1	Running title: Foraging-mediated hydras in epidemics
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3	Virulent disease epidemics can increase host density
4	by depressing foraging of hosts
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26	
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42 Abstract

All else equal, parasites that harm host fitness should depress densities of their hosts. 43 However, parasites that alter host traits may increase host density via indirect ecological 44 interactions. Here, we show how depression of infected host foraging rate can produce such a 45 hydra effect. Using a foraging assay, we quantified reduced foraging rates of a zooplankton host 46 47 infected with a virulent fungal parasite. We then parameterized a dynamical model of hosts, parasites, and resources with this foraging function, showing how foraging depression can create 48 a hydra effect. Mathematically, the hydra arose when increased resource productivity exceeded 49 any increase in resource consumption per host. Therefore, the foraging-mediated hydra effect 50 more likely emerged (1) for hosts which strongly control logistic-like resources and (2) during 51 larger epidemics of moderately virulent parasites. We then analyzed epidemics from 13 fungal 52 epidemics in nature. We found evidence for a foraging-mediated hydra effect: large outbreaks 53 depressed foraging rate and correlated with increased densities of both algae and hosts. 54 Therefore, depression of foraging rate of infected hosts can produce higher host densities even 55 during epidemics of parasites that increase host mortality. Such hydras might prevent collapse of 56 host populations but also could produce higher densities of infected hosts. 57 58

Keywords: foraging depression, host-parasite, host-resources, hydra effect, trait-mediated
indirect effect, density-mediated indirect effect, compensatory population growth, illnessmediated anorexia

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Introduction

63	Disease epidemics can drive declines in host populations (Anagnostakis 1982; Daszak et
64	al. 1999; Frick et al. 2010; Lessios et al. 1984), trigger conservation crises for wildlife such as
65	mammals (Roelke-Parker et al. 1996) and birds (Cooper et al. 2009; Hochachka and Dhondt
66	2000), and even sometimes drive hosts extinct (amphibians: (Vredenburg et al. 2010)). Disease
67	outbreaks can also damage economically valuable crops (Fry and Goodwin 1997) and livestock
68	(Cleaveland et al. 2001). Even worse, climate change can further exacerbate disease epidemics
69	(Altizer et al. 2013; Sanderson and Alexander 2020; Shocket et al. 2018). Therefore, it is
70	imperative to identify when, where, and why parasites depress density of their hosts during
71	epidemics.
72	Typically, we predict that parasites depress host density because infection exacts virulent
73	costs to host fitness. Indeed, infection often can increase mortality rate and/or decrease fecundity
74	of infected hosts. Simple disease models illustrate how those two factors can lower host density
75	relative to disease-free conditions (Anderson and May 1979; Anderson and May 1981).
76	Furthermore, that harm can become amplified by higher transmission of disease (which can lead
77	to higher prevalence of infection). Higher transmission results from higher per capita exposure
78	and/or susceptibility (the product of which is called 'transmission rate' (Dwyer and Elkinton
79	1993; Strauss et al. 2018)). Additionally, higher transmission can occur in more enriched systems
80	that support higher density of hosts (assuming density-dependent spread of disease: (Johnson et
81	al. 2010)). Therefore, we might expect larger absolute and/or relative depression of host density
82	when virulent parasites reach higher prevalence.

83 On the other hand, this above outcome might reverse when infection depresses foraging 84 rate of hosts. Many parasites lower foraging rate of hosts (Hite and Cressler 2019; Hite et al.

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2020; Strauss et al. 2019). At first glance, such foraging depression — whether a defense 85 strategy or fitness cost of infection (Hite et al. 2020) — might seem to exacerbate declines of 86 host density during epidemics. After all, lower intake of energy, when coupled with reduced 87 survivorship and/or fecundity from infection, might harm fitness of hosts even more (all else 88 equal). However, we show that foraging depression can sometimes *increase* host density through 89 90 a hydra effect (Abrams 2009). This outcome requires that hosts must strongly control a dynamic resource, and that the resource must reach highest productivity at intermediate density (e.g., 91 growing logistically or logistic-like). When those conditions are met, foraging reduction can 92 increase density, and hence production, of resources through indirect feedbacks. Those increases 93 in resource production can then compensate for increased energetic demands of hosts (a 94 consequence of virulence). When the shift in resource production exceeds that in resource 95 consumption, a foraging-mediated hydra effect emerges, leading to higher host density in the 96 presence of parasites — even during very large outbreaks. 97

98 Here, we illustrate this foraging-mediated hydra mechanism using a freshwater system with a zooplankton host (*Daphnia dentifera*). This host can strongly depress an algal resource 99 that reaches highest production at intermediate density. This host becomes infected by a fungus 100 101 that predominantly lowers survival (Hall et al. 2010) rather than fecundity (unlike, e.g., the bacterium *Pasteuria*: (Auld et al. 2012)). In this study, we demonstrate that infection also lowers 102 foraging rate of hosts in an experiment, particularly as the final transmission stage of the fungus 103 104 (spores) accumulate within the body cavity. We then parameterized a foraging depression function and incorporated it into a dynamical model. The model revealed how epidemics can 105 drive higher host density (similar to how predators can increase prey density through changes in 106 107 foraging behavior: (Peacor and Werner 2001)). This foraging-mediated hydra effect becomes

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more likely as epidemics become larger (e.g., with higher accrual of spores within hosts and 108 higher carrying capacity of the resource) and with stronger foraging depression. Conversely, it 109 becomes less likely with higher virulence on survival. Finally, a survey of fungal epidemics in 110 lakes showed that larger epidemics (with greater infection prevalence) yielded higher parasite 111 production per host. We estimated the depression in foraging due to disease in those lakes, and 112 113 found that lower foraging correlated with joint increases of algal and zooplankton populations during epidemics. Taken together, this combination of experiments, dynamical modeling, and 114 field surveys demonstrates how foraging depression can increase host density during epidemics 115 of parasites that kill their hosts. 116 117 Study System and a Function for Foraging Depression (a model competition) 118 Disease system 119 The focal host, the zooplankton Daphnia dentifera, strongly grazes on phytoplankton in 120 many lakes throughout the Midwestern USA (Tessier and Woodruff 2002). Hosts ingest 121

infectious propagules (spores) of the parasitic fungus *Metschnikowia bicuspidata* while foraging
on small (< 80 µm) phytoplankton (Hall et al. 2007b). As the parasite fills its host's hemolymph
with spores (Ebert 2005; Green 1974), it reduces host growth, fecundity, and survivorship (Hall
et al. 2009b). Death of the infected host releases spores into the water to then infect new hosts.
Sometimes, epidemics of this fungus reduce host density and indirectly increase density of the
algal resource via a trophic cascade (Duffy 2007; Hall et al. 2011). At other times, host density
remains high during epidemics (Duffy and Hall 2008; Hall et al. 2011).

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130 Foraging rate experiment: methods

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We estimated foraging rate using an experiment, summarized only briefly here (see 131 Appendix Section 1 for details). We measured feeding rate on individuals of cohorts of 132 uninfected and infected hosts. To create a gradient in body size (host length, L_H) and spore 133 accumulation (σ), we measured food consumption by individuals of progressively older (and for 134 the infected class, more infected) age cohorts. Thus, we placed individual hosts into small tubes 135 containing their algal food, allowed them to graze for a short period of time (four hours), 136 measured remaining food using a fluorimeter, and estimated length using a dissecting 137 microscope. We ensured infection status by smashing hosts to release spores contained in their 138 body. These spores represent the final life stage of the parasite; their presence indicates terminal 139 infection (Stewart Merrill et al. 2019). 140

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142 Functions for foraging depression: candidate models

We statistically competed models linking spore accumulation (σ) and body size (host 143 144 length, L_H) to per capita 'foraging' rate, $f(L_H,\sigma)$ for three host genotypes. The candidate models for foraging rate, $f(L_H, \sigma)$, varied in complexity (Table 1). In model 1 (*null*), per capita foraging 145 rate (f) is a single parameter (f). In model 2 (*size only*), a size-specific foraging coefficient (\hat{f}) is 146 multiplied by L_{H}^{2} , proportional to surface area, a common grazing model (Hall et al. 2007b; 147 Kooijman 2010). In model 3 (spores only, linear) and model 4 (spores only, power), foraging 148 rate drops as spores fill body volume, $\propto L_H^3$ (i.e., as σ/L_H^3 increases, governed by coefficients α 149 and γ). Models 3 and 4 both assume foraging does not scale with surface area. The most complex 150 variants, model 5 (size and spores, linear) and model 6 (size and spores, power), combine 151 surface area with the spore-mediated foraging depression. Models 1a-6a were fit assuming a 152 shared foraging coefficient (f or \hat{f}) for both infected and uninfected classes. For models 1b–6b, 153

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we estimated parameters f_j or \hat{f}_j separately for infection class *j*. We inserted the best fitting function, assuming constant host length, into the dynamic epidemiological model (see *Dynamical Model* below; Figs. 2,A3). Additionally, the winning function enabled us to estimate depression of foraging during epidemics (see *Field Survey* below; Figs. 5,A4).

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159 Parameterization and competition of the foraging function

We used maximum likelihood and information theoretic methods to parameterize and compete the foraging models, implemented with Matlab (version 7.8 R2009a; MathWorks). We estimated parameters by fitting a model of algal loss through time due to foraging (Sarnelle and Wilson 2008; Strauss et al. 2019):

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$$\ln(A_t) = \ln(A_0) - f(L_H, \sigma)t_E/V + \varepsilon, \tag{1}$$

where A_t is the concentration of algae remaining at the end of the grazing period of length t, A_0 is the concentration of algae in ungrazed reference tubes at time t_E , $f(L_H, \sigma)$ is one of the foraging models (Table 1), V is the volume in the tube, and errors (ε) were normally distributed. (While it is technically 'clearance rate', the 'foraging' label avoids confusion with the immunological meaning of clearance.) We estimated parameters using maximum likelihood and competed models using standard information criteria (by calculating AIC, Δ AIC_{*i*}, and Akaike weights, w_i for each model: see Table 1 (Burnham and Anderson 2002)).

