



Hybridity, possibility: Degrees of marketization in tradeable permit systems

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Hybridity, possibility: Degrees of marketization in tradeable permit systems

Tradeable permit systems for allocating rights to impact the environment have become an important part of the regulatory toolkit over the last 35 years. Informed by elegant neoclassical economic thought, these regulatory markets promise lowest-cost environmental benefit by delegating decision-making to market participants. This article examines tradeable permits systems, including cap-and-trade carbon markets, tradeable water quality permits, and individual transferable quotas for fisheries, to understand the actual outcomes of attempts to transfer environmental governance to markets. Drawing on heterodox economic thought and the neoliberal natures literature, this article examines the institutional forms tradeable permit systems engender in conjunction with a neoclassical definition of ‘market’ to provide an internal critique of environmental marketization. In general, tradeable permit systems do not facilitate highly liquid financial markets that might signal the increasing importance of regulatory markets as an accumulation strategy for capital. Instead, market-based regulation tends to behave rather differently, acting as performance standards, quotas, or environmental taxes. The article concludes by reflecting on the political possibilities afforded by the recognition that regulatory markets, for the most part, bear little resemblance to high-flying financial markets.

Keywords: Markets, Marketization, environmental policy, tradeable permits, financialization of nature

Over the past 35 years, policymakers, their business allies, and academic economists have increasingly attempted to couple markets with environmental regulation. Proponents claim that markets are the most efficient way of distributing the costs of environmental regulation, allowing the state to define targets for degradation, and then letting markets determine where reductions in degradation are most cheaply accomplished. Market-based regulation through tradeable permit systems (TPS, or ‘cap-and-trade’), nominally meant to address problems ranging from aquatic species loss to ozone depletion, have been deployed across the world. This has contributed to the alarm of many scholars about the potential for the wholesale commodification nature (e.g. Castree 2008; Büscher et. al 2012), signaling a worrying trend towards the marketization of just about everything (Sullivan 2013; Büscher and Fletcher 2015). However, such claims merit scrutiny if we are to have a robust understanding of the capital/nature nexus as it is produced through regulatory programs like TPS. Are TPS turning nature in to a set of interlinking markets such that non-human environments can only be reproduced through the circulation of capital? And does the extension of financial logics into environmental management create new and important conduits for capital? If the answer to these questions is negative, then what sorts of finance/nature entanglements are TPS creating?

The aim of this paper is to evaluate TPS on their own terms, as an avenue to policy engagement in post-political settings that might not otherwise be accessible (Swyngedouw 2014, Bigger and Robertson 2017). I argue that justifications for the use of TPS for environmental regulation that depend on the creation of liquid financial markets for trading the rights to degrade nature are rarely fulfilled in practice. Instead, efforts at their implementation represent a spectrum of modes of environmental governance. These arrangements range from more conventional regulation with added layers of bureaucracy to highly liquid financial markets, but each TPS may fit different categories at different times. This article focuses on existing TPS, particularly the California carbon market where the author conducted 14 months of ethnographic field work. This fieldwork involved observation at more than 50 regulatory hearings and industry events, supplemented with interviews with key regulators, business associations, and environmental NGOs. The

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3 ethnographic observations are coupled with data and secondary literature on Water Quality
4 Trading markets, tradeable fishing rights, and both sympathetic and critical evaluations of other
5 TPS.
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8 The next section surveys literature on the financialization of nature and introduces a theoretical
9 framework that incorporates political and cultural economy approaches. I use this approach to
10 interpret the degree to which TPS create markets conforming to backers' expectations, to
11 examine the possibility for policy hybridity that features market infrastructure, and to understand
12 how actually-existing TPS function political-economically. I discuss the neoclassical economic
13 basis for TPS, and the range of settings to which they have been applied. I present my
14 interpretation of Ben-Porath's stylized (1980) definition of a market, providing a set of emic
15 criteria against which we can interpret the operation of TPS. I use these criteria, along with
16 heterodox economic thought on how facets of the socionatural world are 'rendered economic' to
17 evaluate the degree to which a given TPS is 'marketized' according to neoclassical terms. The
18 subsequent section elaborates three categories of challenges faced by TPS that restrict
19 marketization: lack of participants, challenges with making natures fungible, and regulatory
20 choices. If no, or limited, trading is present, then much of the effort that goes into setting up a
21 TPS is largely for naught and alters how we might interpret these programs within broader
22 capitalist political economy. The promise of TPS is that trading engenders least-cost emissions
23 reductions by allowing rights to degrade nature to circulate among market participants resulting
24 in optimal allocation of both rights and economic activities. This approach is portrayed as
25 beneficial for individual firms who will have the flexibility to choose between reducing their
26 environmental impacts or buying the right to degrade from others. TPS are thus meant to
27 engender least-cost, flexible solutions at levels of firms, sectors, and throughout the economy: in
28 short, maximizing efficiency (Tietenberg 2006; see Felli 2014).
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33 The penultimate section will examine what more conventionally market-like TPS have in
34 common. The majority of tradeable permit systems operate somewhere in between a complete
35 lack of market behavior and highly liquid financial markets. However, TPS tend to engender
36 sprawling state apparatuses to facilitate and manage the circulation of capital and the right to
37 degrade in these programs. That circulation, where it exists at all, is often unidirectional where
38 one of the transacting parties is the regulatory authority. The transfer of capital may not be
39 through market exchange, but instead, circulation is accomplished through purpose-built
40 infrastructure, effectively operating as taxes or subsidies depending on the direction of the flow.
41 The article concludes with reflections about the challenges and opportunities that the continuing
42 roll-out of TPS presents for regulators, the role of finance in environmental regulation, and how
43 geographers might engage with policymakers.
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47 **Performing market governance in capitalist world ecology**

