

1 **Ensembles of multiple spectral water indices for improving surface water**

2 **classification**

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18
19 **Abstract:** Mapping surface water distribution and its dynamics over various environments with robust
20 methods is essential for managing water resources and supporting water-related policy design. Thresholding
21 Single Water Index image (TSWI) with a fixed threshold is a common way of using water index (WI) for
22 mapping water for it is easy to use and could obtain acceptable accuracies in many applications. As more and
23 more WIs are available and each has its distinct merits, the real-world application of TSWI, however, often
24 face two practical concerns: (1) selection of an appropriate WI, and (2) determination of an optimal threshold
25 for a given WI. These two issues are problematic for many users who rely either on trial-and-error procedures
26 that are time-consuming or on their personal preferences that are somewhat subjective. To better deal with
27 these two practical concerns, an alternative way of using WIs is suggested here by transforming the current

28 paradigm into a simple but robust ensemble approach called Collaborative Decision-making with Water
29 Indices (CDWI). A total of 145 subsite images (900×900 m) from 22 Landsat-8 OLI scenes that covering
30 various water-land environments around the world were used to assess the performance of TSWI and the
31 CDWI. Five benchmark WIs were adopted in five TSWI methods and CDWI method: Normalized Difference
32 Water Index (NDWI), the Modified NDWI (MNDWI), the Automated Water Extraction Indices without
33 considering (AWEI0) and with considering (AWEI1) shadows, and the state-of-the-art 2015 water index
34 (WI2015). Two aspects of performance were analyzed: comparing their accuracies (indicated by both
35 F1-scores and Youden's Index) over various environments and comparing their accuracy sensitivities to
36 threshold. The results demonstrate that CDWI produced higher accuracies than the other five TSWI methods
37 for most application cases. Particularly, more samples (indicated by percentage) produced higher F1-scores by
38 CDWI than the other five TSWI methods, i.e. 67% (CDWI) vs. 15% (TSWINDWI), 54% (CDWI) vs. 22%
39 (TSWIMNDWI), 42% (CDWI) vs. 12% (TSWIAWEI0), 57% (CDWI) vs. 17% (TSWIAWEI1), and 34%
40 (CDWI) vs. 12% (TSWIWI2015). Moreover, the F1-score of the CDWI is much less sensitive to the change of
41 thresholds compared with that of the other five TSWI methods. These important benefits of CDWI make it a
42 robust approach for mapping water. The uncertainty of CDWI method was thoroughly discussed and a general
43 guidance (or look-up-table) for selecting WIs was also suggested. The underlying framework of CDWI could
44 be readily generalizable and applicable to other satellite sensor images, such as Landsat TM/ETM+, MODIS,
45 and Sentinel-2 images.

46 **Keywords:** Water index, Threshold, Integrated decision making, Mixed pixels, MNDWI

47 **1. Introduction**

48 Inland water is an important earth resource for providing ecosystem services (Karpatne et al., 2016;
49 Ogashawara et al., 2017), such as being a key habitat for flora and fauna of aquatic ecosystems and support
50 biodiversity conservation (Vörösmarty et al., 2010). It is also a key component of Earth's hydrologic cycle and,
51 as such, can support many aspects of daily life, including drinking water, agricultural irrigation, electricity
52 production, and transportation (Huang et al., 2018). Spatially explicit monitoring of water changes is, therefore,
53 essential for a variety of scientific disciplines and to inform land-use policy and decision-making (Berry et al.,

54 2005; Ma et al., 2010; Pekel et al., 2016).

55 As remote sensing is well recognized for detecting spatiotemporal patterns of land cover, it has been
56 widely used for monitoring water changes with various purposes, such as water resource inventory, flooding
57 and drought assessment, and urban hydrological evaluation (Allen and Pavelsky 2018; Berry et al., 2005; Shao
58 et al., 2019). Generally, the success of mapping water bodies with remote sensing images relies on the distinct
59 reflectance spectra of water in comparison with other land features: water generally show lower reflectance
60 and a decreasing pattern of reflectance from visible to infrared spectral wavelengths (Bukata et al., 2018).
61 Based on such optical characteristics, various types of water classification methods have been developed which
62 can be broadly grouped into indirect and direct strategies.

63 The indirect strategy considers water bodies as one of several broad land cover categories, and the water
64 bodies can be extracted from a land use/land cover map derived from image classification methods, such as
65 deep learning, random forest, support vector machine (Cao et al., 2019). The direct classification strategy is to
66 classify an image into water and non-water (land) categories directly. It is easy to use and widely adopted in
67 practice (Allen and Pavelsky 2018; Berry et al. 2005; Cooley et al. 2017; Guo et al. 2017). One of the most
68 common approaches is called Thresholding Single Water Index (TSWI), in which the water index (WI) is
69 derived from two or more spectral bands with a carefully designed algorithm and water pixels would gain high
70 values and the non-water pixels would gain low values (Ji et al., 2009). In the processing of TSWI, selecting a
71 WI and generating corresponding WI image should be done first, and then pixels in such WI image with their
72 values higher than (or lower than in some cases) a predefined appropriate threshold are categorized as water,
73 otherwise non-water (Huang et al., 2018).

74 As WIs are sensor dependent, only the WIs designed for Landsat images are focused on this research. The
75 Normalized Difference Water Index (NDWI; McFeeters 1996), is considered as the first-generation WI for
76 using TSWI to classify water. It is calculated using the green and near-infrared (NIR) bands of Landsat TM
77 with an equation similar to NDVI which is used for vegetation (Tucker 1979), and the threshold 0 is suggested
78 for thresholding water areas. NDWI was the most widely used index (McFeeters, 2013) before the Modified
79 Normalized Difference Water Index (MNDWI) was introduced by Xu (2006). MNDWI was designed because
80 using NDWI with TSWI cannot efficiently suppress the signal from built-up areas, such that the suggested

81 threshold 0 fails to distinguish water bodies from built-up surfaces accurately. The equation of MDNWI is
82 similar to NDWI, but the NIR band is replaced by the first shortwave infrared (SWIR1) band of Landsat TM
83 imagery. MNDWI is the most widely used WI for a variety of applications, including surface water mapping,
84 land use/cover change analyses, and ecological monitoring research (Allen and Pavelsky 2018; Ji et al., 2009).
85 In certain situations, however, the performance of MNDWI may be relatively poor due to the presence of low
86 reflectance surfaces such as asphalt roads and shadow effects. To overcome such issues, Feyisa et al. (2014)
87 proposed two new WIs, Automated Water Extraction Index with (AWEI1) and without (AWEI0) considering
88 shadows. AWEI0 and AWEI1 are considered highly useful WIs and have been applied with TSWI to extract
89 water bodies from Landsat imagery (Huang et al., 2018; Jiang et al., 2014). Fisher et al. (2016) conducted a
90 comprehensive inter-comparison of the existing WIs and designed the latest water index (WI2015). The
91 WI2015 is derived from linear discriminant analysis and involves all the bands of Landsat TM/ETM+ except
92 for the blue band and it has demonstrated similar accuracy to some of the prevailing WIs.

93 The driving force behind proposing different WIs indicates the fact that water-land environments in the
94 real-world are very heterogeneous and the stability of applying TSWI with any single WI would vary a lot over
95 different environments (Wu et al., 2018, Yang et al., 2018). Therefore, an average user of TSWI would face
96 two basic concerns: (1) *which WI* should be chosen from existed WIs, and (2) what is the *appropriate*
97 *threshold* that should be used for a given WI?

98 In general, the answer to the first concern involves some personal preference because there is no clear
99 guidance of WI selection and a WI performs unsteadily over different water-land environments, such as
100 wetland, mountain, urban, forest, and desert (Fisher et al., 2016; Ji et al., 2009). As a consequence, the same
101 image classified by different TSWI users could produce inconsistent results due to different choices of WIs and
102 the corresponding thresholds (Feyisa et al., 2014; Huang et al., 2018). For the second concern, three types of
103 thresholds have been reported according to the availability of ground reference data, i.e., the real outline of
104 water bodies that were obtained at the same time as the image acquisition time. **Case 1:** If enough reference
105 data is available in an application, the local optimal threshold is suggested because such threshold can be
106 determined (or trained) by the reference data. In most average applications, however, the obtaining of timely
107 reference data could be difficult, especially for highly dynamic water landscapes (e.g., rivers and wetlands

108 during flood events). **Case 2:** If there is no reference data, the locally-adaptive threshold and pre-defined
109 threshold could be the choices. The locally-adaptive threshold is determined by the WI image itself with some
110 segmentation technologies, so that the thresholds can vary self-adaptively for different images (Huang et al.,
111 2018; Li and Sheng 2012; Wen et al., 2020). One obvious shortcoming of locally-adaptive threshold is that it
112 heavily depends on the applied image extent and its land/water ratio, such that threshold can be vastly different
113 for the same location when it is determined from different extents (Zhang et al., 2018). The pre-defined
114 thresholds are often recommended by the original WI inventors or by other experienced authorities. To the best
115 of our knowledge, the pre-defined thresholds are widely used in average water mapping applications for they
116 are super easy to be applied. However, this type of thresholds should be used with caution because they cannot
117 guarantee satisfying results due to the complex water-land environments in the real world (Feyisa et al., 2014;
118 Fisher et al., 2016).

119 In summary, the application of TSWI faces two common concerns as mentioned above and the ways to
120 deal with them are unsatisfied if there is no sufficient reference data. Thus, alternative solutions have been
121 explored over the past few years (Huang et al., 2018), including the construction of new WIs that are robust
122 and relatively insensitive to threshold selection or the development new methods using multiple existing WIs
123 (Sánchez et al., 2018; Wang et al., 2018). The latter is regarded as the most appropriate approach because the
124 combination of multiple WIs could complement their merits and apply to different environments compared
125 with TSWI method (Yang et al., 2015). Such strategy is, to some extent, in line with the collaborative
126 decision-making theory where multiple variables can produce complementary information to support a more
127 robust result than each individual variable (Kacprzyk and Fedrizzi 2012).

