

1 **Carnivore hotspots in Peninsular Malaysia and their**  
2 **landscape attributes**

3 Shyamala Ratnayeke<sup>1\*</sup>, Frank T. van Manen<sup>2</sup>, Gopaldasamy Reuben Clements<sup>1&</sup>, Noor Azleen

4 Mohd Kulaimi<sup>3&</sup>, Stuart P. Sharp<sup>4&</sup>

5 <sup>1</sup>Department of Biological Sciences, Sunway University, Malaysia

6  
7 <sup>2</sup>U.S. Geological Survey, Northern Rocky Mountain Science Center, Interagency Grizzly Bear Study Team, Bozeman,  
8 MT 59715, USA

9  
10 <sup>3</sup>Ex-Situ Conservation Division, Department of Wildlife and National Parks, Malaysia

11  
12 <sup>4</sup>Lancaster Environment Centre, Lancaster University, Lancaster, LA1 4YQ, UK

13  
14  
15  
16 \*Corresponding author

17 Email: shyamalar@sunway.edu.my

18  
19 <sup>†</sup>SR and FTVM are joint senior authors

20 <sup>&</sup>These authors also contributed equally to this work

21  
22

23

24

25

26

27

28

29

30

31

32

33 *Disclaimer: This draft manuscript is distributed solely for purposes of scientific peer review. Its content is*  
34 *deliberative and pre-decisional, so it must not be disclosed or released by reviewers. Because the*  
35 *manuscript has not yet been approved for publication by the U.S. Geological Survey (USGS), it does not*  
36 *represent any official USGS finding or policy.*

## 37 Abstract

38 Mammalian carnivores play a vital role in ecosystem functioning. However, they are prone to  
39 extinction because of low population densities and growth rates, [large area requirements](#), and  
40 high levels of persecution [or exploitation](#). In tropical biodiversity hotspots such as Peninsular  
41 Malaysia, rapid conversion of natural habitats threatens the persistence of this vulnerable  
42 group of animals. Here, we carried out the first comprehensive literature review on 31  
43 carnivore species reported to occur in Peninsular Malaysia and updated their probable  
44 distribution. We georeferenced 375 observations of 28 species of carnivore from 89 unique  
45 geographic locations using records spanning 1948 to 2014. Using the Getis-Ord  $G_i^*$  statistic and  
46 weighted survey records by IUCN Red List status, we identified hotspots of species that were of  
47 conservation concern and built regression models to identify environmental and anthropogenic  
48 landscape factors associated with Getis-Ord  $G_i^*$  scores. Our analyses identified two carnivore  
49 hotspots that were spatially concordant with two of the peninsula's largest and most  
50 contiguous forest complexes, associated with Taman Negara National Park and Royal Belum  
51 State Park. A cold spot overlapped with the southwestern region of the Peninsula, reflecting the  
52 disappearance of carnivores with higher conservation rankings from increasingly fragmented  
53 natural habitats. Getis-Ord  $G_i^*$  scores were negatively associated with elevation, and positively  
54 associated with the proportion of natural land cover and distance from the capital city.  
55 Malaysia contains some of the world's most diverse carnivore assemblages, but recent rates of  
56 forest loss are some of the highest in the world. Concerted efforts to reduce poaching and  
57 maintain large contiguous tracts of lowland forests will be critical, not only for the persistence  
58 of large mammals, but for threatened carnivores in general.

59 **Key words:** predator, tropical rainforest, landscape, hotspot, coldspot, Carnivora

## 60 Introduction

61 Few taxonomic groups elicit as much conservation attention as mammalian carnivores [1-3].  
62 Carnivores of various sizes play a crucial role influencing the composition and dynamics of  
63 ecological communities [4]. The loss of apex predators has been linked to cascading  
64 consequences for smaller herbivores regulated by mid-order predators [5-6], which in turn can  
65 influence plant growth and recruitment via altered patterns of herbivory, seed predation, and  
66 seed dispersal [3,4,7,8]. Charismatic carnivores often serve as conservation flagships [9], and  
67 when their area and resource requirements encompass those of numerous species, they serve  
68 as conservation umbrellas [10-12]. Carnivore presence may be linked positively with  
69 biodiversity [13,14] habitat integrity [15] and ecological processes [4]. Ironically, the very  
70 characteristics that make carnivores such effective conservation surrogates also make them  
71 extinction-prone.

72 Mammalian carnivores are vulnerable to extinction mainly due to habitat loss and human-  
73 induced mortality [16,17]. Carnivores in general occupy the higher region of ecological food  
74 webs, composing a relatively small fraction of ecological biomass and requiring a healthy prey

75 base to maintain viable populations. Large carnivores need substantial areas that support the  
 76 prey they subsist on and some level of functional landscape connectivity for persistence. Loss of  
 77 habitat and prey renders them prone to conflicts with humans [18–21]. Furthermore, carnivores  
 78 are prime targets for poachers seeking valuable body parts or trophies [22–25] and their life  
 79 histories often hinder recovery from population declines [26]. Not surprisingly, many carnivore  
 80 populations across the globe are threatened [27].

81 Carnivore species richness in Peninsular Malaysia is one of the highest in the world, with 31  
 82 species representing seven families recorded to date [28] (Table 1). Sixteen (57%) of the  
 83 remaining 28 species are listed as critically endangered, endangered, vulnerable, or near  
 84 threatened at the global level [27]. The most recent local assessment of the conservation status  
 85 of mammals lists 14 carnivore species as threatened or near threatened in Peninsular Malaysia  
 86 [28].

87 **Table 1. Carnivores of Malaysia with 2015 IUCN conservation status, and Peninsular Malaysia conservation**  
 88 **status in 2007 and 2009 based on percent change in area of occupancy and expert opinion [28].** Although 31  
 89 species are listed, three species may not be indigenous or extant. The highest threat status, based on IUCN  
 90 Red List criteria A–E [29] is reported for each species. EX = extinct, CE = critically endangered, EN =  
 91 endangered, VU = vulnerable, NT = near threatened, LC = least concern.

	Family	Species	Common name	IUCN 2015 Red List status	Peninsular Malaysia 2009 Red List status <sup>a</sup>
1	Canidae	<i>Cuon alpinus</i>	Dhole	EN	NT
2	Felidae	<i>Panthera tigris</i>	Tiger	<del>CE</del> <sup>b</sup> EN	EN
3	Felidae	<i>Panthera pardus</i>	Leopard	NT	EN
4	Felidae	<i>Neofelis nebulosa</i>	Clouded leopard	VU	NT
5	Felidae	<i>Pardofelis marmorata</i>	Marbled cat	NT	LC
6	Felidae	<i>Prionailurus bengalensis</i>	Leopard cat	LC	LC
7	Felidae	<i>Prionailurus viverrinus</i>	Fishing cat <sup>cb</sup>	EN	VU
8	Felidae	<i>Prionailurus planiceps</i>	Flat-headed cat	EN	NT
9	Felidae	<i>Catopuma temminckii</i>	Asian golden cat	NT	LC
10	Herpestidae	<i>Herpestes javanicus</i>	Javan mongoose	LC	LC
11	Herpestidae	<i>Herpestes edwardsii</i> <sup>b</sup>	Indian gray mongoose <sup>de</sup>	LC	EX
12	Herpestidae	<i>Herpestes brachyurus</i>	Short-tailed mongoose	LC	LC

Formatted: Superscript

13	Herpestidae	<i>Herpestes urva</i>	Crab-eating mongoose	LC	EN
14	Mustelidae	<i>Martes flavigula</i>	Yellow-throated marten	LC	NT
15	Mustelidae	<i>Mustela nudipes</i>	Malay weasel	LC	NT
16	Mustelidae	<i>Aonyx cinerea</i>	Asian small-clawed otter	VU	LC
17	Mustelidae	<i>Lutra sumatrana</i>	Hairy-nosed otter	EN	LC
18	Mustelidae	<i>Lutra lutra</i> <sup>c</sup>	Eurasian otter <sup>ed</sup>	NT	EN
19	Mustelidae	<i>Lutrogale perspicillata</i>	Smooth otter	VU	LC
20	Prionodontidae	<i>Prionodon linsang</i>	Banded linsang	LC	NT
21	Ursidae	<i>Helarctos malayanus</i>	Malayan sun bear	VU	VU
22	Viverridae	<i>Viverricula indica</i>	Small Indian civet	LC	NT
23	Viverridae	<i>Viverra zibetha</i>	Large Indian civet	NT	NT
24	Viverridae	<i>Viverra megaspila</i>	Large spotted civet	VU	EN
25	Viverridae	<i>Viverra zibetha</i>	Large Indian civet	NT	NT
26	Viverridae	<i>Cyanogale bennetti</i>	Otter civet	EN	EN
27	Viverridae	<i>Paguma larvata</i>	Masked palm civet	LC	LC
28	Viverridae	<i>Paradoxurus hermaphroditus</i>	Common palm civet	LC	LC
29	Viverridae	<i>Hemigalus derbyanus</i>	Banded civet	NT	LC
30	Viverridae	<i>Arctogalidia trivirgata</i>	Small-toothed palm civet	LC	LC
31	Viverridae	<i>Arctitis binturong</i>	Binturong	VU	LC