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49 Geographers have long been at the forefront of interpreting the neoliberalization of nature, of
50 which TPS is an exemplar (Castree 2008a, 2008b, Bakker 2010). Much of this work has
51 enhanced understanding of the vast universe of market based instruments (MBIs) for
52 environmental policy. This literature includes work on TPS (e.g. Mansfield 2004, 2006; Cooper
53 2015; Carton 2016), but has also focused on forms of market governance like wetland banking
54 (Robertson 2004, 2006), stream mitigation banking (Lave 2012), or payments/compensation for
55 environmental services (PES/CES) (Osborne 2011, McAfee 1999, McAfee and Shaprio 2010).
56 This literature has coincided with the rise of the 'financialization thesis', or the notion that
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finance plays an increasingly important role in the global economy (Krippner 2005, Lapavitsas 2009, French et al 2011, Moore 2011a). However, Christophers (2015a) pushed back this thesis, demonstrating shortcomings, including financialization's propensity to burst its analytical banks, losing purchase by coming to encompass many distinct practices and outcomes. As Ouma (2016, 82) notes, "[f]requently, 'financialization' has been turned into an abstract force sui generis, morphing from explanandum into explanans. In this regard, it has shared the fate of other concepts such as 'capitalism', 'the market', 'globalization' or 'neoliberalism'."

This article seeks to avoid those pitfalls by unpacking the sundry challenges that are inherent in marketizing nature through TPS. It does so by heeding the call of Berndt and Boeckler (2010, 1) for, "intensified efforts to understand how real (as opposed to ideal) markets are produced, stabilized and reshaped, and fall apart." There is a nascent body of literature (to which this article seeks to contribute) that questions both the importance of new natural 'asset classes' to finance, and of finance to reproducing those natures (Dempsey and Suarez 2016, Felli 2014, Loftus and Marsh 2015).

There is a vast body of literature on carbon markets from heterodox economic perspectives in social studies of finance (e.g. Callon 2007; Blok 2010, Deschaneu 2012, Lane 2012) and political economy (e.g. Lohmann 2010; Felli 2015, Smith 2006, Knox-Hayes 2013). I draw on these literatures to move toward a synthetic understanding of the practices and outcomes of making markets via TPS. A few authors have undertaken similar work for understanding related market-based instruments (e.g. Lapeyre et. al 2015; Hrabanski 2015), yet much of this work has been conducted outside geography scholarship. Each of these authors have demonstrated the non-market or not-quite market outcomes of attempts to create circulating financial representations of nature in price-based environmental governance. As Pirard and Lapeyre (2013, i) note in the context of biodiversity banking, "[payments for ecosystems services]... clearly remain a bilateral relationship with very little or no feature of market governance or commodification," while Schomers and Matzdorf (2013, 21), discussing nominally market-based programs for mitigating deforestation in China, show that "the payment is in fact a compensation for legal land-use restrictions and thus not an economic incentive to forest land use changes... it is rather a program to compensate for regulatory interventions." TPS have common intellectual roots, design features, and practical challenges with other MBIs, but remain distinct in part because of their ambition to create liquid financial markets that feature infinitely substitutable or 'screen-tradable' representations of nature. The distinction is important for evaluating the relative import of different mechanisms that seek to bring environmental degradation into formal circuits of capital, as capital attempts to reconcile its environmental contradictions while expanding its productive horizon (Moore 2015). For example, ecosystemsmarketplace.com (2014) tallies US\$12 billion in 'market-based' programs for water quality. However, only US\$41 million of that was transacted through bilateral (non-regulatory) agreements, and only US\$7 million worth of trades took place in water quality trading markets, with the remaining US\$11.95 billion going through subsidies or administrative prices set by regulatory fiat. This paucity of circulation through market trading hardly seems to signal the impending real subsumption of nature by capital.

To analyze TPS, I combine insights from political economy and the social studies of finance, denoted here as 'performativity approaches'. While long portrayed at odds with one another, Christophers (2014) demonstrates the synthetic potential in the study of markets. Attention to

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3 both performative and structural conditions of circulation through markets offers more nuanced
4 understanding of what markets do across spaces and scales. This outlook is shared by an
5 increasing number of approaches, perhaps most directly in the nascent ‘operations of capital’
6 literature set out by Mezzadra and Neilson (2015).
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9 To correct the tendency in performativity accounts to elide structural consequences (Fine 2003),
10 I couple this perspective with recent insights from green and feminist political economy.
11 Moore’s (2015) work on world ecology provides a helpful explanation of the impediments to
12 financializing nature, and can be paired with performative analyses that evaluate the ways that
13 economic theory is enrolled in attempts to overcome those impediments. Moore elaborates a
14 history of capitalism that turns on the selective enrollment of socialnatures (Haraway 1991) into
15 the formal circuits of capital. The main premise of interest is that capital’s growth has been, and
16 can only be, accomplished through the exploitation of nature without compensating for its full
17 cost, economically or environmentally (see also Mies 1986). There is significant nuance in
18 Moore’s work, but the key insight for a discussion of TPS is the importance of the ‘four cheaps’
19 for the continuing expansion of capital - labor, food, energy, and resources (2015). Moore argues
20 we are entering a new period of capitalist crisis marked by the inability of capital to ‘reproduce
21 the conditions of its own reproduction’ (14). We can interpret attempts at ‘internalizing
22 externalities,’ including TPS, as the protective instinct of capital to resolve this tendency through
23 the extension of markets more deeply and broadly in search of new sources of surplus value
24 while ameliorating environmental problems. It is effectively an attempt to extend the
25 spatiotemporal fix that the incursion of financial practices into new settings represents into
26 nature (Harvey 1982, Ouma 2016). The contradiction is that enough ‘free gifts’ (e.g. the
27 atmosphere as a dump) must remain unaccounted for in the extraction of surplus; TPS can thus
28 be understood as a Polanyian double movement embodied in one regulatory instrument (Polanyi
29 1957, Carton 2014). This perspective also attunes us to the central role of the state in attempting
30 to reconcile capitalism’s ecological contradictions as it attempts to facilitate the circulation of
31 capital through nature as a means of safeguarding future conditions of capital accumulation.
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37 We can use this combined performativity/political economy approach to understand the ways in
38 which material and ideological infrastructures are built to facilitate the circulation of capital
39 through nature, and how structural impediments to full-blown marketization come into play
40 alongside the challenges to marketization faced in specific environmental concerns. This is to
41 say, rendering an environmental problem ‘economized’ (Çalışkan and Callon 2009) signals the
42 possibility for market-making, but the latter will not necessarily emerge from the former. It is
43 useful to understand the interstitial spaces these mechanisms occupy and what impacts they have.
44 Indeed, “...[the] dynamics (of marketization) necessitate the invention of new strategies of
45 struggle, alliance, and translation adequate to the production of subjectivities capable of working
46 across these discontinuities and heterogeneities in antagonistic and liberating ways,” (Mezzadra
47 and Neilson, 2015, 7).
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51 **Trading natures**