128 Inspired by these ideas, this research aims to propose a new way of using WIs based on collaborative
129 decision-making theory to deal with the two concerns mentioned above that exist commonly in TSWI method.
130 Such new approach has the advantages of: (1) less concerned about the WIs selection and (2) less sensitive to
131 WIs thresholds than TSWI method. Specifically, the new approach is transforming the current paradigm of
132 using WIs (i.e., TSWI method) into a simple but highly robust ensemble way of using WIs called Collaborative
133 Decision-making with Water Indices (CDWI). The CDWI (the new way of using WIs) was tested in a variety
134 of water-land environments around the world and assessed by comparing its performances with that of TSWI

135 (the common way of using WIs) using five benchmarked WIs.

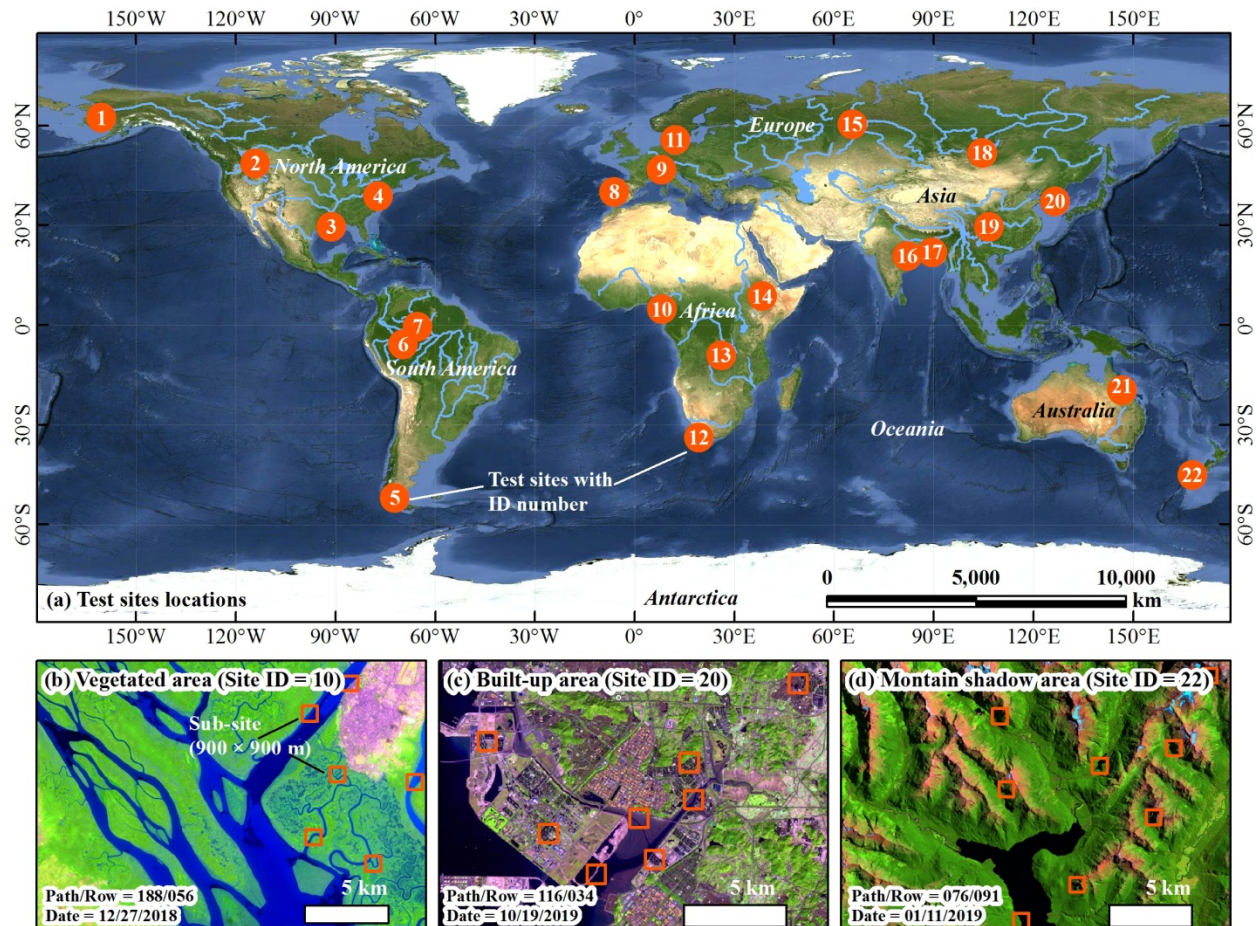
136

137 **2. Test sites and data materials**

138 2.1. Test sites and subsites

139 Performances of water classification methods are generally affected by two error sources: the applied
140 aquatic environments and their surrounding land features(Wu et al., 2018, Yang et al., 2018). The aquatic
141 environments are often characterized by a variety of watercolors (e.g., dark, yellow, red, and brown, etc.) and
142 water types (e.g., river, reservoir, pond, and ditch, etc.). The surrounding land features are usually recognized
143 as vegetation conditions (high-density vegetation, sparse vegetation, etc.), built-up area (road and buildings),
144 and shadows (cloud shadow, building shadows, and terrain shadows). The combinations of these two error
145 sources make the selection of test sites tricky and time-consuming. Fortunately, many test sites have already
146 been used for validating water classification methods in previous studies and such sites can guide us for
147 selecting test sites in this study. Finally, 22 test sites were carefully selected with some come from Yang et al.
148 (2015) and Feyisa et al. (2014) and some newly selected by considering their spatial representativeness (Fig. 1).
149 These sites scattered around the world and covered a variety of water-land environments (Table 1).

150 Among each test site, several subsites with 900×900 m square size each were selected for preparing test
151 data (as exemplified in Figs. 1b, 1c, and 1d). The subsites were manually selected with expert knowledge in true
152 color composite Landsat-8 OIL images (R: Band 4, G: Band 3, B: Band 2) by following two criteria: (1) the
153 subsites should cover both water and land; (2) the subsites should cover as many different types of watercolors,
154 water types, and land features as possible. Overall, 145 subsites were selected from these 22 test site (Table 1).
155 Although various land features have been covered by these subsites, their sample sizes (or area) varied
156 significantly due to their different frequencies of presences in the real world. For example, vegetated land area
157 could be more likely to be sampled than shadowed land near water bodies. To mitigate such imbalanced
158 sample sizes, 35 additional subsites only covered “uncommon” land features (e.g., built-up land, shadowed
159 land) were selected. Finally, a total of 180 subsites were prepared as the test dataset.



160

161 **Fig. 1.** (a) Locations of the 22 test sites representing three types of water-land environments: water bodies
 162 surrounded by vegetated land, built-up land, and shadowed land. The numbers (1 - 22) mark site IDs. (b), (c),
 163 and (d) are examples of test site images (R: Band 6, G: Band 5, and B: Band 4 in Landsat-8 OLI image)
 164 illustrating water bodies surrounded by vegetated area, built-up area, and mountain shadow area, respectively.
 165 The red squares (900×900 m) denote subsites that were extracted for preparing test data. All of the 22 test site
 166 images are shown in the supplementary Fig. S1.

167 **Table 1** Selected 22 test sites and corresponding Landsat-8 images with different environmental conditions.
 168 Watercolors include dark-blue (D), green (G), brown (B), dark-blue-green (DG), dark-blue-brown (DB), and
 169 green-brown (GB). Their typical colors are illustrated in the table's header. Water types include river (R),
 170 lake/reservoir/pond (LPR), and ditch/creek (DC). Background features include high-density vegetation (HV),
 171 moderate-density vegetation (MV), sparse vegetation (SV), built-up area (BA), cloud shadow (CS), building
 172 shadow (BS), and terrain shadow (TS).

Site ID	Path/Row	Image Date	Watercolor						Water type			Land features					
			D	G	B	DG	DB	GB	R	LRP	DC	HV	MV	SV	BA	CS	BS
1	075/016	06/13/2019	•	•					•		•	•	•			•	
2	041/026	09/03/2019	•			•							•	•			•
3	023/039	10/23/2019	•	•	•	•			•	•	•	•	•		•		
4	015/033	10/15/2019	•	•				•				•	•	•	•		
5	230/096	02/03/2019	•		•		•			•			•	•			
6	002/061	03/26/2019			•				•		•	•				•	
7	001/060	08/26/2019	•	•								•					
8	203/032	09/18/2019	•	•		•			•	•			•	•			
9	195/027	07/24/2019		•		•				•		•			•	•	•
10	188/056	12/27/2018			•		•		•		•	•	•		•		
11	195/021	04/19/2019		•		•		•		•			•	•	•		
12	175/083	11/14/2018	•					•		•			•	•	•		•
13	174/066	06/19/2019	•			•				•	•	•	•				
14	168/054	02/01/2019	•	•				•		•			•	•			•
15	162/018	05/30/2019	•			•				•	•	•	•	•			
16	142/045	04/16/2019	•				•			•			•	•	•		
17	138/045	03/03/2019			•		•		•		•	•	•				
18	134/024	08/30/2019	•						•	•			•	•	•		
19	127/040	08/26/2018			•	•		•	•	•		•	•		•		•
20	116/034	10/19/2019			•		•		•	•	•		•	•	•		•
21	095/073	07/12/2019		•		•				•	•			•	•		
22	076/091	01/11/2019	•							•			•	•			•

173 2.2. Data Materials

174 2.2.1 Landsat-8 OLI images

175 A total of 22 Landsat-8 OLI images with each covered one test site and acquired in different seasons were

176 selected (Table 1). They were standard Landsat-8 surface reflectance level-2 products with 30 m spatial
177 resolution and more information of those products can be found in the Product Guide (2018). The images were
178 firstly downloaded from the USGS Earth Resources Observation and Science Center Science Processing
179 Architecture on Demand Interface (<https://espa.cr.usgs.gov/>) and then were clipped into sub-images using
180 subsite-defined square polygons (900 × 900 m, see Fig. 1). Only the pixels that entirely contained by the
181 subsite square polygons were selected. In total, 180 clipped subsite images with 153140 pixels of seven-band
182 surface reflectances (range from 0 to 1 in float) were extracted and stored as integer values by scaling 10,000
183 (any pixels with values less than 0 or greater than 10,000 were masked).

