

1 **The multitaxa functional diversity increases the resilience of biological natural capital in** 2 **the Amazon**

3

4 **Abstract**

5 1. The resilience of biological natural capital in the Amazon is strongly influenced by the
6 functional diversity of multitaxa, which promotes the stability and sustainability of the
7 ecosystem. However, biological natural capital, driven by multiple dimensions of
8 biodiversity, is often underestimated by focusing on taxonomic dimension, such as richness
9 and abundance.

10 2. We applied a multitaxa and trait-based approach to assess how functional diversity
11 supports natural capital. Using a dataset including woody plants, bees, frugivorous
12 butterflies and songbirds, we created two indices: Biological Natural Capital Resilience
13 (BNCR), the proportion of biological assets needed to maintain natural capital integrity, and
14 Biological Natural Capital Uniqueness (BNCU), the proportion of natural capital sustained by
15 species' functional uniqueness.

16 3. BNCR results show that in general more than 80% of biological assets to maintain the
17 integrity of natural capital, and that the BNCR index reached the highest value when all
18 taxonomical groups are combined. These results reveal that natural capital strongly depends
19 on biological assets and underscore the importance of multi-taxa approaches for assessing
20 ecosystem resilience.

21 4. BNCU values show that a substantial portion of natural capital integrity (from 46.22% to
22 64.23%) depends on a small subset of functionally unique species, indicating that the loss of
23 few species, which play irreplaceable ecological roles, rapidly reduces ecosystem
24 functioning.

25 5. Synthesis and applications: This study presents a trait-based framework integrating
26 functional diversity into biological natural capital accounting, capturing both ecosystem
27 resilience and species uniqueness across multiple taxa. It provides ecologically grounded,

28 standardized, and reproducible indicators that move beyond single-taxon or purely
29 taxonomic approaches. By embedding these indicators into natural capital accounts, the
30 framework supports conservation prioritization, restoration planning, impact assessment,
31 trend monitoring, and scenario testing, offering a practical methodology to evaluate and
32 manage the multiple dimensions of biodiversity within natural capital assessments.

33

34 **Keywords:** *Biological Assets; bees; butterflies; birds; flora; Functional Diversity; National*
35 *Forest of Carajás*

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40 **Introduction**

41 The concept of natural capital recognizes the value of natural resources in relation to a
42 product or service, and was explicitly developed in the 1970s and popularized in the 1990s
43 (Gómez-Baggethun & Groot 2010, Ovando 2021). In economics, "capital" refers to the stock
44 of materials or information that generate a flow of services that provide benefits to people
45 (Costanza et al. 1997). Thus, natural capital expresses the stock of renewable and non-
46 renewable natural resources that provide benefits to humanity through ecosystem services
47 (Costanza & Daily 1992, MEA 2005). The natural capital account is a practical approach to
48 environmental management that focuses specifically on understanding and documenting
49 the state of nature and measuring and assigning value to ecosystem services (Oliver 2018,
50 Ovando 2021). It contributes to setting environmental policy goals, because it identifies the
51 status of an environmental asset and shows whether the current use of natural resources is
52 sustainable over time (Bright et al. 2019). In addition, assessing the state of natural capital is
53 a compelling tool for conservation, communication, and awareness of regional natural
54 heritage (Oliver 2018). This approach also places natural heritage at the same level of other
55 forms of capital (e.g., financial, industrial, social, and human) and underscores its

56 importance to the economic sector (Oliver 2018). Natural capital has also become relevant
57 to the private sector (Guerry et al. 2015), as an important tool for assessing natural resource
58 business dependence, risk analysis, and supply chain management (Natural Capital Coalition
59 2016). Therefore, the concept of natural capital has emerged as a valuable framework for
60 understanding and valuing the stock of renewable and non-renewable natural resources,
61 and rises as instrumental to the management of natural resources. However, it is crucial to
62 recognise the potential pitfalls associated with the concept of natural capital, primarily the
63 reduction of nature to monetary values, which leads to the undervaluation of essential
64 assets and the oversimplification of ecosystems (Mace, 2019).

65 The United Nations Statistics Division (UNSD) has been leading the development of the
66 System of Environmental-Economic Accounting - Ecosystem Accounting (SEEA EA)
67 framework, which aims to standardize the quantification of ecosystem asset stocks, such as
68 biodiversity, and ecosystem service flows over time (UNCEEA, 2021). Biodiversity is an
69 essential ecosystem asset of natural capital, since it encompasses all forms of diversity of
70 living components of ecosystems, and its multidimensionality (e.g. taxonomic diversity,
71 functional diversity, genetic diversity and diversity of interactions) makes it a key element at
72 all levels of the ecosystem function hierarchy, both as a regulator of ecosystem services, as a
73 final ecosystem service, or as a good of valuation (Mace et al. 2012). In SEEA EA, biodiversity
74 is defined according to the Convention on Biological Diversity (Secretariat of the Convention
75 on Biological Diversity 2020) as "biological diversity means the variability among living
76 organisms from all sources," and its components are classified into three categories:
77 ecosystem diversity, species diversity, and genetic diversity. In natural capital accounting
78 biological natural capital specifically refers to the stock of biological components, or
79 biological assets, such as species, traits, and ecological functions that contribute to
80 ecosystem services (Mace 2019). Despite its importance for natural capital, current
81 estimates of biological natural capital often focus on biomass or species richness and rarely
82 incorporate other dimensions of biodiversity (Smith et al. 2017). Part of the difficulty in
83 integrating these aspects, particularly dimensions beyond the taxonomic, comes from the
84 fact that their values mainly reflect ecological significance rather than producing estimates
85 directly actionable for management, leading to underestimation of ecosystem resilience and
86 species' contributions to ecosystem functioning (Mace 2019)). Moreover, while SEEA-EA

87 recognizes the many aspects of biodiversity, its primary focus is on ecosystem diversity,
88 giving broad principles for accounting for biodiversity rather than detailed guidance on what
89 sorts of data should be utilized and evaluated (Bradon et al. 2021).