## 92 **Abstract**

93 Mammalian carnivores play a vital role in ecosystem functioning. However, they are prone to  
94 extinction because of low population densities and growth rates, [large area requirements](#), and  
95 high levels of persecution [or exploitation](#). In tropical biodiversity hotspots such as Peninsular  
96 Malaysia, rapid conversion of natural habitats threatens the persistence of this vulnerable  
97 group of animals. Here, we carried out the first comprehensive literature review on 31  
98 carnivore species reported to occur in Peninsular Malaysia and updated their probable  
99 distribution. We georeferenced 375 observations of 28 species of carnivore from 89 unique  
100 geographic locations using records spanning 1948 to 2014. Using the Getis-Ord  $G_i^*$  statistic and  
101 weighted survey records by IUCN Red List status, we identified hotspots of species that were of  
102 conservation concern and built regression models to identify environmental and anthropogenic  
103 landscape factors associated with Getis-Ord  $G_i^*$  scores. Our analyses identified two carnivore  
104 hotspots that were spatially concordant with two of the peninsula's largest and most  
105 contiguous forest complexes, associated with Taman Negara National Park and Royal Belum  
106 State Park. A cold spot overlapped with the southwestern region of the Peninsula, reflecting the  
107 disappearance of carnivores with higher conservation rankings from increasingly fragmented  
108 natural habitats. Getis-Ord  $G_i^*$  scores were negatively associated with elevation, and positively  
109 associated with the proportion of natural land cover and distance from the capital city.  
110 Malaysia contains some of the world's most diverse carnivore assemblages, but recent rates of  
111 forest loss are some of the highest in the world. Concerted efforts to reduce poaching and  
112 maintain large contiguous tracts of lowland forests will be critical, not only for the persistence  
113 of large mammals, but for threatened carnivores in general.

114 **Key words:** predator, tropical rainforest, landscape, hotspot, coldspot, Carnivora

## 115 **Introduction**

116 Few taxonomic groups elicit as much conservation attention as mammalian carnivores [1-3].  
117 Carnivores of various sizes play a crucial role influencing the composition and dynamics of  
118 ecological communities [4]. The loss of apex predators has been linked to cascading  
119 consequences for smaller herbivores regulated by mid-order predators [5-6], which in turn can  
120 influence plant growth and recruitment via altered patterns of herbivory, seed predation, and  
121 seed dispersal [3,4,7,8]. Charismatic carnivores often serve as conservation flagships [9], and  
122 when their area and resource requirements encompass those of numerous species, they serve  
123 as conservation umbrellas [10-12]. Carnivore presence may be linked positively with  
124 biodiversity [13,14] habitat integrity [15] and ecological processes [4]. Ironically, the very  
125 characteristics that make carnivores such effective conservation surrogates also make them  
126 extinction-prone.

127 Mammalian carnivores are vulnerable to extinction mainly due to habitat loss and human-  
128 induced mortality [16,17]. Carnivores in general occupy the higher region of ecological food  
129 webs, composing a relatively small fraction of ecological biomass and requiring a healthy prey

130 base to maintain viable populations. Large carnivores need substantial areas that support the  
 131 prey they subsist on and some level of functional landscape connectivity for persistence. Loss of  
 132 habitat and prey renders them prone to conflicts with humans [18–21]. Furthermore, carnivores  
 133 are prime targets for poachers seeking valuable body parts or trophies [22–25] and their life  
 134 histories often hinder recovery from population declines [26]. Not surprisingly, many carnivore  
 135 populations across the globe are threatened [27].

136 Carnivore species richness in Peninsular Malaysia is one of the highest in the world, with 31  
 137 species representing seven families recorded to date [28] (Table 1). Sixteen (57%) of the  
 138 remaining 28 species are listed as critically endangered, endangered, vulnerable, or near  
 139 threatened at the global level [27]. The most recent local assessment of the conservation status  
 140 of mammals lists 14 carnivore species as threatened or near threatened in Peninsular Malaysia  
 141 [28].

142 **Table 1. Carnivores of Malaysia with 2015 IUCN conservation status, and Peninsular Malaysia conservation**  
 143 **status in 2007 and 2009 based on percent change in area of occupancy and expert opinion [28].** Although 31  
 144 species are listed, three species may not be indigenous or extant. The highest threat status, based on IUCN  
 145 Red List criteria A–E [29] is reported for each species. EX = extinct, CE = critically endangered, EN =  
 146 endangered, VU = vulnerable, NT = near threatened, LC = least concern.

	Family	Species	Common name	IUCN 2015 Red List status	Peninsular Malaysia 2009 Red List status <sup>a</sup>
1	Canidae	<i>Cuon alpinus</i>	Dhole	EN	NT
2	Felidae	<i>Panthera tigris</i>	Tiger	<del>CE</del> <sup>b</sup> EN	EN
3	Felidae	<i>Panthera pardus</i>	Leopard	NT	EN
4	Felidae	<i>Neofelis nebulosa</i>	Clouded leopard	VU	NT
5	Felidae	<i>Pardofelis marmorata</i>	Marbled cat	NT	LC
6	Felidae	<i>Prionailurus bengalensis</i>	Leopard cat	LC	LC
7	Felidae	<i>Prionailurus viverrinus</i>	Fishing cat <sup>cb</sup>	EN	VU
8	Felidae	<i>Prionailurus planiceps</i>	Flat-headed cat	EN	NT
9	Felidae	<i>Catopuma temminckii</i>	Asian golden cat	NT	LC
10	Herpestidae	<i>Herpestes javanicus</i>	Javan mongoose	LC	LC
11	Herpestidae	<i>Herpestes edwardsii</i> <sup>b</sup>	Indian gray mongoose <sup>de</sup>	LC	EX
12	Herpestidae	<i>Herpestes brachyurus</i>	Short-tailed mongoose	LC	LC

Formatted: Superscript

13	Herpestidae	<i>Herpestes urva</i>	Crab-eating mongoose	LC	EN
14	Mustelidae	<i>Martes flavigula</i>	Yellow-throated marten	LC	NT
15	Mustelidae	<i>Mustela nudipes</i>	Malay weasel	LC	NT
16	Mustelidae	<i>Aonyx cinerea</i>	Asian small-clawed otter	VU	LC
17	Mustelidae	<i>Lutra sumatrana</i>	Hairy-nosed otter	EN	LC
18	Mustelidae	<i>Lutra lutra</i> <sup>c</sup>	Eurasian otter <sup>ed</sup>	NT	EN
19	Mustelidae	<i>Lutrogale perspicillata</i>	Smooth otter	VU	LC
20	Prionodontidae	<i>Prionodon linsang</i>	Banded linsang	LC	NT
21	Ursidae	<i>Helarctos malayanus</i>	Malayan sun bear	VU	VU
22	Viverridae	<i>Viverricula indica</i>	Small Indian civet	LC	NT
23	Viverridae	<i>Viverra zibetha</i>	Large Indian civet	NT	NT
24	Viverridae	<i>Viverra megaspila</i>	Large spotted civet	VU	EN
25	Viverridae	<i>Viverra zibetha</i>	Large Indian civet	NT	NT
26	Viverridae	<i>Cyanogale bennetti</i>	Otter civet	EN	EN
27	Viverridae	<i>Paguma larvata</i>	Masked palm civet	LC	LC
28	Viverridae	<i>Paradoxurus hermaphroditus</i>	Common palm civet	LC	LC
29	Viverridae	<i>Hemigalus derbyanus</i>	Banded civet	NT	LC
30	Viverridae	<i>Arctogalidia trivirgata</i>	Small-toothed palm civet	LC	LC
31	Viverridae	<i>Arctitis binturong</i>	Binturong	VU	LC

147[28]

148 [IUCN changed status of tiger from endangered to critically endangered in 2015](#)

149<sup>b</sup>Evidence for an indigenous population in Peninsular Malaysia is inconclusive [30,31].

150<sup>c</sup> Considered introduced with records only from the west coast of the peninsular; no recent records [32].