184 *2.2.2 High spatial resolution images*

185 PlanetScope Analytic Ortho Scene (PSAOS) products were served as reference data for labeling
186 Landsat-8 pixels as water and non-water. PSAOS images have a high spatial resolution (3 m) and very high
187 temporal resolution (1-3 days), which makes them ideal reference data sources. They consist of four bands:
188 blue (455 – 515 nm), green (500 – 590 nm), red (590 – 670 nm), and near-infrared (NIR, 780 – 860 nm).
189 Before distributed to users, they are orthorectified to remove distortion caused by terrain and to eliminate the
190 perspective effect on the ground (not on buildings), as well as to restore the geometry of an image taken at
191 zenith (Planet Labs Inc., 2018).

192 Each PSAOS image was carefully selected in this study such that their acquisition dates matched exactly
193 the same as that of the corresponding Landsat images (Table 1). In other words, both the PSAOS image and
194 corresponding Landsat-8 image were captured on the same day. All the PSAOS images were obtained from
195 Planet Explorer (<https://www.planet.com/explorer/>; Planet Team, 2017) and manually georeferenced to the
196 corresponding Landsat-8 image. The geo-referencing errors of PSAOS images were less than one pixel (30 m),
197 which minimized the geolocation error that could potentially propagate to the final classification results.

198 *2.2.3 Test dataset preparation*

199 Each test pixel (153,140 in total) holds several attributions: location, source image, band reflectance, WIs
200 values, feature type (water or non-water), percentage of water. The first three attributions were directly
201 obtained from the source Landsat-8 image. WIs values were derived from band reflectance with specific

202 algorithms (detailed in Section 3.1). Feature type and percentage of water were identified with the help of the
 203 PSAOS reference images which involved three steps. First, the PSAOS images were displayed in false-color
 204 (R: NIR, G: Red, B: Green) and carefully classified into water (including different watercolors) and non-water
 205 polygons (including vegetated land, built-up land, or shadowed land) through visual digitization with expert
 206 experience. Then, the water area percentage of each corresponding 30 m by 30 m pixel was derived with a
 207 series of spatial analysis functions (e.g., create fishnet, clip, etc.) coded in Python script in ArcGIS 10.5
 208 (version 10.5.0.6491; ESRI, 2016). Finally, all the pixels with water percentage higher than 50% were labeled
 209 as water, otherwise as non-water (Feyisa et al., 2014; Yang et al., 2015). The non-water type was further
 210 identified as vegetated land, built-up land, or shadowed land. In addition, pixels with water percentage equal to
 211 0 (non-water type) or 100% (water) were considered as pure pixels, otherwise as mixed pixels. The numbers of
 212 water pixels, non-water pixels, pure pixels, and mixed pixels are listed in Table 2. The dataset is now available
 213 at Mendeley Data repository (<http://dx.doi.org/10.17632/mfp7jvw7yk.1>).

214 **Table 3** Count numbers of water pixels, non-water pixels, pure pixels, and mixed pixels in the test dataset

	Pure pixels	Mixed pixels	Total
Water pixels	47024	5837	52861
Non-water pixels	93973	6306	100279
Total	140997	12143	153140

215 **3. Methods**

216 3.1. The common way of using spectral water indices: TSWI

217 Although numerous Landsat WIs have been developed over the past three decades, five are prevailing
 218 with distinct merits for different water-land environments: NDWI, MNDWI, AWEI0 (also known as AWEI_{nsh}),
 219 AWEI1 (also known as AWEI_{sh}), and WI2015 (Table 3). The application of TSWI for water classification is
 220 straightforward: applying a pre-defined threshold to a pre-selected single WI image. Pixels with values larger
 221 than the threshold are labeled as water, otherwise, they are labeled as non-water. Please note that the
 222 applications of TSWI method using NDWI, MNDWI, AWEI0, AWEI1, and WI2015, are denoted hereafter as
 223 TSWI_{NDWI}, TSWI_{MNDWI}, TSWI_{AWEI0}, TSWI_{AWEI1}, TSWI_{WI2015}, respectively.

224

225 **Table 3** Five prevailing WIs used in TSWI for mapping water bodies with Landsat-8 OLI images. ρ is surface
 226 reflectance and b_1, b_2, \dots, b_7 are band numbers of Landsat-8 OLI images. The superscript notes “a” and “b”
 227 indicate the pre-defined thresholds suggested by the source authors and Fisher et al. (2016), respectively. Note
 228 that the pre-defined thresholds suggested by Fisher et al. (2016) were also adopted in the proposed CDWI.

Water index	Equation adjusted for Landsat-8 OLI	Source reference	Pre-defined Threshold ^a	Pre-defined threshold ^b
NDWI	$(\rho_{b3}-\rho_{b5}) / (\rho_{b3}+\rho_{b5})$	McFeeters (1996)	0.00	-0.21
MNDWI	$(\rho_{b3}-\rho_{b6}) / (\rho_{b3}+\rho_{b6})$	Xu (2006)	0.00	0.00
AWEI0	$4(\rho_{b3}-\rho_{b6})-0.25\rho_{b5}-2.75\rho_{b7}$	Feyisa et al. (2014)	0.00	-0.07
AWEI1	$\rho_{b2}+2.5\rho_{b3}-1.5(\rho_{b5}+\rho_{b6})-0.25\rho_{b7}$	Feyisa et al. (2014)	0.00	-0.02
WI2015	$1.7204+171\rho_{b3}+3\rho_{b4}-70\rho_{b5}-45\rho_{b6}-71\rho_{b7}$	Fisher et al. (2016)	0.63	0.63

229 3.2. The ensemble spectral water indices: CDWI

230 3.2.1 Principle of CDWI

231 An alternative way of using WIs for water classification is proposed here to handle the common concerns
 232 of using TSWI: WI selection and the corresponding threshold determining. The approach is designed as the
 233 Collaborative Decision-making with Water Indices (CDWI). It combines a group of weighted and thresholded
 234 WI images to generate a new water probability image and a new decision-making probability threshold is
 235 applied to extract water. The rationale of the collaborative decision-making principle is that a group of
 236 variables can provide potentially complementary information to support a more reliable decision than that
 237 based on a single component (Kacprzyk and Fedrizzi 2012). When it comes to handling the concerns of TSWI,
 238 CDWI could provide an alternative way of selecting WIs and a potential stable threshold for extracting water.
 239 The step-by-step procedure of CDWI is as follows (see also Fig. 2) and the ready-to-use Python script is
 240 attached as a supplementary file.

- 241 ● **Step 1:** Select a group of WIs and calculate corresponding WI images. In this study, the five prevailing
 242 WIs were used as listed in Table 3. The reason for selecting these WIs is that they were reported showing
 243 complementary merits in classifying water over different water-land environments. For example,
 244 MNDWI was designed to separate water from vegetated area and built-up area (Ji, et al., 2009; Xu, 2006),

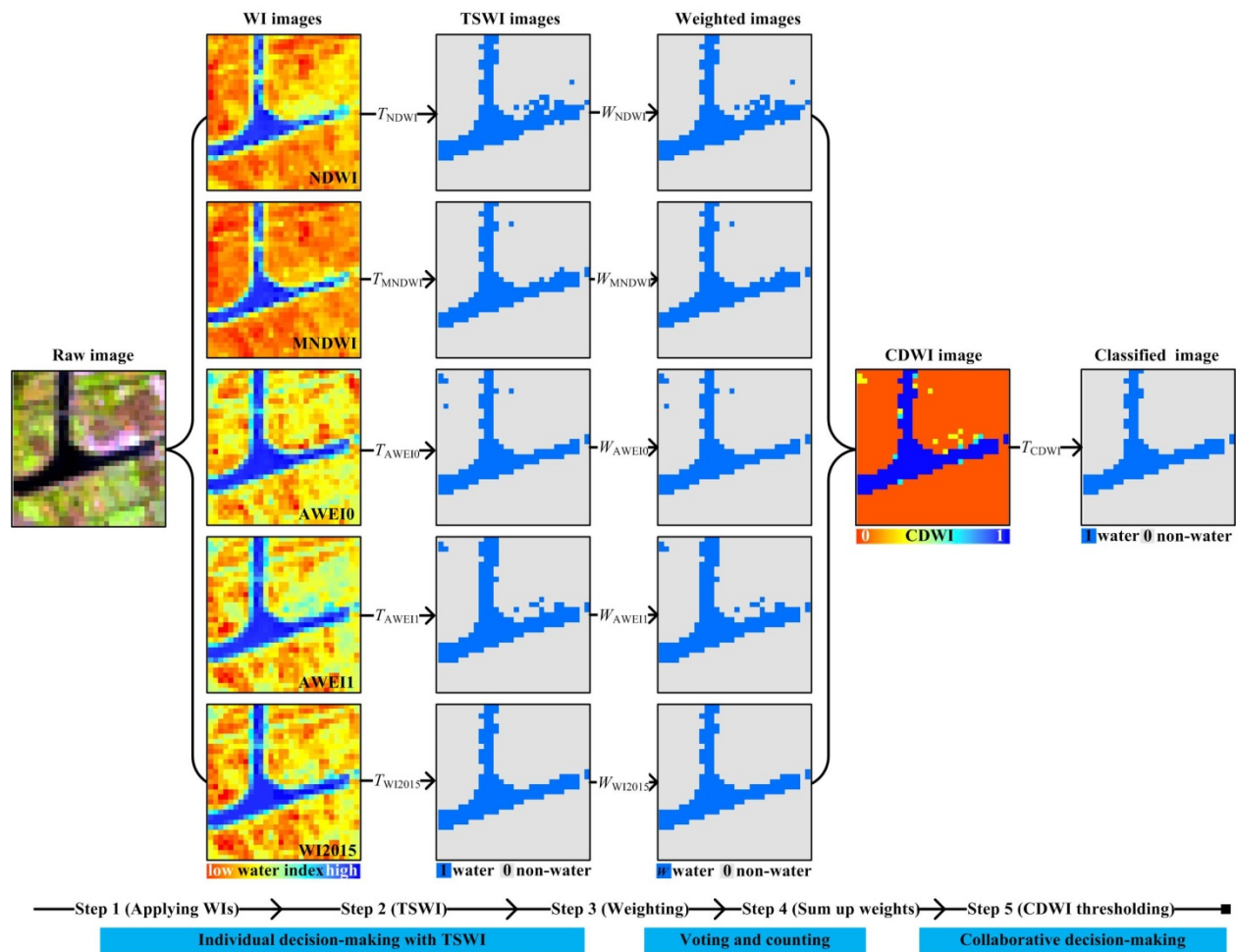
245 AWEI0 performs better than other WIs in differing built-up land from water, and AWEI1 is good at
 246 distinguishing shadow from water (Feyisa et al., 2014).