172 The two best fitting models, 5b and 6b, fit the data equally well ($\Delta AIC < 1$, Table 1, Fig.

A1). We chose the more parsimonious model 5b (*size and spores, linear*) as the winner. We

estimated 95% CI around the parameter of with 10,000 bootstraps (Table A1). We also compared

parameters between host genotypes using 9,999 permutations (Gotelli and Ellison 2004) (Table

176 A1). The slope and intercept of a regression of observed vs. predicted $\left[\ln(A_0/A_t)V/t\right]$ remaining

algae were close to 1 and 0, respectively, indicating good performance (*Observed* = $1.007 \times$ 177 *Predicted* $-0.056 + \varepsilon$; $R^2 = 0.55$) (Piñeiro et al. 2008). 178

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Outcome of the competition among foraging functions: results 180

Parasite infection reduced foraging rate of hosts, particularly during later stages of 181 infection (Fig. 1a-c). In those later stages, fungal spores filled the body cavity of its host (Fig. 182 1d-f). Furthermore, infection stunted body size of sick hosts relative to uninfected hosts of the 183 same age and genotype (Fig. 1d-f). As a result, the best fitting function for parasite-induced 184 foraging depression (the 'size and spores, linear' model 5b; Tables 1,A1 and Figs. 1,A1) was: 185

186 For uninfected hosts:
$$f_S(L_H) = \hat{f}_S L_H^2$$
 (2.a)

187 For infected hosts:
$$f_I(L_H, \sigma) = \hat{f}_I L_H^2 \left(1 - \alpha \left(\frac{\sigma}{L_H^3} \right) \right)$$
 (2.b)

where \hat{f}_S and \hat{f}_I are 'size-specific' (size-independent) per capita foraging rates; L_H is host body 188 length; σ is the spore load per infected host; and α is a linear sensitivity coefficient that governs 189 depression of feeding as spores fill the host's body cavity. These equations (equ. 2.a,b) capture 190 how body size increased foraging. For uninfected hosts, foraging scaled with surface area (\propto 191 L_{H}^{2}) (equ. 2.a; Fig. 1: solid lines, white circles). For infected hosts (equ. 2.b; Fig. 1: dashed lines, 192 black circles), for aging rate also increased with surface area, though at a slower rate (since $\hat{f}_I <$ 193 \hat{f}_{s} ; Table A1) but it decreased as their body volume ($\propto L_{H}^{3}$) filled with spores (σ) (Fig. 1). 194 195

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Dynamical Model: Foraging Depression Can Produce a Hydra During Epidemics Structure of the dynamical model 197

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We inserted the winning foraging function into a dynamical model. This model could

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then delineate conditions leading to foraging-mediated hydras vs. trophic cascades during 199 epidemics (equ. 3, Table 2): 200 Susceptible: $dS/dt = e(f_S S + f_I I) A - dS - u f_S S Z$ (3.a)201 $dI/dt = u f_{\rm s} S Z - (d + v) I$ Infected: (3.b)202 $\frac{dZ}{dt} = \sigma(A) (d + v) I - m Z - (f_S S + f_I I) Z$ Propagules: 203 (3.c)where $\sigma(A) = \frac{\sigma_1 A}{h+A}$ (3.d)204 $dA/dt = r_m A \left(1 - A/K\right) - \left(f_S S + f_I I\right) A$ (3.e)205 Resources: Susceptible hosts (S, equ. 3.a) feed non-selectively at rate f_S on algal resources (A); infected hosts 206 (I) feed at reduced rate f_{I} . Feeding rates followed the best fitting foraging model described above 207 (equ. 2.a,b). For simplicity, we assumed that hosts feed with a linear functional response. 208 Ingested food is converted into offspring with efficiency e. Susceptible hosts (S) then die at 209 background rate d or become infected following exposure (at rate f_S) to spores (Z), with per spore 210 susceptibility *u*. Infected hosts (*I*; equ. 3.b) die from infection (at enhanced rate d + v); they 211 cannot recover. Spores (Z; equ. 3.c) are released from dead hosts; spore yield, $\sigma(A)$, increases 212 213 with algal resources (A) but saturates (with maximum σ_1 , and half-saturation constant h: equ. 3.d). Spores are lost at background rate *m* and via consumption by both host classes. Algae (equ. 214 3.e) grow logistically (at maximum per capita rate r_m and carrying capacity K) and are consumed 215 by both host classes. 216 We simulated the model over a range of algal carrying capacity, K, and sensitivity of 217 spore production to resources, σ_1 . We parameterized it using biologically reasonable values for 218 this system (Table 2) and estimates of \hat{f}_s , \hat{f}_l , and α for the BD-30 genotype (equ. 2, Fig. 1, Table 219 A1; assuming adult size $L_H = 1.4$ mm for both uninfected and infected hosts). Qualitatively 220 Penczykowski et al. 10 Foraging-mediated hydras in epidemics

similar results emerge using parameters from other genotypes. This dynamical model is not 221 analytically tractable, thus we simulated it (using a standard adaptive step integrator in Matlab) 222 for 1000 days. We then averaged densities of the state variables from t = 1000-2000 days. In the 223 focal, biologically relevant region of parameter space shown here, the state variables reached a 224 stable steady state by this time period. We found threshold combinations of K and mortality 225 226 virulence (v) and of K and the sensitivity coefficient (α) that yielded foraging-mediated hydras. In each case, the threshold was found numerically (using a rootfinder) when host density at the 227 boundary equilibrium (equ. A2) equaled host density at the interior equilibrium ($N^* = S^* + I^*$, 228 solved for numerically). Assuming lower baseline size-specific foraging of infected hosts ($\hat{f}_S >$ 229 \hat{f}_{i}), we found threshold levels of virulence mortality (v) at which hydra effects arose, either with 230 sensitivity of foraging to spore accrual ($\alpha > 0$) or not ($\alpha = 0$). Then, at a given level of v, we 231 232 found threshold levels of sensitivity to spore accrual (α) at which hydra effects arose, either due to both mechanisms of foraging depression ($\hat{f}_S > \hat{f}_I, \alpha > 0$) or only due to the effect of spores on 233 foraging ($\hat{f}_{S} = \hat{f}_{I}, \alpha > 0$). 234

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236 Prediction of hydra from the dynamical model: results

Parasite-induced foraging depression can trigger a trait-mediated hydra effect (Fig. 2). Our model (equ. 3, Table 2) predicts that increasing the carrying capacity of algal resources (K, x-axis) or the maximum spore yield per infected host (σ_1 , contours) should increase equilibrial prevalence of infection (Fig. 2a). During larger epidemics, the average per capita death rate of hosts increases due to virulent effects of the parasite on host survivorship (Fig. 2b). Larger epidemics also yield greater density of resources, A, at equilibrium (A^* , Fig. 2c). Since this density is also the minimal resource requirement of hosts, it increases with heightened mortality

of hosts and foraging depression (Figs. A2,A3). More resources fuel greater within-host spore yield, $\sigma(A)$ (equ. 3.d; Fig. 2d). Higher spore yield enhances spread of disease and boosts epidemic size, but it also depresses mean foraging rate of hosts, f (where $f = (1-p^*)f_{S+}p^*f_{I}$; f_S and f_I from equ. 2; Fig. 2e).

The model predicts either trophic cascades or foraging-mediated hydras – the outcome 248 for host density depends on the relative effect of disease on resource production vs. on per capita 249 resource consumption of hosts. The increase of resource density (due to virulent depression of 250 foraging and survival) increases primary production $(PP^* = r_m A^* (1 - A^* / K))$; see Appendix) – as 251 long as K is high enough ($A^* > K/2$ – see Appendix Sections 2.3; Fig. 2f). Food consumption per 252 host, fA^* , also increases with K and σ (Fig. 2g). Host density, N^* , then increases or decreases 253 (relative to disease-free conditions) depending on the tension between responses of PP^* and fA^* 254 (Fig. 2h; see also Appendix Sections 2,3; Fig. A3). At lower K, virulence on survival dominates, 255 decreasing host density. At higher K, foraging depression and higher primary production increase 256 host density. Therefore, the model predicts that larger epidemics may increase host density when 257 parasites reduce feeding rate of their hosts enough in sufficiently enriched systems (see 258 Appendix for more details). Furthermore, the foraging-mediated hydra effect should arise more 259 260 readily when parasites are less lethal to their hosts (lower v; Figs. 2i, 3a,b), especially when infected hosts have lower baseline foraging rates $(\hat{f}_{s} > \hat{f}_{l})$ and their foraging is additionally 261 depressed by within-host spore growth ($\alpha > 0$, Fig. 3a; note $\hat{f}_{s} > \hat{f}_{l}$ is enough to enable the hydra 262 effect even when $\alpha = 0$, Fig. 3b). Also, at a given virulence level (v), the hydra effect is more 263 likely (i.e., can occur at lower K) when spore accrual more strongly suppresses for aging rate 264 (higher α ; Figs. 2j, 3c,d). The hydra effect occurs at lower α when $\hat{f}_S > \hat{f}_I$ (solid line;) than 265 when $\hat{f}_S = \hat{f}_I$ (dashed line; Figs. 3c,d) – therefore, both mechanisms of foraging depression ($\alpha >$ 266

267 $0, \hat{f}_S > \hat{f}_I$) enhance the hydra effect. Finally, depression of host foraging rate may also drive 268 higher infection prevalence (inferred from Fig. A3), through mechanisms involving higher spore 269 production with higher resource density and lower per capita spore consumption (i.e., less 270 removal of spores from the environment) by infected hosts (see Appendix and Fig. A3a-c.) 271

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Field Survey: Evidence for the Hydra During Large Epidemics in Nature

273 Estimation of infection prevalence, algal density, spore yield: methods

We sampled 13 lakes in southern Indiana (Greene and Sullivan Counties, USA) weekly 274 from August until the first week of December 2010. Here, we present data from the epidemic 275 season (end of September through mid-November). On each sampling visit, we pooled three 276 bottom-to-surface tows of a Wisconsin net (13 cm diameter, 153 µm mesh). From this sample, 277 we estimated prevalence infection (p) by diagnosing at least 400 live D. dentifera at 20–50X 278 magnification (Ebert 2005). From this sample, we estimated prevalence of infection in the adult 279 size class only (p_a) . We also measured body length (L_H) of uninfected and infected adult hosts 280 (typically > 20 of each class). Additionally, we estimated the average spore yield (σ) of infected 281 hosts (typically 5 to 40 hosts, pooled together). We estimated host density using preserved (60-282 75% ethanol) samples pooling three additional bottom-to-surface net tows. Finally, we indexed 283 density of 'edible' (< 80 µm Nitex screening) algae in the epilimnion using narrow-band filters 284 on a Trilogy fluorometer (Turner Designs) following chilled ethanol extraction (Webb et al. 285 1992; Welschmeyer 1994). 286

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288 Index of 'foraging depression' and death rate: methods

For each lake population, we calculated an index of disease-induced 'foraging

290depression' of adult hosts using (1) prevalence and spore yield data (Fig. A4a), (2) body size of
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uninfected and infected adults (Fig. A4b), and (3) parameters from the winning foraging function 291 (equ. 2; Table A1: \hat{f}_{S} , \hat{f}_{I} , α ; genotypes labeled 1–3 in Fig. A4b–d). We only summarize this 292 calculation here (see Appendix Section 4 for details). For the infected class, we assumed that 293 each infected adult shared the mean spore yield estimate for that lake-date (σ). With these 294 parameters and data, we calculated mean foraging rate of adults, f_a as mean foraging rate of each 295 infection class of adults $(\overline{f_{a,S}}, \overline{f_{a,I}})$ weighted by prevalence of infection of adults, p_a ; hence, $f_a =$ 296 $(1 - p_a)\overline{f_{a,S}} + p_a\overline{f_{a,I}}$ (see equ. A10.a-c). Next we calculated mean foraging rate of adults 297 assuming differences in mean body size only, f_0 (equ. A10.d). The index of foraging depression, 298 *FD*, was then: $FD = (f_0 - f_a) / f_0 * 100\%$ (equ. A10.e; Fig. A4d). For each lake, we averaged this 299 index, calculated at each sampling date, for each set of genotype-derived parameters (1–3); then, 300 we averaged those three separate genotype-specific estimates to produce one value of FD per 301 lake (see Fig. A4 for sample calculations). We also estimated average death rate of hosts during 302 epidemics using the egg-ratio method (see Appendix Section 4 for details). 303

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5 'Joint algal-host response' index: methods