151<sup>d</sup> No proof that the species existed in Peninsular Malaysia [33], but Azlan and Sharma [34] reported a road kill in

152 Terengganu.

153 Carnivores are difficult to study by direct observation because many are nocturnal and  
154 secretive, and exist at intrinsically low population densities [35]. Early surveys in Peninsular  
155 Malaysia used traps, direct observation, signs, and road kills to infer species presence.  
156 Technological advances such as remote cameras have made it possible for recent surveys to  
157 document a greater variety of carnivore species and make inferences about their behavior,  
158 habitat use, distribution, and community composition [36–39]. All these techniques have their

Formatted: Superscript

159 limitations, but collectively can provide useful information about where a species occurred, its  
160 frequency or rarity of occurrence, and its possible vulnerability or adaptability to land use  
161 change.

162 The demand for tropical forest products or land for agriculture continues to exert enormous  
163 pressure on natural forests in Peninsular Malaysia. The conversion of tropical rainforest  
164 includes small-scale swidden agriculture, rural and urban expansion, and large-scale  
165 commercial agriculture [40,41]. A major cause of tropical forest loss has been the conversion of  
166 secondary forest to industrial plantations including oil palm and rubber [42–45]. Future changes  
167 in land use are inevitable as human populations grow and the country seeks further economic  
168 development through commerce in agriculture and timber extraction. Although Southeast Asia  
169 has few documented carnivore extinctions as a region [46], local extinctions of multiple forest-  
170 dependent species have presumably occurred. Ranges of some species will likely shrink and  
171 fragment, predisposing those remaining populations to even greater extinction risk [47]. For  
172 example, tigers (*Panthera tigris*), a valuable species to gauge the success of landscape  
173 conservation, are experiencing substantial range contraction in Peninsular Malaysia due to high  
174 rates of human-induced changes to the landscape and increased poaching pressure [48,49].  
175 However, we know little about the status and ecological requirements of the vast majority of  
176 carnivores in Peninsular Malaysia, nor where the most sensitive and diverse carnivore  
177 communities are likely to persist.

178 Here, we identify regions of high priority for carnivore conservation in Peninsular Malaysia, and  
179 associated landscape factors. Using data on carnivore species distributions from published  
180 surveys and records in combination with geographic information systems (GIS) data on  
181 landscape variables, we 1) identify priority regions for carnivore conservation and 2) determine  
182 associated environmental and anthropogenic landscape gradients.

183

## 184 **Methods**

### 185 **Study area**

186 Peninsular Malaysia (130,598 km<sup>2</sup>) is located within the Sundaland subregion of tropical East  
187 Asia, which includes Borneo, Sumatra, ~~and~~ [Java, and surrounding islands, including Bali](#) [50]. In  
188 December 2015, human population size was over 24 million with population densities  
189 (excluding Federal territories) ranging from 40 individuals/km<sup>2</sup> in Pahang to 1600/km<sup>2</sup> in Penang  
190 [51]. Malaysia's climate is typical of the tropical Sundaland subregion with abundant rainfall and  
191 warm temperatures that fluctuate little throughout the year. The principal vegetation of tropical  
192 rainforest dominated by Dipterocarps is floristically the richest of all the world's forests [46,52].  
193 The nation's economy is based on minerals, particularly oil and tin, and agricultural produce;  
194 rice and food crops are mainly for domestic consumption, but rubber, palm oil, and timber are  
195 the principal earners of foreign exchange [52]. Conversion of tropical forest to other forms of  
196 land use has been rapid in Malaysia. In a 30-year period, dryland forest declined from 64% of  
197 Peninsular Malaysia's total area to less than 50% by 1990 and swamp forests declined from 14%



198 to 8% [52]. Over a 30-year period (1975–2005), 3.6 Mha of land were converted to oil palm  
199 plantations, resulting in a 20% reduction in forest cover [53]. Rubber plantations that yield both  
200 latex and timber are rapidly expanding to replace natural forests designated for timber  
201 production under sustained yield, and 375,000 ha of monoculture timber are projected to  
202 replace natural rainforest habitat by 2020 [44].  
203

## 204 Literature search and data treatment

205 We first obtained a species list of carnivores in Peninsular Malaysia [28]. Next, we carried out a  
206 literature search for carnivores in the country using scientific and common names, and including  
207 more general search terms (mammal, vertebrate, or carnivore), for all available years up to and  
208 including 2015 and one early 2016 publication ([see S11 Materials1Appendix](#)). We used  
209 Thomson Reuter’s Web of Science to identify indexed papers, and the Malaysian Citation Centre  
210 (<http://www.myjournal.my/public/browse.php>) to search journals in all biological categories. For  
211 non-indexed Malaysian Journals without online search capability, we manually checked journal  
212 contents and excluded papers/records that were not from Peninsular Malaysia. Our final data  
213 set was derived from 85 published papers and reports (Fig 1, [S21 Materials2Appendix](#)) in the  
214 English language with carnivore records based on live captures, direct observations, signs,  
215 remote cameras, or road kills and other reported records from oldest to the most recent (1948  
216 to 2014). Where publications did not provide coordinates of species records, we used an  
217 estimate of the center of the study area for georeferencing. We recorded the date of the study,  
218 location, and principal habitat types. Some studies were conducted in multiple geographic  
219 locations; thus the number of geographic locations ( $n = 89$ ) exceeded the number of papers or  
220 reports, and some geographic locations were surveyed more than once. We mapped recent  
221 (1991–2014) and older (prior to 1991) records by species, family, and IUCN Red List category.  
222 We used 1991 as the cut-off year because most major land-use changes have occurred since  
223 then. We used Kendall’s tau-b to explore associations among the number of records (all years)  
224 per species, body size, global (IUCN) and Peninsular Malaysia threat status [27, 28], and habitat  
225 breadth (number of different habitat types where a species was recorded). We weighted threat  
226 status for each species based on an interval scale of 1 (LC; least concern), 2 (NT; near  
227 threatened), 3 (VU; vulnerable), and 4 (EN ;[or CE; endangered or critically endangered](#)  
228 [respectively](#); see Table 1). We tested the hypothesis that threat status was negatively correlated  
229 with habitat breadth. We assessed eight broad habitat types reported in the literature ([S2 Table](#)  
230 [2](#)) and used species with  $\geq 8$  records to assess associations with habitat breadth. Because  
231 riparian habitats were nested within most other habitats, they were not considered a separate  
232 habitat type for this analysis.

233

234 **Fig 1. Procedure for the selection of studies of mammalian carnivores in Peninsular Malaysia with**  
235 **records collected during 1948–2014.**

236

237 Identifying priority conservation areas

238 We used the georeferenced species data for the period 1948–2014 to identify clusters of  
239 locations (i.e. hotspots) with carnivore assemblages for which conservation priorities were high  
240 [54,55]. Many studies identified in our review were suitable for this objective because they  
241 were broad-based mammal surveys. However, we excluded 25 papers where carnivore species  
242 could not be linked with identifiable locations (a study area or geographic coordinate), or where  
243 records were duplicates from other publications. Thus, we used data from 60 papers for the  
244 hotspot analysis (Fig 1, S2 [MaterialsAppendix](#)).

245 [Our primary aim was to identify regions in the Malay peninsula that had high concentrations of](#)  
246 [species that were globally threatened. Thus, f](#)For the hotspot analysis, we weighted  
247 conservation priority for each species according to IUCN Red List status [27] based on an  
248 interval scale of 1 (LC), 2 (NT), 3 (VU), and 4 (EN) as previously described. Using this scale value  
249 as a weighting factor, we calculated the Getis-Ord  $G_i^*$  statistic in ArcGIS, which is a z-score that  
250 provides a spatial statistic of where high or low values of the weighting factor occur [54]. This  
251 approach allowed us to identify areas where species of greater (high z-scores; hotspots) or  
252 lower (low z-scores; coldspots) global conservation concern were concentrated, which helped  
253 reduce potential bias due to where surveys were conducted [56]. To calculate the z-scores, we  
254 used inverse-squared Euclidean distances to measure spatial relationships among the values of  
255 the weighting factor. This relationship allowed nearby carnivore observations to have greater  
256 influence on computations for a target location than observations further away, with the  
257 influence declining as a quadratic function of distance. The largest distance between two  
258 nearest species records was 85 km so we used that distance as a search radius to ensure that  
259 any unique survey location had at least one neighboring survey location. We used a kernel  
260 density estimator in ArcGIS, again with a search radius of 85 km, to create a continuous surface  
261 map of the z-scores.

