

1 PERSPECTIVE

2 **Coral reef ecology in the Anthropocene**

3

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16

17 **Summary**

18 1. We are in the Anthropocene – an epoch where humans are the dominant force of planetary

19 change. Ecosystems increasingly reflect rapid human-induced, socioeconomic and cultural

20 selection rather than being a product of their surrounding natural biophysical setting. This

21 poses the intriguing question: to what extent do existing ecological paradigms capture and

22 explain the current ecological patterns and processes we observe?

23

24 2. We argue that, although biophysical drivers still influence ecosystem structure and
25 function at particular scales, their ability to offer predictive capacity over coupled social-
26 ecological systems is increasingly compromised as we move further into the Anthropocene.

27

28 3. Traditionally, the dynamics of coral reefs have been studied in response to their proximate
29 drivers of change rather than their underlying socioeconomic and cultural drivers. We
30 hypothesise this is limiting our ability to accurately predict spatial and temporal changes in
31 coral reef ecosystem structure and function.

32

33 4. We propose '*social-ecological macroecology*' as a novel approach to a) identify the
34 interactive effects of biophysical and socioeconomic and cultural drivers of coral reef
35 ecosystems across spatial and temporal scales, b) test the robustness of existing coral reef
36 paradigms, c) explore whether existing paradigms can be adapted to capture the dynamics of
37 contemporary coral reefs, and d) if they cannot, develop novel coral reef social-ecological
38 paradigms, where human dynamics are part of the paradigms rather than the drivers of them.

39

40 5. Human socioeconomic and cultural processes must become embedded in coral reef
41 ecological theory and practice as much as biophysical processes are today if we are to predict
42 and manage these systems successfully in this era of rapid change. This necessary shift in our
43 approach to coral reef science will be challenging and will require truly interdisciplinary
44 collaborations between the natural and social sciences.

45

46 **Key-words:** Anthropocene, coral reef, ecology, macroecology, prediction, scale, social-
47 ecological macroecology, social-ecological systems

48

49 **Introduction**

50 Natural biophysical gradients such as wave energy, primary production, and seawater
51 temperature drive coral reef ecosystem structure and function across multiple scales and
52 trophic levels, from microbes (Kelly *et al.* 2014) and plankton (Gove *et al.* 2016), to corals
53 (Gove *et al.* 2015) and fish assemblages (Heenan *et al.* 2016). However, human impacts such
54 as fishing (Edwards *et al.* 2014), nearshore nutrient enrichment (D'Angelo & Wiedenmann
55 2014), sedimentation (Wolanski, Martinez & Richmond 2009), and the warming and
56 acidifying of our oceans (Albright *et al.* 2016; Hughes *et al.* 2018a) are pushing the
57 environmental boundary conditions defined by natural biophysical drivers on many coral
58 reefs globally. Furthermore, the distal socioeconomic and cultural drivers underlying these
59 proximate impacts, such as trade, consumer demands, human migration, and carbon dioxide
60 emissions, are all predicted to increase (Norström *et al.* 2016; Hughes *et al.* 2017). This
61 presents a new reality where the majority of coral reefs will increasingly reflect human-
62 induced, socioeconomic and cultural drivers rather than being a product of their long-term
63 natural biophysical setting. How we study and describe reef ecology must include this
64 paradigm shift in thinking if we are to predict and manage their dynamics effectively.

65 We propose an approach that will identify how key biophysical, socioeconomic and
66 cultural drivers of reefs interact across scales to drive coral reef ecosystem patterns and
67 processes. In doing so, this approach will arm us with the predictive capacity required to
68 manage coral reef dynamics in this era of rapid change. We start by reviewing how natural
69 biophysical drivers influence reefs, from dictating the dominance, behaviour, and trophic
70 ecology of individual reef organisms, through to governing the spatial ecology of reef
71 communities across the seascape. We then highlight how human socioeconomic and cultural
72 drivers have become an important structuring force of contemporary coral reefs at particular
73 scales. In doing so, we underline our lack of understanding regarding the degree biophysical,

74 or human socioeconomic and cultural drivers dominate depending on the scale of
75 observation, suggesting macroecological approaches as a potential solution. Finally, we
76 question whether traditional coral reef paradigms capture this new interwoven reality and
77 stress the need for a ‘*socio-ecological macroecology*’ approach to develop paradigms for
78 coral reefs in the Anthropocene.

