1 Improving bank erosion modelling at catchment scale by incorporating

2 temporal and spatial variability

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10 Abstract

Bank erosion can contribute a significant portion of the sediment budget within 11 temperate catchments, yet few catchment scale models include an explicit 12 representation of bank erosion processes. Furthermore, representation is often 13 simplistic resulting in an inability to capture realistic spatial and temporal variability in 14 simulated bank erosion. In this study, the sediment component of the catchment 15 scale model SHETRAN is developed to incorporate key factors influencing the 16 spatio-temporal rate of bank erosion, due to the effects of channel sinuosity and 17 channel bank vegetation. The model is applied to the Eden catchment, north-west 18 England, and validated using data derived from a GIS methodology. The developed 19 model simulates magnitudes of total catchment annual bank erosion (617 - 4063 t yr 20 ¹) within the range of observed values (211 - 4426 t yr⁻¹). Additionally the model 21 provides both greater inter-annual and spatial variability of bank eroded sediment 22 generation when compared with the basic model, and indicates a potential 61% 23 increase of bank eroded sediment as a result of temporal flood clustering. The 24 approach developed within this study can be used within a number of distributed 25

hydrologic models and has general applicability to temperate catchments, yet further
development of model representation of bank erosion processes is required.

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29 Keywords

30 Bank erosion, sediment, sinuosity, vegetation, catchment.

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32 Introduction

Sediment erosion and transport are natural geomorphic processes within river 33 34 catchments, but high magnitude events and anthropogenic influences (such as deforestation and over-grazing) can easily disrupt the sensitive equilibrium between 35 them. When these changes result in increased sediment loads, they may have 36 numerous detrimental effects to the river system; increased sedimentation in 37 channels and floodplains affecting land-use and changes in river morphology and 38 behaviour (Owens et al, 2005), flooding (Mcintyre et al, 2012), and disruption to 39 habitats and decreased biodiversity (e.g. salmonid spawning, Soulsby et al, 2001). 40 Furthermore, as sediments act as a transport vector for pollutants such as heavy 41 metals, increased sediment delivery may also change the chemical composition of 42 the river resulting in negative impacts to the ecosystem (eutrophication, Owens and 43 Walling, 2002; and toxicity effects, Mackin et al, 2003). Consequently, information 44 on sediment generation and transport through river systems at a catchment scale, 45 and their temporal and spatial variability is increasingly important to support 46 catchment management. 47

48 Sediment fingerprinting techniques have been applied to a number of catchments 49 worldwide to understand the relative importance of different sources of sediment, 50 including eroded bank material. These suggest that bank erosion contributes significantly to catchment sediment budgets, in some cases representing up to 48% of total sediment supply (Walling, 2005; Walling et al, 2008). Furthermore, where channel banks contain contaminated sediments the contribution of bank erosion to pollutant supply has also been noted to be significant; for example, lead supply from banks of 9 kg m⁻¹ yr⁻¹ (Glengonnar Water, Scotland UK, Rowan et al, 1995) and mercury supply of 2.7 kg km⁻¹ yr⁻¹ (South River, Virginia USA, Rhoades et al, 2009).

The severity of bank erosion is influenced by numerous factors such as the 57 presence of bank vegetation (through both mechanical and hydrological factors) 58 (Micheli and Kirchner, 2002; Bartley et al, 2008; Simon and Collison, 2002); 59 discharge and flow regime (Julian and Torres, 2006; Hooke, 2008; Surian and Mao, 60 2009); lithology (Hooke, 1980); channel confinement (Lewin and Brindle, 1977; 61 Janes et al, 2017); and anthropogenic influences (Winterbottom and Gilvear, 2000; 62 Michalková et al 2011). As such rates of channel bank erosion are both highly 63 temporally and spatially variable (Hooke, 1980; Bull, 1997; Lawler et al, 1999; 64 Couper et al, 2002). 65

Management of sediment and other diffuse pollution issues at a catchment scale 66 is imperative due to the connectivity of the system. Models provide a valuable means 67 of estimating sediment generation and transport at catchment scales, potentially 68 providing insights into the spatio-temporal generation and transport of sediment and 69 70 the system responses to longer term changes such as climate change. However, many existing catchment-scale hydrological and water guality models contain no 71 explicit representation of channel bank erosion processes; CREAMS - Chemicals, 72 Runoff and Erosion from Agricultural Management Systems (Knisel, 1980), 73 ANSWERS - Areal Nonpoint Source Watershed Environment Simulation (Beasley 74 and Huggins, 1980), EPIC - Erosion Productivity Calculator (Sharpley and Williams, 75