247 ● **Step 2:** Apply an appropriate pre-defined threshold to each WI image to initially classify water (labeled 1)
 248 and non-water (labeled 0). Note that this step is also known as applying TSWI for water classification.

249 ● **Step 3:** Apply an appropriate weight to each initially classified TSWI image. The sum of all weights is 1.
 250 TSWI method with better performance needs to be assigned a larger weight to its classified TSWI image.

251 ● **Step 4:** Sum up all weighted images to achieve a new CDWI image. Its pixel values are considered to
 252 represent water probability. The larger CDWI pixel value, the greater confidence of the pixel being
 253 decided as water.

254 ● **Step 5:** Apply a probability decision-making threshold (T_{CDWI}) to binarize the CDWI image and obtain
 255 the final water image.



256

257 **Fig. 2.** The workflow of CDWI exemplified with a Landsat-8 OLI subsite image. T and W stand for threshold
258 and weight, respectively.

259 From the perspective of the collaborative decision-making process, the workflow of CDWI can be
260 understood as following. Consider there is a decision-making committee named CDWI, and the job of which is
261 to decide whether image pixels are water or not. It has several experienced committee members (i.e.,
262 $TSWI_{NDWI}$, $TSWI_{MNDWI}$, $TSWI_{AWEI0}$, $TSWI_{AWEI1}$, and $TSWI_{WI2015}$ in this study) but with different abilities
263 (weights). In the processing of collaborative decision-making, each committee member would independently
264 make an initial decision (water or non-water) first with TSWI method. Then, each member assigns its weight
265 (W) to the corresponding TSWI image. The sum of all weighted TSWI images forms a new CDWI image
266 waiting for the final decision: pixels with values larger than T_{CDWI} are classified as water, otherwise non-water.

267 3.2.2 CDWI parameters estimation

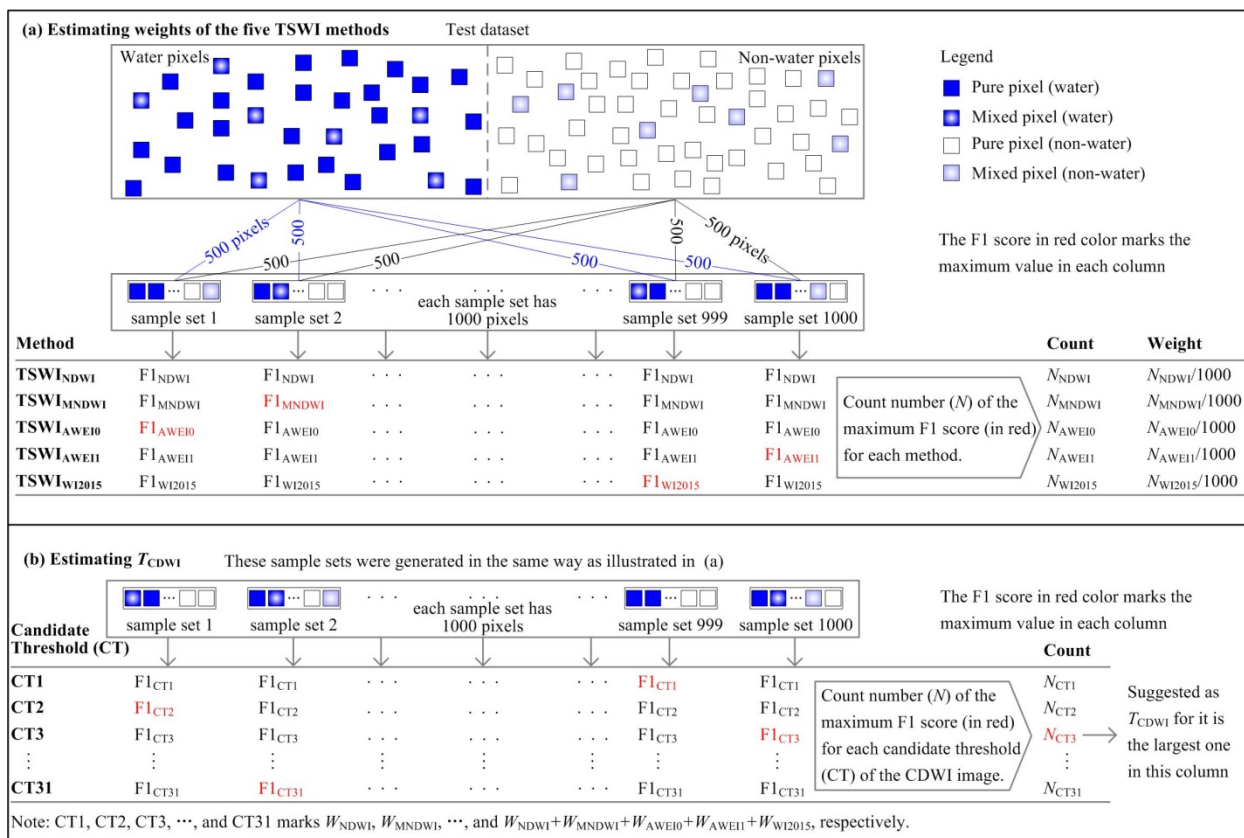
268 The application of CDWI requires three types of parameters: (1) the pre-defined WI thresholds (T_{NDWI} , T_{MNDWI} ,
269 T_{AWEI0} , T_{AWEI1} , and T_{WI2015}) for applying the five TSWI methods, (2) the weights (W_{NDWI} , W_{MNDWI} , W_{AWEI0} ,
270 W_{AWEI1} , and W_{WI2015}) of the five TSWI methods, and (3) the CDWI threshold (T_{CDWI}) for slicing the final
271 CDWI image (Fig. 2). Since the pre-defined thresholds have already been recommended by the previous
272 authors in applying the five TSWI (Table 3), they are directly adopted in this CDWI approach as well. The
273 other two parameters were estimated in the following ways (Fig. 3).

274 (1) Weights of the five TSWI methods

275 According to the principle of CDWI, a TSWI method showing better performance should hold a larger
276 weight. Assessing performances of the five TSWI methods and determining their weights were conducted
277 accordingly as below. First, we prepared 1,000 sample sets with each formed by 1,000 randomly selected
278 pixels from the test dataset: 500 are water and 500 are non-water. Note that the same size of water and
279 non-water pixels can minimize the uncertainty in validation caused by imbalanced sample size (Warmink, et al.,
280 2010). Then, the five TSWI methods with the recommended corresponding pre-defined thresholds (Table 3)
281 were applied to each sample set, and their accuracies were evaluated by F1-score, a harmonic accuracy
282 assessment metric as detailed in Section 3.3.1 (Daskalaki et al., 2006; Zhong et al., 2019). As each sample set

283 produced five F1-scores for the five TSWI methods, and the one holding the maximum F1-score was
 284 considered as performed the best and counted one. After this process went for the entire 1, 000 sample sets,
 285 each WI would get a final count number (N) and the sum of five count numbers equals to 1,000. Finally, the
 286 weight of a TSWI method was determined by the proportion of its count value to the sum of all count values.
 287 In this study, for example, the weight of $TSWI_{NDWI}$ (W_{NDWI} in Fig. 2) was calculated as Eq. (1):

$$288 \quad W_{NDWI} = \frac{N_{NDWI}}{N_{NDWI} + N_{MNDWI} + N_{AWEI0} + N_{AWEI1} + N_{WI2015}} = \frac{N_{NDWI}}{1000} \quad (1)$$



290 **Fig. 3.** The workflow of estimating (a) weights of the five TSWI methods and (b) CDWI threshold (T_{CDWI}).

291 (2) CDWI threshold (T_{CDWI})

292 Since CDWI image is sum of several weighted TSWI images (Fig. 2), any pixel value of such CDWI
 293 image is the sum of one combination weights of TSWI methods. In total, there are 31 different combinations of
 294 weights in the case of this study (Fig. 2): W_{NDWI} , W_{MNDWI} , W_{AWEI0} , W_{AWEI1} , W_{WI2015} , $W_{NDWI}+W_{MNDWI}$,
 295 $W_{NDWI}+W_{AWEI0}$, $W_{NDWI}+W_{AWEI1}$, ..., and $W_{NDWI}+W_{MNDWI}+W_{AWEI0}+W_{AWEI1}+W_{WI2015}$ or 1. Therefore, the final
 296 recommended T_{CDWI} should be determined from this list. The determination process is straightforward. First,

297 we generated 1,000 sample sets in the same way as mentioned above. Each sample set would produce 31
298 F1-scores after applying 31 candidate CDWI thresholds independently. Among these 31 F1-scores, the
299 maximum score and its corresponding threshold was identified and counted. After applying this procedure to
300 all 1,000 sample sets, the threshold which obtained the largest count number was identified as the
301 recommended T_{CDWI} , for it held the most cases of holding the maximum F1-scores than the other candidate
302 thresholds.

