

The crowding out of complex social goods

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2

3 Abstract

4 The valuation of ecosystem services to inform natural resource management and development has
5 gained acceptance in many arenas. Yet, contemporary economic valuation is constrained to the
6 appraisal of *simple goods* that generate benefits that accrue to individuals, neglecting *complex goods*
7 that generate benefits that accrue to society more broadly. Methodological barriers to the valuation
8 of *complex social goods* have led to their frequent omission from natural resource management
9 deliberations. The prevailing valuation paradigm that focuses on *simple individual goods* may erode
10 conservation efforts by crowding out the institutions and behaviours that support socially
11 constructed ecosystem service values. Erosion of these values ultimately harms the environment
12 and society as a whole. The institutionalisation of appropriate methods for estimating the value of
13 *complex social goods* alongside existing methods for valuing *simple individual goods* within
14 international conservation, development and policy-making discourses, is therefore an important
15 evolution for sustainable natural resource management.

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17 Keywords

18 Non-market valuation, Cost Benefit Analysis, Individual Welfare, Social Welfare

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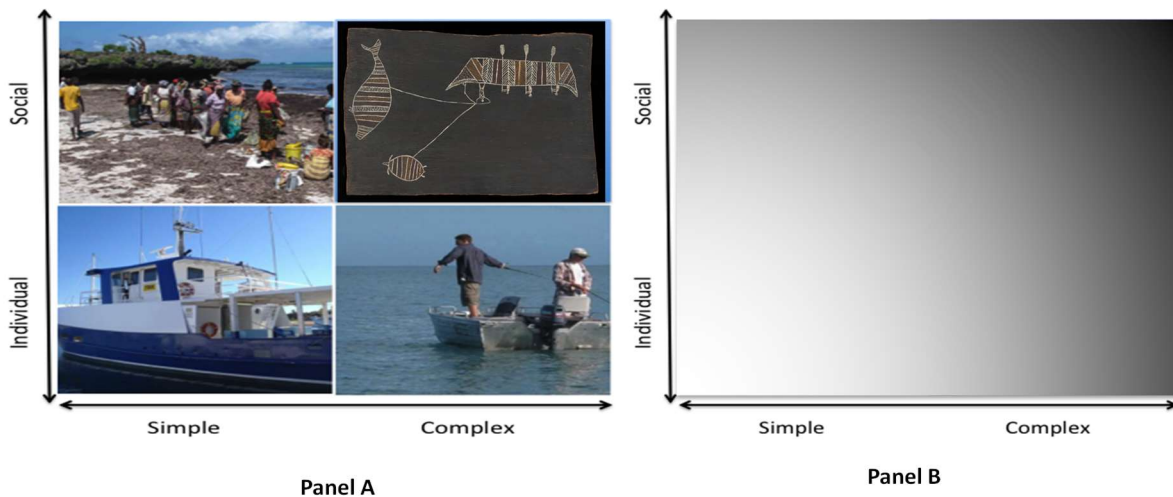
21 **1 Introduction**

22 Natural resource management and development have long been informed by the economic
 23 valuation of various interrelated ecosystem services and over the past two centuries, these valuation
 24 methods have matured significantly. Putting a ‘price’ on nature allows decision makers to quantify
 25 and elucidate the myriad impacts of development. Efforts to “mainstream the values of biodiversity
 26 and ecosystem services into decision-making at all levels” are broadly recognized; non-market
 27 valuation is now firmly established in international policy discourses concerning natural resource use
 28 and conservation and is actively encouraged by state agencies through the publication of guidelines
 29 and handbooks (for example, Commonwealth of Australia, 2006; H.M. Treasury, 2003). There is,
 30 however, growing recognition that dominant valuation approaches make only a subset of benefits
 31 visible (Kenter *et al.*, 2015) and that by focusing entirely on ‘the market’ and on market incentives,
 32 we may undermine the motivations and thus behaviors we seek to promote (Gneezy *et al.*, 2011).

33 An emerging body of literature now recognizes a spectrum of value typologies (Vatn, 2009;
 34 Schwartz, 2015; Kenter *et al.*, 2015). These typologies vary in complexity, but highlight that (at the
 35 simplest level), goods may be thought of as having (at least) two distinct dimensions—depicted as
 36 axes in Figure 1. Along the horizontal axis, goods range from simple, with separable benefits (e.g.,
 37 food), to complex, with multiple, inter-related and inseparable benefits (e.g., wedding banquet).
 38 Along the vertical axis, benefits derived from goods range from individually constructed, reflecting
 39 individual motivations and ‘utility’ (e.g., ambition and success, respectively), to socially constructed,
 40 representing communal norms, relations, and expectations (e.g., social cohesion).