1990), SWAT – Soil and Water Assessment Tool (Arnold et al, 1998), and PSYCHIC 76 - Phosphorus and Sediment Yield Characterisation In Catchments (Davison et al. 77 2008). Additionally, those models which do contain representations of bank erosion 78 only account for few of the numerous aforementioned factors controlling channel 79 bank erosion rates which limits their ability to simulate the observed spatial and 80 temporal variation of sediment generation through bank erosion processes. For 81 example, the semi-distributed INCA-Sed model (Jarritt and Lawrence, 2007) 82 accounts for bank eroded sediment within in-stream sediment sources using a power 83 84 law relationship incorporating discharge and calibration parameters. As acknowledged by the authors, a range of sub-reach scale processes are not 85 included within the model and therefore only a broad range of seasonal trends can 86 be observed, rather than finer temporal and spatial variation. The model SedNet 87 provides a mean-annual sediment budget (Prosser et al, 2001; Wilkinson et al, 88 2009). Riverbank erosion within the model is based on an empirical relationship 89 related to stream power, the extent of channel bank vegetation, and non-erodible 90 surfaces. Whilst this method incorporates some factors influencing the spatial 91 variation of bank erosion rates and provides an estimate of annual sediment 92 generation, it does not account for finer-scale temporal variability or provide an 93 indication of event-based bank erosion. Whilst a dynamic version of the model (D-94 SedNet, Wilkinson et al, 2014) exists, this model disaggregates longer term data to 95 provide daily output this model, meaning the model is unable to fully capture the 96 temporal variability observed in sediment loads. 97

98 Detailed numerical models of bank erosion have been shown to simulate channel 99 migration with reasonable accuracy (Darby et al, 2002, 2007; Duan 2005; Nagata et 100 al; 2000). These models generally incorporate mathematical modelling of hydraulic bank properties, shear stresses acting on channel banks and subsequent erosion. However these models lack simulation of catchment hydrology, and the highresolution data required for such models and their computational requirements limit their application to reach scales. Therefore to provide estimates of bank-eroded sediment at a catchment scale, alternative methods are required.

If models are to provide the more holistic representation of sediment processes at 106 107 a scale that is needed to inform catchment management, further research is needed to improve two key aspects of catchment models; continuous simulation of coupled 108 109 hydrological and sediment processes, and the ability to replicate both temporal and spatial variability of natural systems. This paper therefore describes the further 110 development and application of the Système Hydrologique Européen TRANsport 111 (SHETRAN) model (Ewen et al, 2000) to provide improved spatio-temporal 112 representation of channel bank erosion processes within simulated catchment 113 sediment budgets. The physically based model SHETRAN was chosen due to the 114 ability of the model to represent both spatial and temporal variation of sediment 115 generation through physical representation of these processes and their controlling 116 factors. In particular, the paper shows how the modifications enable improved 117 simulation of the temporal (through representation of bank vegetation removal and 118 bank de-stabilisation associated with high magnitude events, and subsequent 119 120 recovery) and spatial (by taking account of the influence of channel sinuosity) variation of bank eroded sediment generation within the Eden catchment in north-121 west England. 122

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124 Methodology

SHETRAN (Systeme Hydrologique Europeen TRANsport) is a physically-based 125 distributed model for catchment scale simulation of hydrology and transport (Ewen et 126 al, 2000). The model operates using a grid based representation of the catchment, 127 with channel links situated along the edges of the grid cells. An option to include a 128 more comprehensive representation of channel bank hydraulics can also be 129 incorporated, resulting in an additional 10m width grid cell between channel links and 130 the adjacent grid cells. The temporal resolution of the model is typically one hour, 131 although the timestep decreases during storm events to provide an improved 132 representation of rapid infiltration and surface runoff processes. The processes 133 represented within the hydrological and sediment components of the model are 134 shown in Figure 2 and detailed within Birkinshaw et al, 2014 and Elliot et al, 2012. 135 The following section details the development of the bank erosion component of 136 SHETRAN and the application of the developed model is described in the 137 subsequent section. Hereafter, the existing SHETRAN bank erosion model is termed 138 the 'basic' model and the revised model implemented within this study the 139 'enhanced' model. 140

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142 **Description of model improvements**

The representation of bank erosion within the basic model is based on the exceedance of critical shear stress (τ_{bc}) acting on the channel banks. The critical shear stress is calculated using the Shield's curve method (similarly to Simon et al, 2000). Bank erosion (E_b) is calculated as a rate of detachment of material per unit area of bank (kg m⁻² s⁻¹) according to:

$$E_b = BKB.\left(\frac{\tau_b}{\tau_{bc}} - 1\right) \text{ where } \tau_b > \tau_{bc}$$

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where *BKB* is a bank erodibility parameter(kg m⁻² s⁻¹), and τ_b is the shear stress acting on the channel bank (N m⁻²) calculated as:

$$\tau_b = K\tau$$

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where *K* is a proportionality constant calculated from channel width and flow depth and τ is the mean flow shear stress on the bed. Whilst this equation accounts for the influence of varying discharge and hence shear stress acting on channel banks, all other significant factors (including those mentioned in the previous section) are not included. Therefore the natural variation of bank erosion rates both spatially and temporally throughout catchments is likely to be underestimated.