303 3. 3. Performance assessment

304 3.3.1 Accuracy assessment

305 As mentioned in the Section 2.1, there are 145 out of 180 subsite images cover both water and land
306 features (the other 35 out 180 subsite images only cover land features). Therefore, the five TSWI methods and
307 the CDWI method were applied to these 145 subsite images to assess their accuracy stabilities over different
308 water-land environments around the world. As previous studies suggested, both F1-score and Youden's index
309 (YI) were used to assess accuracies of the six methods (Li et al., 2016; Li et al., 2019; Zhong et al., 2019; Wen
310 et al. 2016). The F1-score is the harmonic average of the producer's accuracy and user's accuracy (Daskalaki
311 et al., 2006; Eq. (2)):

$$312 \quad F1\text{-score} = \frac{2 \times \text{Producer's accuracy} \times \text{User's accuracy}}{\text{Producer's accuracy} + \text{User's accuracy}} \quad (2)$$

313 The producer's accuracy is the percentage of correctly classified water pixels from the total number of true
314 water pixels. The user's accuracy is the percentage of correctly classified water pixels from the total number of
315 classified water pixels. F1-score reaches its best value at 1 and worst at 0. It is considered more objective than
316 overall accuracy (the percentage of correctly classified pixels, both as water and non-water, from the total
317 number of pixels) in our binary classification case because a water body mostly covers a small portion of the
318 image under evaluation. The YI was often used for determining local optimal thresholds (Wen et al. 2016), and
319 it was considered as an indicator of water classification accuracy (Eq. (3)). The larger YI value, the smaller
320 sum of omission error and commission error.

$$321 \quad YI = 1 - (\text{Omission error} + \text{Commission error}) \quad (3)$$

322 3.3.2 Sensitivity to thresholds

323 Sensitivity to thresholds, defined as how much the accuracy would change by changing the threshold
324 values for a given method (TSWI methods or CDWI method), is indicated by the slope of a threshold-accuracy
325 curve. A robust classification method should, therefore, be less sensitive (low absolute slope value) to
326 threshold changes. For TSWI methods, such thresholds are the pre-defined WI thresholds; for the CDWI
327 method, such thresholds involve both the pre-defined WI thresholds and T_{CDWI} .

328 For a given TSWI method, classification outcome purely affected by the pre-defined thresholds (Fig. 2).
329 Each pre-defined threshold outputs a classification result and one accuracy (F1-score or YI value). The
330 sensitivity analysis, thus, involves selecting different pre-defined thresholds and calculating their
331 corresponding accuracies. To make such selection more objective, the local optimal thresholds of 145 subsite
332 images were served as candidate pre-defined thresholds. For a subsite image, its local optimal threshold was
333 determined as the threshold at which the YI gained the maximum value (Fisher et al., 2016).

334 For the proposed CDWI method, its accuracy relies on both the five pre-defined WI thresholds (Table 3)
335 and T_{CDWI} (Figs. 2 and 3). To make the sensitivity analysis more clearly, T_{CDWI} was fixed (to the suggested one)
336 in analyzing the sensitivity of CDWI to WI thresholds; while WI thresholds were fixed (to the suggested ones,
337 see Table 3) in analyzing the sensitivity of CDWI method to T_{CDWI} . Each group of the five selected WI
338 thresholds will produce one F1-score of the CDWI. As a WI threshold could be chosen from the 145 candidate
339 local optimal thresholds, 145^5 (=64,097,340,625) different threshold groups could be generated with 145^5
340 accuracies. To reduce this huge computational burden, the 145 candidate local optimal thresholds were split
341 into 15 equal interval groups and the central value of each group was reselected. Finally, there are 15^5
342 (=759375) WI threshold groups and 15^5 corresponding CDWI accuracies are obtained. Each selected WI
343 threshold would generate one accuracy for the corresponding TSWI method but 15^4 (=50625) accuracies for
344 the CDWI method. To make them comparable, the mean accuracies of the CDWI method was used for
345 sensitivity analysis.

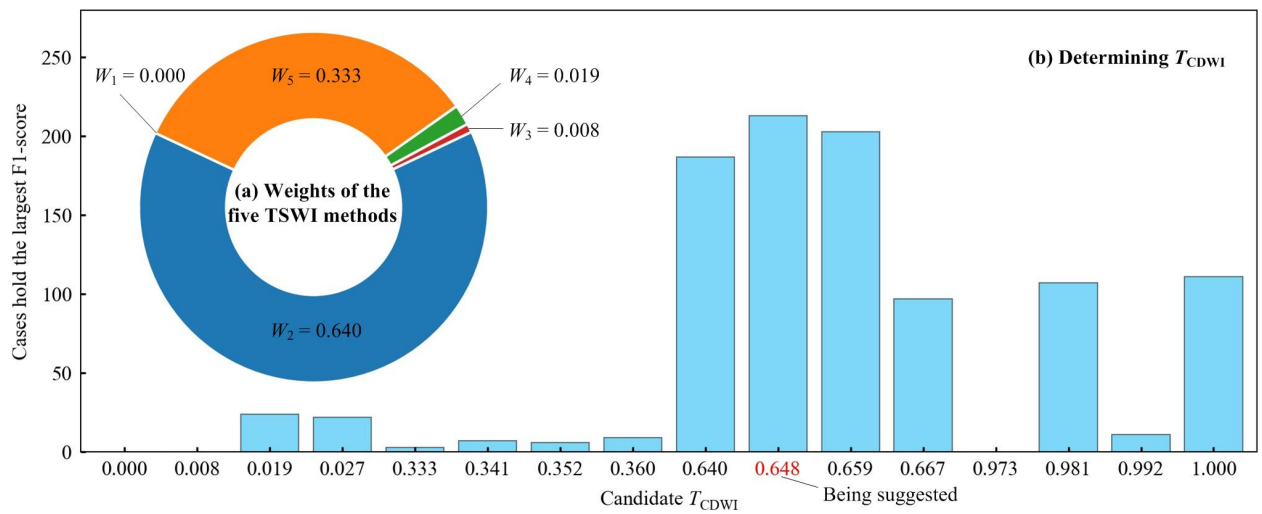
346 4. Results

347 4.1. Suggested CDWI parameters

348 The parameters of applying the CDWI method were estimated carefully (Fig. 3) and are could be directly

349 used in further applications given that they are evaluated by the dataset collected from various different
 350 water-land environment around the world. To estimate the weights of the five TSWI methods, their accuracies
 351 were assessed. Overall, $TSWI_{MNDWI}$ showed the best performance for classifying water and then followed by
 352 $TSWI_{WI2015}$, $TSWI_{AWEI1}$, $TSWI_{AWEI0}$, and $TSWI_{NDWI}$. Accordingly, the suggested five weights the TSWI
 353 methods were estimated as 0.640, 0.333, 0.019, 0.008, and 0.000, respectively (Fig. 4a). Note that $TSWI_{NDWI}$
 354 performed the worst among the five TSWI method and got zero weight, for it held zero cases among 1,000
 355 sample sets that gained the highest F1-scores.

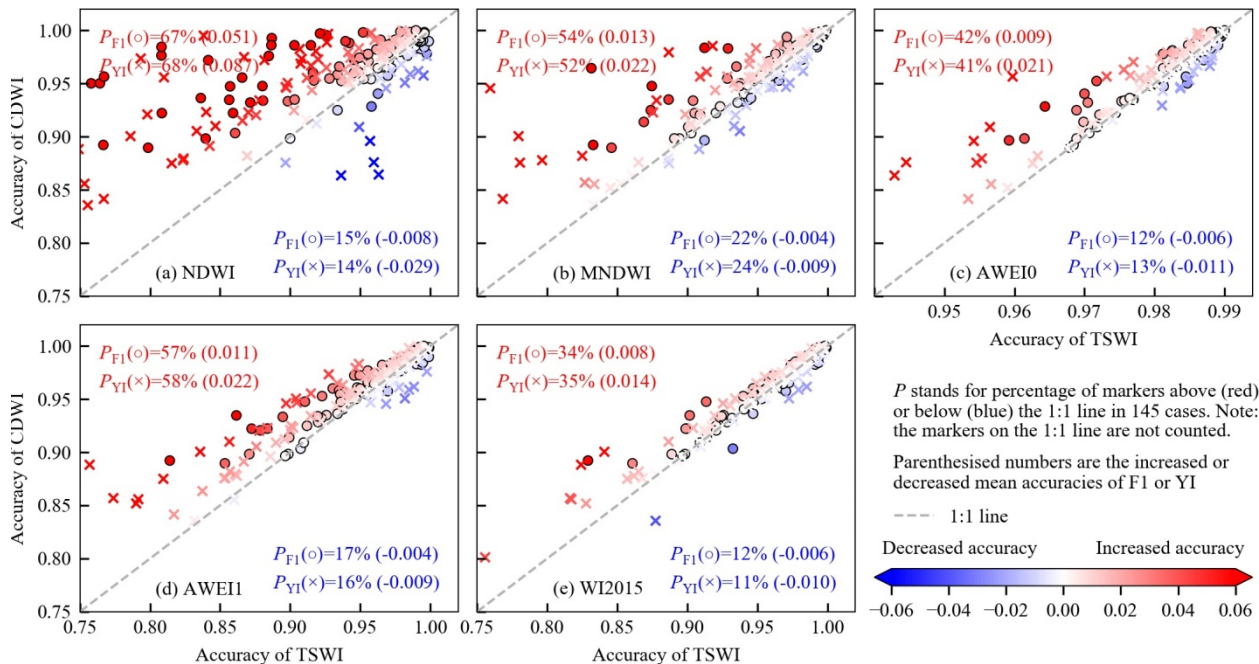
356 With regard to T_{CDWI} , it suggests 0.648 as the best for further applications for it obtained the largest
 357 number of cases that got the maximum F1-score among all the candidate CDWI thresholds (Fig. 4b). The
 358 result means that pixel values larger than 0.648 in the CDWI image (sum of weighted TSWI images) are more
 359 likely to be labeled as water than non-water. Furthermore, this T_{CDWI} is the sum weights of MNDWI ($W_2 =$
 360 0.640) and AWEI0 ($W_3 = 0.008$), which statistically implies that pixels were classified as water by both
 361 $TSWI_{MNDWI}$ and $TSWI_{AWEI0}$ are more likely to be correctly classified than that only classified either by
 362 $TSWI_{MNDWI}$ or $TSWI_{AWEI0}$.