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53 In contrast to the much wider category of MBIs for environmental governance, this article is
54 concerned specifically with regulatory TPS that authorize environmental impacts. The analysis
55 may be applicable to a range of other market mechanisms, from wetlands banking to severable
56 green product certificates (Carton 2016), but TPS are a distinct subset of market-based
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3 instruments (Hrabanski 2015). The label ‘market-based instruments’ encompasses everything
4 from environmental taxes to weather derivatives. As Hrabanski (2015, 143-144) notes, “MBIs
5 constitute an extremely heterogeneous group that make little sense from an economic theory
6 perspective... MBIs as a category look more like an asylum country for all tools with a price
7 component.”
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10 The policy mechanism of interest here are institutional arrangements wherein regulators
11 determine an overarching limit on the kind of behavior it is seeking to regulate, from the
12 discharge of nutrients into a watershed to the total catch in a fishery. The regulator will then
13 distribute the right to engage in that type of behavior as permits to polluters or resource users
14 depending on the environmental concern under management. Regulated entities will ultimately
15 have to control permits commensurate to the amount of that behavior they have engaged in, and
16 surrender them to the state¹. The primary attraction of TPS is that by making tradeable
17 representations of environmental damage, the distributed knowledge of market actors will be
18 harnessed, and desirable outcomes can be achieved at the lowest economy-wide cost (Tietenberg
19 2006). Secondly, the use of TPS is meant to introduce flexibility for regulated polluters, who can
20 choose whether to physically reduce their environmental impact or pay (in the form of
21 purchasing permits) someone else to reduce their share of impacts (Tietenberg 2006; Butraw,
22 Paul, and Woerman 2012). All tradeable permit programs are couched in flexibility for actors
23 engaged in environmental degradation, supposedly creating win/win outcomes of continuing or
24 increased economic activity in a regulated jurisdiction along with desired environmental
25 outcomes, despite the manifold environmental, accounting, and ethical problems that are baked
26 into ‘flexible mechanism’ (Lansing 2012; Bumpus and Liverman 2008; Bumpus 2011;
27 Dalsgaard 2013; Cavanaugh and Benjaminson 2014; Lovell 2010; Pearce 2013).
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32 The use of TPS has its origins in the 1970s for air pollution control in the United States and for
33 fisheries management in New Zealand (Lane 2012), two of the heartlands of the neoliberal
34 counterrevolution (Harvey 2007). In both settings, general antipathy toward traditional regulation
35 was growing along with belief that markets are always the most efficient means of organizing
36 society (Mirowski 2013). MBIs began to proliferate across jurisdictions and environmental
37 problems throughout the 1980s and 1990s, reaching the point in the early 2000s where the
38 approach could be said to be hegemonic across much of the world, led by initiatives such as the
39 European Union Emissions Trading Scheme (Lane 2012). In tallying the extent of TPS
40 (operational or otherwise), US EPA identifies 57 water quality trading markets, more than 60
41 carbon markets are planned or in operation (IETA 2014), individual tradeable quotas (ITQ) for
42 fisheries are in place across in at least 40 countries through 200 programs (Bonzon 2013), and
43 TPS have been deployed in matters as exotic as access to landfills (Barrow 2003).
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47 Environmental management through prices is not a new idea. It can be traced back to Pigou
48 (1920) and the taxation of environmental bads or providing subsidies for desired environmental
49 behaviors. Pigouvian programs have been, and continue to be, a key lever in the regulatory
50 toolbox to achieve environmental goals. The US Department of Agriculture’s soil conservation
51 program has been offering subsidies to farmers to reduce topsoil loss since the dustbowl in the
52 1930s.
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56 ¹ This part of the process demonstrates that the use-value of permits in market governance is regulatory
57 compliance and the potential to realize trading profits rather than attachment to any specific environmental
58 process (Smith 2006).
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3 Partially in response to the growth of Pigouvian taxes and subsidies, Coase (1960) proposed
4 bilateral trading of the right to impose negative outcomes on society among impacted parties.
5 Crocker (1966) and Dales (1968) built the intellectual scaffolding for TPS, reimagining Coaseian
6 bargaining among many polluters and a role for the state beyond the enforcement of property
7 rights and contracts. Dale's (1968) sketch of effluent trading demonstrated how distributed
8 decision-making about prices might achieve state-defined environmental objectives, though he
9 was careful to differentiate this suite of policies from 'conventional' markets. In a Dalesian
10 framework, costs would be minimized because government would inevitably over or undershoot
11 the 'correct' level of Pigouvian taxation, as it was incapable of determining demand for un-
12 degraded nature at any given moment. Instead, markets would have a "prime facie claim to
13 efficiency," (Dales 1968, 804).
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17 The results of attempts to create TPS based on this intellectual heritage can be judged in any
18 number of ways, but my approach here is to evaluate markets by the criteria of economics (see
19 Dempsey and Robertson, 2012). Ben-Porath's (1980, 4) definition, a common reference point for
20 ideal-type, neoclassical markets is "faceless buyers and sellers. . . [meeting] for an instant to
21 exchange standardized goods at equilibrium prices." That is, the key components of
22 marketization on which the logics of TPS rests are the presence and anonymity of buyers and
23 sellers, fully fungible commodities, and standardized contracts at a market-derived price. Brief
24 reflection yields serious challenges to this definition. Nature is not composed of endlessly
25 substitutable goods (Robertson 2006), willing buyers and sellers tend not to appear without state
26 inducement and/or existing supply and demand of underlying products (USEPA 2008), and
27 environmental regulation and market-making is nothing if not a protracted process of negotiation
28 (Pearce 2013). There are numerous treatments of the conceptual and practical flaws in TPS in
29 individual instances (e.g. Lohmann 2012, Mansfield 2004, Cooper 2015), but even so, some
30 cases bear a closer resemblance to the highly liquid, screen-tradable TPS envisioned by
31 contemporary proponents. This version of TPS itself represents a conceptual *cum* ideological
32 slippage. TPS were originally drawn up to be more purely regulatory programs operating over a
33 relatively small spatial extent with a relatively high number of participants, so trading would be
34 facilitated by regulated entities' familiarity with one another and with the regulator (Dales 1968;
35 Lane 2012). It is only alongside the mania for liquidity and faith in efficient markets over the
36 neoliberal period that market proponents have come to envision TPS engendering markets that
37 conform to Ben-Porath's market definition (Lane 2012). These examples are very much in the
38 minority, and yet, market-based environmental governance continues to proliferate. Thus, the
39 institutional and economic forms that result from their implementation necessitate closer
40 examination.
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47 In order to implement market-based environmental governance of any kind, an environmental
48 situation must be rendered a matter of concern (Latour 2004). If an environmental issue has not
49 been recognized as something to be dealt with, it is illegible to notions of scarcity and
50 distribution, and thus impossible to be treated as an economic input. Failure to render an issue a
51 matter of concern does not mean that it is not a problem, but that it has not been recognized as
52 such by those with sufficient power (scientific, political, or economic) to begin to adjudicate
53 access or behaviors (Callon 1998b). Structurally it represents the incapacity or unwillingness of
54 capital (and/or the capitalist state) to recognize, in this case, an environmental situation as a
55 threat to future accumulation.
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3 The first step on the continuum that environmental concerns can be managed are conventionally
4 defined command-and-control measures like quotas (non-tradeable) or technology standards that
5 allocate without reference to markets. We can think of this first step as 'economization', or "the
6 assembly and qualification of actions, devices and analytical/practical descriptions as 'economic'
7 by social scientists and market actors" (Çalışkan and Callon 2009, 369). In the modern policy
8 making process, beset as it is with statutory requirements for cost/benefit analysis,
9 economization is unavoidable. However, the details and modalities may differ significantly
10 depending on the matter of concern that is under consideration (Çalışkan and Callon 2009).