90 In order to increase the valuation of biodiversity in natural capital accounting, researchers
91 have been advocating for approaches that integrate multiple elements of biodiversity, such
92 as interaction diversity, functional diversity, and genetic diversity, as fundamental
93 components of natural capital accounting (Coates et al. 2020). In this sense, Brand (2009)
94 proposed that among the dimensions that should be considered in natural capital
95 accounting are those related with the ecological resilience and the vulnerability of the
96 natural capital assets. Functional traits and functional diversity are key to ecological
97 resilience, with similarities among species creating functional redundancy that buffers
98 ecosystems against disturbances, and unique traits indicating vulnerable functions.
99 Together, redundancy and uniqueness offer valuable tools for assessing biodiversity,
100 ecosystem integrity, and adaptive capacity, often overlooked in natural capital accounting
101 (Ricotta et al., 2016; Díaz & Cabido, 2001; Brand et al., 2009; La Notte et al., 2017).
102 Incorporating functional diversity into natural capital accounting potentially provides critical
103 insights into ecosystem integrity, the capacity to resist and adapt to global changes, and the
104 sustainable long-term use of natural resources, while improves the assessment of
105 ecosystem contributions - enabling more accurate evaluations of the resilience and
106 vulnerability of natural assets for conservation and management decisions.

107 Amazon forest is the world's largest tropical forest, housing more than a third of the world's
108 biodiversity and providing essential ecosystem services to humanity at local, regional and
109 global scales (Mittermier et al. 2003). Despite this importance, high deforestation and
110 habitat degradation rates have been causing biodiversity loss in Amazon, affecting the forest
111 resilience to human activities and climate change (Levine et al. 2016), compromising the
112 forest's capacity to sequester carbon (Gatti et al. 2022) and changing its hydrological cycle,
113 with potential negative consequences for the climate at global scale (Lovejoy & Nobre
114 2019). Amazon biodiversity contains a large diversity of unique traits in the community, as a
115 consequence of the evolutionary strategy of strong differentiation between species (Kraft et
116 al. 2008). Studies in tropical rainforest ecosystems, including the Amazon forest, have

117 shown that rare and highly functionally unique species often contribute disproportionately
118 to ecosystem functioning, meaning that the loss of a few functionally unique species can
119 lead to a significant reduction in ecosystem stability and natural capital integrity (Leitão et
120 al. 2016; Dee et al. 2019).

121 Given the high biodiversity and ecological complexity of the Amazon, this study uses a multi-
122 taxa, trait-based approach to evaluate how functional diversity reflect the resilience and
123 uniqueness of Amazonian natural capital under species loss scenarios. Specifically, we
124 investigate the extent to which functional redundancy buffers ecosystem integrity and how
125 functional uniqueness shapes the vulnerability of these communities. Therefore, we
126 propose the following testable hypotheses: (1) the resilience of natural capital in Amazon is
127 mainly sustained by biodiversity, meaning most of biological assets need to be maintained
128 to conserve natural capital integrity; (2) due to the high uniqueness of these Amazon
129 communities, a substantial portion of natural capital integrity relies on a relatively high
130 number of unique species, making natural capital in Amazon particularly vulnerable to
131 species loss.

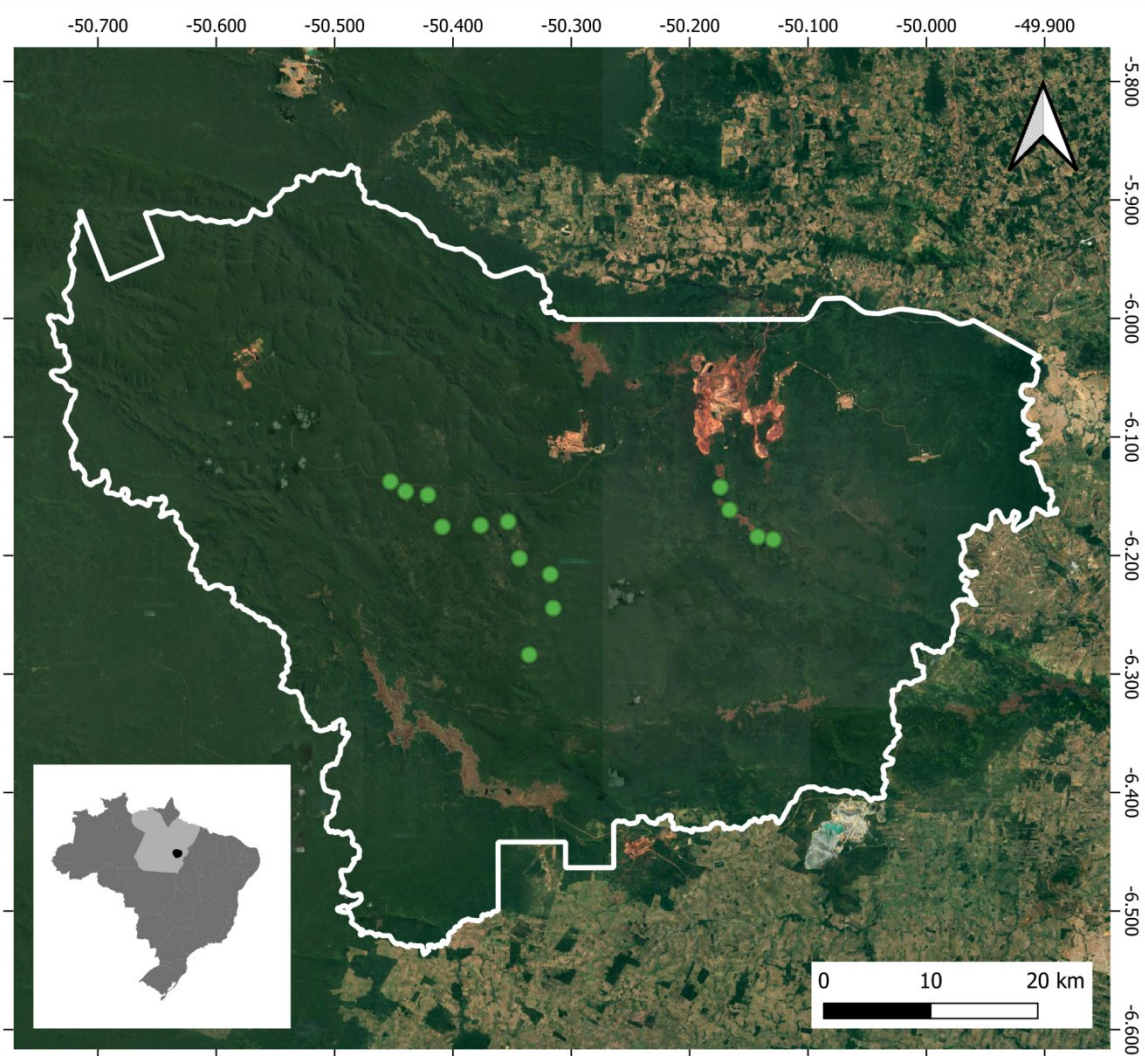
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133 **Material and Methods**