262 Finally, we examined relationships between the z-scores and the landscape variables to gain  
263 insights into which landscape gradients may be associated with areas where carnivore species  
264 with high conservation rankings are concentrated as opposed to depleted. We examined  
265 whether the z-scores were associated with the following environmental and anthropogenic  
266 landscape gradients: elevation, natural land cover, human population density, proximity to  
267 nearest town or village, and density of primary roads ([S1 Dataset](#)). We obtained elevation (m)  
268 data from the Consortium for Spatial Information (<http://srtm.csi.cgiar.org/>). We reclassified  
269 land-cover data from the Global Land Cover Database  
270 (<http://forobs.jrc.ec.europa.eu/products/glc2000/legend.php>) into a binary layer to represent  
271 all natural land cover types, excluding urban, cultivated, and managed areas. We then used a  
272 neighborhood analysis to calculate the proportion of natural land cover within a radius of 15  
273 km. We chose 15 km to reflect the large scale of our analysis and to ensure that values covered  
274 the full range of very low up to 100% natural land cover. We obtained human population data  
275 (counts per 30-arc grid cell, or approximate density/km<sup>2</sup>) from a Global Population Distribution  
276 database (<http://www.ciesin.org/>). We calculated proximity to the geographic center of the  
277 nearest town or village digitized from Google Maps. Finally, using the line density function in

278 ArcGIS, we calculated density of improved roads (km/km<sup>2</sup>; digitized from Google Maps) based  
279 on a moving window with a 15-km radius. Land cover and human population data were from  
280 2000, which was the approximate mid-point of the period during which most carnivore  
281 observations were recorded. In addition to these environmental and anthropogenic variables,  
282 we considered a variable that may have affected the sampling distribution, namely proximity to  
283 the capital, Kuala Lumpur. Because of logistical considerations, many early surveys were  
284 conducted in relatively close proximity (~100 km) to the capital (we used the GPS coordinates  
285 of the headquarters of the Department of Wildlife and National Parks as our reference point).  
286 This area has relatively high densities of improved roads, therefore we added an interaction  
287 effect between road density and proximity to headquarters to every model to account for  
288 potential sampling bias. Given the large spatial scale of our assessment, we set the resolution of  
289 all data layers to 30-arc seconds for Peninsular Malaysia.

290 To explore potential relationships between the Getis-Ord  $G_i^*$  z-scores and landscape variables,  
291 we used ordinary least squares linear regression in ArcGIS to examine a set of models with  
292 different combinations of the environmental and anthropogenic variables to assess their  
293 relative influence. We used proximity to Kuala Lumpur, improved road density and their  
294 interaction as the basis for model building, to account for spatial sampling biases and reduce  
295 spatial autocorrelation [57]. We used the bias-corrected Akaike's information criterion ( $AIC_c$ ) for  
296 model selection and considered models within  $2 \Delta AIC_c$  values to be parsimonious [58]. To  
297 reduce skewness in the data, we log-transformed human population and proximity to Kuala  
298 Lumpur and square-root transformed improved road density. We tested for normal distribution  
299 of residuals using the Jarque-Bera statistic. We used Koenker's studentized Bruesch-Pagan  
300 statistic to determine if explanatory variables had a consistent relationship with Getis-Ord  $G_i^*$   
301 z-scores in geographic space and data space; if this test was significant, we calculated robust  
302 standard errors,  $t$ -values, and probabilities for beta values. Finally, we tested whether model  
303 residuals showed spatial autocorrelation based on Moran's  $I$  statistic.

304

## 305 Results

### 306 Records of distribution and habitat

307 Observation records spanned the period 1948–2014 with 96% collected during the last 50 years  
308 and 50% collected after 1991 (Fig 2, S2 Table). We mapped all survey locations by family and  
309 species (S1–S4 Figs) and by threat category (S5–S7 Figs). Recent survey records (i.e., since 1991)  
310 in largely primary rainforest in northern Perak revealed high carnivore species richness. In  
311 Selangor, 75% of carnivore records preceded 1991, thus fewer surveys may have influenced the  
312 relative paucity of recent versus older carnivore records (S1–S6 Figs). Records were few (<5) for  
313 10 species, almost all of which were small to medium-sized carnivores (Fig 3) and there were no  
314 recent records of the endangered otter civet (*Cyanogale bennettii*). The number of records  
315 tended to be greater with species' body size (Kendall's tau-b = 0.24,  $z = 1.78$ ,  $P = 0.038$ ), but not  
316 with IUCN global or Peninsular Malaysia conservation scores.

Formatted: No underline

317 **Fig 2. Distribution of surveys ( $n = 133$ ) of carnivores among different states in Peninsular Malaysia**  
318 **with records collected during 1948–2014.** Data were based on 60 published papers and reports that  
319 used conventional trapping, direct observation, signs, remote cameras, or road kills. Some publications  
320 compiled data from several surveys and some geographic locations were surveyed more than once.  
321 [Boundary layers: Esri, Garmin International \(formerly DeLorme Publishing Company, Inc.\), Inc.;](#) [Inset](#)  
322 [map: U.S. Central Intelligence Agency \(The World Factbook\).](#)

323  
324 **Fig 3. Number of records of Carnivora species in Peninsular Malaysia.** Data were obtained from surveys  
325 that used conventional trapping, direct observation, sign, remote cameras, or road kills collected during  
326 1991–2014. Species are grouped by family and ranked by number of records.

327 Surveys (or specimens collected) in forest reserves, wildlife reserves and national parks  
328 consisting mostly of dry-land forest comprised 75% of the reports. The remaining reports were  
329 from rice fields (12%), peat swamp/mangrove forest (6%), oil palm plantations (3%), mangrove  
330 forests (2%) and human inhabited areas (2%). We used carnivore species presence data from 89  
331 geographic locations to examine habitat types associated with species records (S2 Table).  
332 Habitat breadth was associated with the number of records per species (Kendall's tau-b = 0.554,  
333  $z = 3.03$ ,  $P = 0.001$ ), but not with species' IUCN global or Peninsular Malaysia conservation  
334 scores, or with body size.

### 335 Priority conservation areas

336 A region in the northeastern portion of the peninsula had the greatest concentration of  
337 carnivores with high conservation status, within which two areas were particularly prominent:  
338 the forest complex associated with Royal Belum State Park in the northern portion of this region  
339 and, southeast of it, an area associated with Taman Negara National Park (Fig 4). Notably, we  
340 also identified a concentrated area with carnivore observations and diversity associated with  
341 the southern half of Selangor and the adjacent region in Pahang, including Krau Wildlife  
342 Reserve, but the presence of carnivores with high conservation status was much lower  
343 compared with other areas.

344  
345 **Fig 4. Locations of mammalian carnivore surveys and kernel density surface of Getis-Ord  $G_i^*$  z-scores**  
346 **of weighted ranking of IUCN red list categories for recorded species locations in Peninsular Malaysia,**  
347 **1948–2014.** [Hillshade layer derived from Shuttle Radar Topography Mission \(STRM\) 90-m Digital](#)  
348 [Elevation Data from Consultative Group on International Agricultural Research \(CGIAR\) and reprinted](#)  
349 [under a CC BY license, with permission from International Center for Tropical Agriculture \(CIAT\), original](#)  
350 [copyright 2004.](#) Protected areas mentioned in the text are labeled; [reprinted from World Database on](#)  
351 [Protected Areas \(<http://www.protectedplanet.net>\) under a CC BY license, with permission from the](#)  
352 [United Nations Environmental Programme-World Conservation Monitoring Centre, original copyright](#)  
353 [2010.](#)

354 Model selection of ordinary least squares regressions showed the best-fitting model included  
355 elevation, proportion of natural land cover, improved road density, proximity to Kuala Lumpur,  
356 and the interaction between the latter 2 variables (adjusted  $R^2 = 0.62$ ; S1 Table). The second-

357 best model was within 2  $\Delta AIC_c$  values and contained human population density as an additional  
358 variable. However, the 95% confidence interval of that variable overlapped zero so we focused  
359 our interpretation on the top model. The Jarque-Bera (*JB*) statistic indicated the residuals of the  
360 model did not deviate from normality (*JB* = 0.459, 2 df, *P* = 0.797). Getis-Ord  $G_i^*$  z-scores were  
361 negatively associated with elevation ( $\beta$  = -0.00124, SE = 0.00057, *t* = -2.169, *P* = 0.042) and  
362 positively associated with proportion of natural land cover ( $\beta$  = 1.899, SE = 0.711, *t* = 2.671, *P* =  
363 0.009) and distance to Kuala Lumpur ( $\beta$  = 1.811, SE = 0.339, *t* = 5.344, *P* < 0.001). Thus, areas  
364 where observations of species with higher conservation ranks were spatially clustered generally  
365 coincided with areas at lower elevations, with greater proportion of natural land cover, and  
366 tended to be more distant from Kuala Lumpur. Human population density and proximity of the  
367 nearest town or village did not show an association with the Getis-Ord  $G_i^*$  z-scores. There was  
368 some evidence of spatial autocorrelation among the residuals (Moran's *I* = 0.461, *z* = 2.018, *P* =  
369 0.044).