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13 Once rendered a matter of concern and sufficiently economized to be legible to the practices of
14 'economists in the wild' (Callon 1998), an environmental issue can become subject to the
15 assignment of prices. This can take place in many ways, such as equivalency in a non-capitalist
16 economy or administrative prices as in Pigouvian taxes. Examples include TPS where prices
17 ostensibly arise from market behaviors but are better understood as administrative prices
18 (discussed below). While price is a condition of market exchange, it does not necessarily signal
19 the presence or even possibility of a market, much less the highly liquid variety envisioned by
20 TPS backers.

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23 At this point, arrangements begin to look more like 'normal' markets. Prices may be influenced
24 by supply and demand relationships in illiquid markets. In many settings, illiquid markets will be
25 difficult to distinguish from bilateral (Coasean) arrangements, save the more visible role of the
26 state acting as a broker between parties, or more likely, acting as one of the parties, collecting
27 rents or distributing subsidies. There may be capital circulating between private sector actors, but
28 it is just as likely that the program will revert to Coasean bargaining or administrative prices, or
29 behave as a command-and-control mechanism with administrative prices and environmental
30 caps.

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33 Finally, we reach what proponents envision as marketization proper: highly liquid financial
34 markets that conform to Ben-Porath's definition, and subsequently, what we can think of as
35 financialization. Here, trading occurs without reference to underlying use value, it is a pure M-
36 M' circuit; the only reference is exchange values derivative from underlying financial products.
37 Cases like these are rare in the world of TPS, but do exist at some periods in time, such as the
38 outset of the European Union Emissions Trading Scheme. The economic theory underlying TPS
39 aims to create 'normal' markets, but the operation of these programs demonstrates that this is
40 difficult to achieve, much less sustain. The next section explores the reasons that TPS tend
41 toward command and control rules, administrative prices, or illiquid trading regimes rather than
42 the neoclassical markets envisioned by promoters because of that wobbly use value that turns on
43 the multiple, at times contrasting, motivations of the state to resolve socio-ecological
44 contradictions.