We calculated a 'joint algal-host response' index to test qualitative predictions of the 306 dynamical model (i.e., hosts and resources should both increase during epidemics, particularly 307 during larger ones). To quantify this index, we first estimated the linear slopes of hosts and algal 308 resources through time (e.g., Fig. 4a,b). The 'joint response' index is the cross-product of these 309 two vectors (Fig. 4c,d), estimated after standardizing their slopes (by the standard deviation in 310 slope vectors among lakes). When both algae and hosts increased through time, the cross-product 311 was positive (e.g., the large epidemic in Goodman Lake: Fig. 4a,c), consistent with a hydra 312 effect. However, if only one of these (algae or hosts) increased through time, the cross-product 313

was negative (e.g., the small epidemic in Long Lake: Fig. 4b,d). Densities of algae and hosts 314 never both decreased through time (i.e., positive values only arose from two positive slopes, not 315 two negative ones). 316

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Signature of the trait-mediated hydra effect in the field: results

319 In the field survey, we detected the hydra pattern anticipated by the dynamical model. Infected hosts yielded more spores in lakes with larger outbreaks (Fig. 5a) and more algal 320 resources (shown previously (Civitello et al. 2015)). For lakes with greater spore loads, in turn, 321 we estimated stronger depression of foraging by adult hosts (Figs. 5b,A4). Lakes with stronger 322 foraging depression then had greater values of a dynamical index of algal resources and hosts 323 through time (Figs. 4,5c). In this 'joint algal-host response' index, positive numbers indicate a 324 hydra (see above). As predicted then, the signal of the foraging-mediated hydra effect arose 325 during larger epidemics (higher prevalence) with stronger foraging depression (Fig. 5c,d). In 326 contrast, mean death rate (estimate of d + vp) during epidemics was not correlated with the index 327 of foraging depression (R = 0.23, P = 0.42) or the joint algal-host index (R = 0.26, P = 0.35). 328

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Discussion

Undeniably, large epidemics of virulent parasites can depress host densities. However, 331 here we show that indirect feedbacks between hosts and resources can drive the opposite pattern: 332 333 increased host density during outbreaks. More specifically, parasites that virulently depress infected host foraging rate can indirectly produce more hosts under certain conditions. Using a 334 zooplankton-fungus-algal system, we show how infection by a virulent parasite depresses 335 336 foraging of infected hosts. Then, using a dynamical model of host-parasite-resource interactions,

we show when the foraging-mediate hydra effect should and should not arise. The model
predicted hydras during larger epidemics that strongly depress foraging of hosts while, at the
same time, not depressing fitness too much. We then turned to naturally occurring epidemics,
finding support for a foraging-mediated hydra effect. During larger epidemics, more spores
accumulated in host bodies, which depressed foraging. Reduced foraging, in turn, correlated with
a joint increase in hosts and algal resources – a signature of the hydra effect.

How and why does the foraging depression mechanism work? In the model, it works via 343 two components of host density: the ratio of resource production and resource consumption per 344 host. Both components start with an increase in the minimal resource requirement of hosts (an 345 indirect effect). Hosts require enough resources to offset increased mortality (resulting from 346 parasite virulence) with reproduction (extending logic from (Grover 1997)). Reduced foraging 347 further increased this requirement. The subsequent increase of resource density can increase 348 resource production (Case 2000). However, higher food density compensates for slower feeding, 349 yielding no net change in resource consumption. Therefore, foraging depression alone enhances 350 the likelihood of hydra effects during epidemics. In this system, foraging depression arose in 351 multiple ways. Infected hosts had lower size-specific feeding rate; infected hosts were smaller 352 353 (reducing size-dependent feeding further); and spore accumulation in host bodies substantially diminished foraging. Higher density of resources should exacerbate this spore-accumulation 354 effect (Civitello et al. 2015; Hall et al. 2009b). Finally, these hosts slow feeding when contacting 355 356 parasite propagules (Hite et al. 2017; Strauss et al. 2019). Hence, in this plankton system, multiple mechanisms produce foraging depression. Since parasite-mediated foraging depression 357 arises commonly in other systems as well (Hite and Cressler 2019; Hite et al. 2020), this trait-358 359 based mechanism for a hydra may apply quite broadly.

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Even with foraging depression, hydra effects may still not arise unless additional 360 conditions are met. First, hosts must strongly control their resource. While *Daphnia* famously 361 depresses its resources, not all hosts can (Borer et al. 2005; Shurin and Seabloom 2005). Second, 362 the subsequent increase in resource density must enhance resource productivity. Some resources 363 follow a more donor-controlled, chemostat-style supply (Polis et al. 1997); in these cases, 364 productivity drops as resource density increases, eliminating the hydra (see Appendix). Notably, 365 many experiments impose donor control, which means foraging-mediated hydras cannot occur. 366 Furthermore, sufficient enrichment is needed for higher density to increase resource productivity. 367 Third, parasites cannot depress survivorship or fecundity (Hall et al. 2007a; Lafferty and Kuris 368 2009) too strongly. Those forms of virulence increase per host resource consumption, potentially 369 overwhelming any increase in resource productivity. Fourth, epidemics must become large 370 enough to trigger the requisite indirect effects to densities and traits. While this is a lengthy set of 371 requirements, our results suggest they happen in at least this planktonic system. It remains to be 372 373 determined how many other systems can also produce a foraging-mediated hydra. Where does this foraging-mediated hydra result fit within other behaviors of host-374 parasite-resource systems? First, hydras can arise via other mechanisms (Abrams 2009; Abrams 375 376 and Matsuda 2005; Cortez and Abrams 2016). Increased mortality of hosts during epidemics could stabilize oscillatory host-resource cycles to increase host density. Here, the linear 377 functional response yielded stable dynamics, obviating evaluation of this mechanism. Yet, we 378 379 found no relationship between mean per capita death rate and the joint algal-host index (but see (McIntire and Juliano 2018) for an example of increased mortality driving higher density in 380 mosquitoes). Second, parasites can drive trophic cascades (Buck and Ripple 2017). In our model, 381 382 cascades were more likely at lower productivity, for less sensitive foragers, and for more virulent

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parasites; trophic cascades have been found in our plankton system, too (Duffy 2007). Third,
parasites can trigger 'biomass-overcompensation' in their host. This outcome, assuming certain
trait asymmetries between life stages of hosts, can increase biomass of the life stage most readily
infected (de Roos and Persson 2013; Preston and Sauer 2020; Schröder et al. 2009). Hopefully, a
coherent theory will emerge that synthesizes these possibilities for hydras, cascades, and biomass
overcompensation during epidemics.

Moving one step further, the foraging-mediated hydra effect should be integrated into a 389 broader theory for the community ecology of disease. First, foraging depression by parasites 390 should stabilize host-resource oscillations, providing another mechanism to produce a hydra 391 effect (e.g., (Hilker et al. 2009; Hurtado et al. 2014)). Second, other food web interactors might 392 stifle this foraging-mediated hydra. For instance, competitors of hosts could fix resources at their 393 own minimal resource requirement (analogous to systems with inedible producers: (Grover 394 1995)). Therefore, competition might prevent hydras. Third, hosts can evolve during epidemics 395 (Boots et al. 2009; Duffy and Forde 2009). This Daphnia host shows foraging-mediated 396 relationships between fecundity and transmission rate (Auld et al. 2013; Hall et al. 2010) and 397 between feeding rate and sensitivity to contact with spores (Strauss et al. 2019). Such 398 399 relationships could interact interestingly with foraging-mediated hydras as hosts evolve during epidemics. Therefore, integration of the foraging-mediated hydra effect awaits future 400 401 developments.

The foraging-mediated hydra effect means that large outbreaks may not depress host density. Parasite-mediated foraging depression occurs in a diverse array of systems (Hite and Cressler 2019; Hite et al. 2020). Yet, the foraging-mediated hydra here rests on a number of requirements, including that hosts strongly control resources, that resource productivity

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406	increases, and that infection only moderately increases mortality. It remains unknown how many
407	other systems meet these conditions. However, it is important to note that these foraging-
408	mediated hydra effects may produce desirable or undesirable outcomes. Hydras might prevent
409	worrisome collapses in host density during large outbreaks. Yet, they also increase density of
410	infected hosts, potentially elevating disease risk to humans (via contact with infected hosts) or
411	spillover to other hosts. Future efforts should evaluate the frequency and magnitude of foraging-
412	mediated hydra effects and their influence on disease and communities.
413	
111	A
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414 415 416	Acknowledgements We thank J.J. Potter for lab assistance. K. Boatman, Z. Brown, D. Grippi, J. Hite, C. Searle, and A. Smith helped in the field and lab. C. Gowler, J. Marino, C. Shaw, and C. Wood provided
414 415 416 417	Acknowledgements We thank J.J. Potter for lab assistance. K. Boatman, Z. Brown, D. Grippi, J. Hite, C. Searle, and A. Smith helped in the field and lab. C. Gowler, J. Marino, C. Shaw, and C. Wood provided comments on earlier drafts of the manuscript. This project was supported by NSF (DEB-0841679
414 415 416 417 418	Acknowledgements We thank J.J. Potter for lab assistance. K. Boatman, Z. Brown, D. Grippi, J. Hite, C. Searle, and A. Smith helped in the field and lab. C. Gowler, J. Marino, C. Shaw, and C. Wood provided comments on earlier drafts of the manuscript. This project was supported by NSF (DEB-0841679 to MAD and DEB-0841817 and 1120316 to SRH, and Graduate Research Fellowships to RMP
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Online Appendix

423 **Overview**

In the first section of this Appendix, we provide more results from the foraging assay 424 (Fig. 1a-c). The spore yield and length data (Fig. 1d-f) were used to parameterize the various 425 competing functions of foraging (as visualized in Fig. A1). All of the details of the winning 426 427 model (Table A1) and of the competition itself (Table 1) are also shown here. Second, we provide an in-depth analysis of the response of host density to depression of foraging rate in the 428 absence of disease (Fig. A2). Since it analytically conveys key logic for the more complex model 429 in the main text, we describe it in some depth. Third, we study how parasite-induced foraging 430 depression affects equilibrial densities of resources and hosts during epidemics, using key 431 comparisons (Fig. A3). We contrast the dynamical model from the main text (equ. 3; Fig. 2), 432 where parasites reduce both host survival and foraging, with two 'virulence variants' (described 433 below). Finally, we describe methods and illustrate calculations used to describe foraging 434 depression and mortality during epidemics (Fig. A4). 435

436

(1) More methods and results from the foraging rate experiment, and the parameterization and competition of foraging functions

439 Additional methods

We measured foraging rate, body size (host length, L_H), and spore density per host (σ) to parameterize the foraging models for uninfected and infected hosts of three genotypes. Hosts and parasites were originally from lakes in Barry County, MI, USA, except one host genotype, Beaver Dam 30 (BD-30), which was from Greene County, IN, USA. To standardize maternal effects, each genotype was reared in Artificial *Daphnia* Medium (ADaM (Klüttgen et al. 1994))

mixed with filtered water from Lake Lanier (Georgia, USA), and fed 0.9 µg C mL⁻¹ day⁻¹ (a 445 standard, non-limiting level) of a nutritious green alga (Ankistrodesmus falcatus). We generated 446 cohorts of 8-, 10-, 12-, 14-, 16-, 20-, and 24-day-old animals by collecting them within a 24-h 447 period (grouped as 10 per 150-mL beaker, kept at 20 °C in a 16:8 h light:dark cycle, then later 448 spread to six per beaker at six days old). Exposed beakers then received parasites (900 spores 449 mL⁻¹). We transferred all hosts to fresh medium after the 24-h exposure, and then every 4 days 450 until the day of the foraging rate assay. For the assay, hosts were placed singly into 15-mL 451 centrifuge tubes. For each treatment, sample size exceeded 12 for each age \times infection \times 452 genotype combination except n = 3 for 24-day-old infected STD hosts (most of these hosts had 453 already died of infection by then). Hosts grazed on 0.45 µg C mL⁻¹ of A. *falcatus* for 4 h; tubes 454 were inverted every 20 min to ensure algae stayed suspended. Hosts were then removed from 455 each tube and measured from the middle of the eye to the base of the tail spine at 40X. We 456 quantified food remaining in the tube using a Trilogy fluorometer (in vivo module, Turner 457 Designs, Sunnyvale, CA, USA). 458 To estimate foraging rate of the infected class, we only used hosts that developed 459 infections that reached the ascospore stage (Stewart Merrill and Cáceres 2018). Spores were 460 461 visually apparent once infected hosts were 16+ days old (i.e., 10+ days post-exposure). To estimate spore yield, we transferred hosts to microcentrifuge tubes, gently smashed each 462

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individual using a pestle, and counted the released spores using a hemocytometer at 200X

et al. 2014), we assumed they contained none during the assay, but diagnosed them later,

retaining only successfully infected hosts in the analysis. After removing hosts that did not

magnification. Since infected hosts less than 16 days old typically contain very few spores (Auld

ultimately develop infections, each treatment had at least 9 replicates, except the 24-day-old
infected STDs (discussed above).