41 The various possible combinations of goods and benefits may be illustrated by different types of
 42 fisheries (Panel A, Figure 1), described in Table 1. As discussed in section 2 economists’ ability to
 43 estimate values along these axes is variable. Hence, a limited range of goods has been assessed, with
 44 significant policy implications (discussed in section 3).

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47 **Figure 1:** Panel A: Various ecosystem services and their associated benefits may be visualised in two dimensions.
 48 Along one axis, goods range from simple to complex, illustrated here using different types of fisheries. Along the
 49 other, benefits range from individual to social constructs. Individually constructed benefits reflect individual
 50 ‘utility.’ Socially constructed benefits relate to social norms, relations, and expectations. Panel B: Established
 51 methodology sheds light on values relating to simple-individual goods (such as commercial fisheries) with other

52 types of goods left mostly in the dark (e.g. recreational fisheries, subsistence fisheries, and in particular, traditional
 53 fisheries).

54 **Table 1: Simplified typology of goods and benefits** (derived from insights provided by Vatn 2009;
 55 Schwartz, 2015; Kenter *et al.* 2015)

		Type of good	
		Simple	Complex
Type of Benefit	Individual	Commercial fisheries are an example of a simple good bearing individually constructed benefits. Although the identify of commercial fishers is often tied to their occupation (highlighting the importance of cultural values to these fishers), the primary motivation for commercial fishing (particularly incorporated companies) tends to be economic. The benefits associated with this industry (e.g., contribution to GDP) are the sum outcome of individual motivations and incentives to procure income.	Recreational fishing is an example of a complex good that bears individually constructed benefits. Recreational fishing makes a contribution to individual utility – although the contribution comprises complex and intertwined cultural/recreational and economic benefits.
	Social	Small-scale fisheries exemplify simple goods bearing socially constructed benefits. They often provide an economic safety net for vulnerable members of a community. Such fisheries tend to feature strong norms of fair access (Hicks et al., 2014), reciprocity and taboos concerning trade-offs that promote the fair distribution of benefits (Daw et al., 2015). In many fisheries along the coast of the western Indian Ocean, catch is distributed amongst a range of stakeholders, including fishers, traders and small-scale processors. When the take is plentiful, this arrangement is predominantly financial. However, when catches are low, the most vulnerable in society (often widowed women) are assured subsistence without financial exchange (Hicks et al., 2014; Daw et al., 2015).	Traditional Indigenous fisheries are complex goods bearing socially constructed benefits. While the basic acts of hunting and fishing may generate simple, individual benefits like nutrition and income, the relations and meanings formed through these practices’ associated norms, traditions, stories, and ceremonies bear socio-cultural benefits which are often more significant than the individual benefits (Delisle, 2013; Watkin Lui et al., 2016).

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58 **2 A very brief history of non-market valuation**

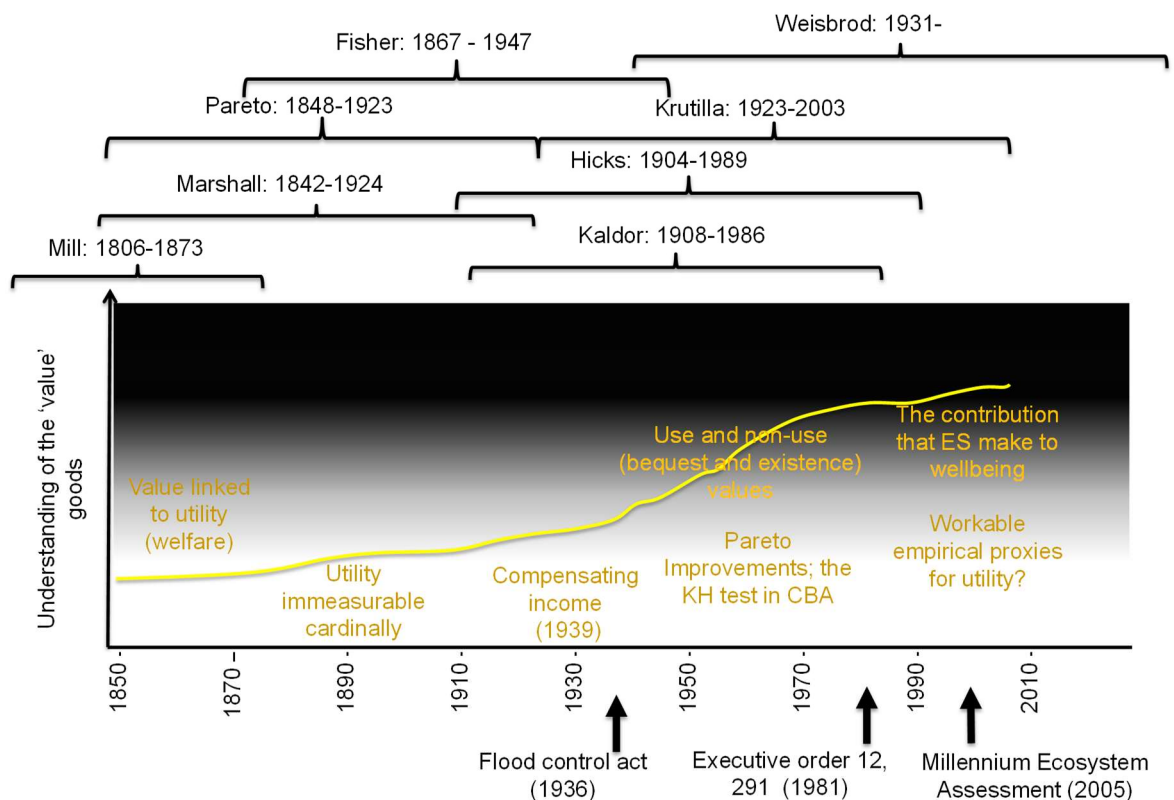
59 Although contemporary discourse about the ‘value’ of environmental goods and services is
 60 dominated by the concept of price (exchange value), this does not mean that economists believe
 61 price reflects value; rather, non-market valuation methods are firmly grounded in the broader idea
 62 of *utilitarianism*—i.e., value is a matter of usefulness (ideas largely attributable to Mill and Marshall).
 63 For most economists, the term *utility* represents the satisfaction that people gain from the use (or
 64 *non-use*) of a good or service; economists are thus concerned about people’s satisfaction. For
 65 economists, maximizing someone’s utility thus implies making that person as satisfied as possible.
 66 Many welfare and environmental economists have generalized this notion of utility to the level of
 67 society, whereby ‘social utility’ may be thought of as a measure of social welfare.