45 46 47 48 **Troubling markets**

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50 While there are numerous treatments of the conceptual and practical flaws in TPS in individual
51 instances (e.g. Lohmann 2012, Mansfield 2004, Cooper 2015), I do not contest the limited
52 potential for TPS to behave in a way that resembles the economic theory that animates them.
53 These arrangements are very much the exception, as evidenced by the many reasons that TPS fail
54 to engender the circulation of capital through price-clearing markets; the replicability and
55 durability of these arrangements are dubious. I explore the conditions for full marketization in
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3 the next section, but first look at the reasons that trading fails to materialize in the first place, or
4 that TPS revert to somewhat less-than fully marketized conditions after trading has taken place.
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6 *Lack of trading options*

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8 The first set of reasons why TPS may fail to circulate capital is that there are not enough parties
9 to transact with one another. The most straight-forward example of this kind of failure is a lack
10 of regulated entities (Atkinson and Tietenberg 1991). In this case, there is little reason to develop
11 screen-tradeable permits when they would still result in Coasean bargaining at best. Even where
12 the potential for trading exists, larger actors may prefer to invest in emissions reductions
13 technology while smaller actors fear becoming subject to the market power of their larger
14 counterparts (Suter et. al 2013). In this setting, larger regulated entities will have no possibility to
15 trade for compliance reasons because the smaller emitters cannot reduce their emissions enough
16 to create a surplus large enough to fulfil the needs of the larger regulated entity. Conversely, this
17 means that smaller regulated entities could become beholden to the prices that larger sellers can
18 set, leading all actors to treat a market as an emissions standard (Suter et. al 2013).
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22 One reason that a TPS may have a limited number of participants is the localized nature of many
23 environmental problems. While the indirect causes of environmental degradation are structural
24 (e.g. O'Connor 1994, Moore 2011a), the immediate causes and effects are often proximate,
25 requiring rigorous boundaries to achieve environmental outcomes, limiting the number of
26 potential trading partners. This is not an issue for the largest TPS, such as air emissions and
27 carbon markets that may have hundreds or thousands of regulatory participants, but presents a
28 serious challenge to marketization in, for example, watershed-scale water quality trading
29 markets. However, larger TPS (in addition to their smaller counterparts) face another spatial
30 problem called 'leakage', or the condition where an emitter flees a regulated jurisdiction to avoid
31 the costs associated with compliance. Leakage has the double impact of reducing demand for
32 permits and, in altering the supply/demand balance, reducing incentives for emissions reductions
33 among other polluters and attendant secondary trading activity (Cullenward 2014). It is, in effect,
34 a spatial fix to move where cheap natures remain available.²
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38 *Finite substitutability*

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40 Another factor that discourages trading are limitations on how fungible the natures in question
41 must be to be tradeable. This is true of situations where regulators attempt to create fungibility
42 between the emissions of point and non-point sources. For example, discharges in a watershed
43 with both wastewater treatment plants (point) and farmers (non-point) which means regulators
44 must account for potential inaccuracies in measurement from diffuse agricultural runoff
45 (Greenhalgh and Selman 2012). Many water quality trading programs have experimented with
46 trading ratios, where a more easily quantified volume of discharge must be compensated for by
47 two or three times that amount from a less easily surveilled emitter (Greenhalgh and Selman
48 2012). Further, equivalency may be complicated by the way liability is assigned for
49 environmental failures. In California's carbon market, if an offset is invalidated because, for
50 example, the forest that was planted to sequester carbon burns down, the buyer of the offset that
51 is liable for replacing the permits that represent the now-emitted greenhouse gases (California
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56 ² This is why market proponents (e.g. the International Emissions Trading Association (2014)), consistently
57 advocate for globally harmonized marketization, in addition to the potential for added liquidity owing to so many
58 mandated participants, as discussed below.
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3 2015). This makes the development of screen-tradable offsets more challenging because
4 purchasers incur costs for evaluating the quality and conditions of production of the offsets.
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6 *State-markets*