469

470	Additional results	
471	Individuals from the three host genotypes grew through time when uninfected, but	
472	growth tended to slow or plateau once spores accumulated in their bodies (i.e., 10+ days post-	•
473	exposure to the parasite, when hosts were 16+ days old; Fig. 1d-f). The best fitting models (5b):
474	size and spores, linear, and 6b: size and spores, power) explained the drop in foraging of	
475	infected hosts (Fig. 1,A1; Table 1). Parameter estimates for the winning model varied among	the
476	three genotypes (Table A1), and that variation was included in calculations of the index of	
477	foraging depression in the lakes (Fig. 5b,c; Fig. A4c,d).	
478		
479	(2) Theoretical insights from the disease-free subsystem of the dynamical model	
480	The disease-free subsystem only has hosts (all of whom are uninfected, or susceptible,	, <i>S</i>)
481	and resources (A). Imagine that the host feeds with a linear functional response (at foraging ra	te
482	f; following the model in the main text) while the algal resource grows logistically, yielding:	
483	$dS/dt = efAS - dS \tag{A1.a}$	
484		
-	$dA/dt = r_m(1 - A/K)A - fAS, \tag{A1.b}$	
485	$dA/dt = r_m(1 - A/K)A - fAS$, (A1.b) where population growth rate of susceptible hosts (dS/dt , equ. A1.a) increases with foraging rates	ate
485 486	$dA/dt = r_m(1 - A/K)A - fAS$, (A1.b) where population growth rate of susceptible hosts (dS/dt , equ. A1.a) increases with foraging rate f and conversion of consumed resources into offspring (with efficiency e) but decreases at a	ate

equilibrium. The growth rate of the resource (dA/dt, equ. A1.b) is logistic as governed by

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maximal growth rate r_m and carrying capacity *K*. Without grazing, the algal resource would reach its carrying capacity (*K*) at the boundary equilibrium; with grazing, the interior equilibrium is:

491
$$A^* = d/(ef)$$
 (A2.a)

492

$$S^* = \left[r_m \left(1 - \frac{A^*}{K} \right) \right] A^* / (fA^*) = \frac{PP}{fA^*}$$
(A2.b)

At this equilibrium, resource density $(A^*, equ. A2.a)$ is equal to the minimal resource requirement 493 of the host (where low death rate, d, high conversion efficiency, e, and/or high foraging rate, f, 494 lead to strong control over the resource, or lower A^*). Here, per capita death rate (d) of the host 495 equals its per capita birth rate, b (where again, $b = e f A^*$). Equilibrial host density (S^{*}, equ. A2.b) 496 is the ratio of two key quantities (written in a particular way here to maximize meaning below). 497 The numerator of this ratio is primary production of the algal resources, $PP = r(A^*) A^*$; that is, 498 per capita productivity of the resource, $r(A^*) = r_m (1 - A^* / K)$ (in square brackets of equ. A2.b) 499 times equilibrial algal density, A^* . Primary productivity, $r(A^*) A^*$, follows the familiar, unimodal 500 501 hump of the logistic model with increasing A; thus it is maximized at K/2. The denominator is per host consumption of the resource, fA^* (which itself is proportional to host per capita birth 502 rate, b). Host density, S^* , then depends on how primary productivity (PP, the numerator) is 503 partitioned among grazers (the denominator, with each grazer taking portion fA^*). 504

How will this interior equilibrium (equ. A2) respond to depressed foraging rate, f? We can see that a slower forager ('lower f') can reach higher density than a faster forager ('higher f') if carrying capacity is above a certain threshold (this threshold level of *K* increases with death rate [Fig. A2a,b]). More insight arrives from a tiny bit of calculus. Not surprisingly, resource density will *always* increase if *f* drops since:

510

$$-\partial A^*/\partial f = A^*/f. \tag{A3}$$

(Notice the negative partial derivative here to denote foraging rate, f, shrinking.) Hence, a slower 511 foraging host needs a higher minimal resource requirement. However, a drop in foraging rate (f)512 can either elevate or depress density of the host, S^* . The outcome depends on how primary 513 productivity, $r(A^*) A^*$, responds to the increase in resource density, A^* (given lowered foraging 514 rate; equ. A3) and how each host's rate of resource consumption, fA^* , changes. This latter rate 515 does not change with decrease in f alone (i.e., $-\partial (fA^*)/\partial f = 0$) because the depression of foraging 516 rate is exactly offset (compensated) by an increase in resource density (given equ. A3). Thus, the 517 response of host density to foraging depression solely hinges on how primary productivity 518 519 changes with increased crowding of resources. Formally,

520
$$-\frac{\partial S^*}{\partial f} = \frac{r}{f^2 K} (K - 2A^*)$$
(A4.a)

521
$$-\frac{\partial PP^*}{\partial f} = \frac{rA^*}{K}(K - 2A^*)$$
(A4.b)

which says that hosts, S^* , will increase with a drop in foraging, f (i.e., $-\partial S^*/\partial f > 0$) if resource 522 density (A^*) is less than half of the resource's carrying capacity (K/2; equ. A4.a). That threshold 523 happens when A^* is below peak primary productivity (K/2) – notice how primary production also 524 525 increases with a drop in foraging in this same range (equ. A4.b). Therefore, higher primary productivity can support higher density of the more slowly foraging hosts (yellow region, ' $S\uparrow$ '; 526 higher f-K space in Figs. A2c,d; more f-d space at higher K in Figs. A2e vs. A2f). This case is 527 most likely when the host strongly controls the resource (i.e., when its minimal resource 528 requirement, A^* , is well below K due to high initial foraging rate or in more productive systems 529 (higher K; Figs. A2c,d). Alternatively, when $A^* > K/2$, total primary productivity drops as 530 resources increase due to foraging depression. Lower primary productivity supports fewer hosts. 531 This case arises when hosts do not control their resources strongly (i.e., then they cannot depress 532 A^* below K/2 due to low foraging rate or in less productive systems (white regions, S_{\downarrow} ; lower f-533 Penczykowski et al. 24 Foraging-mediated hydras in epidemics

K space in Figs. A2c,d; more *f*-*d* space at lower *K* in Figs. A2e vs. A2f). Note also that a resource that is donor-controlled / has chemostat-style renewal would prohibit an increase of hosts with declining feeding rate. For instance, imagine that resource renewal was $a (A_S - A)$ in equation A1.b (with dilution rate *a* [day⁻¹] and supply point A_S [mg/L]) instead of $r_m A(1-A/K)$. In this case, the minimal resource requirement would not change ($A^* = d/(ef)$), and primary production ($PP = a [A_S - A^*]$) would always decline with decreasing f ($-\partial PP/\partial f = -a A^*/f$) as would host density [$C^* = (a/f)(A_S/A^* - 1)$; $-\partial C^*/\partial f = -a/f^2$].

This equilibrial analysis of the susceptible host/grazer (*S*)–algal resource (*A*) subsystem prompts the following qualitative predictions for a more complicated system with a parasite that depresses foraging rate of its host (equ. 2-3).

(1) Algal density, A^* , should always increase when foraging rate of the host, *f*, drops. (2) However, the response of host density depends upon how primary productivity, $r(A^*)$ A^* , responds to this higher density of resources (i.e., whether A^* is higher or lower than peak primary productivity, $A^* = K/2$).

(3) Primary productivity should increase, and hence host density should increase, with depressed foraging rate when hosts strongly control their resource (relatively low A^*) or in more productive systems (higher *K*; case 1 in Fig. A2b). Here, we see a joint increase ($A^*\uparrow$,

551 $S^*\uparrow$) caused by foraging depression.

(4) Primary productivity declines and host density drops with foraging depression when
grazers cannot strongly control their resource (relatively high *A*^{*}, due to high *d*, low *e*, and/or

- low baseline f; case 3 in Fig. A2a) or in less productive systems (lower K; case 3 in Fig.
- 555 A2b). Then, we expect a trophic cascade-like pattern $(A^*\uparrow, S^*\downarrow)$.
- (5) However, per capita resource consumption, fA^* , and hence per capita birth rate (b = e

 fA^*), should not change with a decline in foraging rate, *f*, alone. In other words, the response of host density to depressed foraging does not involve per capita birth rate. This aspect is critical, since host response depends on how decreasing *f* affects primary production (predictions 3 and 4) as well as per capita resource consumption (equ. A2.b). (6) This dependence on per capita resource consumption explains why host density

562

If death rate increases:

564

563

$$\partial A^* / \partial d = A^* / d \tag{A5.a}$$

should decline with increased death rate, d, i.e., if the parasite virulently depresses survival.

565
$$\partial S^* / \partial d = -\frac{rA^*}{dfK}$$
 (A5.b)

thus, resource density should increase (equ. A5.a) but grazer density should decrease (equ. 566 A5.b) at higher d. The response of producer density to higher d qualitatively echoes that seen 567 for foraging depression (equ. A3). However, a little bit of calculus shows that host density 568 always declines with increasing d – even if primary production increases (i.e., $A^* < K/2$) – 569 because per capita consumption by grazers has to increase too much with higher d (since e f 570 $A^* = b = d$ at equilibrium, by definition). In other words, with increasing d, higher 571 consumption demands per host overwhelm any primary productivity response. This insight 572 explains why the same f-K combination at low d produces more hosts with lower f but fewer 573 hosts at higher d (contrast dot in Fig. A2c [lower d] vs. A2d [higher d]). Similarly, at low K, 574 an *f*-*d* combination that would produce less hosts with foraging depression yields more hosts 575 at higher K (contrast dot in Fig. A2e [lower K] vs. A2d [higher K]) 576 (7) Hence parasite-mediated hydra effects become more likely at higher productivity for 577 hosts which control their resources (guaranteeing $A^* < K/2$). Furthermore, they are more 578

579 likely when foraging depression is large (boosting *PP*) but when parasites are not too virulent

580

581

(lower v, which would increase food consumption, fA^* , per host too much, overwhelming gains in PP).