134 *Study area*

135 The study was carried out in the National Forest of Carajás, Eastern Amazon (Pará State).
136 This forest is a sustainable use protected area covering more than 3,930 km² and its
137 vegetation is characterized mainly by an Open Ombrophilous Forest, with local variations
138 related to changes in soil and relief (Barbosa-Silva et al. 2022). In the steep areas, the "forest
139 with lianas" predominates - characterized by a medium biomass with low density, with more
140 light penetration in its interior. The region of the study area stands out for its significant
141 natural capital in terms of biodiversity, carbon stock, water regulation, climate regulation
142 and food security (Neugarten et al. 2015) . However, little is known regionally about its role
143 in natural capital, highlighting both its great potential for climate regulation via
144 evapotranspiration (Pontes et al. 2017) and the vulnerability of its biodiversity-driven
145 ecosystem services under scenarios futures (Costa et al. 2018). Our study does not aim to

146 comprehensively sample or quantify all biodiversity or natural capital of the whole Amazon.
147 Instead, we focus on a representative local pristine dataset to explore fundamental
148 mechanisms linking functional diversity to the stability of biological natural capital.
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150
151 Figure 1. Study area and sampling locations. The white line delineates the boundaries of the
152 Carajás National Forest, and the green dots represent the sampling points. The background
153 image is a satellite view from Google Earth. The inset map (bottom left) shows the location
154 of the study area within Brazil: the dark grey area represents Brazil, the light green indicates
155 the state of Pará, and the black dot marks the location of the Carajás National Forest.

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157

158 *Biodiversity sampling*

159 In order to evaluate the biodiversity assets in the natural capital account, we conducted a
160 comprehensive survey of four groups of organisms across 14 pristine forest sites that were
161 at least 1.5 km apart in November 2019 and March, April, May, August and September
162 2022. These dates were selected avoiding the coldest period of the year ,when biodiversity
163 levels tend to be lower due to reduced biological activity, and respecting safety protocols
164 due to the COVID pandemic during the research period. The selection of these groups was
165 based on their significance in providing and sustaining ecosystem functions. Bird
166 communities were examined by detecting their songs through the use of Audiomoth 1.0.0
167 digital audio recordings (Hill et al, 2019) at each of the 14 sampling sites. The devices were
168 positioned in the undergrowth of the forest, approximately 1.5 m above ground level. For
169 bird identification, we selected minutes between 5:00 and 9:00 and between 17:00 and
170 19:00 from the recordings, and vocalizations were identified by a specialist. Frugivorous
171 butterflies (Nymphalidae) were sampled using Van Someren-Rydon model traps baited with
172 a fermented mixture of banana and beer for 48 hours, at the same 14 sampling sites. We set
173 up two traps by sampling site, positioned in both the understory (ca. 1.5m above ground)
174 and near the canopy (ca. 20-30m above ground). Each trap was monitored every 48 hours
175 and kept at each location for 6 days, following the methodology outlined by Uehara-Prado
176 et al. (2005). After identification, the specimens were placed in the entomological collection
177 of the Museu Paraense Emílio Goeldi (MPEG) and identified with the help of literature that
178 included plates with color photographs, as well as "Butterflies of America" (Warren et al.,
179 2016). To sample bees, we utilized three sampling methods: aromatic traps, honey traps and
180 flower visitors. Aromatic traps were used to attract orchid bee males with four artificial
181 essences (Eucalyptol, Vanillin, Eugenol and Methyl Salicylate). Honey traps were used to
182 attract foraging bees from both the understory and the forest canopy. A honey and water
183 mixture (1:1) was sprayed in the vegetation and all bees collected. To sample flower visitors
184 we observed all open flowers and collected all visitors for two non-consecutive hours per
185 sample in each site. Three samples were carried out in each site. Bees were identified by
186 specialists and deposited in MPEG. Lastly, to sample woody plants, we conducted species

187 surveys in 20 m x 100 m plots at six of the 14 sampled sites followed by random collections
188 in the vicinity of the plots to supplement data on fruiting plants, which are essential for
189 accurate identification by taxonomists. All plants have their diameter at height breast (DBH)
190 and canopy height measured, and we kept the records of those with DBH > 10cm. Plant
191 exsiccatae were also deposited at the.

192 For all taxonomic groups, we aggregated all spatial and temporal sampling data into a single
193 pooled community, representing the total set of species detected across all sampling points
194 and dates. While these communities do not aim to represent the full regional species pool
195 of each taxonomic group, it represents from 71% to 93% of the estimated total species
196 richness for the surveyed taxa (Supplementary Material I). This approach allows us to
197 analyze overall functional patterns at the landscape scale, while acknowledging the
198 limitations inherent to sampling effort, detectability, and seasonal variation.

199 *Functional Diversity Indexes*

200 We use functional diversity measures to infer the biological resilience and uniqueness of
201 natural capital maintained by biological assets. These measures are based on the diversity of
202 functional traits of communities, which in turn represent the diversity of ecosystem
203 functions and services performed by communities in the ecosystem. For each taxonomic
204 group sampled, we construct a database of functional traits, using both sampled species
205 morphological measures and information from a free database. For bees, the selected traits
206 were body mass, tongue length, distance between the inner margins of the eyes, spectrum
207 of pollen hosts, level of sociality, and recognition of the species as a crop pollinator. For
208 Lepidoptera, the selected traits were body size, proboscis length, forewing width, length
209 from thorax to abdomen, presence of camouflage colours in the wings, presence of
210 blemishes in the wings, iridescence in the wings or dorsal body parts, distance from the top
211 to the bottom of the eye, and number of host plant species. For birds, the selected traits
212 were body mass, Hand-Wing Index, habitat score, dietary composition, and foraging vertical
213 distribution. For flora, the selected traits were type of dispersal and pollination syndromes,
214 specific leaf area, wood density, and leaf phosphorus and nitrogen concentrations. Trait
215 information for bees was gathered from open databases and literature, except for
216 intertegular distance (used to estimate body size via allometric equations), which was

217 measured by researchers. For birds and flora, all trait information was obtained from open
218 databases, while for butterflies, all traits were measured directly by researchers, except for
219 the number of host plant species, which was obtained from the literature (Supplementary
220 Information II).