363
 364 **Fig. 4.** Suggested parameters of CDWI: (a) wights of the five TSWI methods (W_1 , W_2 , W_3 , W_4 , and W_5 , see
 365 also in Fig. 2), and (b) T_{CDWI} . The red-colored threshold (0.648) in (b) marks the suggested T_{CDWI} for it holds
 366 the most cases that obtained the maximum F1-score among all the candidate CDWI thresholds.

367 4.2. Accuracy assessment over different environments

368 The accuracies of the six methods were applied to 145 individual subsite images to compare their
 369 accuracies over different water-land environments (Fig. 5). All the TSWI methods and CDWI method obtained
 370 high accuracies for their F1-scores and YI values greater than 0.9 for most subsites (Fig. 5). Although they all
 371 performed relatively well, the differences in their performances can be observed. In general, the number of
 372 subsites with their accuracies improved by the CDWI method was much greater than the number of subsites
 373 with their accuracies that decreased by the CDWI method. For example, 54% subsite images classified by the
 374 CDWI method produced higher F1-scores than that produced by the $TSWI_{MNDWI}$ method, and only 22%
 375 subsite images got lower F1-scores by using CDWI than $TSWI_{MNDWI}$ method (Fig. 5b). Moreover, the absolute
 376 mean value of decreased accuracies was smaller than that of the increased accuracies. Take YI as an example,
 377 such a pattern can be observed as: $|-0.029|$ vs. 0.087 in Fig. 5a, $|-0.009|$ vs. 0.022 in Fig.5b, $|-0.011|$ vs. 0.021 in
 378 Fig. 5c, etc. This finding shows that the CDWI method could be more likely to obtain a better water
 379 classification result than any TSWI method in general.

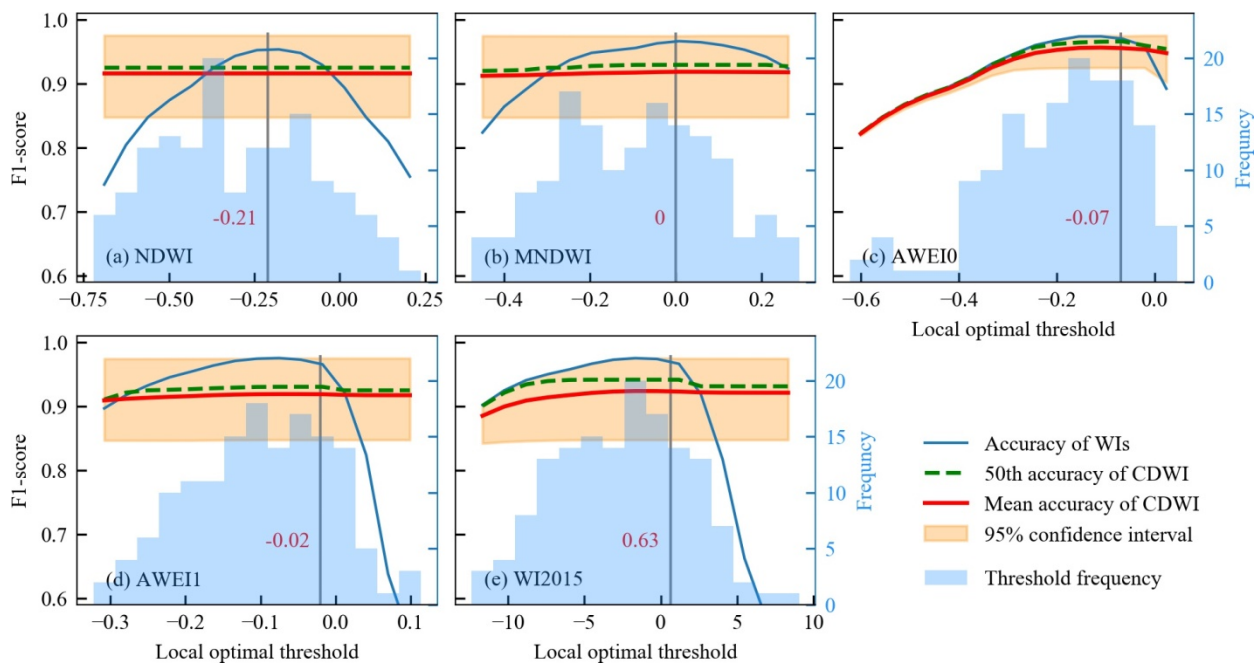


380
 381 **Fig. 5.** Accuracy (indicated by F1-score and YI value) comparisons between CDWI and the TSWI method
 382 using five WIs: (a) NDWI, (b) MNDWI, (c) AWEI0, (d) AWEI1, and (e) WI2015. Decreased accuracies (blue
 383 dots) and increased accuracies (red dots) represent the accuracy difference between the CDWI and TSWI
 384 method.

385 4.3. Sensitivity to threshold

386 4.3.1 Sensitivity to pre-defined WI thresholds

387 Each subsite image can obtain a local optimal threshold. For all subsite images, their local optimal
 388 threshold varied significantly, as shown in Fig. 6. Generally, the histograms of those local optimal thresholds
 389 approximately follow Gaussian distributions. The F1-score of any TSWI method changed dramatically with
 390 different pre-defined WI thresholds were used (the blue lines in Fig. 6). Overall, sensitivity curves of all the
 391 TSWI methods are in unimodal patterns and peak at their thresholds around the suggested pre-defined
 392 thresholds that we used in this study (see Table 3). These sensitivity curves can be broadly categorized into
 393 three types: high sensitivity with a steep slope, moderate sensitivity with a moderate slope, and low sensitivity
 394 with roughly flat slope. The further distance of a threshold to the suggested pre-defined threshold, the higher
 395 sensitivity of a TSWI method to such threshold can be observed (Fig. 6).



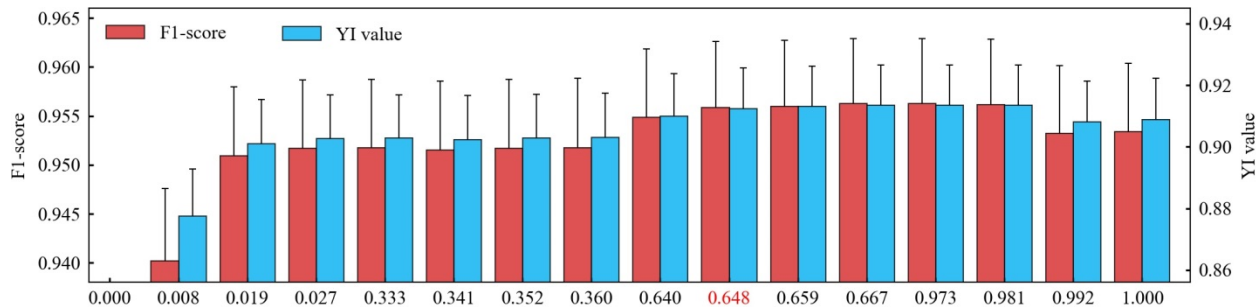
396
 397 **Fig. 6.** The sensitivity of F1-score to threshold for the five TSWI methods and the CDWI method. The
 398 sensitivities are indicated by the slope of the sensitivity curve: the threshold-against-F1 curve. The red colored
 399 values are the pre-defined WI thresholds suggested by the literatures as listed in Table 3.

400 In contrast, the proposed CDWI method showed the least sensitive to threshold. That is, no matter what
 401 threshold were used, the accuracies of the CDWI method changed slighter than those of any TSWI method.
 402 For example, when the threshold changed from -0.45 to 0.26, the F1-score of $TSWI_{MNDWI}$ changed from 0.82

403 to 0.97, whereas the mean F1-score of CDWI method changed from 0.912 to 0.918 (Fig. 6b). This low
 404 sensitivity-to-threshold of the CDWI method indicate that the uncertainties related to threshold determination
 405 can be significantly reduced compared to the TSWI methods. Such characteristics of CDWI method could
 406 make users less worrying about whether the selected thresholds are the optimal ones or not in applications
 407 without reference data.

408 *4.3.2 Sensitivity to T_{CDWI}*

409 Overall, the sensitivity of CDWI accuracy to T_{CDWI} is relatively low (Fig. 7). The mean F1-score of the
 410 CDWI method changes from 0.940 to 0.956 as the T_{CDWI} changing from 0.008 to 1.000. It generally shows a
 411 “∩” pattern with short increasing, long-flatten, and a slightly decreasing trend in order. In terms of YI value, it
 412 also shows a similar sensitivity-to- T_{CDWI} pattern as of F1-score. It is noteworthy that the accuracy produced by
 413 combined T_{CDWI} (i.e., summed by two or more TSWI weights) is overall larger than that produced by single
 414 T_{CDWI} (i.e., single TSWI weight), which is explained here. All the T_{CDWI} values are denoted by the x-axis
 415 ticklabels in Fig. 7, the single T_{CDWI} values are 0.000 (W_1), 0.008 (W_2), 0.019 (W_3), 0.333 (W_4), and 0.640 (W_5);
 416 the rest are combined T_{CDWI} values. It is observed that the mean F1-score produced by the 0.027 (W_2+W_3) is
 417 larger (0.952) than the mean F1-score produced either by 0.940 (W_2) or 0.951 (W_3). This observation goes for
 418 our suggested T_{CDWI} (0.648, W_2+W_5) in the study (Fig. 4): its mean F1-score and YI value are both larger than
 419 that produced by corresponding single T_{CDWI} : 0.008 (W_2) and 0.640 (W_5).