582

583 (3) Further discussion of the dynamical model of the full host–parasite–resource system

In addition to the modeling results presented in the main text, we examined two other 584 'virulence variants' to better understand the predicted response of hosts to either foraging 585 depression alone (lower f) or higher virulence on survival (v) alone (Fig. A3). These examples 586 build on the intuition from the disease-free subsystem (section 2). The first variant (left column, 587 Fig. A3) extends the example in the main text: during epidemics, hosts experience foraging 588 depression (sensitivity coefficient $\alpha > 0$; see equ. 2.b) and virulence on survival ($\nu > 0$). 'Variant 589 2' features the same virulence on survival but no foraging depression ($\alpha = 0, \nu > 0$; middle 590 column). 'Variant 3' models only foraging depression without virulence on survival ($\alpha > 0$, $\nu =$ 591 0; right column). 592

These three 'virulence variants' disentangle the effects of decreased foraging and survival 593 on epidemiology (i.e., disease prevalence at equilibrium) as well as densities of resources and 594 hosts. **Disease prevalence** (proportion infected, p^*) at equilibrium is quantitatively different 595 596 among the variants (Fig. A3a–c). At a given carrying capacity (K) and maximal spore yield (σ_1), prevalence is typically greater in 'variant 1' and 'variant 3' (which include foraging depression), 597 compared to 'variant 2' (which only includes virulence on survival). Therefore, all else equal, 598 parasite-driven foraging depression promotes larger epidemics through a combination of greater 599 total host density (contrast Fig. A3m,o vs. A3n) and less removal of spores from the environment 600 by already-infected hosts. Disease prevalence in 'variant 1' is also enhanced by greater resource 601 density (contrast Fig. A3d vs. A3e) – and thus spore yield of infected hosts – relative to 'variant 602

603 2.' The **resource response** is striking but unidirectional. Resource density, A^* , is:

604
$$A^* = \frac{d + vp^*}{e[(1 - p^*)f_S + p^*f_I]}$$
(A6)

which is the ratio of per capita mortality, $d + vp^*$ to per capita, per resource birth rate, i.e., 605 conversion rate, e, times mean feeding rate of the population, taking into account proportion 606 susceptible, $(1-p^*)f_S$, and infected, p^*f_I (equ. A6). In 'variant 1' (Fig. A3d), A^* shows synergy 607 between the indirect effects of virulence on survival (Fig. A3e) and foraging depression (Fig. 608 A3f) on resource density. The effects of foraging depression alone on A^* are actually small 609 (given the parameters). **Primary production**, *PP*, largely mirrors the algal density response (Fig. 610 A3g-1). As described in the second section of this supplement (Theoretical insights from the 611 *disease-free subsystem* above), primary production is $r_m A^*(1 - A^*/K)$, and it determines the 612 numerator of host density. Over much of the K range, primary production increases during 613 epidemics (more subtly with only foraging depression [Fig. A3j], more for virulence on survival 614 [Fig. A3h], and synergistically for both [Fig. A3g]). Food consumption, $fA^* = (1-p^*)f_S + p^*f_I$, 615 follows a relatively similar pattern. It is highest when both $\alpha >0$ and $\nu >0$ (Fig. A3j), lower when 616 infection only imposes mortality (Fig. A3k), but does not change when it only imposes foraging 617 depression (because foraging is compensated for in the minimal requirement exactly; Fig. A31). 618 **Total host density**, N^* , is: 619

620
$$N^* = \frac{PP^*}{fA^*} = \frac{r_m A^* (1 - A^*/K)}{[(1 - p^*)f_S + p^*f_I]A^*}$$

the ratio primary production to food consumption (equ. A7). It reflects tension between the two
sources of virulence. With only virulent effects on survival, host density decreases (Fig. A3n). In
contrast, over the vast majority of the *K* gradient (except for very low *K*), foraging depression
alone indirectly increases host density during epidemics (Fig. A30), given that these hosts
strongly control their algal resource (i.e., they have low minimal resource requirement, *A**).
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(A7)

Thus, the response of host density to a combination of foraging depression and virulence on 626 survival depends on carrying capacity. At lower K, the host-decreasing effect of virulence on 627 survival dominates; at higher K, the host-increasing effect of foraging depression prevails (Fig. 628 A3m). This result reminds us that the response of host density during epidemics does not merely 629 follow an increase in primary production. The per capita foraging consumption (fA^*) by hosts 630 needed to 'break even' determines how many hosts the primary production can support. Hosts 631 suffering higher virulence on survival require higher resource consumption to break even; hosts 632 experiencing foraging depression do not (using the logic in section 2 above). 633

In summary, this comparison of model variants clarifies the response of hosts and their 634 resources during epidemics. If the parasite only virulently lowers survival of hosts, the model 635 predicts only a trophic cascade, where resources increase while host density declines relative to 636 disease-free conditions. (If the consumer-resource system could oscillate, host density might 637 increase with lower survival, under some conditions (Abrams 2009). This possibility was not 638 modeled here.) In contrast, if the parasite only lowers foraging rate, a host with these Daphnia-639 like traits (i.e., exerting strong control over its resource; Table 2) should typically increase during 640 epidemics. In other words, both host and resources increase, relative to disease-free systems. 641 642 Parasites that depress host feeding rate should also typically have larger epidemics compared to parasites that reduce survival only. For parasites that both depress survival and foraging rate, the 643 host response depends on the relative strength of effects of survival vs. foraging rate on host 644 645 density. This balance can shift with carrying capacity of the resource, as increases in host density during epidemics are more likely when carrying capacity is higher. 646

647

648 (4) Field survey: more methods and sample calculations of indices describing data from the

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649 field survey

650 *Methods for calculating death rate*

In the field survey, we calculated temperature-dependent death rate in a way that 651 incorporates diel migration of the host. This species of host typically migrates below the 652 thermocline (into the 'metalimnion') of lakes during day into deeper, colder, but still oxygenated 653 $(> 1.0 \text{ mg/L dissolved O}_2 \text{ [DO]})$ waters. Then, at night, it moves above the thermocline into 654 upper, warmer habitat (the 'epilimnion') (e.g., (Duffy et al. 2005; Hall et al. 2005). Therefore, 655 using temperature data, we calculated depth of the thermocline (during periods of stratification) 656 by: (1) converting temperature data into densities (following (Chen and Millero 1977)); (2) then 657 calculating buoyancy frequency, $N = (g/\rho(d\rho/dz))^{1/2}$ [where g is acceleration due to gravity, ρ 658 is mean density, and $d\rho/dz$ is the vertical density gradient], at 0.1 m depths by differentiating 659 piece-wise cubic splines fit through the density-depth data (with pchip.m in Matlab); and (3) 660 finding the thermocline as the depth of maximum buoyancy frequency. We found the 661 oxygenation threshold (1.0 mg DO) using cubic splines fit through DO-depth data. With 662 temperature, thermocline depth, and oxygenation threshold information, we calculated mean 663 development time in the oxygenated metalimnion (day, D_M) and epilimnion (night, D_E). 664 $D_M = \exp[\ln(a) + b \ln(T_M) + c (\ln(T_M))^2]$ (A7a) 665 $D_E = \exp[\ln(a) + b \ln(T_E) + c (\ln(T_E))^2]$ (A7b) 666

where T_M and T_E are mean temperatures in the metalimnion and epilimnion, respectively, and coefficients $\ln(a) = 3.4$, b = 0.22, and c = -0.3 come from (Bottrell et al. 1976). Mean development time at each lake-date, D_{ave} , is then just the weighted average of D_E and D_M :

$$D_{ave} = \varphi_M D_M + \varphi_E D_E \tag{A8}$$

671 where φ_M and φ_E are the proportion of time per day spent in the metalimnion and epilimnion, Penczykowski et al. 30 *Foraging-mediated hydras in epidemics*

respectively (taking into account waning of daylight as autumn progresses). 672

Then, we calculated birth rate using the egg ratio methods. We calculated the average 673 weighted egg ratio, E_{ave} , using data on infected and uninfected adult host classes. Next we 674 calculated the population-level egg ratio, E_p , by multiplying E_{ave} times the percentage of asexual 675 females in the population. We calculated the per capita birth rate, *b*: 676

677
$$b = \ln(E_p + 1)/D_{ave}$$
 (A9)

where D_{ave} follows equ. A8. We calculated the instantaneous rate of increase, $r = \ln (N_{t+\tau}/N_t)/\tau$, 678 where $N_{t+\tau}/N_t$ is the ratio of host density between sampling intervals τ . Then, death rate during 679 epidemics is d + pv = b - r. 680

681

Example calculation of index of foraging depression 682

The 'index of foraging depression' becomes more tangible with an illustrative example 683 from one of the 13 studied lakes. As an epidemic unfolded in Goodman Lake, spore yield lagged 684 behind prevalence through time (Fig. A4a), and mean size of uninfected and infected host 685 changed slightly (Fig. A4b). The size-only effect had a modest influence on mean foraging of 686 adult hosts ('only size' solid lines in Fig. A4c, as parameterized for the three laboratory-assayed 687 688 genotypes [Table A1]). However, foraging dropped considerably once infection was modeled ('spore-depressed' dashed lines in Fig. A4c). The index of foraging depression comes from the 689 difference between lines 'only size' and 'spore-depressed' lines (Fig. A4d). The index of foraging 690 691 depression, FD, for a given lake-date-genotype combination is:

$$\overline{f_{aS}} = \left(\sum_{i}^{n_S} f_S(L_H)_i\right) / n_S \tag{A10.a}$$

694

$$\overline{f_{-i}} = \left(\sum_{i}^{n_{I}} f_{I}(L_{H},\sigma)_{i}\right)/n_{I}$$
(A10.b)

$$f_a = (1 - p_a)\overline{f_{a,s}} + p_a\overline{f_{a,l}}$$
(A10.c)

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(A10.b)

695
$$f_0 = \left(\sum_{i}^{n_S + n_I} f_S(L_H)_i\right) / (n_S + n_I)$$
(A10.d)

696

$$FD = (f_0 - f_a) / f_0 * 100\%.$$
(A10.e)

Here, mean feeding rate of the sample of susceptible (uninfected) adults, $\overline{f_{a.s.}}$, is calculated for n_s 697 individuals using foraging function $f_S(L_H)$ (equ. 2.a) for individual *i* given its body length L_H 698 (equ. A10.a). Similarly, mean feeding rate of infected adults, $\overline{f_{q_I}}$, calculated for n_I individuals 699 using foraging function $f_I(L_H,\sigma)$ (equ. 2.b) for individual j given its body length L_H and the mean 700 spore yield per infected host on that sampling date, σ (equ. A10.b). The mean foraging rate, f_a , is 701 then the average of these mean rates weighted by prevalence of infection in adults, p_a (equ. 702 A10.c). For comparison, we calculated mean adult foraging rate assuming only length (L_H) 703 influenced it, f_0 (equ. A10.d). Specifically, to calculate f_0 we used the foraging function for 704 susceptibles, $f_S(L_H)$ (equ. 2a) for each susceptible and infected individual *i* (summed over the 705 total sample, $n_S + n_I$). (This calculation therefore assumed size-specific foraging rate for infected 706 hosts, \hat{f}_I , was set to that of susceptible hosts, \hat{f}_S , and that spore accrual did not suppress feeding, 707 so $\alpha = 0$; note that this is equivalent to the 'size only' model 2a from Table 1). The index of 708 foraging depression, FD, was then the relative depression of foraging due to disease (equ. 709 A10.e). In a given lake, we calculated three separate values of FD each sampling date: one for 710 each set of genotype-specific parameter estimates from the foraging rate experiment (Table A1; 711 these produced the three lines in Fig. A4.d). We calculated the temporal mean for each of the 712 three genotype-specific FD values, and then averaged across those three temporal means to 713 produce one value of FD per lake (plotted in Fig. 5b,c). 714

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715	Table A1. Statistical description of the winning foraging function, '5b: size and spores, linear'
716	(from Table 1; also presented as equ. 2). (a) Best-fit parameter estimates (with bootstrapped
717	lower and upper 95% CI) for size-specific foraging rates of uninfected hosts ($\hat{f}_S \times 10^{-2} \text{ L} \cdot \text{mm}^-$
718	² ·day ⁻¹) and infected hosts ($\hat{f}_I \times 10^{-2} \text{ L} \cdot \text{mm}^{-2} \cdot \text{day}^{-1}$), and for the linear sensitivity coefficient ($\alpha \times$
719	10^{-5} mm ³ ·spore ⁻¹). (b) <i>P</i> -values of permutation tests comparing those parameter estimates
720	between host genotypes (9,999 randomizations per contrast; asterisks indicate significant
721	pairwise differences after Holm–Bonferroni correction). These parameters generate the curves
722	shown in the text (Fig. 1) and were also paired with field data in the calculation of the index of
723	foraging depression (Fig. A4; Fig. 5b,c).