221 Species sampled whose functional traits were not measured or found in the database were
222 excluded from the community for the calculation of functional diversity. For continuous
223 functional traits, we used the mean value of the trait for each species, both in the case of
224 measurements on the collected specimens and in the case of multiple measurements per
225 species in the literature. From these trait databases, we calculate the Functional Richness
226 (FRic), which represents the amount of functional trait space occupied by the community
227 (Villéger et al. 2008). Species with available functional trait data represented 88.69%,
228 84.12%, 96.29%, and 91.29% of the total species richness of bee, bird, frugivorous butterfly,
229 and tree communities, respectively. In terms of total abundance, these species accounted
230 for 90.31% (bees), 80.40% (birds), and 98.86% (frugivorous butterflies). For trees, species
231 with trait data contributed 65.81% of the total basal area, calculated based on the diameter
232 at breast height (DBH) of each individual.

233 To calculate FRic, for each group we first normalized the continuous traits and log
234 transformed them to avoid the influence of variables with high standard deviation and
235 extreme values (Legendre & Legendre 2019). We then performed a Principal Coordinate
236 Analysis (PCoA) with three dimensions (scores) for each taxonomic group to standardize the
237 numbers dimensions among taxonomic groups. In all groups, we used the modified Gower
238 distance (Pavoine et al. 2009), with equal weights for all traits, to calculate the PCoA,
239 considering the different nature of the variables (continuous, proportional, or categorical).
240 Categorical variables were transformed into binary variables using the package "ade4" (Dray
241 & Dufour 2007) and proportional variables were considered as fuzzy variables. The PCoA
242 procedure allowed us to integrate numerical and categorical variables and to reduce and
243 standardize the number of dimensions between taxonomic groups. For all subsequent
244 analyses, we use the three PCoA scores as functional traits, as proxies for the communities'
245 functional space.

246 *Biological resilience and uniqueness of natural capital*

247 From the functional diversity indices calculated for the study sites, we derive the biological
248 resilience and uniqueness of the natural capital through simulations of the loss of biological
249 assets. Both resilience and uniqueness measures were based on the relation between
250 biological assets and the integrity of natural capital sustained by biodiversity. For this we
251 assumed the species richness as a proxy for the biological assets of natural capital, once it
252 represents the biodiversity stock that sustains ecosystem functions and services, and is
253 widely adopted in ecological assessments and policy frameworks (Coates et al. 2020). We
254 also defined the integrity of natural capital as the capacity of biodiversity to support a broad
255 range of ecosystem functions and services. To quantify this, we used Functional Richness
256 (FRic) as a proxy of the integrity of natural capital, since it captures the extent of functional
257 trait diversity within a community - higher FRic reflects a wider occupation of the functional
258 niche space, indicating greater ecological role diversity and, consequently, a more
259 functionally robust and resilient system (Villéger et al. 2008).

260 Through a random simulation of species loss, the resilience index proposed, named
261 Biological natural capital Resilience (BNCR), depicts the percentage of total biological assets
262 (species richness) that should be maintained to preserve the integrity of natural capital. This
263 index was computed individually for each taxonomic group sampled, considering all species
264 as the whole collection of biological assets. To do this, we consider that the values of FRic
265 calculated for the whole community, representing 100% of the integrity of the natural
266 capital supported by the biodiversity of each taxonomic group. We also consider, for each
267 taxonomic group, the total number of species as 100% of the biological assets of natural
268 capital.

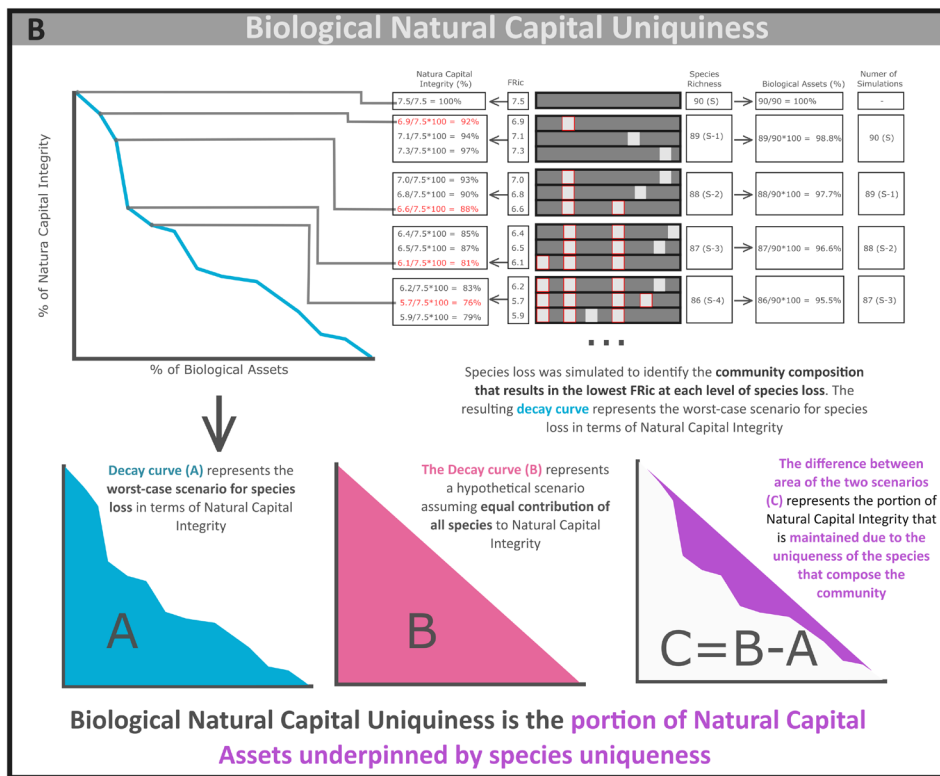
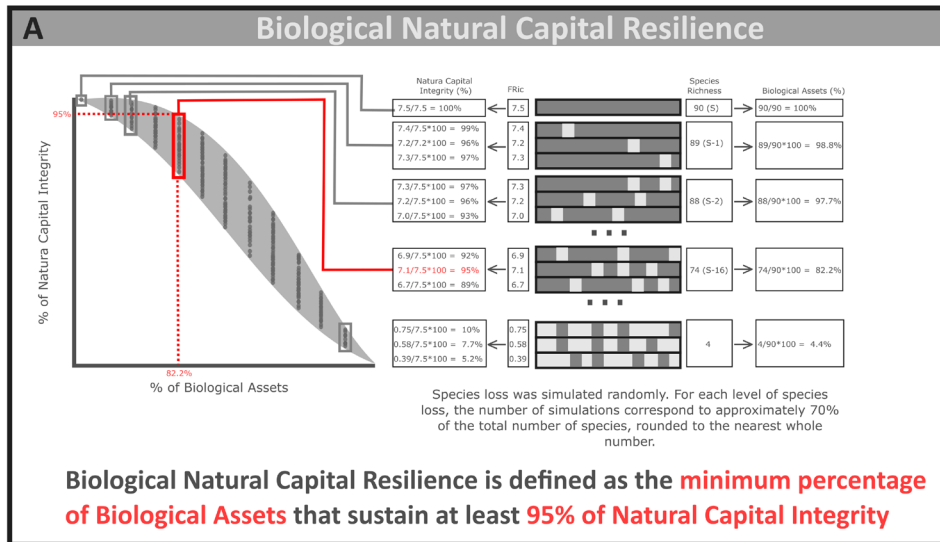
269 To assess BNCR, for each taxonomical group, we simulated communities with random
270 biological asset loss (species richness) and for each simulated community we calculated
271 functional richness (FRic) standardized by the global FRic (FRic calculated with all species)
272 for comparison (See Supplementary Information III, for simulation details). For each
273 simulated community, the number of remaining species was converted to the percentage of
274 total species and considered as the percentage of natural capital integrity. We then define
275 Biological Natural Capital Resilience (BNCR) as the minimum percentage of biological assets