370

## 371 Discussion

372 Peninsular Malaysia contains possibly the greatest number of native species of Carnivora within  
373 Sundaland, and more than half are globally threatened or near threatened. Using data compiled  
374 from the first comprehensive review of publications with carnivore records, we identified two  
375 regions that overlapped with protected areas, Taman Negara National Park and Royal Belum  
376 State Park, as hotspots for carnivore species of greatest conservation concern.

377 Both these protected areas are considered priority areas for tiger conservation in Malaysia [59].  
378 Established in 1938, Taman Negara (4343 km<sup>2</sup>) is Malaysia's oldest national park [60] and  
379 comprises portions of ~~three the S~~states of Pahang, Terengganu, and Kelantan. It contains  
380 Malaysia's largest continuous tract of primary forest, of which nearly 60% consists of low  
381 elevation (75–300 m) rainforest. Royal Belum State Park, however, was gazetted ~~in relatively~~  
382 ~~recently-2007~~ [61] and is part of the Belum-Temengor Forest Complex (3546 km<sup>2</sup>) located in  
383 northern Perak; it shares its northern boundary with Thailand, where it connects with two  
384 protected areas, Hala Bala Wildlife Sanctuary and Bang Lang National Park. The combined  
385 extent of protected areas and forest reserves in this forest complex, which consists of lowland  
386 and hill dipterocarp forests from 130 to 1500 m ~~[61]~~, is said to rival that of Taman Negara [62].  
387 The number of carnivore species reported in Taman Negara and Belum-Temengor were 19 and  
388 22, respectively, each with eight threatened and five near-threatened species.

389 A crucial finding was the relative scarcity of reports of carnivores of conservation concern in the  
390 southwestern region of the Peninsular encompassing the state of Selangor and the adjacent  
391 region in the state of Pahang, despite frequent surveys in that area. The surveys were  
392 conducted within a 50- to 60-km radius of Kuala Lumpur, where several small forest reserves  
393 and areas (12–200 ha) of secondary forest have existed within the city limits for decades, with  
394 more extensive lowland and hill dipterocarp forests in peri-urban areas [63]. Surveys in this

395 region occurred over a long time span, with over half the records collected prior to 1991. The  
396 distinct paucity of records of carnivores of conservation concern suggests that many of these  
397 species cannot persist in small fragmented habitats, or even in larger extents of habitats close to  
398 urbanization. Krau Wildlife Reserve (603 km<sup>2</sup>), situated within a 1556-km<sup>2</sup> forested area [64],  
399 was the largest protected area in this coldspot. Krau Wildlife Reserve is surrounded by  
400 agriculture and settlements, but its northeastern boundary is < 50 km south of the large  
401 forested landscape of Taman Negara. It was thus considered a secondary priority site for tiger  
402 conservation in Malaysia [65]. Carnivore species richness in Krau ( $n = 20$ ) was similar to that  
403 reported at Belum-Temengor and Taman Negara, although with fewer threatened ( $n = 5$ ) and  
404 near threatened species ( $n = 4$ ).

405 Carnivore hotspots were associated with large extents of natural land cover, lower elevations,  
406 and greater distances from Kuala Lumpur, within the state of Selangor. Selangor (800,000 ha),  
407 the most populous state in Malaysia with 5.8 million people [66] has the highest per capita GDP,  
408 and has experienced the most rapid growth in the manufacturing sector in the last five decades.  
409 Urban and agricultural development has been responsible for most of the state's change in land  
410 use with the expansion of oil palm plantations at the expense of peat swamp forest [42, 67].  
411 Considering that 75% of the surveys in Selangor were conducted before 1991, [and our human](#)  
412 [population and land use data were derived more recently](#), the status of carnivore populations in  
413 this state may be more critical than the data suggest.

414 A common consequence of urbanization and development is habitat fragmentation and the  
415 extirpation of large apex predators. Laidlaw's [68] survey of seven sites (70 to >10,000 ha) in  
416 Peninsular Malaysia suggested that large tracts of natural forest were the most important  
417 predictor of mammal species richness and large carnivore presence. Woodroffe [16]  
418 demonstrated a strong positive relationship between reserve size and the persistence of large  
419 carnivores and concluded that smaller habitat patches increased the potential for human-  
420 carnivore conflicts with subsequent extirpation of local carnivore populations. Many small and  
421 mid-sized carnivores also rely on larger habitat patches suggesting that factors other than body  
422 size, such as resource specialization, behavior, and social structure, may play an important role  
423 in this dependency [16, 69]. Smaller habitat patches could mean the loss of suitable habitat,  
424 new barriers to movement, or competition with species better adapted to disturbed  
425 environments [70]. Proximity to urbanization and primary roads, even where habitat is  
426 sufficiently large, limits dispersal and enhances the risk of road mortality and illegal hunting  
427 [71–73].

428 Low-elevation habitats with natural forest cover may be one of the most valuable habitats for  
429 carnivores in tropical regions. We found that all but two species of Carnivora were reported in  
430 lowland forests (S2 Table). Notably, the number of species of Carnivora reported in lowland  
431 swamp forests ( $n = 17$ ) was high, considering the relatively few surveys ( $n = 16$ ). In Southeast  
432 Asia, lowland equatorial forests support the vast majority of species [46] and in Peninsular  
433 Malaysia, lowland forests support almost 90% of mammal species with 61% occurring only in

434 lowland and hill forests below 1000 m [74]. Malaysia has lost nearly 40% of its original forest  
435 cover [75] and recent annual deforestation rates in the peninsula (0.9% annually from 2000 to  
436 2010 [46]) show little sign of abatement.

437 With the exception of the otter civet (one record in 1987), records since 1991 exist for the  
438 remaining 27 species in the peninsula. Records were few for nine species, mostly small  
439 carnivores, including four species of Viverridae and all three species of Herpestidae native to  
440 the Malay Peninsula. We found only one record of the Javan mongoose (*Herpestes javanicus*)  
441 and one of the small Indian civet (*Viverricula indica*) since 1991; these species are neither  
442 globally threatened nor near threatened. Conversely, records were greater for larger species  
443 such as the tiger, sun bear (*Helarctos malayanus*), and leopard (*Panthera pardus*). In an  
444 extensive review of carnivore research effort, [76] reported a strong association between body  
445 size and research effort in the Carnivora, with the Viverridae and Herpestidae among the four  
446 least studied of the carnivore families. Larger species leave more definitive signs and range over  
447 larger areas, thus increasing the probability of detection. Also, the rarely recorded Javan  
448 mongoose and small Indian civet favor open, less forested habitats (77,78); apart from rice  
449 fields, these habitats are rare in Peninsular Malaysia. The dearth of ecological studies on  
450 smaller carnivores in peninsular Malaysia may predispose them to early extinction, when  
451 efforts for their conservation are less costly than for large-bodied species, and more likely to  
452 succeed [79].

453 Large body size confers greater vagility and thus the ability to use a wide array of habitats but  
454 we found no association between habitat breadth and body size. Also, species that use a wide  
455 range of habitats may be more tolerant of habitat loss and fragmentation [80]. Although there  
456 may be some sampling bias given that species with more records were reported in more  
457 habitats, habitat breadth was not associated with global (IUCN) or local (Peninsular Malaysia)  
458 threat status. To illustrate, three small carnivores, the common palm civet (*Paradoxurus*  
459 *hermaphroditus*; least concern), the leopard cat (*Prionailurus bengalensis*; least concern), and  
460 the flat-headed cat (*Prionailurus planiceps*; endangered) were reported in a wide variety of  
461 habitats (S2 Table), including small forest patches in urban landscapes. The flat-headed cat is  
462 adapted for feeding on aquatic prey, thus the presence of wetland habitat, which is abundant in  
463 Peninsular Malaysia, may be more important for its persistence than forest cover. Locally, the  
464 flat-headed cat is considered near threatened [28], in contrast with its global endangered status  
465 [27], which may reflect its ability to persist in a variety of habitats associated with freshwater.