Within the enhanced model, spatial variation of bank erosion is represented by 159 way of the non-linear influence of local channel sinuosity on bank erosion. This is 160 incorporated within the model by categorising channel sinuosity in to one of three 161 groups (similarly to channel curvature ratio categories as detailed by Crosato, 2009); 162 channel links with low sinuosity (<1.2) have low erosion rates, moderately sinuous 163 channels (1.2-1.5) have the highest erosion rates, and highly sinuous channels 164 (>1.5) have erosion rates slightly lower than that of moderately sinuous channels 165 (Janes, 2013). 166

Temporal variation of bank erosion as a result of the changing channel bank vegetation is represented within the model by varying the bank erodibility coefficient (*BKB*) between minimum and maximum values over time (see Figure 3). When channel discharge at a location in the catchment exceeds a threshold value (Q_{Thresh}) for that location the bank erodibility coefficient at that location increases to a maximum value (*BKB_{max}*). Q_{Thresh} represents the discharge at which vegetation within

some parts of the reach is expected to be removed, and hence bank erodibility is 173 increased. For outer-bends with little vegetation this increase in erodibility represents 174 de-stabilisation of channel banks. Q_{Thresh} at the catchment outlet is set by the user 175 (based on flood recurrence interval), and then each link is given a unique value of 176 Q_{Thresh} calculated from the value of Q_{Thresh} at the outlet (the methodology used is 177 detailed in the model application section). For all subsequent time steps of the model 178 where the threshold value is not exceeded, the bank erodibility coefficient gradually 179 decreases over time to the minimum value (BKB_{min}) at a rate set by the recovery 180 181 factor (R):

$$BKB_t = BKB_{max}$$
 where $Q \ge Q_{Thresh}$

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$$BKB_t = BKB_{t-1}$$
. R where $BKB_t > BKB_{min}$

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The difference in the magnitude of *BKB_{min}* and *BKB_{max}* represents the stabilising influence of vegetation on channel banks. The seasonal climate also influences the recovery factor (R), which reflects the potential rate of re-growth of bank vegetation and subsequent bank protection and stabilisation. R is calculated from the potential evapotranspiration (as a proxy for plant development) assuming that bank-side vegetation are not water-limited due to the shallow depth to the watertable:

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$$R = 1 - \left(k. \partial t. \left(\frac{PE_{obs}}{PE_{max}}\right)\right)$$

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where PE_{max} represents the maximum daily potential evapotranspiration (mm s⁻¹), PE_{obs} (mm s⁻¹) is the observed potential evapotranspiration and ∂t is the length of the time-step (seconds). The parameter k controls the time-scale of vegetation recovery and should reflect the type of vegetation in the catchment. Higher values of k, leading to a quicker recovery times, are appropriate for species with the ability of rapid re-growth, such as willow (Salix fragilis). Table 1 shows the input parameters required for the developed bank erosion model.

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200 Application of the enhanced model

The model was applied to the 2400km² predominately rural Eden catchment in north west England, UK (see Figure 4). Topographical variation across the catchment (788m AOD at the highest point, to 15m at the outfall at the Sheepmount gauge) results in significant variation of average annual rainfall; the lower Eden receives approximately 800mm yr⁻¹ whilst upper reaches receive in excess of 2800 mm yr⁻¹(Mayes et al, 2006).

The model was applied with a grid resolution of 1km² (and bank cells with a 207 length of 1km and width of 10m) with a maximum hourly temporal resolution. A 1km² 208 grid resolution reasonably captured the OS (Ordnance Survey - UK national 209 mapping agency) blue line channel network. The model was set-up using 30m Digital 210 elevation model (Ordnance Survey, 2009), land-use (CEH, 2007), and soils (Wosten 211 et al, 1999). A daily 1km² gridded daily rainfall product from 1990-2007 (Perry et al, 212 2009) was used to specify the spatial rainfall, with tipping bucket rain gauge data 213 then used to disaggregate the daily data to an hourly resolution to capture the 214 shorter duration intensities. A simple nearest neighbour approach was applied to 215 disaggregate the daily totals to hourly; for each grid cell, the shape of the nearest 216 available hourly record was used to distribute the daily total to hourly intervals (see 217 Lewis et al, 2016 for further details). 218