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8 The third, and most extensive, set of factors that inhibit high-level marketization arise from the
9 regulatory nature of markets. State-induced supply and demand is critical to understanding why
10 many programs fail to perform as envisioned. Blignaut and van der Elst (2014) note that political
11 uncertainty and financial risk are inextricably tethered in the creation of regulatory markets. The
12 range of policy uncertainties that can accompany any TPS are equal to the range of decisions that
13 go into crafting the rules of the program. These can pose risks for would-be participants,
14 including overpayment for permits that will subsequently erode in exchange values, or the
15 cancelation of the program without compensation. This uncertainty is common, as the US Clean
16 Air Act expressly notes that permits do not constitute a property right (Teinenberg 2006). The
17 non-property definition has filtered out into other TPS, including California's carbon market
18 (California 2015), damaging a permit's potential use value as representing the potential to extract
19 profit, leaving only regulatory compliance and discouraging secondary market activity.
20 Regulatory uncertainty is present across the full spectrum of MES and PES schemes, but is
21 particularly pronounced in TPS since the fundamental ambit of these programs is to induce
22 capital circulation which is inhibited by uncertainty.
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27 In addition to being the potentially unstable product of regulatory experimentation, these market
28 institutions feature prohibitions on behaviors that would be common in non-administrative
29 markets. For example, in the California carbon market there are price controls, permit holding
30 limits, restrictions on borrowing from future emissions allowances, and limitations on how
31 proceeds from trading gains can be spent for some market participants (California 2015). These
32 restrictions flag the variety of state motivations that are embodied in regulatory programs,
33 seeking to address environmental issues by facilitating capital circulation - just not too much of it
34 or in the wrong direction. Many of these prohibitions are a means of encouraging market
35 stability by limiting price volatility or the expressions of market power—in short, risks that come
36 with financial markets. However, these restrictions, particularly holding limits, significantly
37 constrain trading behaviors and are off-putting to speculators.
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40 Even if the infrastructure of a market is completely developed and participants feel confident
41 about the robustness of the program, there must be regulatory will and capacity to administer
42 them. This can prove challenging in smaller jurisdictions and for less high-profile environmental
43 issues. As Newburn and Woodward (2012, 157) explain in the context of water quality trading,
44 “[f]ailure to establish credible bounds on total nutrient load may also limit demand... [which]
45 has been a significant reason why even established programs have failed to generate many
46 trades.” This is particularly true of environmental regulations implemented during tough
47 economic times when regulatory stringency is perceived as being at odds with economic growth
48 (Lockhart 2016), rendering the goal of pricing cheap natures secondary to that of fulfilling
49 capital's expansionary imperative.
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53 Even in more liquid arrangements, regulatory uncertainty impedes market behaviors by
54 restricting demand. This case is playing out in California currently, as courts hear challenges to
55 the overarching legality of the program (Kasler 2017). Thus, participation in state-run auctions
56 for emissions allowances have dropped precipitously and prices, already at the regulatory-
57 defined floor price, have stayed constant, effectively creating a state-defined carbon price that
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3 behaves as a fee with limited secondary market activity. So even while the potential for
4 circulation exists, in practice the program is mostly administrative. Another example of this
5 complication is the US State of Virginia's experience rewriting rules, brokering agreements, and
6 making environmental compromises to its Chesapeake Bay effluent discharge program to attract
7 a new paper mill (Pipkin 2017). This experience also demonstrates the challenges to
8 marketization posed by the entrepreneurial state and the impetus toward growth, whereby
9 marketized nature is subordinated to wider political economic imperatives.
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12 An additional impediment to inducing trading behavior is the entirely state-decided supply-
13 demand relationship. One of the most frequent causes of 'incomplete' marketization is a weak
14 cap, like early phases of the EU emissions trading scheme, is set above the level of actually-
15 existing emissions within the regulated jurisdiction (Keohane 2009). In this case, TPS are de
16 facto licenses to pollute at business-as-usual levels, one of the ways the EU emissions trading
17 systems has operated, even while also functioning (sometimes) as financial market (Lohmann
18 2009). Where the cap is set below actual emissions levels, it may act as a performance standard
19 for polluters who prefer the certainty of reducing emissions rather than bear the risks of trading
20 (Suter et al 2013), a situation that is emphatically not the financialization of nature.
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23 A frequent topic of concern among program participants and their industrial organizations in the
24 development of California cap-and-trade was the level of oversight required of them (MSCG
25 2009). This oversight is exerted not only by the primary regulator, but includes surveillance by at
26 least six different regulators or contractors (California 2016). Oversight occurs at every stage of
27 market participation, from requiring multiple attestations for anyone involved with trading at
28 regulated companies to the reporting of every trade with regulators to ensure compliance with
29 permit holding limits. This level of scrutiny and the transaction costs it entails is off-putting to
30 potential speculators who would inject liquidity and, theoretically, move closer to the 'correct'
31 price of pollution (californiacarbon.info 2014).
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34 The regulatory supply/demand balance is further impacted by non-market regulations that
35 interact with the need to participate in markets. Especially in conjunction with unambitious caps,
36 the use of command and control mechanisms, pollution fees, or technology mandates all
37 potentially stifle trading. Extensive use of non-market mechanisms betrays a lack of total faith in
38 markets to rectify existing market failures (effectively the existence of cheap natures) on the part
39 of regulators. Some market proponents point to these regulatory interactions as a reason that TPS
40 are illiquid and, as per the neoliberal playbook, will claim that the only way to fix markets is
41 through more and 'freer' markets (Mirowski 2013). However, there is significant disagreement
42 among professional economists about regulatory interactions between overlapping policies and
43 the need to constrain market behaviors to ensure the stability of marketization (Gayer and
44 Horowitz 2006). Nevertheless, the urge to 'unleash' markets is one of the key rhetorical tools for
45 the ideological project of implanting market logics into environmental regulation in the first
46 place, which give proponents space to, "inscribe a distinction between economy and
47 culture/society and to conceptually separate an abstract perfect *Market* from concrete imperfect
48 *markets*. This allows market apologists to blame unwelcome external infringements (social,
49 cultural, political etc.) for "market failure" immunizing them from criticism" (Berdnt and
50 Boeckler 2009, 535).
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53 This approach to market governance tends to elide the fact no market exists in a regulatory
54 vacuum. TPS are grafted onto existing regulatory frameworks or implemented alongside other
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3 kinds of regulations in cases where an environmental issue is newly rendered a matter of
4 concern. For example, in fisheries regulations the state may try to achieve multiple
5 environmental and social goals by including temporal restrictions on fishing, by-catch
6 restrictions, hook limits, bans on certain fishing techniques, and export and processing rules
7 (Grafton and McIlgorm 2009). Similarly, California's cap-and-trade program is only one of
8 around 20 climate related regulations. These regulations range from other instruments that rely
9 on varying levels of marketization like the trading system for production of lower carbon fuels
10 that have tradeable certificates, to administrative measures like performance standards and green
11 energy quotas (CARB 2014). The result of these multifaceted interactions can reduce demand in
12 the aggregate, and for individual entities, result in a system of regulations in which trading is
13 only one route, and often the more complicated one, to compliance (Zhang et al 2015).
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18 **Higher order marketization**