466 We acknowledge several caveats in our study. Despite our attempt to obtain as complete a set  
467 of published studies for our analysis as possible, at least three papers with nine additional  
468 records of leopard [81,82] and one record of a flat-headed cat [83] escaped our attention. ~~Of~~  
469 ~~However, if we included these 10 records, including~~ 29 recently released records of threatened  
470 and near threatened carnivores [84–87], ~~82% 82% of records~~ occurred within the hotspots  
471 identified in our analysis, confirming the importance of these regions for carnivore  
472 conservation. We caution, however, that despite demonstrating distinct landscape associations

473 with the distribution of carnivores as weighted by their conservation rankings, we could not  
474 fully account for spatial autocorrelation and our data were not derived from standardized,  
475 probabilistic, or systematic coverage of the entire peninsula. Thus, our inference is weaker in  
476 areas with fewer surveys and published records. For example, the data included few surveys for  
477 the southern region of the peninsula, including the Endau Rompin Forest Complex (~2389 km<sup>2</sup>).  
478 ~~This area with comprises~~ substantial low-elevation rainforest ~~that likely supports with the~~  
479 ~~potential to support~~ a diversity of indigenous carnivores ~~despite its~~. ~~Endau Rompin was~~  
480 ~~considered low priority for tiger conservation owing to~~ highly fragmented surroundings areas  
481 and poor connectivity with large, forested landscapes [65]. ~~Nevertheless,~~ A recent remote  
482 camera survey reported the presence of six felid species, including tigers [88].  
483

## 484 Conclusion

485 Peninsular Malaysia supports several species of globally threatened carnivores and our study  
486 underscores the importance of natural forest cover for their persistence. We show that  
487 carnivores of greatest conservation concern are less likely to persist in small, fragmented  
488 habitats or habitats close to urban areas. Recent (2000–2012) changes in global forest cover  
489 indicate that Malaysia lost 14% of its forest cover, a rate of loss that exceeded any other  
490 country [89]. Oil palm and industrial timber plantations replaced most of the lost forest [90]  
491 and trends point to their continued expansion. Surveys and targeted ecological studies of  
492 carnivores in habitat types other than primary and secondary forests will thus be important to  
493 elucidate their status and capacity to persist in the face of progressive habitat alteration.  
494 Recent studies in oil palm estates and commercial forest plantations suggest that these altered  
495 habitats may serve as ecological corridors and shelter valuable elements of biodiversity [91–93],  
496 but primarily when interspersed with large (>1000 ha) stands of natural, secondary forest [94].  
497 Ultimately, reducing poaching and habitat loss within large, contiguous stands of rainforest will  
498 be crucial for the persistence of Malaysia’s most threatened carnivores and consequently the  
499 broader ecological communities that carnivores influence.

500

## 501 Supporting information

502 [S1–7 Figures. Recent \(1991–2014\) and older \(1948–1990\) records of carnivores by family and](#)  
503 [IUCN threat status in Peninsular Malaysia.](#)

504 [\(.docx\)](#)

505 [S1 Table. Model selection results to identify landscape variables associated with spatial](#)  
506 [clustering of carnivore records based on weighted ranking of IUCN red list categories.](#)

507 [\(PDF\)](#)

508 [S2 Table. Carnivora species reported in Peninsular Malaysia and associated habitats, 1948–](#)  
509 [2014.](#)

510 [\(PDF\)](#)



511 [S1 Appendix. Search terms and sources for carnivore records and habitats in Peninsular](#)  
512 [Malaysia.](#)  
513 [\(PDF\)](#)  
514 [S2 Appendix. Records of Carnivora by species, locations, and year.](#)  
515 [\(.xlsx\)](#)  
516 [S1 Dataset. Geo-referenced TIFF files for spatial data layers used in landscape analysis.](#)  
517

## 518 Acknowledgments

519 We thank Anusha Krishnan at the Sunway University Library for assistance with obtaining  
520 articles. Lim Boo Liat provided a complete list of his publications on Malaysian vertebrates, and  
521 valuable opinions and insights on the occurrence of the fishing cat, European otter, and gray  
522 mongoose in Malaysia. We thank the director of the Biodiversity Institute and Suzilawati Binti  
523 Ramzan for access to the Bukit Ringitt Museum. [We thank Joseph D. Clark for review](#)  
524 [comments provided as part of the U.S. Geological Survey's Fundamental Science Practices. Any](#)  
525 [use of trade, firm, or product names is for descriptive purposes only and does not imply](#)  
526 [endorsement by the U.S. Government.](#)  
527

## 528 Literature cited

- 529 1. Gittleman JL, Funk SM, Macdonald DW, Wayne RK. Why 'carnivore conservation'? In: Gittleman JL,  
530 Funk SM, Macdonald DW, Wayne RK, editors. Carnivore conservation. Cambridge, UK: Cambridge  
531 University Press; 2001.
- 532 2. Ripple WJ, Estes JA, Beschta RL, Wilmers CC, Ritchie EG, Hebblewhite M, Berger J, Elmhagen B,  
533 Letnic M, Nelson MP, Schmitz OJ. Status and ecological effects of the world's largest carnivores.  
534 Science. 2014;343(6167): 1241484.
- 535 3. Terborgh J, Lopez L, Nuñez P, Rao M, Shahabuddin G, Orihuela G, Riveros M, Ascanio R, Adler GH,  
536 Lambert TD, Balbas L. Ecological meltdown in predator-free forest fragments. Science. 2001;294:  
537 1923–1926.
- 538 4. Sergio F, Caro T, Brown D, Clucas B, Hunter J, Ketchum J, McHugh K, Hiraldo F. Top predators as  
539 conservation tools: ecological rationale, assumptions, and efficacy. Annu Rev Ecol Evol Syst.  
540 2008;39: 1–19.
- 541 5. Prugh LR, Stoner CJ, Epps CW, Bean WT, Ripple WJ, Laliberte AS, Brashares JS. The rise of the  
542 mesopredator. BioScience. 2009;59: 779–791.
- 543 6. Estes JA, Terborgh J, Brashares JS, Power ME, Berger J, Bond WJ, et al. Trophic downgrading of  
544 planet Earth. Science. 2011;333: 301–306.
- 545 7. Redford KH. The empty forest. BioScience. 1992;46: 412–422.
- 546 8. Duffy JE. Biodiversity loss, trophic skew and ecosystem functioning. Ecol Lett. 2003;6: 680–687.
- 547 9. Simberloff D. Flagships, umbrellas, and keystones: Is single species management passé in the  
548 landscape era? Biol Conserv. 1998;83: 247–257.
- 549 10. Noss RF. Indicators for monitoring biodiversity: a hierarchical approach. Conserv Biol. 1990: 355-  
550 364.
- 551 11. Ratnayeke S, van Manen FT. Assessing sloth bears as surrogates for carnivore conservation in Sri  
552 Lanka. Ursus. 2012;23: 206-217.

- 553 12. Di Minin E, Slotow R, Hunter LT, Pouzols FM, Toivonen T, Verburg PH, Leader-Williams N, Petracca L,  
554 Moilanen A. Global priorities for national carnivore conservation under land use change. *Sci Rep.*  
555 2016; 6: 23814
- 556 13. Berger J, Stacey PB, Bellis L, Johnson MP. A mammalian predator–prey imbalance: grizzly bear and  
557 wolf extinction affect avian neotropical migrants. *Ecol Appl.* 2001;11: 947–960.
- 558 14. Caro T, Engilis A, Fitzherbert E, Gardner T. Preliminary assessment of the flagship species concept at  
559 a small scale. *Anim Conserv.* 2004;7: 63–70.
- 560 15. Crook KR. Relative sensitivities of mammalian carnivores to habitat fragmentation. *Conserv Biol.*  
561 2002;16: 488–502.
- 562 16. Woodroffe R. Strategies for carnivore conservation: lessons from contemporary extinctions. In:  
563 Gittleman JL, Funk SM, Macdonald DW, Wayne RK, editors. *Carnivore conservation.* Cambridge  
564 University Press; 2001. pp. 61–92.
- 565 17. Cardillo M, Purvis A, Sechrest W, Gittleman JL, Bielby J, Mace GM. Human population density and  
566 extinction risk in the world’s carnivores. *PLoS Biol.* 2004;2:e197.
- 567 18. Saberwal VK, Gibbs JP, Chellam R, Johnsingh A. Lion-human conflict in the Gir Forest, India. *Conserv*  
568 *Biol.* 1994;8: 501–507.
- 569 19. Treves A, Karanth KU. 2003. Human-carnivore conflict and perspectives on carnivore management  
570 worldwide. *Conserv Biol.* 2003;17: 1491–1499.
- 571 20. Ikanda D, Packer C. Ritual vs. retaliatory killing of African lions in the Ngorongoro Conservation Area,  
572 Tanzania. *Endanger Species Res.* 2008;6: 67–74.
- 573 21. Inskip C, Zimmermann A. 2009. Human-felid conflict: a review of patterns and priorities worldwide.  
574 *Oryx.* 43:18–34.
- 575 22. Ellis R. 2005. *Tiger bone & rhino horn: the destruction of wildlife for traditional Chinese medicine.*  
576 Island Press; 2005.
- 577 23. Chapron G, Miquelle DG, Lambert A, Goodrich JM, Legendre S, Clobert J. The impact on tigers of  
578 poaching versus prey depletion. *J Appl Ecol.* 2008;45: 1667–1674.
- 579 24. Shepherd C, Shepherd L. The poaching and trade of Malayan sun bears in Peninsular Malaysia.  
580 *Traffic Bulletin.* 2010;23: 49–52.
- 581 25. Liberg O, Chapron G, Wabakken P, Pedersen HC, Hobbs NT, Sand H. Shoot, shovel and shut up:  
582 cryptic poaching slows restoration of a large carnivore in Europe. *Proc R Soc Lond [Biol].* 2012; 279:  
583 910–915.
- 584 26. Purvis A, Gittleman JL, Cowlishaw G, Mace GM. Predicting extinction risk in declining species. *Proc.*  
585 *R. Soc. London.* 2000;B267: 1947–1952.
- 586 27. International Union for Conservation of Nature [IUCN]. The IUCN Red List of Threatened Species.  
587 Version 2015. Available from: <<http://www.iucnredlist.org>>. Cited 15 Oct 2015.
- 588 ~~27.~~
- 589 29. International Union for Conservation of Nature. 2001. 2001 IUCN Red List Categories and Criteria  
590 version 3.1. International Union for Conservation of Nature and Natural Resources, IUCN, Gland,  
591 Switzerland.
- 592 30. Van Bree P, Khan M, Khan M. On a fishing cat, *Felis (Prionailurus) viverrina* Bennett, 1833, from  
593 continental Malaysia. *Z Saugetierkd.* 1992;57: 179–180.
- 594 31. Kawanishi K, Sunquist M. Possible new records of fishing cat from Peninsular Malaysia. *Cat News.*  
595 2003;39: 3–5.
- 596 32. Medway, L. 1969. *The wild mammals of Malaya and offshore islands including Singapore.* Oxford  
597 University Press. 1969. Pp.1-127.
- 598 33. Sivasothi N, Burhanuddin HMN. 1994. A review of otters (Carnivora: Mustelidae: Lutrinae) in  
599 Malaysia and Singapore. *Hydrobiologia.* 1994;285: 151–170.
- 600 34. Azlan JM, Sharma DSK. Mammal diversity and conservation in a secondary forest in Peninsular  
601 Malaysia. *Biodivers Conserv.* 2006;15: 1013–1025.
- 602 35. Boitani L, Ciucci P, Mortelliti A. Designing carnivore surveys. In: Boitani L, Powell RA, editors.  
603 *Carnivore Ecology and Conservation.* Oxford University Press; 2012. Pp. 8–30.