219 The parameter Q_{Thresh} , which determines the discharge that leads to significant bank de-stabilisation and erosion, was derived in a three stage process and has a 220 unique value for each link scaled from the value of Q_{Thresh} at the outlet. Firstly, the 221 model was run using the long term average daily rainfall (temporally constant, but 222 spatially variable across the catchment) to derive steady state simulated discharge at 223 the catchment outlet, from which scaling factors were calculated for all links based 224 on the ratio of local link flow to the outlet discharge. Secondly, the discharge 225 magnitude at the catchment outlet for a flood of a return interval to represent Q_{Thresh} 226 227 event was calculated using the annual maximum (AMAX) dataset (CEH, 2015) covering 46 hydrological years (1966-2012), the median of annual maximum values 228 (Q_{med}) and a Generalised Logistic growth curve (estimated using L-moments, see 229 230 Flood Estimation Handbook, Faulkner 1999). For a given return period T:

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 $Q_T = x_T \cdot Q_{MED}$

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where Q_T is the discharge for an event with return interval ($_T$), x_T is the growth factor (the value of the growth curve at a given return period). Finally the corresponding Q_{Thresh} values throughout the catchment were calculated by multiplying Q_{Thresh} valueat the catchment outlet by the scaling factors.

All channel links within SHETRAN representations are located between two channel bank cells and have a default sinuosity of 1. Therefore a GIS-based channel network was used to estimate sinuosity for each link. Sinuosity was measured across the catchment using WFD river waterbodies data (Environment Agency, 2012) and GIS; a channel network polyline was split into reaches of equal length, and sinuosity calculation for each reach was calculated as the channel distance divided by the straight-line distance between reach start and end points. As the value
of sinuosity is dependent on the reach length at which it is measured, this process
was repeated for a range of length scales. The length scale with the largest peak in
variance of sinuosity (measurement length of 975m) was used as this best captured
the variation of sinuosity across the catchment.

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Model calibration and validation

After a one year 'start-up' period in which groundwater levels tended to an 251 252 equilibrium, the model was run from 1991-2001 for parameter calibration, and 2001-2007 for validation. Similarly to previous studies using SHETRAN (Bathurst et al, 253 2006; Lukey et al, 2000; Elliott et al, 2012) calibration parameters included the 254 overland and channel flow resistance coefficients, with calibration conducted 255 manually due to the computational requirements of the model. The hydrological 256 component of the model was compared with hourly and daily hydrological data from 257 the National River Flow Archive (CEH, 2015) gauging stations and HiFlows data sets 258 (see Figure 4). From this a range of parameter value sets were derived (see Table 3) 259 based on parameters to which the simulated flows were most sensitive (Lukey et al, 260 2000 Bathurst et al, 2006). The simulation outputs were then superimposed on each 261 other, providing an envelope of minimum and maximum model estimates of river 262 flows. 263

Analysis of peak-over-threshold (POT) events was also conducted as part of the validation process to ensure the model could accurately reproduce high-magnitude events, using POT data from the NRFA (CEH, 2015). For each POT event the observed event maximum discharge was compared with the maximum simulated discharge within 24 hours either side of the event timing. The average percentage error of simulated POT events was then calculated within the calibration/validationperiods for each gauging station.

The bank erodibility parameters (see Table 2) were calibrated by comparison with 271 observed bank erosion values derived using an historical map overlay methodology 272 in GIS, further details of which can be found in Janes et al (2017). Channel banklines 273 were digitised for the Eden and main tributaries Caldew, Irthing, Lyvennet, Eamont 274 and Petteril from Historical OS maps for the 5 available years (1880, 1901, 1956, 275 1970, and 2012) with consecutive banklines overlaid to provide an area of bank 276 277 erosion. As smaller tributaries are often represented on OS maps as a single line (particularly on older maps) it is not possible to calculate bank erosion values for 278 these channels using this methodology. To account for potential geo-referencing and 279 mapping errors within the data, the eroded area was calculated using the simple 280 overlay procedure, and also applying a buffer of 3.5m to the older channel, providing 281 upper and lower erosion estimates respectively. Minimum and maximum bank height 282 estimates were calculated from the two bank heights provided within the RHS survey 283 data, to account for error within the estimate. Minimum and maximum estimates of 284 annual bank eroded sediment were estimated for each sub-catchment using this 285 procedure. Whilst alternative methods of data collection such as erosion pin 286 methodologies can provide estimates of bank eroded sediment at a finer temporal 287 resolution (event scale), these methods are limited spatially and cannot provide 288 catchment wide estimates of bank erosion and are therefore unsuitable for this 289 study. 290

291 Preliminary magnitudes of differences in erosion rates between vegetated and 292 non-vegetated banks, and parameters influencing the length of recovery time were 293 based on literature of riparian growth rates of vegetation types found in the area

(Environment Agency, 1998). The recovery factor was calibrated as 3 months during 294 summer according to bank vegetation growth rates in Environment Agency, 1998. 295 The return period of an event used to calibrate the Q_{Thresh} parameter was guided by 296 literature evidence and was based on an event with return period of greater than 12 297 years. The variation of bank erodibility with channel sinuosity was parameterized 298 based on Janes et al (2013); bank erosion rates at channel sinuosities around the 299 threshold value of sinuosity (~1.5) are approximately 2.75 times greater than straight 300 channels (low sinuosities), and in highly sinuous channels (>1.5) approximately 2 301 302 times greater.