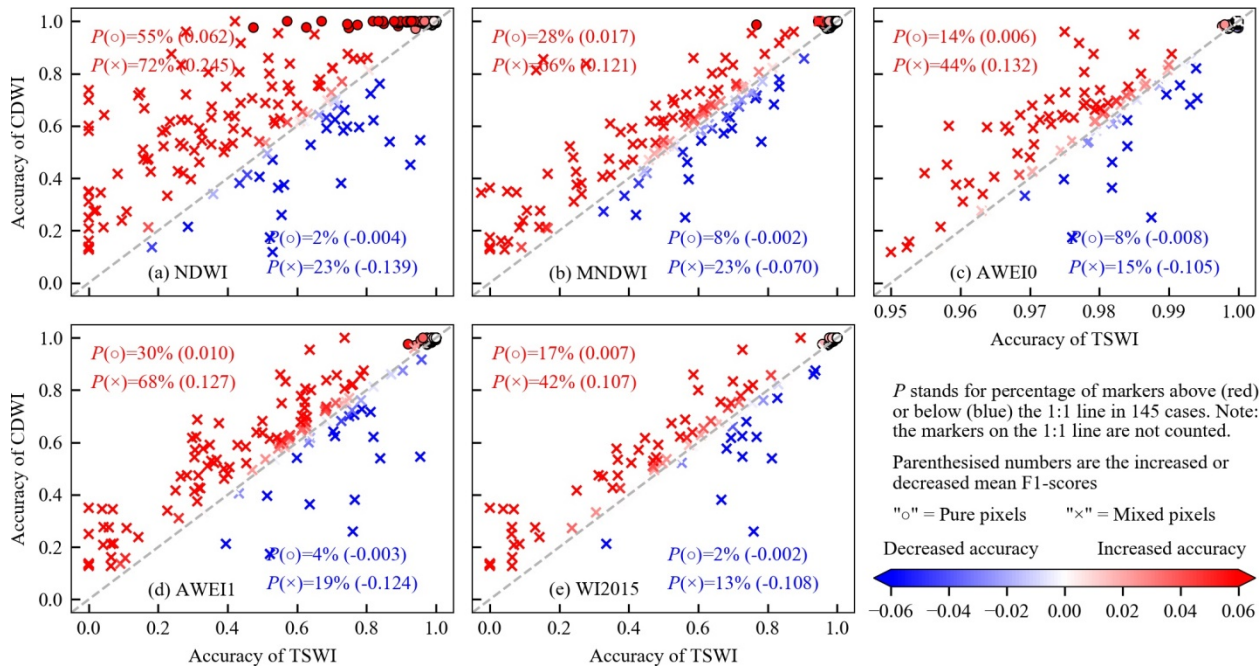
420
 421 **Fig. 7.** The sensitivity of CDWI accuracy (F1-score and YI value) to the T_{CDWI} . The red-colored threshold
 422 (0.648) marks the suggested T_{CDWI} in this study (see Fig. 4).

423 **5. Discussion**

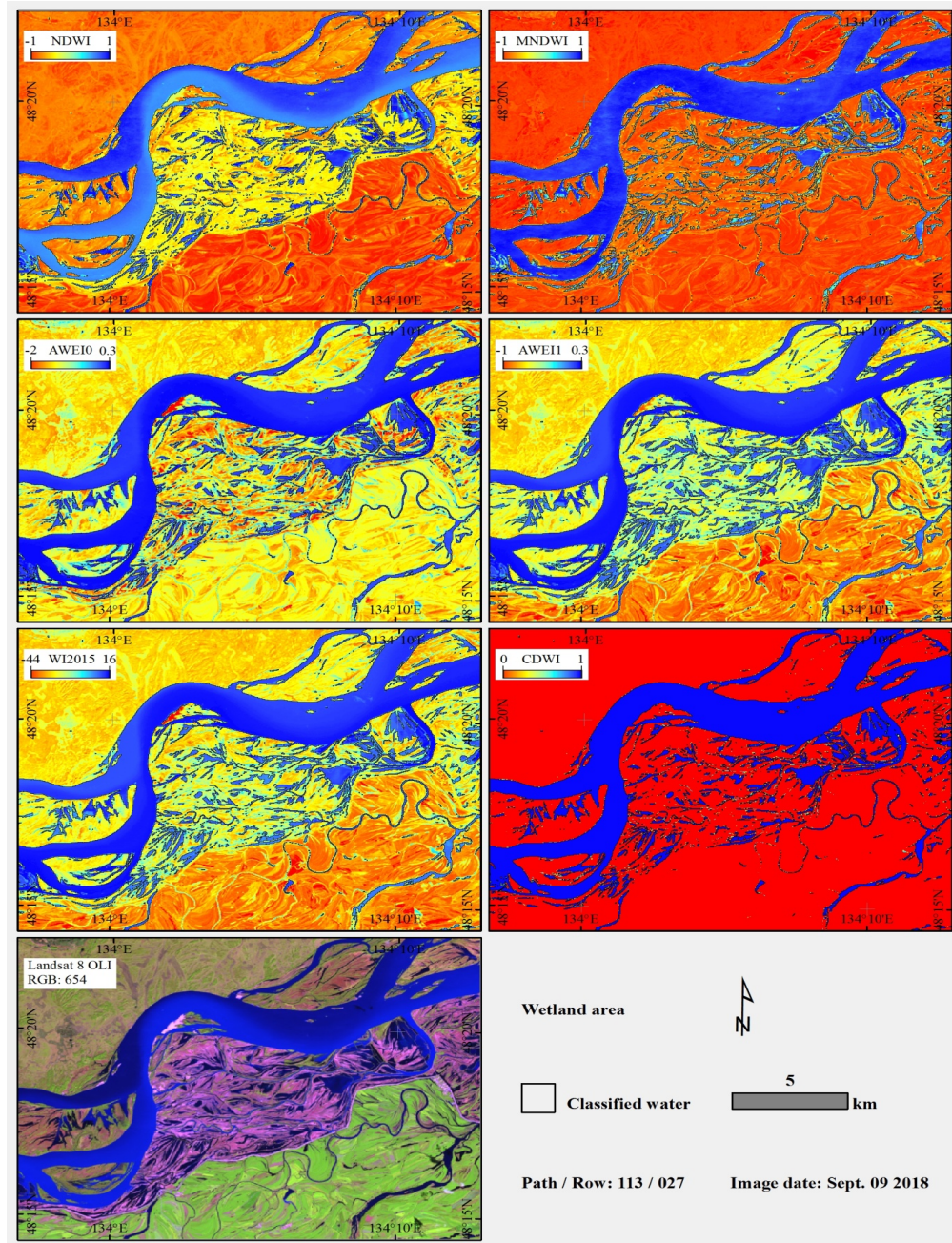
424 5.1 Uncertainty analysis

425 5.1.1 Pure pixels vs. mixed pixels

426 One commonly recognized uncertainty of a water classification method may come from water-land mixed
 427 pixels or water-land boundary pixels (Comber et al., 2012; Yang et al., 2015). To better understand how the
 428 CDWI works, we compared the performances of the six methods in classifying both pure pixels and mixed
 429 pixels of the 145 subsite images (Fig. 8). It is observed that all the TSWI methods and CDWI method
 430 performed worse for mixed pixels than for pure pixels. Because the TSWI methods were developed based on
 431 the principle that water and land features have distinct reflectance properties: water shows a decrease in
 432 reflectance from the visible to infrared wavelengths, while land features (e.g., vegetation) often do not show
 433 such reflectance pattern(Xiong et al., 2018). Moreover, those WI methods are “hard” classification methods
 434 using a Boolean set (i.e., 0 or 1) to restrict each pixel to either water or non-water types (Yang, et al., 2015).
 435 Therefore, classifying mixed pixels often introduce more errors to the result than classifying pure pixels with
 436 TSWI methods, due to the averaging of the reflectance properties of the water and non-water components
 437 (Fisher et al., 2016). How to reduce the class uncertainty of mixed pixels in classifying water is accordingly a
 438 research topic for many researchers.



439 **Fig. 8.** Accuracy (indicated by F1-score) comparison between CDWI and the five TSWI methods for both pure
 440 pixels (○) and mixed pixels (×) of subsite images: (a) NDWI, (b) MNDWI, (c) AWEI0, (d) AWEI1, and (e)
 441 WI2015. Decreased accuracies (blue dots) and increased accuracies (red dots) represent the accuracy
 442 difference between the CDWI and an individual TSWI method.



444

445 **Fig. 9.** An example application of using the five TSWI methods and CDWI method in a heterogeneous wetland
 446 environment (More examples are illustrated in Supplementary Figs. S2-S6).