276 that retains 95% of total FRic - which implies that a community that maintains 95% of its
277 functional diversity is considered as intact as a community with 100% of its functional
278 diversity (Figure 2A). While this 5% threshold is extensively supported by statistical
279 literature (Cowles & Davis 1982, Murtaugh 2014) - based on frequentist statistics where a
280 difference of less than 5% between populations of values represents a non-significant
281 difference - it is possible to establish other values according to the objectives of natural
282 capital accounting. This includes raising the threshold to higher values (more conservative
283 goal) or lower values (more permissive goal).

284 In addition to BNCR for each taxonomic group, we also calculated the multitaxon BNCR by
285 integrating all taxonomic groups and performing random simulation. For each taxonomic
286 group both species richness and FRic were expressed as a percentage of the group's total
287 values, thereby standardizing contributions from groups with different richness and
288 functional scales. For each level of species loss, we combined the previously within-group
289 simulations across all taxonomic groups, using the percentage of remaining species as the
290 proportion of biological assets and the corresponding sum of FRic as the percentage of
291 natural capital integrity. The multitaxon BNCR was then defined as the minimum proportion
292 of biological assets required across all groups to maintain at least 95% of total natural
293 capital integrity, based on these standardized percentages. This approach preserves the
294 relative contribution of each group to overall functional diversity while avoiding bias caused
295 by differences in absolute FRic values.

296 To assess Biological Natural Capital Uniqueness (BNCU), we ran simulations of species loss
297 again, this time not randomly, but assuming a worst-case scenario of natural capital
298 integrity loss. Thus, for each level of biological assets loss (species loss), we calculated the
299 FRic value for all possible communities and selected the one with the lowest FRic value. This
300 value was standardized by global FRic of the taxonomic group and considered as the lowest
301 percentage of natural capital integrity for that level of loss of biological assets. The species
302 excluded at a given level of biological asset loss was also excluded at subsequent loss levels.
303 Considering then the decay curve between the percentages of biological assets and natural
304 capital integrity lost we then calculated the area under the curve (AUC) of the decay curve.
305 This calculated area is then subtracted from the AUC of a hypothetical decay curve in which

306 all species contribute equally to the integrity of natural capital, implying that the loss of a
307 specific quantity of biological assets results in the same amount of natural capital integrity
308 loss. We consider the difference between this and the AUC to represent the ratio of natural
309 capital integrity supported by the uniqueness of species composing the natural capital
310 assets. In terms of calculation, once both natural capital integrity and biological assets are
311 represented as percentages, this hypothetical AUC corresponds to an isosceles right-angled
312 triangle with an area of 0.5. Therefore, NBCU was calculated as 0.5 minus the AUC of the
313 worst-case scenario, then multiplied by 2 to standardize the values between 0 and 1, where
314 values close to 0 indicate an equal contribution of species to maintaining natural capital
315 integrity, while values close to 1 indicate a high uniqueness contribution of species in
316 sustaining natural capital integrity (Figure 2B).



317

318 Figure 2. Metrics of Biological natural capital Resilience and Uniqueness. Top: Resilience is
 319 the minimum proportion of species loss that still maintains 95% of natural capital integrity
 320 (82.2% in this example). Bottom: Uniqueness measures functional integrity loss under the
 321 worst-case species loss scenario (blue) versus equal contribution scenario (pink). The purple
 322 area (C) quantifies the integrity attributed to unique species.