Model simulations with the sediment component were conducted across the 303 range of hydrological parameters specified in Table 3, so that the simulated 304 suspended sediment load and bank erosion values incorporate the effects of the 305 hydrological parameter uncertainty. Similarly to the hydrological component of the 306 model, minimum and maximum parameter values were set for sensitive sediment 307 parameters, and simulations were conducted using a range of parameter values 308 within this range (see Table 3). Simulated annual sediment loads were calculated 309 and compared to those predicted by sediment rating curves, derived using grab 310 samples and turbidity data collected from several locations between November 2006 311 and March 2009 (see Figure 4) by the CHASM (Catchment Hydrology And 312 313 Sustainable Management) project (Mills, 2009). These were then used in conjunction with either gauging station data or simulated discharge to provide estimates of 314 annual sediment loads at these locations. 315

The sensitivity of the enhanced model to temporal flood clustering was analysed with respect to the magnitude of bank eroded sediment. To do this the model was run with a one year start-up period, and then three days of rainfall (taken from the

January 2005 event, 6/01/2005 - 8/01/2005 inclusive with a peak discharge at 319 Sheepmount of 1516.3 m³s⁻¹, as this was a notable high magnitude event). A 320 temporally constant rainfall was then used for one week before a second smaller 321 rainfall event that did not exceed Q_{Thresh}. The model was then re-run with 2, 4, 6, 8 322 and 12 week gaps between the two events. Constant temporal rainfall input between 323 the two events was used to ensure identical antecedent hydrological conditions prior 324 to the second event so that simulated differences in the magnitude of bank eroded 325 sediment were due solely to event timing. 326

327

328 **Results**

329 Hydrological assessment

Table 4 shows the average hourly hydrological performance statistics of the model for the validation period (and daily statistics at Kirkby Stephen where hourly flow data were unavailable). All hourly NSE and R² values are above 0.55 and 0.7 respectively, indicating satisfactory model performance at all sites (Moriasi et al, 2007). The simulated absolute percentage bias is below 25% at all gauging stations (indicating satisfactory model performance according to Moriasi et al, 2007) and at 5 of the 8 stations is less than 8%.

The POT analysis indicates the model's ability to predict high-magnitude events (see Figure 5 and Table 5). Although the model under-estimates event peak flow at most locations, as is common with other hydrological models (Butts et al, 2004; Van Liew et al, 2003), 65% of POT events were within the simulated uncertainty range at the catchment outlet at Sheepmount (Table 4 and Figure 5). It should be noted that the gauging station on the Irthing at Greenholme is often affected by backwater from the Eden at medium-high flows, which could partially explain the lower peak overthreshold simulation accuracy observed at this location (Table 5).

345 Bank erosion

The GIS overlay methodology indicates the total mass of sediment generated 346 through bank erosion processes within the catchment is between 539-2346 t yr⁻¹ 347 (Table 6). The estimates from both GIS methodologies provide an uncertainty range 348 between 211-4426 t yr⁻¹. Total annual simulated bank erosion in Table 7 is higher 349 than the most recent observed average annual bank erosion rates (1970-2012 -350 Table 6) but within the observed uncertainty range over the historical. Additionally, 351 352 Table 7 indicates the enhanced model simulates a greater inter-annual variability of average annual bank erosion rates than the basic model. The enhanced model 353 simulates a greater range of spatial variation of bank erosion throughout the 354 catchment than the basic model. The basic version of the model was parameterised 355 so that the total catchment average annual mass of bank eroded sediment 356 generation was similar to the enhanced model to enable comparison of spatial bank 357 erosion simulation in Figure 6. The observed data used for comparison here is taken 358 from the upper estimate. The basic version of the model (Figure 6A) simulates a 359 fairly spatially constant magnitude of bank erosion throughout the catchment in 360 comparison to the enhanced model (Figure 6B) and the observed data (Figure 6C). 361 The model was also validated at a sub-catchment scale using Water Framework 362 Directive sub-catchment boundaries by correlating the total simulated bank eroded 363 sediment of the basic and enhanced versions of the model with the observed data. 364 Correlations between simulated and observed data indicate the enhanced model 365 provides a more accurate spatial estimation of bank erosion at the sub-catchment 366 level (R=0.500, p=0.007) compared to the basic model (R=0.367, p=0.048). These 367

368 correlation values indicate an improvement in the spatial variability of bank erosion
 369 simulated by the developed model, but nevertheless the overall predictive ability of
 370 the spatial variability is poor due to reasons detailed within the discussion.