- 604 36. Hedges L, Clements GR, Aziz S, Yap W, Laurance S, Goosem M, Laurance W. 2013. Small carnivore  
605 records from a threatened habitat linkage in Terengganu, Peninsular Malaysia. *Small Carniv*  
606 *Conserv.* 49:9–14.
- 607 37. Hedges L, Lam WY, Campos-Arceiz A, Rayan DM, Laurance WF, Latham CJ, Saaban S, Clements GR.  
608 Melanistic leopards reveal their spots: Infrared camera traps provide a population density estimate  
609 of leopards in Malaysia. *J Wildl Manag.* 2015;79: 846–853.
- 610 38. Rayan DM. Tiger Monitoring Study in Gunung Basor Forest Reserve, Jeli, Kelantan. Unpublished  
611 Report. WWF-Malaysia, Petaling Jaya, Malaysia. 2007.
- 612 39. Sasidhran S, Adila N, Hamdan MS, Samantha LD, Aziz N, Kamarudin N, Puan CL, Turner E, Azhar B.  
613 Habitat occupancy patterns and activity rate of native mammals in tropical fragmented peat swamp  
614 reserves in Peninsular Malaysia. *Forest Ecol Manag.* 2016;363:140–148.
- 615 40. Phua MH, Tsuyuki S, Furuya N, Lee JS. Detecting deforestation with a spectral change detection  
616 approach using multitemporal Landsat data: A case study of Kinabalu Park, Sabah, Malaysia. *J*  
617 *Environ Manage.* 2008;88: 784–795.
- 618 41. Koh LP, Kettle CJ, Sheil D, Lee TM, Giam X, Gibson L, Clements GR. Biodiversity state and trends in  
619 Southeast Asia. *Encyclopedia of biodiversity.* 2013;1: 509–527.
- 620 42. Abdullah SA, Nakagoshi N. Changes in agricultural landscape pattern and its spatial relationship with  
621 forestland in the State of Selangor, peninsular Malaysia. *Landsc Urban Plan.* 2008; 87:147–155.
- 622 43. Koh LP, Wilcove DS. Is oil palm agriculture really destroying tropical biodiversity? *Conserv Lett.*  
623 2008;1 :60–64.
- 624 44. Aziz SA, Laurance WF, Clements R. Forests reserved for rubber? *Front Ecol Environ.* 2010;8: 178–  
625 178.
- 626 45. Shevade VS, Potapov PV, Harris NL, Loboda TV. 2017. Expansion of industrial plantations continues  
627 to threaten Malayan tiger habitat. *Remote Sense.* 2017; 9: 747; doi:10.3390/rs9070747.
- 628 46. Corlett RT. *The ecology of tropical East Asia.* Oxford University Press. 2014.
- 629 47. Frankham R, Briscoe DA, Ballou JD. *Introduction to conservation genetics.* Cambridge University  
630 Press; 2002.
- 631 48. Clements RD, Rayan M, Zafir AWA, Venkataraman A, Alfred R, Payne J, Ambu L, Sharma DSK. Trio  
632 under threat: can we secure the future of rhinos, elephants and tigers in Malaysia? *Biodivers*  
633 *Conserv.* 2010;19: 1115–1136.
- 634 49. Clements GR, Lynam, AJ, Gaveau D, Yap, WL, Lhota, S, Goosem M, Laurance S & Laurance WF.  
635 Where and how are roads endangering mammals in Southeast Asia's forests? *PLoS One.* 2014;9:  
636 e115376.
- 637 50. Hock SS. 2007. *The Population of Peninsular Malaysia.* Institute of Southeast Asian Studies;2007.
- 638 51. Department of Statistics Malaysia. *Malaysia population by state and ethnic group.* Federal  
639 Government Administrative Centre, Putrajaya, Malaysia.2015a. Available from:  
640 [https://web.archive.org/web/20160212125740/http://pmr.penerangan.gov.my/index.php/info-](https://web.archive.org/web/20160212125740/http://pmr.penerangan.gov.my/index.php/info-terkini/19463-unjuran-populasi-penduduk-2015.html)  
641 [terkini/19463-unjuran-populasi-penduduk-2015.html](https://web.archive.org/web/20160212125740/http://pmr.penerangan.gov.my/index.php/info-terkini/19463-unjuran-populasi-penduduk-2015.html).
- 642 52. Brookfield H, Byron Y. Deforestation and timber extraction in Borneo and the Malay Peninsula: The  
643 record since 1965. *Glob Environ Change.* 1990;1: 42–56.
- 644 53. Wicke B, Sikkema R, Dornburg V, Faaij A. Exploring land use changes and the role of palm oil  
645 production in Indonesia and Malaysia. *Land Use Policy.* 2011;28: 193–206.
- 646 54. Getis A, Ord JK. The analysis of spatial association by use of distance statistics. *Geogr Anal.* 1992;24:  
647 189–206.
- 648 55. Scott LM, Janikas MV. Spatial statistics in ArcGIS. In: *Handbook of applied spatial analysis.* 2010. pp.  
649 27–41.
- 650 56. Shekhar S, Evans MR, Kang JM, Mohan P. 2011. Identifying patterns in spatial information: a survey  
651 of methods. *WIREs Data Mining and Knowledge Discovery.* 2011; 1: 193–214.
- 652 57. Thayne JB, Simanis JM. Accounting for Spatial Autocorrelation in Linear Regression Models Using  
653 Spatial Filtering with Eigenvectors. *Ann Assoc Am Geogr* 2012. DOI:10.1080/00045608.2012.685048,  
654 <http://dx.doi.org/10.1080/00045608.2012.685048>.