447 Various techniques have been developed in attempts to reduce the uncertainty of mixed pixels in water
 448 classification. Some are based on the idea of “soft” classification such as sub-pixel classification and fuzzy
 449 classification (Dewi et al., 2016; Xiong et al., 2018). Some use machine learning techniques by taking mixed
 450 pixels into the training process (Foody and Mathur 2006). In this study, however, the CDWI achieved higher

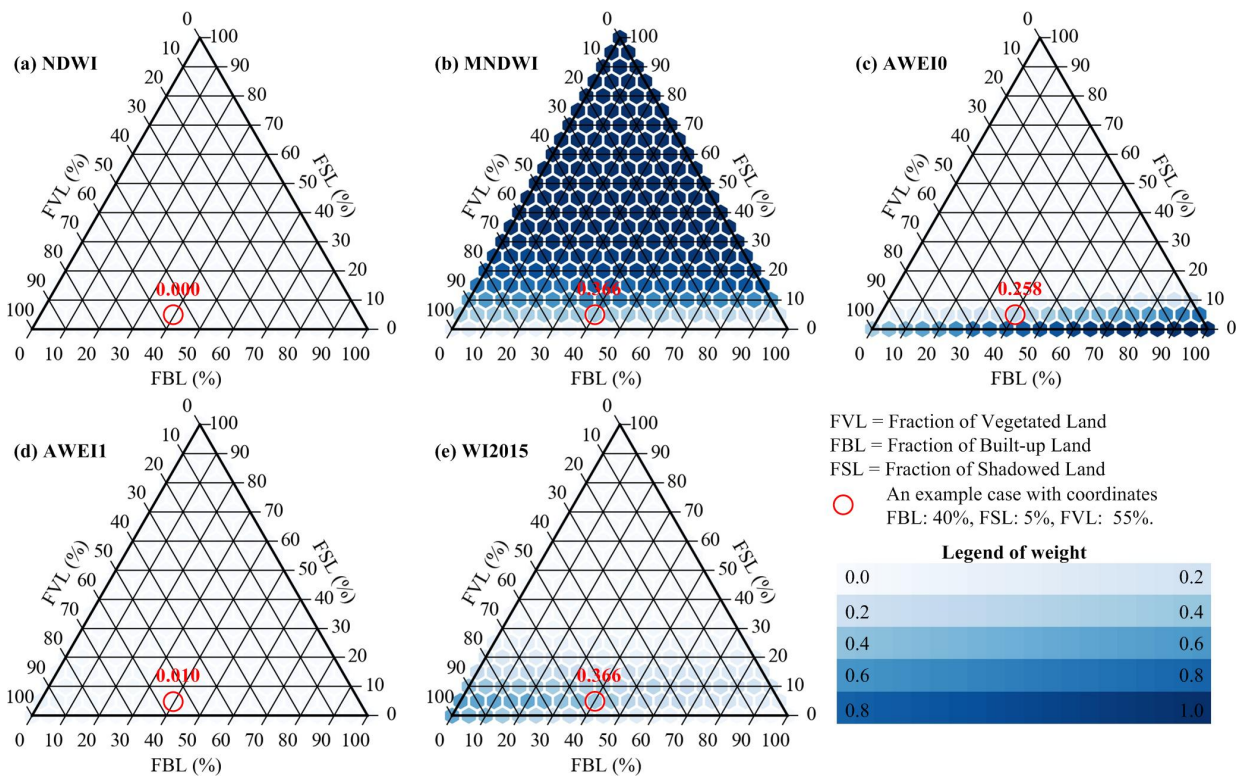
451 performance than the other TSWI methods for classifying water from mixed pixels (Fig. 8). It looks like this
452 study provides an alternative way of reducing the uncertainty of mixed pixels. For a mixed pixel labeled as
453 water (i.e., water percentage larger than 50%), the processing of CDWI could be considered as accumulating
454 the probability of a water pixel that being correctly classified. That is, the decision of a mixed pixel be water or
455 non-water is not only based on a single result of an individual TSWI method (except it has large weight) but
456 collectively decided by the results of several TSWI methods. Based on these understandings, it is highly
457 recommended to apply CDWI to the cases where mixed pixels are very common, such as small water bodies
458 (e.g., pond), or water bodies with large perimeters-area ratios (e.g., dike, creek, tide channel, and mountainous
459 reservoir) as shown in Fig. 9.

460 *5.1.2 Different compositions of land features*

461 We observed in some subsites that CDWI performed worse than TSWI methods as illustrated in Fig. 5
462 (the blue dots below the 1:1 line). One reason could be the parameters of CDWI were estimated from a
463 simulated general scenario, not from the specific scenario of each subsite. Such a general scenario was
464 simulated by 1, 000 sample sets, with each of them formed by 1,000 randomly selected water and non-water
465 pixels from the test dataset. Since the dataset collected from various water-land environments around the world
466 (Fig. 1 and Table 1), a general scenario could consist of water with different colors, and land features with
467 most covered by vegetation and some parts covered by built-up land and shadows. However, for some specific
468 scenarios, the proportion of land components may differ a lot from the general scenario. For example, an urban
469 is mostly occupied by built-up land and building shadows and a small portion of vegetated land. In such a case,
470 the suggested parameters of CDWI in Fig. 4 could not perform well than the ones carefully designed for an
471 urban area, like AWEI0 and MNDWI (Feyisa et al., 2014).

472 To explore more application scenarios, we first simulated a variety of land environments that consisted of
473 different fractions of three typical land features, namely vegetated land, built-up land, and shadowed land. For
474 each simulated land environment, the corresponding five WI weights are estimated in the same way as that for
475 the general scenario (Fig. 10). Overall, the performances (indicated by TSWI weights) of both $TSWI_{INDWI}$ and
476 $TSWI_{AWEI1}$ are not sensitive to any kind of land environment and gained the lowest weights (0 or near to 0).
477 When the fraction of shadowed land larger than 10%, $TSWI_{MNDWI}$ gained the largest weights than the other

478 TSWI methods. It implies that in the scene with a large portion of shadows, image classified by $TSWI_{MNDWI}$
 479 method should be assigned dominate weight than that by the other TSWI methods in applying CDWI method;
 480 or if one just wants to use TSWI method, the $TSWI_{MNDWI}$ should also be suggested for guiding WI selection in
 481 applying TSWI method. It also shows that the $AWEI0$ is sensitive to the fraction of built-up land: the more
 482 built-up land in an application, the higher weight of the $TSWI_{AWEI0}$ gains (Feyisa et al., 2014; Fisher et al.,
 483 2016). In an extreme scenario such as in urban areas, it is suggested to assign the largest weight to $TSWI_{AWEI0}$
 484 than other TSWI methods in applying the CDWI method. We recommend that the above findings and Fig. 10
 485 could be served as a general guidance or a look-up-table for selecting WI in water classification applications
 486 using either TSWI or CDWI methods.



493 (Feyisa et al., 2014; McFeeters 1996; Xu 2006), the CDWI could also be considered as a new framework that
494 could readily be used in many applications involving different sensors. First, both the number and the form of
495 TSWI methods involved in the CDWI are not fixed and can be adjusted according to practical conditions. For
496 example, the existing water indices that are not used in this study, such as TCW (Crist 1985), WRI (Rokni et
497 al., 2014), TSUWI (Wu et al., 2018) and MBWI (Wang et al., 2018), could be integrated readily into the CDWI
498 method in further applications. Likewise, as newly designed water indices become available, they can be
499 brought into the framework of CDWI. Moreover, any water classification maps either obtained by TSWI
500 methods or by more sophisticated methods (e.g., Random Forest and Support Vector Machine; see Acharya et
501 al., 2016; Ireland et al., 2015) can be included in the CDWI method to determine the final water classification
502 results. Second, although the proposed CDWI method is tested and demonstrated on Landsat-8 OLI images, it
503 is also suggested for application to Landsat TM/ETM+ images because the TSWIs used here were all
504 originally designed for Landsat TM/ETM+ images (Huang et al., 2018). Since these TSWI methods work well
505 on the Landsat-8 OLI images in this study, they should be suitable for Landsat TM/ETM+ images as well.
506 Third, the framework of the CDWI method can be applied to other types of images with different bands than
507 the Landsat images, such as MODIS (Sharma et al., 2015), Sentinel-2A/B (Du et al., 2016), and HJ-1A/B
508 images (Lu et al., 2011). Because the image bands of these images are very different from those of the Landsat
509 images, their sensor-dependent water indices should be carefully selected before using the CDWI method.

510 In summary, the proposed CDWI method has four critical potential advantages:

- 511 (1) The operation procedure of CDWI is straightforward, applied with basic raster algebra. Users can expand
512 any TSWI methods into the CDWI framework.
- 513 (2) The robustness of the CDWI is higher than that of the TSWI methods making it suitable for a wide range
514 of applications over different water-land environments.
- 515 (3) The accuracy of the CDWI is less sensitive to the threshold (both pre-defined WI thresholds and T_{CDWI})
516 selection compared to the TSWI methods, such that the need for tedious parameter tuning of the threshold
517 is reduced or avoided.
- 518 (4) The framework underlying the CDWI is not WI dependent and sensor dependent. It has the potential to be
519 applied to other indices (e.g., impervious surface index) and other sensors (e.g., Landsat TM/ETM+,

520 MODIS, and Sentinel-2).

521 **6. Conclusions**

522 The TSWI methods are widely adopted in water mapping applications due to their potential ease-of-use
523 and generally acceptable performances. However, two concerns need to be carefully considered before
524 applying them in practice: the selection of WI and the determination of an appropriate threshold for the given
525 WI. In practice, answers to these two concerns could be affected by several subjective factors, such as
526 experiments and personal preference. To overcome these two concerns, a new ensemble way of using WIs for
527 water mapping approach that integrates five widely used WIs is proposed, namely the CDWI, based on the
528 collaborative decision-making principle.

529 A total of 145 subsite images were selected representing different geographical areas with distinct
530 water-land environments and different seasonal patterns. The performances of the CDWI method and the five
531 TSWI methods were assessed in terms of accuracy and robustness. It was found that (1) the CDWI produced
532 higher or comparable accuracies than the five benchmark TSWI methods for most cases, making it less
533 sensitive to application scenarios and, thus, suitable for more different water-land environments. (2) The
534 accuracy of the CDWI is much less sensitive to the pre-defined WI thresholds chosen for the TSWI methods;
535 (3) The underlying framework of CDWI has great potential for transferability and further application. For
536 example, it can be modified readily by adding new WIs in the future. Moreover, the principle underlying the
537 CDWI method is not sensor-dependent and, thus, the proposed CDWI can be applied to different types of
538 images, such as Landsat TM/ETM+, MODIS, Sentinel-2A/B and HJ-1A/B images in future applications.

539 **Acknowledgments**

540 This work is supported by the Open Fund of the State Laboratory of Information Engineering in
541 Surveying, Mapping and Remote Sensing, Wuhan University (No: 18R07), the Open Fund of the State Key
542 Laboratory of Lake Science and Environment, Chinese Academy of Sciences (No: 2018SKL006), and by the
543 National Natural Science Foundation of China (No: 41501096, 51779241, and 41401051). The authors thank
544 the anonymous reviewers for providing constructive comments and feedbacks on this paper.

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