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20 Despite all the attempts to 'internalize' the costs of degradation and come to grips with the
21 impending end of cheap nature through tradeable permit systems, there are cases of liquid TPS
22 that engender capital circulation through the creation of secondary and derivatives markets.
23 However, these examples remain relatively small and are very tightly regulated in comparison to
24 the wider world of financial markets. Globally, regulatory³ carbon markets were worth around
25 US\$50 billion in 2015, or about 15% of ExxonMobil's market capitalization - hardly the
26 behemoth that Richard Sandor, the founder of the Chicago Climate Exchange, predicted when he
27 said that, "[c]arbon will be the biggest commodity in the world by 2020" (in Erlich 2010).
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30 The examples of TPS that do work more closely to their own terms can be found among
31 tradeable quotas for large-scale fisheries management and the largest carbon and air pollution
32 markets. What these examples have in common are the large spatial extents over which they
33 work, the high number of regulatory-mandated participants, the relative fungibility of the
34 underlying matter of concern, and already existing demand for the commodities to which they
35 are tied. Both fish and carbon credits (of the non-offset variety) are highly commoditized,
36 reducing transaction costs among market participants. Further, particularly for fish, aggressive
37 caps spread over a large spatial extent and are tightly enforced which can induce demand for
38 permits that is absent in many other TPS. One reason for this is because these things (fish,
39 greenhouse gases, and urban air pollution) are related to existing product markets with definitive
40 relations of exchange. This contrasts with a water temperature market where the object of
41 management is a proxy environmental service for ameliorating other environmental problems,
42 and the emitters are a blend of public and private actors. Public actors have motivations beyond
43 profit maximization and regulatory compliance, meaning they may engage in effluent reduction
44 more willingly than private sector entities (USEPA 2008). Additionally, public goods are a more
45 challenging setting than private ones for engendering circulation in general (Costanza 2006;
46 Bakker 2009).
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51 In addition to being spatially extensive programs built on product markets with underlying
52 dynamics of supply and demand there are other factors that improve the chances for inducing
53 circulation in TPS. In addition to already being commodified, it helps if the object of
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56 ³ In distinction to voluntary carbon markets or 'baseline-and-credit' markets like the Clean Development
57 Mechanism of the Kyoto Protocols, which share many features of regulatory carbon markets but are distinct in
58 that neither has a state-defined cap, among other difference.
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3 management is plentiful and (relatively) easily made fungible. For example, the US acid rain
4 trading program encompassed a small range of pollutants of which there were many emitters
5 putting out many tons; there was little controversy over the one-to-one nature of individual
6 molecules (Stavins 1998). Some fish species are similarly commoditized and caught in sufficient
7 numbers to facilitate circulation of permits to catch them, and even in carbon markets that
8 encompass as many as six different greenhouse gases that have different properties, sufficient
9 work has gone into ‘making things the same’ (MacKenzie, 2008) such that equivalences are
10 relatively uncontroversial. Meanwhile, it would be difficult to achieve such a settlement if one
11 wanted to build a TPS for highly biodiverse urban boreal wetlands. There would be no
12 underlying ‘products’, boundaries to ensure fungibility would need to be tightly drawn, scientific
13 controversy over the bundle of ecosystems services provided would remain contested (or ‘hot’
14 (Callon 2009)), and supply and demand would be limited even if a sufficient ‘cap’ were
15 imposed. In short, the types of environmental degradation possible to confront with TPS are
16 extremely limited and has resulted in the development of other kinds of ‘market-like’ regulatory
17 programs (Laperye and Pirard 2013).
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22 **Hybridity, possibility**

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24 Given the variety of reasons that TPS may fail to engender neoclassical markets, and the
25 specificity of criteria that seems to encourage their function, it is unsurprising that the
26 deployment of TPS results in diverse economic and regulatory forms. In the context of
27 biodiversity offsetting, a different form of market-based instrument, Wynn-Jones (2013, 78)
28 shows that the outcome is often a “hybrid model of governance, blending market principals with
29 existing regulatory frameworks.” This is particularly true of environmental issues that are
30 already subject to non-market regulation. Higgins et al (2014, 464) describe this arrangement as
31 “a layering of instruments and approaches, whereby policymakers seek to inject more market
32 discipline into established procedures for procuring and contracting environmental services,
33 reshaping existing instruments to accord more closely with neoliberal notions of competition and
34 efficiency.” Higgins et al (2014, 464) go on to say that the process can also proceed in the other
35 direction, requiring “the reshaping of market-based policies instruments to make them workable
36 and acceptable within particular contexts.”
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40 For TPS, there is a third process by which hybridity occurs. In TPS, regulators build market
41 infrastructure, but they may not engender trading behaviors amongst their participants in ways
42 that result in ideal-type neoclassical markets. Alternatively, they may actual result in market-like
43 arrangements for a time, only to revert to administrative prices or performance/effluent
44 standards. For example, the California carbon market saw substantial trading activity at the
45 outset of the program, only to become more-or-less an administrative carbon price that acts much
46 like a carbon fee under a cap with significant regulatory interaction. This demonstrates that even
47 where the conditions are largely favorable to the use of TPS for making markets that correspond
48 to ideal-type economic theory, it is challenging to create durable conduits for capital circulation.
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51 This situation, in and of itself, is not a problem, and indeed is likely desirable if key provisions
52 are in place. The cap does the environmental work, a price on emissions ought to act as a
53 disincentive for emitting, especially when bolstered by a regulatorily-determined price floor, and
54 the infrastructure of TPS requires the state to develop surveillance mechanisms that may not
55 have existed (Felli 2015; Dempsey and Robertson 2012). However, for the proponents of
56 marketization, the creation of unidirectional transfers or the state being placed at the center of
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3 circulation violates one of the aims of TPS, which is to facilitate the circulation of capital
4 between emitters and through capital markets more broadly. Instead, TPS often result in
5 unidirectional transfers of capital to the state (tax), or a transfer from one emitter to another with
6 the state acting as a broker (tax and subsidy), and in this way, cheap nature is accounted for to
7 some degree, even if not to a degree that effects material changes in the economy.
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10 The critique of TPS as a unidirectional payment in the form of a tax has continuously been
11 leveled at the California cap-and-trade market (Carrol 2016). Regulators and legislators claim
12 that cap-and-trade is not a revenue program for the state; however, the program has brought in
13 more than US\$4 billion in the first three years of the program, which the state then distributes in
14 the form of subsidies and infrastructure spending to meet its climate goals (Carrol 2016). The
15 development of a TPS is an arduous way to implement a revenue program for the state. Stephson
16 and Shabman (2011, 21) argue that, “pursuit of revenue through trading is likely to divert policy
17 makers’ time and attention from more effective means of financing... reductions,” a sentiment
18 that rings true across the world of TPS because of the associated administrative costs. However,
19 the widespread deployment of TPS over these objections highlights the power of market
20 discourse over environmental regulation, and so regulators find themselves going through the
21 motions of market-making to achieve any level of environmental protection.
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25 Enormous work goes into building infrastructure for TPS, whatever their outcome. This
26 resonates with Graeber’s (2015, 7) ‘Iron Law of Liberalism’, which states that rather than
27 reducing bureaucracy, every attempt to create markets results in the expansion of the
28 administrative apparatus. TPS demonstrate that creating a liquid financial market is not
29 necessary for bureaucratic expansion—the discourse alone is sufficient. In California, the
30 creation of the cap-and-trade market required the hiring or repurposing of two hundred
31 environmental regulators along with countless hours of work from NGOs, regulated industry,
32 legislative assistants, academic economists, third party data providers, financial infrastructure
33 firms, and industry associations.
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36 Some regulators feel as though their time would be better spent directly regulating whatever
37 matter of environmental concern they are tasked with. In the context of controlling nutrient
38 runoff through WQT, a USEPA report says that, “the most efficient tactic may be to simply
39 direct funds to non-point source control, either through grants to farmers (as currently exists) or
40 through development of a regulatory program that explicitly controls agricultural runoff,” (US
41 EPA 2008, 3-27). This is a common theme across TPS even though one of the justifications for
42 the use of TPS is administrative simplicity. For example, in tradeable fisheries quotas, “the
43 OECD found that management costs rose in 18 out of 23 fisheries that implemented ITQs”
44 (Grafton and McIlgorm 2009, 715).
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47 While this analysis has focused on the processes of deploying economic thought for the
48 marketization of environmental problems, the *raison d’être* of TPS is to achieve environmental
49 goals. Despite the time and money that have gone into developing TPS for water quality in the
50 United States, one out of 57 existing programs improved the watershed (USEPA 2008, 3-10),
51 while falling emissions under the European carbon trading market are mainly attributable to the
52 financial crisis (Bailey et. al 2011). SO_x and NO_x trading, TPS backers’ shining example,
53 benefited greatly from changing political economic conditions, and what reductions were
54 achieved occurred slowly compared to those accomplished by other countries that implemented
55 regulations without trading (Lipow 2007).
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Conclusion