(a) Estimate (95% CI) for each host genotype				
Parameter	A4-4	BD-30	STD	
\hat{f}_S	1.94 (1.83, 2.05)	1.78 (1.62, 1.96)	1.69 (1.53, 1.85)	
\hat{f}_I	1.54 (1.41, 1.67)	1.31 (1.17, 1.46)	0.90 (0.80, 1.00)	
α	2.49 (2.18, 2.85)	2.86 (2.23, 3.35)	0.92 (0.22, 1.59)	
(b) <i>P</i> -value for	comparison between ge	notypes		
Parameter	A4-4 vs. BD-30	BD-30 vs. STD	STD vs. A4-4	
\hat{f}_S	0.028	0.26	0.0002 *	
\hat{f}_I	0.0017 *	< 0.0001 *	< 0.0001 *	
α	0.095	< 0.0001 *	< 0.0001 *	

724 APPENDIX FIGURE LEGENDS

725

726	Figure A1. Functions of spore-dependent foraging rate, $f(L_H, \sigma)$, in models 1b-6b (see
727	Table 1). These functions were fit to observed algal data for uninfected (f_s , white) and infected
728	(f_I , black) hosts from three genotypes but visualized with calculated foraging rate (mean ± 1
729	SE). (a-c) No size dependence: Foraging functions 1b (solid line), 3b (dashed), and model 4b
730	(dotted), respectively, do not scale with host surface area (L_H) ; in panel a, 3b and 4b overlap. (d-
731	f) Size dependence: Function 2b (solid), 5b (dashed; this function is equ. 2 and also plotted in
732	Fig. 1a-c), and 6b (dotted) depend on surface area; in panel f, 5b and 6b overlap. Note that for
733	uninfected hosts with σ =0, models 1b, 3b, and 4b are equivalent and models 2b, 5b, and 6b are
734	equivalent (f_s , solid). Genotypes: A4-4 (panels a,d), Beaver Dam-30 ('BD-30'; b,e), and
735	standard ('STD'; c,f).
736	Figure A2. Graphical response of hosts (S^*) to depressed foraging rate (<i>f</i>) without
737	disease. (a,b) Host density at lower f (dashed; 0.0175 L·day ⁻¹) and higher f (solid; 0.0350 L·day ⁻¹)
738	¹). In the yellow region, hosts with lower f are more abundant, illustrated for (a) lower mortality
739	$(d=0.03 \text{ day}^{-1})$ and (b) higher mortality $(d=0.06 \text{ day}^{-1})$. (c)-(d) Regions of carrying capacity (K)
740	and feeding rate (<i>f</i>) in which hosts increase with lower feeding rate (yellow; $S\uparrow$; $-\partial S^*/\partial f > 0$;
741	equ. A4.a), hosts decrease with lower feeding rate (white; 'S \downarrow '; $-\partial S^*/\partial f < 0$), or hosts cannot
742	persist ('S=0'; $K < A^*$), for (c) lower mortality ($d=0.03 \text{ day}^{-1}$) and (d) higher mortality ($d=0.06$
743	day ⁻¹). (e)-(f) Regions of death rate (d)-feeing rate (f) parameter space supporting those same

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three states (S \uparrow , S \downarrow , S=0). Other parameters follow Table 2. (Dots in panels c-f are referred to in text).

746	Figure A3. More results from the dynamical model, simulated under 'virulence variants'.
747	Left column: both foraging depression and virulence on survival ($\alpha > 0$, $\nu > 0$; shown in Fig. 2).
748	Middle column: only virulence on survival ($\alpha = 0, \nu > 0$). Right column: only foraging
749	depression ($\alpha > 0$, $\nu = 0$). (a-c) Equilibrial prevalence of infection, p^* . (d–f) Resource density, A^* .
750	(g–i) Primary production, $PP = r_m A^*(1 - A^*/K)$. (j-l) Food consumption per host, fA^* . (m-o) Total
751	host density, $N^* = S^* + I^*$. Arrows point along contours of increasing maximum spore yields, σ_1 .
752	Disease-free conditions ('0') noted with thick solid contours. Parameters follow Table 2.
753	Figure A4. Illustration of the 'foraging depression' index, calculated with adult hosts
754	(shown in Fig. 5). (a) An example of a large fungal epidemic in Goodman Lake (Fig. 4a,c):
755	prevalence of infection of adults (percentage infected; black diamonds) and spore yield per
756	infected host (σ , grey squares). (b) Mean size of infected and uninfected adults (L_H). (c)
757	Components of the foraging rate (f) depression index (equ. A10), calculated for clonal genotypes
758	1 (A4-4), 2 (BD-30), and 3 (STD; parameters in Table A1). The 'only size' lines (solid) calculate
759	foraging rate based on host size alone. 'Spore-depressed' lines (dashed) assume different size-
760	specific foraging rates for infected (\hat{f}_I) and uninfected (\hat{f}_S) hosts, and spore-mediated foraging
761	depression (proportional to α). (d) Percentage decrease from the 'only size' to 'spore-depressed'
762	estimates. This calculation shows that spore accumulation within hosts strongly depresses mean
763	foraging rate of adults in this population of <i>Daphnia</i> .

Fig. A1







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Fig. A3



Fig. A4



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764 **References**

- Abrams, P. A. 2009. When does greater mortality increase population size? The long history and
 diverse mechanisms underlying the hydra effect. Ecology Letters 12:462-474.
- Abrams, P. A., and H. Matsuda. 2005. The effect of adaptive change in the prey on the dynamics
- of an exploited predator population. Canadian Journal of Fisheries and Aquatic Sciences62:758-766.
- Altizer, S., R. S. Ostfeld, P. T. J. Johnson, S. Kutz, and C. D. Harvell. 2013. Climate change and
 infectious diseases: from evidence to a predictive framework. Science 341:514-519.
- Anagnostakis, S. L. 1982. Biological control of chestnut blight. Science 215:466-471.
- Anderson, R. M., and R. M. May. 1979. Population biology of infectious diseases: Part I. Nature
 280:361–367.
- Anderson, R. M., and R. M. May. 1981. The population dynamics of microparasites and their

invertebrate hosts. Philosophical Transactions of the Royal Society of London. B,Biological Sciences 291:451-524.

- Auld, S. K. J. R., S. R. Hall, and M. A. Duffy. 2012. Epidemiology of a *Daphnia*-multiparasite
 system and its implications for the Red Queen. Plos One 7:6.
- Auld, S. K. J. R., S. R. Hall, J. H. Ochs, M. Sebastian, and M. A. Duffy. 2014. Predators and
- patterns of within-host growth can mediate both among-host competition and evolution
 of transmission potential of parasites. American Naturalist 184:S77-S90.
- Auld, S. K. J. R., R. M. Penczykowski, J. H. Ochs, D. C. Grippi, S. R. Hall, and M. A. Duffy.
- 2013. Variation in costs of parasite resistance among natural host populations. Journal of
 Evolutionary Biology 26:2479-2486.

Penczykowski et al.

Foraging-mediated hydras in epidemics

- Boots, M., A. Best, M. R. Miller, and A. White. 2009. The role of ecological feedbacks in the
 evolution of host defence: what does theory tell us? Philosophical Transactions of the
 Royal Society B-Biological Sciences 364:27-36.
- Borer, E. T., E. W. Seabloom, J. B. Shurin, K. E. Anderson, C. A. Blanchette, B. Broitman, S. D.
- Cooper et al. 2005. What determines the strength of a trophic cascade? Ecology 86:528-537.
- Bottrell, H. H., A. Duncan, Z. M. Gliwicz, E. Grygierek, A. Herzig, A. Hillbricht Ilkowska, H.
 Kurasawa et al. 1976. A review of some problems in zooplankton production studies.
- Buck, J. C., and W. J. Ripple. 2017. Infectious agents trigger trophic cascades. Trends in
 Ecology & Evolution 32:681-694.
- Burnham, K. P., and D. R. Anderson. 2002, Model selection and multimodel inference: a
 practical information-theoretic approach, 2nd ed. New York, Springer-Verlag.
- Case, T. J. 2000, An Illustrated Guide to Theoretical Ecology. New York, Oxford University
 Press.
- Chen, C. T., and F. J. Millero. 1977. Use and misuse of pure water PVT properties for lake
 waters. Nature 266:707-708.
- Civitello, D. J., R. M. Penczykowski, A. N. Smith, M. S. Shocket, M. A. Duffy, and S. R. Hall.

2015. Resources, key traits and the size of fungal epidemics in *Daphnia* populations.
Journal of Animal Ecology 84:1010-1017.

- 805 Cleaveland, S., M. K. Laurenson, and L. H. Taylor. 2001. Diseases of humans and their domestic
- mammals: pathogen characteristics, host range and the risk of emergence. Philosophical
- Transactions of the Royal Society B-Biological Sciences 356:991-999.

Penczykowski et al.

Foraging-mediated hydras in epidemics

- 808 Cooper, J., R. J. M. Crawford, M. S. De Villiers, B. M. Dyer, G. J. G. Hofmeyr, and A. Jonker.
- 2009. Disease outbreaks among penguins at sub-Antarctic Marion Island: a conservation
 concern Marine Ornithology 37:193-196
- 811 Cortez, M. H., and P. A. Abrams. 2016. Hydra effects in stable communities and their
- implications for system dynamics. Ecology 97:1135-1145.
- Daszak, P., L. Berger, A. A. Cunningham, A. D. Hyatt, D. E. Green, and R. Speare. 1999.
- Emerging infectious diseases and amphibian population declines. Emerging Infectious
 Diseases 5:735-748.
- de Roos, A. M., and L. Persson. 2013, Population and Community Ecology of Ontogenetic
- 817 Development. Princeton, New Jersey, USA, Princeton University Press.
- Duffy, M. A. 2007. Selective predation, parasitism, and trophic cascades in a bluegill-*Daphnia* parasite system. Oecologia 153:453-460.
- Duffy, M. A., and S. E. Forde. 2009. Ecological feedbacks and the evolution of resistance.
- Journal of Animal Ecology 78:1106-1112.
- Duffy, M. A., and S. R. Hall. 2008. Selective predation and rapid evolution can jointly dampen
 effects of virulent parasites on *Daphnia* populations. American Naturalist 171:499-510.
- Duffy, M. A., S. R. Hall, A. J. Tessier, and M. Huebner. 2005. Selective predators and their
- parasitized prey: Are epidemics in zooplankton under top-down control? Limnology and
 Oceanography 50:412-420.
- Dwyer, G., and J. S. Elkinton. 1993. Using simple models to predict virus epizootics in gypsy
 moth populations. Journal of Animal Ecology 62:1-11.
- Ebert, D. 2005, Ecology, Epidemiology and Evolution of Parasitism in *Daphnia*. Bethesda, MD,
- 830 National Library of Medicine (US), National Center for Biotechnology Information.