79

80 **Biophysical drivers: setting natural bounds on coral reef ecosystems**

81 By studying coral reefs in remote locations with limited direct human influence, we
82 have learnt how coral reefs respond to, and are shaped by gradients in biophysical drivers
83 such as wave energy, primary production, and seawater temperature (**Fig. 1**). High wave
84 energy environments, for instance, can promote the dominance of low-lying benthic
85 organisms such as turf algae, crustose coralline algae, and encrusting corals that are less
86 vulnerable to physical dislodgement (Geister 1977; Gove *et al.* 2015). In contrast, lower
87 wave energy environments tend to favour more structurally complex benthic communities,
88 dominated by three-dimensional calcifying corals and upright macroalgae (Williams *et al.*
89 2013; Aston *et al.* 2018). Such increases in substrate complexity are often positively related
90 to reef fish density and biomass (Graham & Nash 2012) due to increased refuge from
91 predation (Rogers, Blanchard & Mumby 2014), and as such waves can indirectly mediate
92 predator-prey dynamics on reefs. Across the Pacific Ocean, for example, the biomass of
93 grazing herbivorous fishes peaks at islands with moderate wave exposure where the largest
94 edible algal mass for these fishes tends to occur (Heenan *et al.* 2016). Wave energy can also
95 influence reef fish community structure through interactions with fin morphology and
96 swimming performance (Fulton, Bellwood & Wainwright 2005), with high wave energy
97 environments capable of impacting the feeding success of some fishes and thus key
98 ecosystem functions like herbivory (Bejarano *et al.* 2017).

99 Natural gradients in nutrient concentrations and primary production have predictable
100 effects on coral reef ecosystems (Gove *et al.* 2016). For example, tropical islands located in
101 more productive regions of the Pacific Ocean support a greater number of microbes with
102 nutrient-related metabolisms (e.g., nitrate and nitrite ammonification) (Kelly *et al.* 2014), an
103 increased cover of calcifying benthic organisms (Williams *et al.* 2015a), and a greater
104 biomass of grazing herbivorous, planktivorous, and top-predatory fishes (Nadon *et al.* 2012;
105 Williams *et al.* 2015b; Heenan *et al.* 2016). Gradients in nutrients and primary production
106 also exist at smaller scales around individual islands. For example, when deep subsurface
107 waves interact with bathymetry around islands they break and can pump water up through the
108 thermocline. These so-called ‘internal waves’ can raise nutrient concentrations in the
109 shallows (Leichter, Stewart & Miller 2003; Wang, Dai & Chen 2007; Aston *et al.* 2018),
110 which in turn can promote heterotrophic feeding and growth rates in corals (Leichter &
111 Salvatore 2006; Fox *et al.* 2018; Williams *et al.* 2018), and ultimately drive broad spatial
112 transitions in benthic functional group dominance around islands (Aston *et al.* 2018). Coral
113 reefs are also hydrodynamically connected by additional physical processes such as lagoonal
114 outflow and surface downwelling that can move allochthonous nutrient sources between reef
115 habitats (Williams *et al.* 2018) and, in the absence of confounding local human impacts,
116 enhance reef productivity and function (Graham *et al.* 2018).

117 Seawater temperature is another key determinant of coral reef persistence and
118 function. Most coral reef ecosystems occur in waters with a seasonal minimum sea-surface
119 temperature of 18°C (Kleypas, McManus & Menez 1999). Marginal reef communities can
120 form in waters below 22°C, and this can be explained by the interacting effect of temperature
121 with light, nutrients, and aragonite saturation (Couce, Ridgwell & Hendy 2012). Bounded
122 within these temperature limits, gradients in seawater temperature influence the dominance
123 and life history of individual reef organisms. For example, hard coral cover

124 decreases at lower temperatures, while macroalgae become more prevalent with latitudinal
125 and cyclical seasonal drops in temperature (Glenn, Smith & Doty 1990; Fulton *et al.* 2014;
126 Williams *et al.* 2015a). Browsing herbivorous fishes become more dominant in cooler waters,
127 while warmer waters support an increased biomass of detritivorous fishes (Floeter *et al.*
128 2005; Hoey, Pratchett & Cvitanovic 2011; Heenan *et al.* 2016). Fish body size also varies
129 predictably with temperature. Body size is inversely related to temperature due to the
130 increased growth rate, earlier maturation, and shorter life span of individuals at higher
131 temperatures (Atkinson 1994; Trip *et al.* 2008; Taylor, Trip & Choat 2018).

