% of total number of individuals captured in each gear is illustrated in
 11 double arrows displayed at far right bottom corner.



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13 Fig. 5. Distribution of functional entities is shown in functional space from a Principal
 14 Coordinate Analysis on traits based on all catch. 163 computed functional entities
 15 (black dots) plotted in the first two dimensions (four total) of functional space defined
 16 by six traits: body size (arrow indicating increasing body length), diet; mobility (red
 17 text); time of activity (sun and moon); social grouping (arrow indicating increasing
 18 size of fish school); and position in the water column (blue text). Illustrations and text
 19 show the position of average trait levels in the functional space. Distribution of
 20 functional entities is shown in functional spaces for each gear from a Principal

21 Coordinate Analysis on traits (bottom convex hulls). Colour filled points are FEs
22 present in the catch of each gear while grey filled points are FEs absent in the 75%
23 catch of each gear. The total convex hull, including the 245 species split into 163
24 FEs, is enclosed by grey continuous lines joining vertices of the convex hull that
25 shape edges. Continuous coloured lines outline the FV (FV) determined by coloured
26 points representing the most abundant FEs comprising 75% of the catch for each
27 gear. Coloured text in the bottom convex hulls show FV for the first two dimensions
28 whereas black text indicate FV considering all four dimensions. FV is expressed as a
29 percentage relative to all fish caught. Black points are FEs representing the most
30 abundant FEs comprising 50% of the catch for each gear.

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