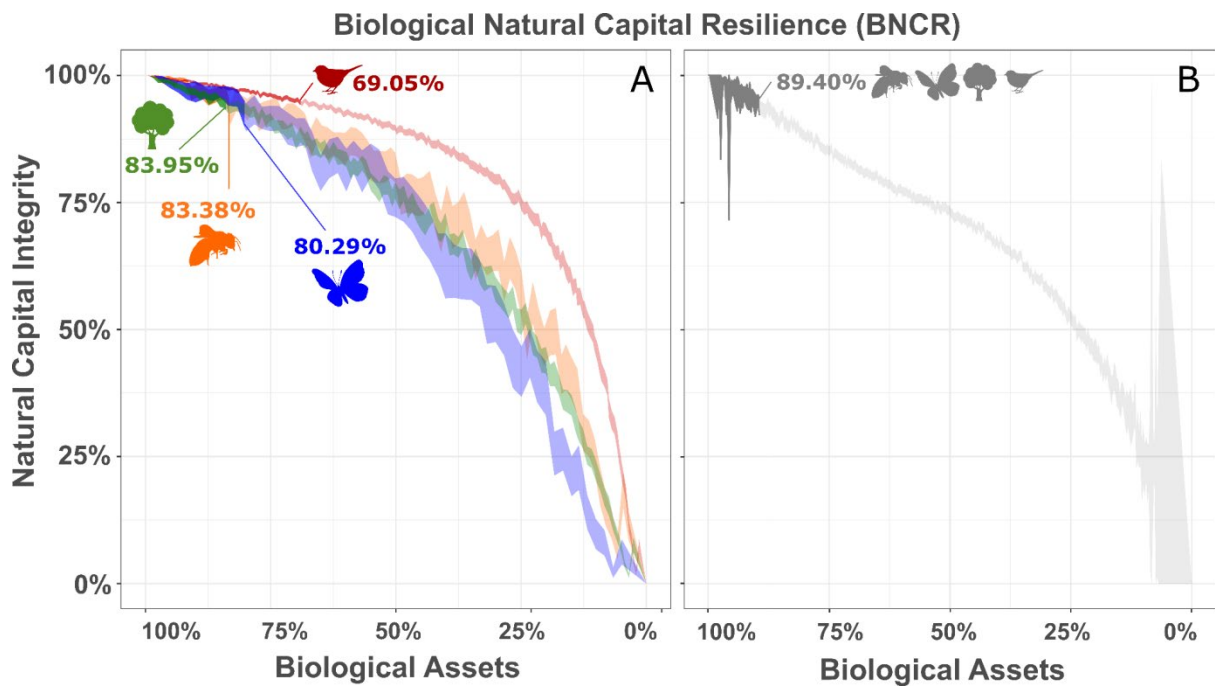
323 **Results**

324 We constructed indices of resilience and uniqueness of natural capital supported by
325 biodiversity based on woody plants, bees, frugivorous butterflies, and songbirds. Our
326 dataset comprises a total of 3642 specimens, including 1886 woody plants, 154 frugivorous
327 butterflies, and 1580 bees. Additionally, we recorded 4895 birdsongs. The total of 667
328 species sampled includes 287 woody plant species, 259 bird species, 54 butterfly species,
329 and 121 bee species.

330 For taxonomic groups, the BNCR index, was higher in woody plants (83.6%), followed by
331 bees (83.3%), frugivorous butterflies (80.2%), and birds (69.05%), while the multi taxa
332 analysis of BNCR showed the highest value of resilience (89.55%; Figure 3).

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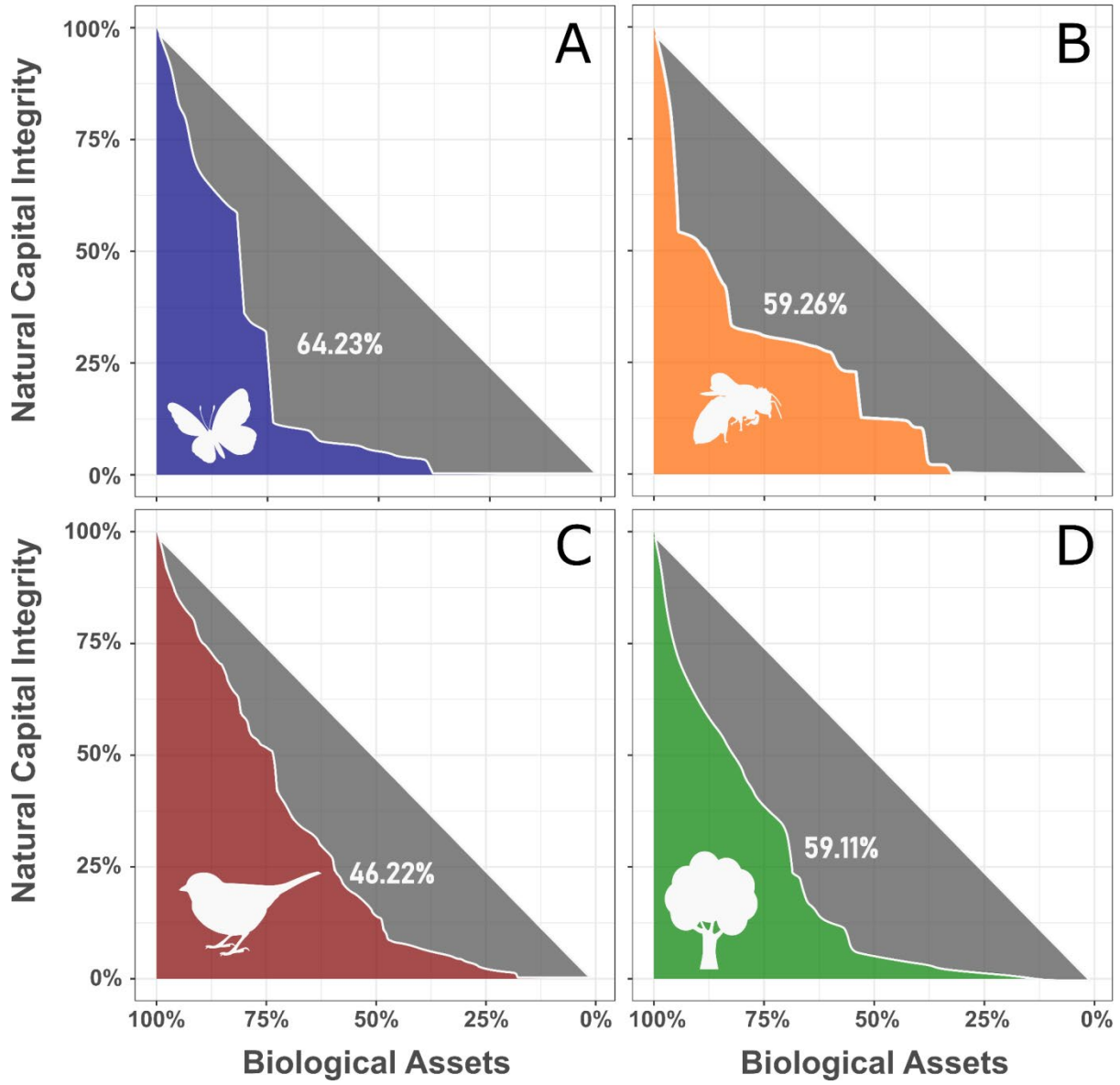
336 *Figure 3. The biological natural capital resilience (BNCR). For each taxonomic groups (A) and*
337 *multi taxa (B). In each graph, colored polygons indicate the range of the percent natural*
338 *capital integrity maintained by communities with a random simulated percentage of*
339 *biological assets loss. Dark colored polygons indicate communities that maintain at least*
340 *95% of the natural capital integrity, while light colored polygons indicate communities that*
341 *maintain less than 95% of the natural capital integrity. The percentage indicated in each*
342 *graph and the boundary between light and dark polygons indicate the minimum percentage*
343 *of biological assets that sustain 95% of the natural capital integrity, named Biological*
344 *natural capital Resilience (BNCR).*

345

346 For BNCU, the frugivorous butterflies were the biological group with the highest BNCU index
347 value (64.2%), followed by bees (59.2%), woody plants (59.1%) and birds (46.2%) (Figure 4).