132 These natural constraints on a reef's biophysical and functional form do not act in
133 isolation and appear predictable in the absence of confounding local human impacts
134 (Williams *et al.* 2015a). However, anthropogenic activities have become a dominant driver of
135 coral reef ecosystems across a broad range of socioeconomic and cultural contexts, increasing
136 the complexity of drivers and their interactions that govern ecosystem state (**Fig. 2**).

137

138 **Socioeconomic and cultural drivers: a new reality for coral reefs**

139 The footprint of human activity is evident on coral reefs at all trophic levels. Fishing
140 has dramatically reduced overall fish biomass on coral reefs (Williams *et al.* 2011; MacNeil
141 *et al.* 2015; Graham *et al.* 2017), with an emphasised loss of herbivores (Edwards *et al.* 2014)
142 and top predators (Sandin *et al.* 2008; Valdivia, Cox & Bruno 2017; Cinner *et al.* 2018) and
143 thus the key ecosystem functions they perform. Fishing can also disrupt the basic physiology
144 and behavior of target species, including the sex change dynamics (Taylor 2014) and flight
145 responses of reef fishes (Januchowski-Hartley *et al.* 2012), both of which have the potential
146 to affect overall reef ecosystem function (Madin *et al.* 2010).

147 Land-use change alters sedimentation regimes and nutrient input to reefs (Wolanski,
148 Martinez & Richmond 2009). In conjunction with fishing (McClanahan *et al.* 2003), these

149 effects can favour the competitive superiority of fleshy algae (Barott *et al.* 2012) to ultimately
150 promote their overall dominance (Smith, Hunter & Smith 2010; Smith *et al.* 2016). Dredging
151 and plastic pollution are increasing coral disease prevalence on reefs (Pollock *et al.* 2014;
152 Lamb *et al.* 2018), which in turn contributes to a loss of live coral cover and reduced reef
153 calcification rates. Humans are also re-structuring coral reef microbial communities (Kelly *et*
154 *al.* 2014), and promoting the abundance of disease-causing bacteria and viruses (Dinsdale *et*
155 *al.* 2008). Remarkably, human-introduced invasive rats can lower fish growth rates and levels
156 of herbivory on reefs by preying on seabirds that would otherwise deliver offshore nutrient
157 subsidies to shallow waters bordering the islands (Graham *et al.* 2018).

158 Globally, human-induced warming of the ocean is resulting in increasingly frequent
159 mass coral bleaching events (Hughes *et al.* 2018a) that are transforming coral assemblages
160 (Hughes *et al.* 2018b) and in some cases causing regime shifts to fleshy macroalgae (Graham
161 *et al.* 2015). In combination with human-induced ocean acidification (Albright *et al.* 2016),
162 these shifts in benthic composition have broader ecosystem effects, from compromising the
163 growth of reef structures (Perry *et al.* 2013; Perry *et al.* 2018) to changing the diversity,
164 abundance, and behaviour of other reef-associated organisms (Keith *et al.* 2018; Richardson
165 *et al.* 2018; Stuart-Smith *et al.* 2018). Hence, myriad interconnected human drivers are
166 rapidly changing the structure and function of reefs (Pendleton *et al.* 2016).

167 The proximate human impacts to reefs described above are, themselves, ultimately
168 dictated by underlying distal socioeconomic and cultural drivers, such as global trade,
169 markets and finance, as well as the movement and behavioral choices of people and their
170 associated demands on coastal resources (Kittinger *et al.* 2012; Hicks *et al.* 2016a; Norström
171 *et al.* 2016) (**Fig.2**). While the coral reef research community has made significant advances
172 in measuring the response (decline) of coral reef ecosystems to these distal socioeconomic
173 and cultural drivers, we have not done so intentionally in an *a priori* manner. Instead, we