371 Sediment load accuracy

Table 8 shows observed annual sediment loads with upper and lower 95% 372 confidence intervals (calculated from the coefficient of the rating curve equations 373 from Mills, 2009), and simulated annual sediment loads with upper and lower bounds 374 based on the parameter set used for simulation. The confidence intervals of the 375 376 observed sediment loads incorporate both hydrological and sediment parameterisation uncertainty and are of a similar magnitude to the uncertainty 377 bounds of simulated sediment loads. Furthermore, the ranges of simulated and 378 observed sediment loads overlap at all locations. 379

380 Sensitivity to temporal flood clustering

Values of bank eroded sediment generation for each of the five temporal flood 381 cluster scenarios was calculated by summing the total catchment bank erosion for 31 382 days, starting from the date of the second rainfall event (see Table 9). The model 383 indicates bank eroded sediment generated from a single flood event may be up to 384 61% greater if the event occurs within 2 weeks of a large flood event. As the 385 temporal separation of the two flood events increases the magnitude of bank erosion 386 caused by the second event decreases. Once channel bank vegetation has 387 recovered from the first event, subsequent events below the threshold discharge do 388 not result in increased magnitudes of bank erosion. 389

390

391 **Discussion**

Observed bank erosion rates within this study determine the significance of channel bank erosion as a sediment source within the Eden catchment, Cumbria. Based on average annual simulated sediment load at Sheepmount, the data collected indicate that bank erosion represents 5-11% of the annual catchment sediment budget. This value is at the lower end of the range observed within other UK catchments (Walling, 2005; Walling et al 2006; Bartley et al 2007) which could be partly due to the predominance of grassland within the catchment.

The GIS dataset also indicates significant temporal variability of average annual 399 bank erosion rates between the four time-periods analysed, but does not fully 400 capture the inter-annual variability. Several previous studies have noted significant 401 inter-annual variability of bank erosion processes (Hooke, 2008; Kronvang et al, 402 403 2013). Simulated bank eroded sediment generation using the enhanced model shows greater inter-annual variation of bank erosion rates than those of the basic 404 model (Table 7), with the highest values during the year 2005. This is expected as 405 the largest event discharge recorded during the study period (and 2nd largest to date) 406 at this station occurred during the January of this year $(8/1/2005 \ 1516.3 \ m^3 s^{-1})$. 407 Previous studies have indicated the significance of high magnitude events to bank 408 erosion (Hooke, 1979; Julian and Torres 2006; Henshaw et al, 2012; Palmer et al, 409 2014). The developed representation of bank erosion processes enables model 410 411 sensitivity to high magnitude events, and therefore replication of observed temporal (inter-annual) variability of sediment generation. 412

The observed average annual bank erosion rates for the years 1970-2012 shown in Table 6 are lower than average simulated values for 2001-2006. The observed data present an average annual bank erosion value across several years and interannual variation within time periods, as a result of flood rich and poor years, is not

represented. The average annual maximum discharge recorded at Sheepmount from 417 1970-2012 was considerably lower than between 2001-2006 (647m³s⁻¹ and 764m³s⁻¹ 418 respectively). Therefore bank erosion rates between 2001-2006 would be expected 419 420 to be higher than the 1970-2012 average. Furthermore, observed data show total bank erosion within 6 main channels of the Eden catchment, additional smaller 421 tributaries have not been included, yet simulated values include the whole catchment 422 as represented by the model. The lower estimates of observed bank erosion are 423 taken from the GIS overlay methodology with a 3.5m buffer applied to account for 424 425 errors within the mapping process, which for more recent maps (such as 1970 and 2012) should be less significant than for earlier maps. Therefore the lower estimate 426 of actual bank erosion for the 1970-2012 time-period is potentially a significant 427 428 underestimate of reality.

The enhanced model simulates sensitivity to flood clustering, by incorporating an 429 element of catchment recovery following a large event. The results indicate bank 430 eroded sediment generation for an event of the same magnitude may vary 431 depending on the event timing. Previous studies have noted the importance of 432 antecedent conditions to bank erosion processes; Hooke (1979) noted that whilst 433 event-based bank erosion at certain sites was correlated with discharge of the 434 previous peak, the influence of this variable is complex. Previous high flows can 435 436 weaken banks by undercutting but can also remove loose bank material leaving the bank more resistant to subsequent high flows. Thorne (1982) observed that mass 437 failure of banks can result in an increase in bank stability due to supply of sediment 438 to the basal zone, unless critical shear stress for removal of this basal material is 439 exceeded. The enhanced model developed in this study provides an additional 440 element of catchment memory for bank erosion and enables simulation of the effects 441

of event clustering, and influence of antecedent conditions. The frequency of high magnitude events within the UK is expected to increase with projected climatic changes (Bell et al, 2012; Kay et al, 2014; Madsen et al, 2014). Therefore, to enable climate-proof catchment management practices models will be required to represent the effects of flood clustering.