348

Biological Natural Capital Uniqueness (BNCU)



351 *Figure 4. The biological natural capital uniqueness (BNCU) of the four taxonomic group.*
352 *Frugivorous butterflies (A), bees (B), birds (C) and woody plants (D). In each graph, the*
353 *colored polygon represents the decay curve of biological assets in the worst-case scenario of*
354 *loss of integrity of natural capital, while the gray triangle (part covered by the colored*
355 *polygon) represents the decay curve of biological assets in a scenario where all species*
356 *contribute equally to the integrity of natural capital. Gray polygons represent the difference*
357 *between the two scenarios, and its percentage in relation to the total area of the gray*
358 *triangle (expressed numerically in the graph), reflects the portion of natural capital that is*
359 *maintained by the unique contribution of species (BNCU).*

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363 **Discussion**

364 Our findings show that both resilience and uniqueness metrics reveal critical patterns about
365 how biodiversity supports natural capital. BNCR values were consistently high across
366 taxonomic groups, supporting our prediction that natural capital in these systems are mainly
367 underpinned by biodiversity, and that most biological assets need to be maintained to
368 conserve natural capital integrity. Also aligned to our hypothesis, BNCU values were also
369 elevated, indicating that a substantial portion of natural capital integrity relies on a
370 relatively high number of unique species, expressing the high irreplaceability of many
371 species within the studied communities. These results indicate limited functional
372 redundancy among species, confirming our expectation of high uniqueness and showing
373 that the loss of just a few species leads to an abrupt decline in natural capital integrity.
374 Importantly, rather than offering generalized patterns for the entire Amazon biome, the
375 study provides ecologically meaningful insights for the focal region and a replicable
376 framework for other biodiversity-rich areas. These outcomes underscore the value of
377 incorporating trait-based biodiversity measures into conservation frameworks, particularly
378 those focused on assessing natural capital. Ultimately, the observed BNCR and BNCU values
379 across distinct taxonomic groups indicate that the integrity of natural capital relies heavily

380 on preserving both the quantity and the uniqueness of biological assets. These metrics
381 provide a relative assessment of resilience and irreplaceability, highlighting that even small
382 losses of unique species may disproportionately affect natural capital integrity.

383 BNCR values varied among taxonomic groups, reflecting ecological and functional
384 differences inherent to each group. Woody plants exhibited the highest BNCR (83.95%),
385 followed closely by bees and butterflies, while birds had the lowest (69.05%). These
386 differences likely stem from a combination of factors, including trait composition, species
387 richness, and ecological roles. For instance, woody plants, shaped by diverse ecological
388 adaptations to environmental pressures such as drought and nutrient variability, exhibit a
389 wide range of functional traits that make their contributions to ecosystem integrity unique
390 (Swenson et al. 2012, Rossel et al. 2022), reflected in the high BNCR value, where even small
391 losses in biological assets can disproportionately impact ecosystem stability. Similarly, Bees
392 and butterflies, especially due to their species-level specialization in foraging and resource
393 partitioning, such as the reliance of butterflies on distinct host plants or habitats, showed
394 high BNCR values, reflecting a strong dependency on multiple functionally distinct species to
395 sustain key ecological processes (Ye et al. 2024, Ramos et al. 2021). In contrast, birds,
396 particularly in species-rich tropical regions, often exhibit high functional redundancy,
397 wherein many species share overlapping traits and ecological roles, allowing key ecosystem
398 functions to be maintained even if some species are lost, which is reflected in their
399 comparatively lower BNCR values (Cooke et al. 2019, Jarzyna et al. 2021). The multitaxa
400 BNCR exhibited the highest functional resilience value, suggesting that integrating multiple
401 taxonomic groups may reveal greater vulnerability of the overall system. This pattern likely
402 reflects emergent properties of ecosystem resilience, supporting previous calls for multi-
403 taxa analyses to fully understand ecological vulnerability (Capdevila et al 2021, Seddon et al.
404 2021). Differences in trait composition, redundancy, and species richness among groups can
405 reduce apparent redundancy when combined, while integrating multiple ecological scales
406 highlights the disproportionate impact of losing highly unique species (Fischer & de Bello
407 2023) . Part of the high value may also reflect the large sample size and standardization
408 across groups, which allows meaningful comparisons but can influence absolute
409 percentages (Dalerum et al 2012).

410 BNCU values revealed that a considerable share of natural capital integrity relies on a
411 relatively small subset of highly unique species. This index simulates a worst-case scenario in

412 which species are lost sequentially from the most functionally distinct to the least, an
413 unlikely but critical lens for assessing ecological vulnerability. Under this pessimistic
414 assumption, ecosystem functions degrade rapidly, as the initial species lost are precisely
415 those contributing the most unique functional roles. Our results show BNCU values ranging
416 from 46.2% in birds to 64.2% in frugivorous butterflies, underscoring significant variation in
417 functional distinctness among groups. Frugivorous butterflies likely exhibit high functional
418 distinctness due to unique adaptations feed and flight shaped by environmental filters along
419 gradients (Henriques et al. 2022). These adaptations reflect their specialization to specific
420 habitat conditions and resource availability, making some species disproportionately
421 important for maintaining natural capital integrity, as captured by their high contributions
422 to the BNCU index. Similarly, bees showed high BNCU values, reflecting strong species-level
423 specialization in foraging and pollination interactions (Mokkapati et al. 2024). In contrast,
424 birds exhibited the lowest BNCU, reflecting the high functional redundancy typical of
425 tropical assemblages, where many species share similar ecological roles. This pattern may
426 arise from dense niche packing and substantial trait overlap, meaning that increases in
427 species richness do not necessarily translate into broader functional diversity (Pigot et al.
428 2016, Cooke et al. 2019). As a result, the loss of even functionally distinct bird species may
429 be partially buffered by the presence of ecologically similar species, leading to a more
430 gradual decline in natural capital integrity. Woody plants had intermediate BNCU values,
431 likely reflecting a balance between high species richness, which can lead to functional
432 redundancy, and the presence of species with distinct traits that contribute uniquely to
433 ecosystem functioning in tropical environments (Sun et al. 2024). The decay curves derived
434 from BNCU simulations highlight the steep consequences of losing unique species: a 25%
435 loss in biological assets led to declines of up to 88.5% in natural capital integrity. This
436 reinforces the need to preserve functionally irreplaceable species, as their loss triggers rapid
437 functional collapse. Importantly, functional uniqueness offers a complementary perspective
438 to richness-based conservation, allowing the identification of species whose marginal
439 contributions to ecosystem functioning are disproportionately high. Thus, BNCU provides a
440 valuable framework not only for estimating worst-case functional loss, but can also
441 informing conservation prioritization by quantifying each species' irreplaceability within the
442 natural capital system.