174 have indirectly measured their effect by studying their emergent proximate impacts, such as
175 commercial and recreational fisheries (Fig. 2). We hypothesise this is limiting our ability to
176 accurately predict spatiotemporal changes of contemporary reef ecosystems. We further
177 suggest that these human socioeconomic and cultural drivers can combine to become such a
178 dominating structuring force of reef ecosystem state that they overwhelm any influence of a
179 reefs' surrounding natural biophysical setting. Williams *et al.* (2015a) tested this hypothesis
180 by quantifying the relationship between coral reef benthic communities and gradients in
181 biophysical drivers across Pacific islands that spanned a gradient of human density. At
182 island-mean scales, they demonstrated that biophysical drivers were able to strongly predict
183 coral reef ecosystem state when human density was low, but that these relationships were lost
184 or fundamentally altered when human population density increased. We propose that this loss
185 of predictive power over reef ecosystem state will be regained when human socioeconomic
186 and cultural variables are instead fully integrated into analyses and used as predictors in the
187 modeling framework. Further, we argue that implementing a multi-scaled macroecology
188 approach will provide a more nuanced understanding of how biophysical, socioeconomic,
189 and cultural drivers interact across spatial and temporal scales to influence coral reef patterns
190 and processes.

191 Work has begun to address the crucial data and knowledge gaps linking the structure
192 and function of natural ecosystems to the distal socioeconomic and cultural drivers that
193 underpin their proximate drivers of change. Examples include the socioeconomic drivers of
194 biodiversity loss and societal response capacities of hyperdiverse tropical ecosystems
195 (Barlow *et al.* 2018), quantitative data on land grabbing and the international trade of coral
196 reef resources (Norström *et al.* 2016), and the increasing amount of social science
197 quantitative indicators people can use in social-ecological systems research and sustainability
198 science (Hicks *et al.* 2016b). These types of data can improve our ability to predict the

199 dynamics of natural ecosystems (Hicks *et al.* 2016a), including coral reefs. For example,
200 distance to markets is a better predictor of the condition of reef fisheries than local human
201 population densities in the vicinity of the reefs (Cinner *et al.* 2013). Further, combining travel
202 times to reefs, as a measure of their accessibility (Maire *et al.* 2016), with human population
203 sizes within a given distance, produces a metric known as ‘gravity’, which is a stronger
204 predictor of fisheries exploitation on any given reef than human population density alone
205 (Cinner *et al.* 2016; Cinner *et al.* 2018). When reef fisheries are quantified as either doing
206 better (bright spots) or worse (dark spots) than expected given their natural biophysical
207 bounds, it is human socioeconomic and cultural data such as customary taboos, marine
208 tenure, and levels of local engagement in management, that are able to better predict the two
209 outcomes (Cinner *et al.* 2016). We highly advocate these recent approaches and anticipate
210 that unless we start to more routinely monitor, decipher, and account for socioecological links
211 across scales we will become unable to predict spatiotemporal changes to coral reef
212 ecosystem dynamics. We need to move to a point where we are integrating this thinking in to
213 new ecological theories and paradigms that explicitly insert humans in to the equation across
214 scales. As such, we require a new multi-scaled macroecological approach to coral reef
215 ecology that is aligned with our current time, i.e., the Anthropocene.

216

217 **Looking to the future: coral reef ecology in the Anthropocene**

218 The past century has seen an evolution in ecological thinking, with theories and
219 frameworks continuously updated and refined based on our ever-increasing understanding of
220 natural systems. Coral reef science is no exception. Early theories and descriptions of the
221 origins, structure, and distribution of coral reefs (Darwin 1842) were extended to encompass
222 a more mechanistic, ecological, and process-based understanding of these diverse ecosystems
223 (Odum & Odum 1955). Concurrently, the broad field of ecology was evolving across

224 multiple terrestrial and aquatic systems. The longstanding Clementsian view of unidirectional
225 ecological succession (Clements 1936) gave way to an appreciation of more complex
226 interacting processes governing ecosystem dynamics and non-equilibrium theory (Odum
227 1969; Whittaker 1970). Concepts of ecological resilience then developed (Holling 1973) and
228 were later directly applied to non-equilibrium systems like coral reefs (Connell 1978;
229 Nyström, Folke & Moberg 2000). More recently, resilience theory has expanded to embrace
230 a social-ecological systems framework that explicitly treats humans as internal rather than
231 external to the system (Berkes & Folke 1998; Biggs *et al.* 2012). These works have given rise
232 to a range of ecological paradigms that have formed our views on what defines coral reef
233 ecosystems, what shapes them, and how they function.