The spatial variation of bank erosion simulated by the basic model was controlled 447 solely by flow variation (and hence variation of shear stress) throughout the 448 catchment. As shown in Figure 6A this resulted in little variation of simulated bank 449 450 erosion across the catchment. Significant spatial variation was observed from the GIS analysis within this study (Figure 6C), and has been observed within several 451 additional UK catchments (Bull, 1997; Lawler et al, 1999). The inclusion of sinuosity 452 within the enhanced model enables simulation of some spatial variability of bank 453 erosion rates within the catchment (Figure 6B). Correlation of sub-catchment totalled 454 bank erosion rates indicate that bank erosion predicted by the enhanced model is 455 more accurate than the basic model, yet still provides a weak fit of the observed 456 bank erosion rates throughout the catchment. Several factors such as anthropogenic 457 influences, lithology, channel confinement, bank height, and slope influence bank 458 erosion rates resulting in the significant observed spatial variability within 459 catchments. Whilst sinuosity is known to be one factor influencing the spatial 460 variation of bank erosion (Janes 2013; Micheli and Kirchner 2002) many of these 461 additional factors are not included within the developed model due to current limited 462 understanding of their behaviour, complex interactions, and lack of spatial data 463 coverage. Therefore some differences between the simulated and observed bank 464 erosion rates are to be expected due to the omission of many of these factors and 465 the widely recognised difficulty of capturing the naturally high variability in bank 466

467 erosion rates. Comparisons of observed and model simulated bank erosion values 468 such as those in Figure 6 are rarely performed but these types of analyses are 469 required if models are to be judged useful in management at the local scale. The 470 model can be used to assist identification of areas where bank erosion would be 471 expected to occur naturally, and comparison with observational data can indicate 472 areas where bank erosion is prevented/accelerated due to anthropogenic factors not 473 included within the model.

The observed bank erosion data within this study provides an estimate of annual bank eroded sediment generation with greater spatial resolution and over a longer timescale than is possible using field-based techniques (such as erosion pins). However, it is not possible to accurately estimate event-based bank eroded sediment using data derived from this methodology. Further data (such as LIDAR analysis of bank migration at a finer temporal scale) and analysis is required to calibrate the model and assess performance during individual events.

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482 **Conclusions**

Channel bank erosion contributes a significant proportion of catchment sediment 483 budgets and yet is commonly excluded or overly simplified within catchment scale 484 models. In this study, the bank erosion component within the physically-based 485 486 SHETRAN model has been further developed to incorporate both temporal and spatial variability of bank erosion by inclusion of additional controlling factors; 487 removal of bank vegetation and bank collapse after a flood event and subsequent 488 recovery, and channel sinuosity. The developments within this study improve the 489 representation of natural processes influencing bank erosion rates, and enable 490 representation of catchment sensitivity to flood event clustering. 491

The model has been successfully applied to the Eden catchment, north-west 492 England, and validated using hydrological, bank erosion and suspended sediment 493 data. The enhanced model has been shown to simulate improved inter-annual and 494 spatial variability of catchment scale bank eroded sediment generation when 495 compared with the basic model, yet it is noted that the developed model still provides 496 a weak fit with observed data. Differences between the spatial variation of observed 497 and simulated bank erosion rates are attributed to additional factors not included 498 within the model due to limitations in current understanding and data availability. 499 500 Simulated sediment loads were compared with observational data, and whilst uncertainty in both observed and predicted sediment loads is large, values were 501 found to overlap throughout the catchment, indicating reasonable accuracy of model 502 503 simulations. Whilst the accuracy of spatial bank erosion simulations is currently insufficient to support application of the model for management purposes the study 504 represents a contribution to the research need for continuing development of 505 sediment models. The developed representation of bank erosion processes that 506 have been applied to the SHETRAN model in this study could also be applied to a 507 number of existing physically based models. 508

The developed representation of sediment source estimation within the model provides a more holistic representation of sediment processes throughout the catchment. The resultant model provides an improved representation of the spatial and temporal variability of sediment loads, yet further development of such models is required to provide estimates of sediment loads with sufficient accuracy to support management of diffuse pollution.

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516 Acknowledgments

We would like to thank the two anonymous reviewers for their helpful and 517 constructive comments that assisted in improving this manuscript. This work was 518 funded by EPSRC as part of the FloodMEMORY project EP/K013513/1. Additional 519 thanks go to the Environment Agency for the provision of channel survey and bank 520 height data, Cranfield University for LandIS soil data, and the CHASM project and 521 Carolyn Mills for sediment data. The associated metadata/data presented in this 522 research can be accessed using the following DOIs: 10.17862/cranfield.rd.4300220, 523 10.17862/cranfield.rd.4300202. 524

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741 Table 1: Model user input parameters required for the developed bank erosion

model. Parameter Q_{Thresh} is scaled to the outlet value.