- 655 58. Burnham KP, Anderson DR. Model selection and multimodel inference: a practical information-  
656 theoretic approach. 2nd edition. New York, USA: Springer-Verlag; 2002.
- 657 59. Department of Wildlife and National Parks (DWNP), Malaysia. 2008. National tiger action plan for  
658 Malaysia.
- 659 60. Sen YH. A special issue to commemorate the golden jubilee of Taman Negara. *J Wildl Parks*,  
660 Malaysia. 1990;X: 1–152.
- 661 61. Lim KC. Belum-Temengor Forest Complex, north Peninsular Malaysia. *BirdingASIA*. 2010;14: 15–22.
- 662 62. Schwabe KA, Carson RT, DeShazo JR, Potts MD, Reese AN, Vincent JR. Creation of Malaysia's Royal  
663 Belum State Park: a case study of conservation in a developing country. *J. Environ. Dev.* 2015; 24:54–  
664 81.
- 665 63. Webb R. Urban forestry in Kuala Lumpur, Malaysia. *Arboric J.* 1998;22: 287–296, DOI:  
666 10.1080/03071375.1998.9747211.
- 667 64. Yusof E, Sorenson KW. Krau Wildlife Reserve: protected area management experiences. *J Wildl*  
668 *Parks*. 2000;18: 3–13. Available from:  
669 <http://www.wildlife.gov.my/images/document/penerbitan/jurnal/Jil182000.pdf>
- 670 65. Lynam AJ, Laidlaw R, Wan Noordin WS, Elagupillay S, Bennett EL. Assessing the conservation status  
671 of the tiger *Panthera tigris* at priority sites in Peninsular Malaysia. *Oryx*. 2007;41: 454–462.
- 672 66. Department of Statistics Malaysia. Malaysia @ a Glance. Federal Government Administrative Centre,  
673 Putrajaya, Malaysia. 2015b. Available from:  
674 [https://www.statistics.gov.my/index.php?r=column/cone&menu\\_id=ZmVrN2FoYnBvZE05T1AzK0RLcEtiZz09](https://www.statistics.gov.my/index.php?r=column/cone&menu_id=ZmVrN2FoYnBvZE05T1AzK0RLcEtiZz09)  
675
- 676 67. Abdullah SA, Nakagoshi N. Changes in landscape spatial pattern in the highly developing State of  
677 Selangor, peninsular Malaysia. *Landscape Urban Plan.* 2006;77: 263–275.
- 678 68. Laidlaw RK. Effects of habitat disturbance and protected areas on mammals of peninsular Malaysia.  
679 *Conserv Biol.* 2000;14: 1639–1648.
- 680 69. Kiviat E, MacDonald K. 2002. Biodiversity patterns and conservation in the Hackensack  
681 Meadowlands, New Jersey. *Urban Habitats*. 2002;2: 28–61.
- 682 70. Shochat E, Lerman SB, Anderies JM, Warren PS, Faeth SH, Nilon CH. Invasion, competition, and  
683 biodiversity loss in urban ecosystems. *BioScience*. 2010;60: 199–208.
- 684 71. Fahrig L, Rytwinski T. Effects of roads on animal abundance: an empirical review and synthesis. *Ecol*  
685 *Soc.* 2009; 14(1): 21.
- 686 72. Van Langevelde F, van Dooremalen C, Jaarsma CF. Traffic mortality and the role of minor roads. *J*  
687 *Environ Manage.* 2009;90: 660–667.
- 688 73. Haines AM, Elledge D, Wilsing LK, Grabe M, Barske MD, Burke N, Webb SL. Spatially explicit analysis  
689 of poaching activity as a conservation management tool. *Wildl Soc Bull.* 2012;36: 685–692.
- 690 74. Lim BL. Critical habitats for the survival of Malayan mammals in Peninsular Malaysia. *J Sci Technol*  
691 *Tropics*. 2008;4: 27–37.
- 692 75. Laurance WF. Forest destruction in tropical Asia. *Curr Sci.* 2007;93: 1544–1550.
- 693 76. Brooke ZM, Bielby J, Nambiar K, Carbone C. Correlates of research effort in carnivores: body size,  
694 range size and diet matter. *PLoS One*. 2014;9:e93195.
- 695 77. Wells D. Notes on the distribution and taxonomy of Peninsular Malaysian mongooses (*Herpestes*).  
696 *Nat Hist Bull Siam Soc.* 1989;37: 87–97.
- 697 78. Choudhury A, Duckworth JW, Timmins R, Chutipong W, Willcox DHA, Rahman H, Ghimirey Y,  
698 Mudappa D. *Viverricula indica*. The IUCN Red List of Threatened Species. 2015: Available from  
699 <http://dx.doi.org/10.2305/IUCN.UK.2015-4.RLTS.T41710A45220632.en>. Downloaded on 22 May  
700 2017.
- 701 79. Fisher DO. Cost, effort and outcome of mammal rediscovery: neglect of small species. *Biol Conserv.*  
702 2011;144: 1712–1718.
- 703 80. Swihart RK, Gehring TM, Kolozsvary MB, Nupp TE. Responses of 'Resistant' Vertebrates to Habitat  
704 Loss and Fragmentation: The Importance of Niche Breadth and Range Boundaries. *Diversity and*  
705 *Distributions*. 2009; 9:1-18.

- 706 81. Kawanishi K, Sunquist ME, Eizirik E, Lynam AJ, Ngoprasert D, Wan Shahrudin WN, Rayan DM,  
707 Sharma, DSK, Steinmetz R. Near fixation of melanism in leopards of the Malay Peninsula. *J. Zool.*  
708 2010;282: 201–206.
- 709 82. Tan CKW, Moore J, Saaban S, Campos-Arceiz A, Macdonald DW. 2015. The discovery of two spotted  
710 leopards (*Panthera pardus*) in Peninsular Malaysia. *Trop. Conserv. Sci.* 2015;8: 732-737.
- 711 83. Wadey J, Fletcher C, Campos-Arceiz A. First photographic evidence of flatheaded cats (*Prionailurus*  
712 *planiceps*) in Pasoh Forest Reserve, Peninsular Malaysia. *Trop. Conserv. Sci.* 2014;7: 171–177.
- 713 84. Adyla MNN, Ikhwan Z, Ngah MZ, Shukor MN. Diversity and activity pattern of wildlife inhabiting  
714 catchment of Hulu Terengganu Hydroelectric Dam, Terengganu, Peninsular Malaysia. *AIP*  
715 *Conference Proceedings* 2016;:1784, 060038-~~(2016)~~; doi: <http://dx.doi.org/10.1063/1.4966876>.
- 716 85. Rayan DM, Linkie M. Conserving tigers in Malaysia: A science-driven approach for eliciting  
717 conservation policy change. *Biol Conserv.* 2016;204: 360–366.
- 718 86. Rostro-García S, Kamler JF, Ash E, Clements GR, Gibson L, Lynam AJ, McEwing R, Naing H, Paglia S.  
719 Endangered leopards: range collapse of the Indochinese leopard (*Panthera pardus delacouri*) in  
720 Southeast Asia. *Biol. Conserv.* 2016;201: 293–300.
- 721 87. Tan CKW, Rocha DG, Clements GR, Brenes-Mora E, Hedges L, Kawanishi K, et al. Habitat use and  
722 predicted range for the mainland clouded leopard *Neofelis nebulosa* in Peninsular Malaysia. *Biol*  
723 *Cons.* 2017; 206: 65–74.
- 724 88. Gumal M, Salleh A, Yasak M, Horng LS, Lee BPHY, Pheng LC. et. al. Small-medium wild cats of Endau  
725 Rompin Landscape in Johor, Peninsular Malaysia. *CATnews Special Issue.* 2014;8: 10–18.
- 726 89. Hansen MC, Hansen, Potapov PV, Moore R, , Hancher M, Turubanova SA, Tyukavina A, et al. High-  
727 Resolution Global Maps of 21st-Century Forest Cover Change. *Science.* 2013;342: 850–853.
- 728 90. Agus F, Gunarso P, Sahardjo BH, Harris N, van Noordwijk M, Killeen TJ. ~~(2013)~~ Historical CO2  
729 emissions from land use and land use change from the oil palm industry in Indonesia, Malaysia and  
730 Papua New Guinea. Roundtable on Sustainable Palm Oil, Kuala Lumpur. 2013. Available from:  
731 <http://www.worldagroforestry.org/sea/Publications/files/report/RP0296-13.pdf>. Cited 23 Oct  
732 2017.
- 733 91. McShea WJ, Stewart C, Peterson L, Erb P, Stuebing R, Gimán B. The importance of secondary forest  
734 blocks for terrestrial mammals within an Acacia/secondary forest matrix in Sarawak, Malaysia. *Biol*  
735 *Conserv.* 2009;142: 3108–3119.
- 736 92. Azhar B, Lindenmayer DB, Wood J, Fischer J, Zakaria M. Ecological impacts of oil palm agriculture on  
737 forest mammals in plantation estates and smallholdings. *Biodivers Conserv.* 2014;23: 1175–1191.
- 738 93. Sollmann R, Mohamed A, Niedballa J, Bender J, Ambu L, Lagan P, Mannan S, Ong RC, Langner A,  
739 Gardner B, Wilting A. Quantifying mammal biodiversity co-benefits in certified tropical forests.  
740 *Divers Distributions.* 2017;23: 317–328.
- 741 94. Edwards DP, Fisher B, Wilcove DS. High Conservation Value or high confusion value? Sustainable  
742 agriculture and biodiversity conservation in the tropics. *Conservation Letters.* 2012;5: 20–27.
- 743

Commented [FTvM1]: No parentheses?