234 What is now unequivocal is that human imprints can be observed at all biophysical
235 scales, across all levels of biological organisation, and in the processes upon which ecological
236 theories rest, such as species dispersal, colonisation, invasion, extinction, isolation, tolerance,
237 and competition (Ellis 2015). Acknowledging that humans have emerged as a significant
238 force in nature, “natural” biophysical processes that previously determined the assembly,
239 dynamics, structure and functional ecology of ecological communities, may now be
240 overwhelmed by anthropogenic activities (**Fig. 2**). This new situation poses an intriguing
241 question: *to what extent do traditional ecological paradigms capture and explain the*
242 *ecological patterns and processes we observe in the Anthropocene?*

243 The unprecedented breakdown of isolation by human migration and trade has caused
244 dramatic changes to the dispersion and diversity of species globally (Meyerson & Mooney
245 2007; Westphal *et al.* 2008; Banks *et al.* 2014), with both positive and negative impacts to
246 ecosystem services (Charles & Dukes 2007; Pejchar & Mooney 2009; Schlaepfer, Sax &
247 Olden 2011). This loss of isolation is potentially compromising the explanatory and
248 predictive power of traditional ecological models. For example, when Helmus *et al.* (2014)

249 investigated the species–isolation relationship for anole lizards among Caribbean islands they
250 found that anole biogeography reflects anthropogenic processes, such as economic isolation
251 of human populations, rather than geographic processes postulated in traditional island
252 biogeography theory. Similar perturbations to the effectiveness and relevance of traditional
253 ecological models and paradigms are likely occurring in the ocean.

254 In the marine environment, humans have influenced species biogeography by the
255 unintentional and intentional introduction of species through transport and trade. Examples
256 include ballast water release from cargo ships, aquaculture, and the aquarium industry
257 (Padilla & Williams 2004). Moreover, we have created artificial ‘islands’ that are no longer
258 static stepping-stones, but instead, float and move. Human-derived flotsam is providing a
259 dispersal mechanism for tropical Atlantic fishes to cross the deep-water Mid-Atlantic Barrier
260 (Luiz *et al.* 2012) and is facilitating alien species invasions (Gregory 2009). Floating plastic
261 waste harbours distinct microbial assemblages, the so-called ‘Plastisphere’ (Zettler, Mincer &
262 Amaral-Zettler 2013), with this unique biotope providing a mechanism by which disease-
263 causing pathogens of reef corals spread in the Anthropocene (Lamb *et al.* 2018). Recent
264 biophysical dispersal models have even offered the provocative suggestion that human
265 infrastructure, such as oil and gas installations across the North Sea, can form a highly inter-
266 connected regional network of coral ecosystems capable of supplying larvae to natural
267 populations downstream (Henry *et al.* 2018). In these instances, to fully understand and be
268 able to predict the observed ecological dynamics at play requires a new strategy. The human
269 socioeconomic and cultural processes governing such modifications to species dispersal and
270 diversity must become an integral part of ecological theory and practice as much as
271 biological and geophysical processes are today (Ellis 2015; Österblom *et al.* 2017).

272 The pervasive global influence of humans in governing the spatial dynamics of
273 ecological systems requires new theoretical advances to study, define, and sustainably

274 manage them (Herrick & Sarukhán 2007; Hulme-Beaman *et al.* 2016; Rose *et al.* 2016;
275 Cadotte *et al.* 2017). Coral reefs are no exception; they face a new reality with their dynamics
276 governed by cross-scale interacting biophysical and human socioeconomic and cultural
277 drivers (Norström *et al.* 2016; Hughes *et al.* 2017) (**Fig. 2**) and we question whether
278 traditional coral reef paradigms accurately capture this complexity.

279 Moving forward, humans (and their activities) must become an integral part of coral
280 reefs and their dynamics. For this purpose, we propose ‘*social-ecological macroecology*’ as a
281 novel concept for studying coral reefs in the 21st century. This approach embeds
282 macroecology – the study of organism-environment relationships at large spatial and
283 temporal scales (Brown & Maurer 1989; Keith *et al.* 2012; Heffernan *et al.* 2014), within a
284 social-ecological systems framework. In doing so, social-ecological macroecology explicitly
285 inserts the presence, behaviour, dynamics, and ecology of the human species into the
286 equation, and does so across spatial and temporal scales. We stress the critical role of a
287 macroecology approach – the scale of observation directly influences the ecological
288 finding(s), their interpretation, and their subsequent use in guiding coral reef management.