	Parameter	Units	Description
	BKB min	kg m⁻¹ s⁻¹	Minimum bank erodibility
	BKB max	kg m⁻¹ s⁻¹	Maximum bank erodibility
			Threshold discharge at which BKB for the link increases from
	Q _{Thresh}	m ³ s⁻¹	BKBmin to BKBmax
			Vegetation recovery speed (high values = rapid growing
	k	N/A	vegetation types)
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762 Table 2: Calibrated parameter values of the bank erosion model.

Parameter	Calibrated value
Return period of Q _{Thresh}	12
k	0.03
Factoral difference between BKB _{min} and	
BKB _{max}	20

	Sinuosity	Straight channels <1.2	Meandering channels 1.2-1.5	Highly sinuous channels >1.5
	BKB _{min}	3.5E-11	9.6E-11	7.0E-11
765	BKB _{max}	7.0E-10	1.9E-09	1.4E-09
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782	Table 3: Validated parameter values for the Eden catchment model.								
	Parameter/function	Low value	High value						
	Hydrological Strickler overland flow resistance coefficient Saturated hydraulic conductivity in channel soil (mm day ⁻¹) Channel bank Strickler coefficients (x and y directions)	1 0.1 20	3 60 30						
783	Sediment Overland flow erodibility (kg m ⁻² s ⁻¹) Raindrop impact erodibility (J ⁻¹)	0.02 2E-12	0.05 1E-11						
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Table 4: Average performance statistics from the simulation of hourly flows
across the Eden catchment (with the exception of Kirkby Stephen based on
daily flows) during the validation period.

	Catchment/sub- catchment	Gauging station	Upstream area (km²)	NSE	R ²	PBIAS (%)
		Sheepmount	2286	0.901	0.911	3
		Great Corby	1373	0.857	0.869	3
	Eden	Temple	616			_
		Sowerby	00	0.857	0.873	8
		KIRKDY Stophon*	69	0 0 1 0	0 070	11
	Irthing	Groonbolmo	334	0.040	0.070	20
	numg	Harrahy	160	0.720	0.009	20
	Petterill	Green	100	0.630	0.796	-16
	Caldew	Cummersdale	244	0.830	0.835	8
	Eamont	Udford	396	0.598	0.713	-3
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- Table 5: Percentage of peak over threshold events within the simulated range
- 814 during the validation period, and average percentage error of simulated peak
- 815 discharge.

Percentage of simulated	
events within 15% of the Average error of	event
Channel Location observed event discharge simulation	(%)
Sheepmount 91 -1	
Great Corby 88 -1	
Eden Temple	
Sowerby 47 -19	
Kirkby Stephen 22 -44	
Irthing Greenholme 8 -51	
Petterill Harraby Green 38 19	
Caldew Cummersdale 31 -37	
Eamont Udford 60 28	
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- Table 6: Observed bank erosion rates (t yr⁻¹) from each overlay time period.
- Values shown are averages from all methodological estimates,

	Channel	1880-1901	1901-1956	1956-1970	1970-2012
	Eden Petteril Caldew Irthing Lyvenet Eamont	1329 136 412 356 55 58	682 58 187 216 26 17	1612 209 439 487 59 44	198 29 117 166 12 16
833	Total	2346	1186	2849	539
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Table 7: Annual bank erosion for the whole catchment as simulated by both
the basic and enhanced models during the validation period. Values are in t yr⁻
¹.

			2001	2002	2003	2004	2005	2006
		Minimum	721	1655	617	1686	2842	622
	Enhanced	Maximum	4063	2833	2219	2682	3898	2784
		Average	2331	2120	1401	2093	3350	1400
	Desta	Minimum	1951	3170	1542	2907	2356	2943
	Basic	iviaximum Averade	2126	3355 3231	1728	3129 2072	2539 2404	3183 3013
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Table 8: Observed and simulated average annual sediment loads (t yr⁻¹).

Location	Observed average	Simulated average	Observed 95% Confidence range	Simulated
Groat Corby	21068	21254	10325-43277	11366-31056
Tomple Sowerby	16016	0101	6096 26106	1071 12654
Appleby	150010	9121		407 1-13034
	15364	JOZ1	1229-10/4/	3110-0774
Great Musgrave	5126	4263	1794-7945	2197-6479
Kirkby Stephen	1794	1528	/36-3086	758-2362
Smardale	444	739	164-719	368-1147
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Table 9: Model sensitivity to temporal sequencing of flood events. Bankerosion values shown are summed from the whole catchment over a period of31 days, starting from the beginning of the second rainfall event.

Time between flood events (weeks)	Monthly bank erosion during second event (t)
1	851
2	681
4	547
6	536
8	530
12	528