- Frick, W. F., J. F. Pollock, A. C. Hicks, K. E. Langwig, D. S. Reynolds, G. G. Turner, C. M.
- Butchkoski et al. 2010. An emerging disease causes regional population collapse of a
- common North American bat species. Science 329:679-682.
- Fry, W. E., and S. B. Goodwin. 1997. Re-emergence of potato and tomato late blight in the
- United States. Plant Disease 81:1349-1357.
- Gotelli, N. J., and A. M. Ellison. 2004, A Primer of Ecological Statistics. Sunderland, MA,
- 837 Sinauer Associates, Inc.
- Green, J. 1974. Parasites and epibionts of *Cladocera*. Transactions of the Zoological Society of
 London 32:417-515.
- Grover, J. P. 1995. Competition, herbivory, and enrichment: nutrient-based models for edible
 and inedible plants. The American Naturalist 145:746-774.
- --. 1997, Resource Competition: Population and Community Biology Series. Boston, MA,
 Springer.
- Hall, S. R., C. Becker, and C. E. Cáceres. 2007a. Parasitic castration: a perspective from a model
 of dynamic energy budgets. Integrative and Comparative Biology 47:295-309.
- Hall, S. R., C. R. Becker, M. A. Duffy, and C. E. Cáceres. 2010. Variation in resource
- acquisition and use among host clones creates key epidemiological trade-offs. American
 Naturalist 176:557-565.
- —. 2011. Epidemic size determines population-level effects of fungal parasites on *Daphnia* hosts. Oecologia 166:833-842.
- Hall, S. R., C. R. Becker, J. L. Simonis, M. A. Duffy, A. J. Tessier, and C. E. Cáceres. 2009a.

Friendly competition: evidence for a dilution effect among competitors in a planktonic

host-parasite system. Ecology 90:791-801.

Penczykowski et al.

- Hall, S. R., M. A. Duffy, A. J. Tessier, and C. E. Cáceres. 2005. Spatial heterogeneity of
 daphniid parasitism within lakes. Oecologia 143:635-644.
- Hall, S. R., J. L. Simonis, R. M. Nisbet, A. J. Tessier, and C. E. Cáceres. 2009b. Resource
- ecology of virulence in a planktonic host-parasite system: an explanation using dynamic
 energy budgets. American Naturalist 174:149-162.
- Hall, S. R., L. Sivars-Becker, C. Becker, M. A. Duffy, A. J. Tessier, and C. E. Cáceres. 2007b.
- Eating yourself sick: transmission of disease as a function of foraging ecology. EcologyLetters 10:207-218.
- Hilker, F. M., M. Langlais, and H. Malchow. 2009. The Allee effect and infectious diseases:
- extinction, multistability, and the (dis-)appearance of oscillations. American Naturalist
 173:72-88.
- Hite, J. L., and C. E. Cressler. 2019. Parasite-mediated anorexia and nutrition modulate virulence
 evolution. Integrative and Comparative Biology 59:1264-1274.
- Hite, J. L., R. M. Penczykowski, M. S. Shocket, K. A. Griebel, A. T. Strauss, M. A. Duffy, C. E.
- Cáceres et al. 2017. Allocation, not male resistance, increases male frequency during
 epidemics: a case study in facultatively sexual hosts. Ecology 98:2773-2783.
- Hite, J. L., A. C. Pfenning, and C. E. Cressler. 2020. Starving the enemy? Feeding behavior
 shapes host-parasite interactions. Trends in Ecology & Evolution 35:68-80.
- Hochachka, W. M., and A. A. Dhondt. 2000. Density-dependent decline of host abundance
- resulting from a new infectious disease. Proceedings of the National Academy of
- Sciences of the United States of America 97:5303-5306.

Penczykowski et al.

Foraging-mediated hydras in epidemics

- Hurtado, P. J., S. R. Hall, and S. P. Ellner. 2014. Infectious disease in consumer populations:
- dynamic consequences of resource-mediated transmission and infectiousness. Theoretical
 Ecology 7:163-179.
- Johnson, P. T. J., A. R. Townsend, C. C. Cleveland, P. M. Glibert, R. W. Howarth, V. J.
- McKenzie, E. Rejmankova et al. 2010. Linking environmental nutrient enrichment and
 disease emergence in humans and wildlife. Ecological Applications 20:16-29.
- Klüttgen, B., U. Dulmer, M. Engels, and H. T. Ratte. 1994. ADaM, an artificial freshwater for
 the culture of zooplankton. Water Research 28:743-746.
- Kooijman, S. A. L. M. 2010, Dynamic Energy Budget Theory for Metabolic Organisation. Great
 Britain, Cambridge University Press.
- Lafferty, K. D., and A. M. Kuris. 2009. Parasitic castration: the evolution and ecology of body
 snatchers. Trends in Parasitology 25:564-572.
- Lessios, H. A., D. R. Robertson, and J. D. Cubit. 1984. Spread of *Diadema* mass mortality
 through the Caribbean. Science 226:335-337.
- McIntire, K. M., and S. A. Juliano. 2018. How can mortality increase population size? A test of
 two mechanistic hypotheses. Ecology 99:1660-1670.
- Overholt, E. P., S. R. Hall, C. E. Williamson, C. K. Meikle, M. A. Duffy, and C. E. Cáceres.
- 2012. Solar radiation decreases parasitism in *Daphnia*. Ecology Letters 15:47-54.
- Peacor, S. D., and E. E. Werner. 2001. The contribution of trait-mediated indirect effects to the
- net effects of a predator. Proceedings of the National Academy of Sciences of the United
 States of America 98:3904-3908.
- Piñeiro, G., S. Perelman, J. P. Guerschman, and J. M. Paruelo. 2008. How to evaluate models:
- 897 Observed vs. predicted or predicted vs. observed? Ecological Modelling 216:316-322.

Penczykowski et al.

- Polis, G. A., W. B. Anderson, and R. D. Holt. 1997. Toward an integration of landscape and food
 web ecology: The dynamics of spatially subsidized food webs. Annual Review of
 Ecology and Systematics 28:289-316.
- Preston, D. L., and E. L. Sauer. 2020. Infection pathology and competition mediate host biomass
 overcompensation from disease. Ecology 101.
- Roelke-Parker, M. E., L. Munson, C. Packer, R. Kock, S. Cleaveland, M. Carpenter, S. J. Obrien
 et al. 1996. A canine distemper virus epidemic in Serengeti lions (*Panthera leo*). Nature
 379:441-445.
- Sanderson, C. E., and K. A. Alexander. 2020. Unchartered waters: Climate change likely to
- 907 intensify infectious disease outbreaks causing mass mortality events in marine mammals.908 Global Change Biology. In press.
- Sarnelle, O., and A. E. Wilson. 2008. Type III functional response in *Daphnia*. Ecology 89:17231732.
- Schröder, A., L. Persson, and A. M. de Roos. 2009. Culling experiments demonstrate size-class
 specific biomass increases with mortality. Proceedings of the National Academy of

Sciences of the United States of America 106:2671-2676.

- 914 Shocket, M. S., A. T. Strauss, J. L. Hite, M. Sljivar, D. J. Civitello, M. A. Duffy, C. E. Cáceres et
- al. 2018. Temperature drives epidemics in a zooplankton-fungus disease system: a trait-
- driven approach points to transmission via host foraging. American Naturalist 191:435-
- 917 451.
- Shurin, J. B., and E. W. Seabloom. 2005. The strength of trophic cascades across ecosystems:
- predictions from allometry and energetics. Journal of Animal Ecology 74:1029-1038.

- Stewart Merrill, T. E., and C. E. Cáceres. 2018. Within-host complexity of a plankton-parasite
 interaction. Ecology 99:2864-2867.
- 922 Stewart Merrill, T. E., S. R. Hall, L. Merrill, and C. E. Cáceres. 2019. Variation in immune
- defense shapes disease outcomes in laboratory and wild *Daphnia*. Integrative and
- 924 Comparative Biology 59:1203-1219.
- Strauss, A. T., A. M. Bowling, M. A. Duffy, C. E. Cáceres, and S. R. Hall. 2018. Linking host
 traits, interactions with competitors and disease: Mechanistic foundations for disease
 dilution. Functional Ecology 32:1271-1279.
- Strauss, A. T., D. J. Civitello, C. E. Cáceres, and S. R. Hall. 2015. Success, failure and ambiguity
 of the dilution effect among competitors. Ecology Letters 18:916-926.
- 930 Strauss, A. T., J. L. Hite, D. J. Civitello, M. S. Shocket, C. E. Cáceres, and S. R. Hall. 2019.

Genotypic variation in parasite avoidance behaviour and other mechanistic, nonlinear
components of transmission. Proceedings of the Royal Society B-Biological Sciences
286.

- Tessier, A. J., and P. Woodruff. 2002. Cryptic trophic cascade along a gradient of lake size.
 Ecology 83:1263-1270.
- 936 Vredenburg, V. T., R. A. Knapp, T. S. Tunstall, and C. J. Briggs. 2010. Dynamics of an

emerging disease drive large-scale amphibian population extinctions. Proceedings of the
National Academy of Sciences of the United States of America 107:9689-9694.

- Webb, D. J., B. K. Burnison, A. M. Trimbee, and E. E. Prepas. 1992. Comparison of chlorophyll
- 940 *a* extractions with ethanol and dimethyl sulfoxide/acetone, and a concern about
- 941 spectrophotometric phaeopigment correction. Canadian Journal of Fisheries and Aquatic942 Sciences 49:2331-2336.

Penczykowski et al.

- 943 Welschmeyer, N. A. 1994. Fluorometric analysis of chlorophyll *a* in the presence of chlorophyll
- *b* and pheopigments. Limnology and Oceanography 39:1985-1992.

Table 1. Results of the model competition to estimate foraging rate, $f(L_H, \sigma)$. Models 1a–6a fit a common 'foraging' parameter (f or size-specific \hat{f}) to infected and uninfected hosts together for each genotype. In models 1b–6b, foraging parameters (f_j or \hat{f}_j) were estimated separately for uninfected (f_S or \hat{f}_S) and infected (f_I or \hat{f}_I) hosts in each genotype. Body length, L_H , and spore yield, σ , were measured empirically (Fig. 1d-f), and we estimated the linear sensitivity coefficient (α , mm³·spore⁻¹) and power coefficient (γ) for each genotype.