289 Taking a social-ecological macroecology approach to studying coral reefs will require
290 some innovative thinking and we suggest four core pathways to this approach:

291

292 1. **Identify the interactive effects of biophysical and human socioeconomic and**
293 **cultural drivers of coral reef ecosystems across spatial and temporal scales.** This
294 will improve our predictive capacity, such that we understand how changing one
295 parameter, either biophysical, socioeconomic or cultural, at any specific scale
296 interacts with other drivers at other scales to alter coral reef ecosystem structure and
297 function.

298

299 2. **Test the robustness of classic coral reef paradigms.** We need to revisit and test
300 whether classic ecological paradigms developed, in many instances, outside of the
301 social-ecological systems framework, are still able to capture the dynamics of
302 Anthropocene reefs accurately.

303

304 3. **Adapt current coral reef paradigms.** If classic paradigms fail to capture the
305 spatiotemporal dynamics of reefs today accurately, we should explore whether
306 adapting these paradigms, by including human dynamics as drivers, substantially
307 improve their predictive capacity.

308

309 4. **Develop novel coral reef social-ecological paradigms.** In some cases, adapting
310 existing coral reef paradigms may not be enough; we will need to develop novel rules
311 and theories to create '*social-ecological paradigms*,' where human dynamics are part
312 of the paradigms rather than the external drivers of them.

313

314 We will need to continually re-visit and test the performance of any of the adapted or
315 novel paradigms developed under this approach due to the unprecedented rate of social and
316 ecological change in the Anthropocene. In following these guidelines, coral reef ecologists
317 should be able to identify, at any given spatial or temporal scale of observation, which
318 interacting predictors (biophysical, socioeconomic or cultural) offer the best predictive
319 capacity over coral reef ecosystem structure and function.

320

321 **Conclusions**

322 We remain convinced that human social, cultural and economic processes must
323 become an integral part of ecological theory and practice as much as biological, geological

324 and physical processes are today. This warrants a revisiting of traditional coral reef
325 ecological paradigms and theories and either adapting them so that they capture
326 contemporary dynamics of intertwined social-ecological systems, or developing novel social-
327 ecological theories. This will be challenging and will require truly interdisciplinary
328 collaborations between researchers in the natural and social sciences.

329

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335 The authors declare that they have no conflict of interest.

336

337 **Author contributions**

338 G.J.W, N.A.J.G, J-B.J, A.V.N, and M.N conceived the research idea and led its development
339 with all the other authors providing input. G.J.W led the writing with all authors contributing
340 to the discussion and editing of the paper.

341

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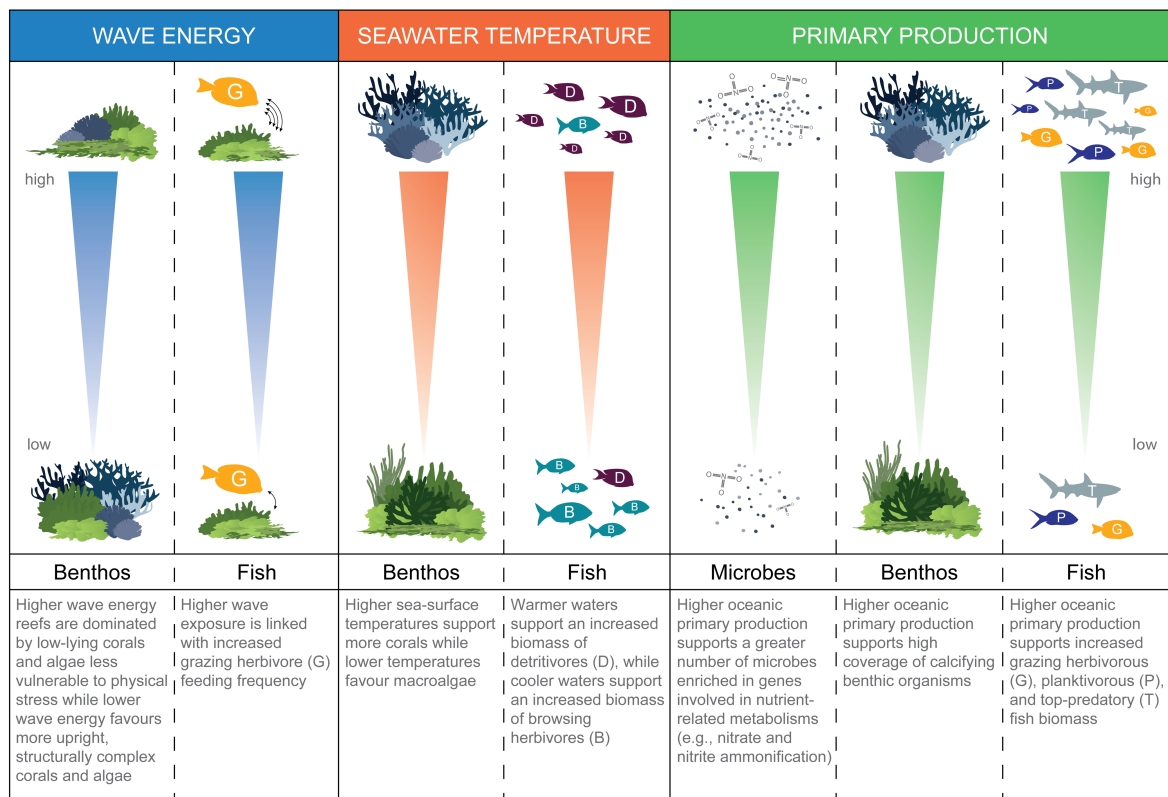
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649 **Figures and figure legends**