TPS rarely result in the highly liquid, price clearing financial markets we might envision when we think of, for example, warnings about subprime carbon (Chan 2009) or the real subsumption of nature by capital (Smith 2006). While TPS do not need to generate the circulation of capital to achieve their (generally modest) environmental goals, trading is the primary selling point of using market mechanisms because of the flexibility it affords regulated polluters and the economic efficiencies gained by allowing capital to flow to least-cost emissions reductions. Instead, these programs often behave more like command and control regulations or administrative prices, failing to meet their emic goals of market-determined prices of environmental degradation through liquid trading and circumventing inefficient state decision making. Despite the ongoing challenges facing TPS, proponents continue to advocate for these policy mechanisms as the best way to achieve environmental outcomes at the lowest economy-wide cost (IETA 2014). As capital, and most capitalist states, comes to recognize the magnitude of the threat to future accumulation posed by our sundry environmental crises, policy experimentation with mechanisms that promise to resolve environmental contradictions while functioning as new circuits of capital are likely to remain appealing. As such, it behooves us to not only understand how these programs fit within the economy as a whole, but also the ways TPS function in unique settings, but with key similarities.

By toggling between performativity approaches and green political economy, I have tried to demonstrate how structural impediments to building markets in environmental degradation manifest in different ways and in different kinds of nature. Regulators often must balance goals when designing programs to find solutions to capital's environmental contradictions, the biophysical properties of the concern under management may be recalcitrant to commodification, and the regulatory products of the application of neoclassical theory to those biophysical phenomena are difficult to stabilize to the extent they achieve a use value other than regulatory compliance. While different TPS fail to create liquid financial markets for different, and at times idiosyncratic reasons, the overarching lesson is that, as a mechanism, TPS are having a difficult time coming to grips with the end of cheap nature and capitalism's environmental contradictions.

However, even if TPS do not result in institutional forms that facilitate the circulation of capital through nature, or do not result in the circulation of private capital if they do achieve some level of circulation, they can potentially discipline targeted polluters to technocratic, if inefficient, management regimes, providing environmental regulators with mechanisms by which to pursue their goals within the constraints in which they find themselves (Dempsey and Robertson 2012) and create rents for the state (Felli 2014). Building on Callon (2007), Berndt and Boeckler (2010, 7) note that practices of making markets leads to "the possibility of co-construction of economy and politics... due to their incompleteness and dynamics markets trigger the emergence of matters of concern... spur the proliferation of new social identities and trigger the creation of unexpected social communities." Geographers could use the incompleteness of market making to insert themselves into policy processes, demonstrating ways that the mechanisms developed for TPS can be used to other ends (Bigger and Robertson 2017). To be sure, there are cases where financial practices are influencing our relationships to non-human nature in ways that look like the markets of neoclassical environmental economists, as in the growing market for green debt (Bracking 2015). Wall Street is indeed a way of organizing nature, and it is increasingly trying to account for the end of cheap natures (Moore 2011b, 2015). We must remain attuned to ways

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3 nature and capital are being (re)produced through financial and regulatory practices—it just may
4 be that TPS are not the mechanism by which this occurs.
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