	Foraging	D h	ars ^b AIC	ΔAIC ^c	Akaike
Model	rate, $f(L_H, \sigma)^a$	Pars			weight $(w_i)^d$
(6b) Size and spores, power	$\hat{f}_j L_H^2 \left(1 - \alpha \left(\frac{\sigma}{L_H^3}\right)^{\gamma}\right)$	18	-213.5	0.0	0.62
(5b) Size and spores,	$\hat{f}_j L_H^2 \left(1 - \alpha \left(\frac{\sigma}{L_H^3} \right) \right)$	15	-212.5	0.9	0.38
(3b) Spores only, linear	$f_j\left(1-\alpha\left(\frac{\sigma}{L_H^3}\right)\right)$	15	-96.3	117.2	2.2×10^{-26}
(4b) Spores only, power	$f_j \left(1 - \alpha \left(\frac{\sigma}{L_H^3}\right)^{\gamma}\right)$	18	-94.3	119.2	$8.2 imes 10^{-27}$
(2b) Size only	$\hat{f_j} \left. L_{\scriptscriptstyle H} ight.^2$	12	-74.2	139.3	$3.5 imes 10^{-31}$
(6a) Size and spores, power	$\hat{f} L_{H}^{2} \left(1 - \alpha \left(\frac{\sigma}{L_{H}^{3}}\right)^{\gamma}\right)$	12	-71.3	142.1	8.5×10^{-32}
(5a) Size and spores, linear	$\hat{f} L_{H}^{2}\left(1-\alpha\left(\frac{\sigma}{L_{H}^{3}}\right)\right)$	9	-67.8	145.6	$1.5 imes 10^{-32}$
(1b) Null	f_j	12	-6.1	207.4	5.7×10^{-46}

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(3a) Spores only, linear
$$f\left(1-\alpha\left(\frac{\sigma}{L_{H}^{3}}\right)\right)$$
 9 136.5 350.0 6.2×10^{-77}
(2a) Size only $\hat{f} L_{H}^{2}$ 6 143.6 357.1 1.8×10^{-78}
(4a) Spores only, power $f\left(1-\alpha\left(\frac{\sigma}{L_{H}^{3}}\right)^{\gamma}\right)$ 12 144.7 358.2 1.0×10^{-78}
(1a) Null f 6 262.1 475.6 3.3×10^{-104}

952 ^{*a*} Per capita rates; technically, this is "clearance rate". Units for f_j : L·day⁻¹; for size-specific \hat{f}_j :

953 $L \cdot mm^{-2} \cdot day^{-1}$

^b Number of parameters estimated for hosts of three genotypes, including a variance parameter

estimated for each infection class (models 1b–6b) and genotype (all models). Parameters α and γ

were fixed at zero (i.e., not estimated) for uninfected hosts in models 3b–6b.

^c The winning model has $\Delta AIC = 0$. Models with $\Delta AIC > 10$ have essentially no support.

 d The probability that the model is the best among those under consideration.

Table 2. Variables, parameters, and functions used in the dynamical host-parasite-resource

Symbo	ol Meaning	Default values/range	Units
A	density of resource of host		mg C·L ⁻¹
Ι	density of infected hosts		host·L ⁻¹
Ν	density of hosts: $N = S + I$		host·L ⁻¹
S	density of susceptible hosts		host·L ⁻¹
t	time		day
Ζ	density of parasite propagules (spores)		spore·L ⁻¹
d	background death rate of hosts	0.03	day ^{-1 a}
е	conversion efficiency of host	8.5	mg C ^{-1 <i>b</i>}
L_H	body length of host	1.4	mm
\hat{f}_{s}	size-specific foraging rate for susceptible	0.0178	L·mm ⁻² ·day ⁻¹
	hosts		
\hat{f}_I	size-specific foraging rate for infected	0.0131	L·mm ⁻² ·day ⁻¹
	hosts		
α	sensitivity coefficient of foraging of	$2.86 \times 10^{\text{-5}}$ (Fig. 2h:	mm ³ ·spore ⁻¹
	infected hosts	0.35-2.8 x 10 ⁻⁵)	
f	mean foraging rate of hosts: $f = (1-p)f_S +$		$L \cdot day^{-1}$
	pfi		
f_S	foraging rate of susceptible hosts (equ. 2a)	0.035	$L \cdot day^{-1}$
f_I	foraging rate of infected hosts (equ. 2b)	0 (bounded)-0.035	L·day ⁻¹

model (equ. 3), with default values or ranges (when applicable).

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h	half saturation constant, spore yield	0.015	mg C·L ^{-1 c}
K	carrying capacity of resources	0.1-3.0	$mg \ C \cdot L^{\text{-}1}$
т	mortality of spores, Z	0.8	day ^{-1 d}
р	prevalence of infection: <i>I</i> /(<i>S</i> + <i>I</i>)		unitless
PP	primary production: $PP = r_m A(1-A/K)$		mg $C \cdot L^{-1} \cdot day^{-1}$
r _m	max. per capita growth rate of A	0.8	day ⁻¹ a
и	susceptibility	0.0004	host·spore ⁻¹ ^e
v	virulence on survival	0.07 (Fig. 2g: 0.03-	day ⁻¹
		0.15)	
$\sigma(A)$	spore yield (equ. 3.d)		spore · host -1
σ_1	maximum spore yield	$1.0-2.5 \text{ x } 10^3 \text{ (Fig.}$	spore∙host⁻
		2g,h: 1.5 x 10 ⁴)	$^{1} \cdot mg C^{-1}$

- ^{*a*} Reasonable value for this host and algae.
- ^b Yields a reasonable instantaneous birth rate, b, of 0.30 day⁻¹ for uninfected hosts at 1.0 mg C L⁻¹
- 963 (where $b = ef_{S}A$) (Hall et al. 2010).
- ^{*c*} Reasonable value for this host (Strauss et al. 2015).
- d A high loss rate due to solar radiation (Overholt et al. 2012) and other sources (e.g.,
- consumption by non-focal hosts (Hall et al. 2009a)).
- 967 *e* Yields an infection risk (transmission rate, β) of 1.4 x 10⁻⁵ L·spore⁻¹·day⁻¹ (where $\beta = uf_S$).

968

FIGURE LEGENDS

Fig. 1. Parasites depress host foraging rate, f, as functions of host length, L_{H} , and spore 969 yield, σ . Foraging rate (a-c): Foraging rate (f, mean ± 1 SE) of uninfected (f_s, white circles) and 970 infected (f_I , black circles; exposed to spores when six days old) individuals of three genotypes of 971 a zooplankton host with the best-fitting foraging function (f_s , equ. 2.a, solid lines; f_l , equ. 2.b, 972 dashed lines). (a,b) For genotypes A4-4 and BD-30, foraging rate increased with age (thus, body 973 974 size) of uninfected hosts and those at early stages of infection. Foraging then dropped as infected 975 hosts filled with spores. (c) Infection reduced foraging rate earlier for the STD genotype. Spore yield and host length (d-f): Host length (L_H) of uninfected (white circles) and infected (black 976 circles) hosts and spore yield (σ , grey squares) of three host genotypes: (d) A4-4, (e) BD-30, (f) 977 978 STD. Spore yield also increased with age (noting a few [N = 3], smaller STD hosts at 24 days). *P*-values are from GLM-based tests of age (A), infection (I), and their interaction ($A \times I$) on 979 length, and of age on spore yield (p < 0.001***, p < 0.01**, p < 0.05*). Asterisks above body 980 length points indicate significant post-hoc pairwise differences (Tukey's) between infection 981 classes. Letters denote significant post-hoc differences in spore yield between age classes. 982 Points: means ± 1 SE. 983

Fig. 2. *A fully dynamical model reveals a trait-mediated hydra effect through depression of foraging rate.* Equilibrial density of hosts, N^* , can increase during epidemics of a virulent parasite over gradients of carrying capacity, *K* (x-axis). Disease-free states are denoted by thick contours. For epidemics, arrows across contours show increasing values of maximal spore yield σ_1 (spore-host⁻¹·mg C⁻¹×10⁴; panels a-f), virulence, *v* (day⁻¹; panel *g*), or sensitivity coefficient of foraging of infected hosts, α (mm³·spore⁻¹; panel h). (a) Equilibrial disease prevalence (proportion infected, p^*); (b) mean per capita death rate (*d*+*vp*^{*}); (c) algal resources (*A*^{*}); (d)

spore yield ($\sigma(A)$); (e) mean foraging rate ($f = [1 - p^*] f_S + p^* f_I$; (f) primary production ($PP = r_m A^*(1 - A^*/K)$; (g) resource consumption per host (fA^*); and (h) total host density (N^*). Hydras arise at higher K (N^* higher with disease [thin] than without) and become larger with higher σ_1 . The hydra effect was accentuated by (i) lower virulence on survivorship (ν [day⁻¹]) and (j) higher feeding sensitivity (α [×10⁻⁵ mm³·spore⁻¹]). Therefore, hydras were more likely with higher carrying capacity of the resource (K) and for parasites that depress mortality less strongly (lower ν) and foraging more strongly (higher α).

Fig. 3. Parameter space predicting trophic cascades (host density decreases, $N \downarrow$) or 998 foraging-mediated hydra effects ($N\uparrow$) over gradients of carrying capacity (K) of the resource. 999 (a,b) Foraging-mediated hydras occur at a given K if virulence mortality, v, is not too high 1000 (below solid lines). Scenarios assuming susceptible hosts feed faster than infecteds ($\hat{f}_{s} > \hat{f}_{l}$): (a) 1001 foraging is sensitive to spores ($\alpha > 0$) and (b) is not ($\alpha = 0$). (c,d) Foraging-mediated hydras are 1002 predicted, at a given K, when the sensitivity coefficient, α , exceeds a threshold, which is smaller 1003 when susceptible hosts feed faster than infected hosts even without spore build up (i.e., $\hat{f}_S > \hat{f}_I$, 1004 1005 equ. 2; solid line, white and yellow region) than when they feed at the same rate without spores $(\hat{f}_{S} = \hat{f}_{I}, \text{ dashed line, yellow region alone}).$ (c) Higher virulence on mortality ($v = 0.07 \text{ day}^{-1}$); (d) 1006 lower virulence ($v = 0.03 \text{ day}^{-1}$). All parameters follow defaults in Table 2. 1007

Fig. 4. *Changes in hosts and algal resources create a 'joint algal-host response'*. (a) During the large epidemic in Goodman Lake (max. infection prevalence: 48.6%; see also Fig. A4), both hosts and algal resources increased through time. (b) During Long Lake's small epidemic (max. prevalence: 5.2%), hosts increased but algal resources declined. (c,d) The joint algal-host response index for (c) Goodman and (d) Long is calculated using cross products and the standardized temporal slopes (vectors). The 'joint algal-host response' index is the cross

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product of these vectors, i.e., their product (area), illustrated as grey rectangles. (This joint indexis presented in Fig.5c,d.)

Fig. 5. *Evidence for a joint increase in densities of zooplankton (Daphnia) hosts and algal resources during natural fungal epidemics (a hydra).* (a) Infected hosts produced more spores (σ), during larger epidemics (higher maximum prevalence of infection, p_{max}). (b) These greater spore loads depressed average per capita feeding rate of adult hosts, *f* (calculated using length data; Fig. A4). (c) Stronger parasite-induced depression of foraging rate correlated with a larger index of joint algal–host response (Fig. 4c,d) through time during epidemics. (d) The joint algal–host response index was larger during bigger fungal outbreaks. Points are lake means.

Pearson correlation coefficients (r) are accompanied by corresponding P-values.



Figure 1

Figure 2



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Figure 3



Figure 4



Figure 5