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652 **Figure 1.** Examples of the natural bounds set by gradients in biophysical drivers on coral reef

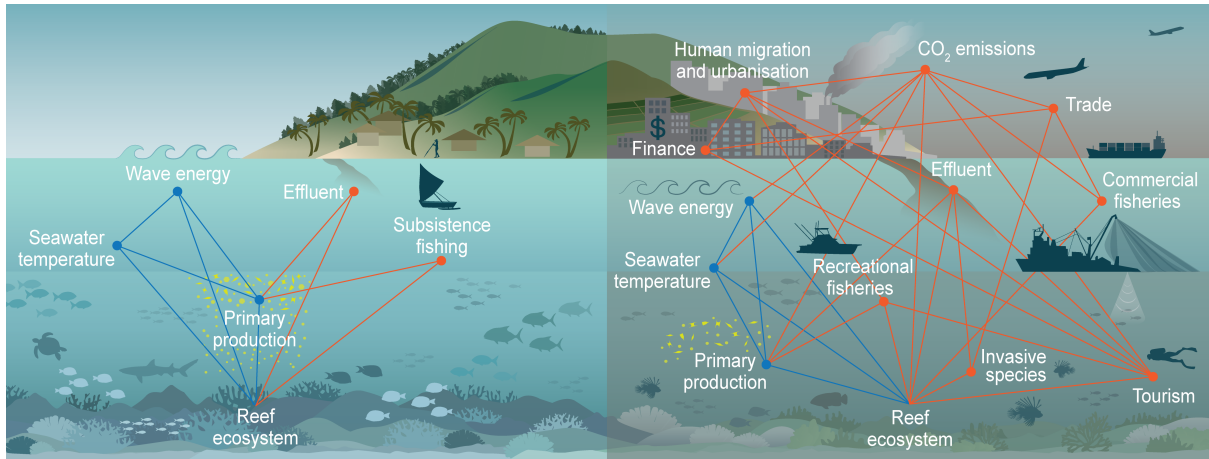
653 ecosystem structure and function across trophic levels, from microbes to sharks.

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660 **Figure 2.** Drivers of coral reef ecosystems pre- and post- Anthropocene. Before coral reefs
 661 entered the Anthropocene, their ecosystem state was heavily governed by natural biophysical
 662 drivers, even in the presence of small subsistence-based human populations (left). This is still
 663 the case for some remote, uninhabited coral reef islands and atolls that are far removed from
 664 direct human impacts. However, many coral reefs today are impacted by local human drivers,
 665 such as commercial and recreational fishing and effluent discharge from land (right).
 666 Importantly, these proximate drivers of reef ecosystem state are themselves ultimately dictated
 667 by a complex network of underlying socioeconomic and cultural drivers (right). The
 668 biophysical drivers are still present on Anthropocene reefs, but their relative influence in
 669 governing reef ecosystem state is likely greatly reduced. Because of this, we propose the need
 670 for ‘*social-ecological macroecology*’ which embeds macroecology – the study of organism-
 671 environment relationships at large spatial and temporal scales, within a social-ecological
 672 systems framework.

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