443 Integrating functional diversity into natural capital accounting addresses a key limitation of
444 traditional methods, which often rely solely on taxonomic diversity and overlook the varying
445 contributions of species to ecosystem function. Functional traits offer a process-based
446 perspective, emphasizing the roles species play in key ecosystem processes such as nutrient
447 cycling, productivity, and resilience. Our approach, aligned with recent calls to incorporate
448 these dimensions (Mace et al. 2019; Ricotta et al. 2016), employs interpretable metrics
449 (BNCR and BNCU) that are applicable at local or regional scales without requiring reference
450 areas. By converting functional diversity metrics into measures of natural capital, our
451 approach allows for a more direct accounting of how biodiversity supports ecosystem
452 services, by selecting functional traits related to these relevant ecosystem services (for
453 example, wood density for timber production, or bee proboscis length for flower
454 reproduction). An essential step in a functional approach involves the careful selection of
455 functional traits, taking into account the specific ecological processes, services, and
456 responses to which these traits are related (Violle et al., 2007). This represents an
457 opportunity for a targeted functional approach to natural capital accounting, as it allows for
458 the selection of traits associated with ecosystem services and highly valued benefits, as well
459 as those that indirectly support ecosystem services and processes that are difficult to
460 express in monetary terms (Mace et al., 2019). It is particularly important, especially in a
461 multi-taxon approach, to carefully consider the number, resolution, and nature of functional
462 attributes, aiming for a balanced representation across taxonomic groups. In this regard, we
463 also recommend analyzing metric sensitivity to functional traits (Supplementary Information
464 IV). For example, the relatively low BNCR values of birds may partially result from coarser
465 trait resolution in global databases. Whenever possible, we recommend using locally
466 collected, high-resolution traits, which better capture functional variation and improve the
467 accuracy and interpretability of functional diversity metrics.

468 Another valuable advantage of the proposed approach lies in its relevance beyond
469 traditional diversity measures. The indices used are ecologically meaningful and easily
470 interpretable within the context of natural capital - expressing both the percentage of
471 biological assets required to sustain natural capital and the proportion of capital supported
472 by the uniqueness of these assets - enhancing their integration into broader natural capital
473 frameworks, bridging ecological insight with economic valuation. Their expression as

474 percentages also allows assessment of a single area independently, without requiring
475 external comparisons. In this context, it is important to interpret BNCR and BNCU values
476 within a relative framework rather than relying on absolute thresholds. For BNCU, a
477 conceptual reference scenario corresponds to a community where all species contribute
478 equally to ecosystem functioning, similar to assumptions made by species-richness based
479 metrics. Values deviating from this scenario indicate elevated functional uniqueness. For
480 BNCR, however, no neutral scenario exists: when species are not functionally redundant,
481 any loss of species proportionally reduces functional richness, making high or low BNCR
482 inherently context-dependent. Thus, “high” or “low” values are interpreted relative to the
483 observed patterns within each community, allowing us to assess the resilience and
484 vulnerability of natural capital in a descriptive and methodologically meaningful way. It is
485 important to note that accurate interpretation of these indices also depends on sufficient
486 sampling completeness, as incomplete surveys may underestimate functional richness or
487 uniqueness, and assess sampling coverage (Supplimentary Information I) ensures that
488 pooled communities reasonably represent the local species pool, increasing confidence in
489 the derived BNCR and BNCU values. Despite these advantages and the benefits of
490 standardization, it is essential to be cautious when comparing values across taxonomic
491 groups or regions. While natural capital accounting is primarily designed for spatially explicit
492 assessments of ecosystem stocks and service flows within defined territories (Hein et al.
493 2020; Arguillo et al. 2022), it can also be used to compare natural capital across regions
494 under specific conditions. Such comparisons require standardized and ecologically
495 meaningful metrics, consistent data quality and taxonomic resolution across sites, and
496 harmonized definitions of ecosystem boundaries and functional units (Maechler & Graz
497 2020).

498 Our study demonstrates that the proposed metrics, BNCR and BNCU, effectively capture
499 critical dimensions of ecological resilience and species uniqueness within the natural capital
500 framework. The variation observed among taxonomic groups likely reflects underlying
501 ecological filters and evolutionary adaptations that shape species’ functional traits and
502 roles. This variation allows the approach to differentiate groups that are more or less
503 susceptible to biodiversity loss and functional decline, providing nuanced insights into
504 community vulnerability. Importantly, adopting a multitaxa perspective reveals the

505 compounded fragility of natural capital integrity when multiple taxonomic groups are
506 considered together, emphasizing the need for integrative conservation efforts that
507 transcend single-taxon focus. The BNCR and BNCU metrics complement each other by
508 quantifying both the resilience of ecosystem functions to species loss and the
509 irreplaceability of unique functional contributions, respectively. Their standardized and
510 interpretable nature enables their incorporation into natural capital accounting, bridging
511 ecological theory with applied management. By integrating functional diversity into natural
512 capital assessments, this approach expands the scope beyond traditional taxonomic
513 measures, which often underestimate the value of biological assets. This opens
514 opportunities to include a wider array of functional traits and other biological assets,
515 providing a more realistic and comprehensive valuation of natural capital. Ultimately,
516 embedding functional ecology within natural capital frameworks supports more informed
517 and adaptive biodiversity conservation and land-use planning. It enables prioritization
518 strategies that consider not only species richness but also the functional roles critical for
519 ecosystem resilience and sustainability, advancing the integration of ecological complexity
520 into natural capital management and policy.

521

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526 **Author contributions**

527 Felipe Martello and Tereza Cristina Giannini conceptualized the research; Felipe Martello
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529 Amanda Paracampo; Rafael Cabral Borges; Ulysses Madureira Maia; Sidnei M. Dantas and
530 Leonardo de Sousa Miranda conducted the investigation; Tereza Cristina Giannini
531 supervised the research, Felipe Martello and Tereza Cristina Giannini wrote the main
532 manuscript; Caroline Oliveira Andrino; Rafael Gomes Barbosa-Silva, Ulysses Madureira Maia,

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539 The authors declare that there are no conflicts of interest

540

541 Data availability statement

542 Data available from the Oxford University Research Archive repository (ORA)

543 <https://ora.ox.ac.uk/objects/uuid:a2ee43d8-9fef-47fd-87f4-8f672e591e8c>

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