68 Environmental economists often seek to highlight the contribution that environmental goods and
 69 services make to individual or social welfare (utility/wellbeing) and have developed numerous
 70 methods of doing so. Below, we provide a brief (and non-definitive) summary of key events relevant
 71 to the development of these methods and of Cost-Benefit Analysis (CBA), an institutionalised
 72 framework for weighing up a variety of market and non-market costs and benefits (see Hanley and
 73 Spash, 1993, for a more comprehensive overview). We use Figure 2 to depict key contributing

74 economists, concepts and policies chronologically, providing a notional measure of the growth in our
75 ‘knowledge’ of values and valuation methods with a yellow line (with undefined units of ‘knowledge’
76 measured on the vertical axis).

- 77 • Late 1800s/early 1900s. Mill and Marshall clarify the distinction between value and price, with
78 the notion of *utility*—linked to the idea of individual, and potentially also social ‘welfare’ (or
79 wellbeing)
- 80 • Early 1900s. Economists and social scientists collaborate frequently on research into individual
81 and collective wellbeing (*utility/welfare*).
- 82 • 1930’s. Influential economists—namely, Irving Fisher and Vilfredo Pareto—successfully argue
83 that utility cannot be measured in a way that facilitates meaningful interpersonal comparisons
84 (Kristoffersen, 2010). From this point, for almost a century, economists and social scientists
85 largely pursue different research agendas relating to wellbeing/utility/welfare.
- 86 • The US Flood Control Act of 1936 (The Act). This act legislated construction of an unprecedented
87 litany of civil infrastructure works and obligated the government to undertake flood control
88 efforts in the interest of the “general welfare,” so long as a development project’s “benefits to
89 whomever they may accrue are in excess of the estimated costs” (US Congress, 1936, p. 1570).
90 The Act established a concrete policy imperative for research into the social costs and benefits of
91 natural resource development. Critically, the Act established that the improvement of social
92 welfare necessarily required one to account for a diverse range of potentially competing
93 interests, but it did not specify how this was to be done.
- 94 • Late 1930s. Hicks (1939) demonstrated that the welfare gains or losses arising from the addition
95 or removal of goods and services can be articulated in terms of ‘income-equivalent’
96 compensations. That is, the cost (or benefit) to an individual of the loss (or gain) in their access
97 to a good can be calculated by determining how much remuneration would be required to
98 compensate them for that loss (formally, to keep their utility constant).
- 99 • Late 1930s—early 2000s
 - 100 • The Total Economic Value (TEV) framework slowly developed, helping to clarify ways in
101 which people benefit from the environment. Simplistically, the contemporary version of
102 the framework identifies three broad categories: ‘direct use values’, where people
103 benefit from the direct use of environmental goods and services (e.g., agriculture);
104 ‘indirect use values’, where people benefit indirectly from the environment (e.g., from
105 soil substrate water filtration—see Ciriacy-Wantrup, 1955, for an early discussion of
106 these benefits); and ‘non-use values’, where people benefit from the environment
107 without using it—as for existence values (e.g., knowing the environment is there) and
108 bequest values (e.g., leaving the environment intact for future generations) (Krutilla,
109 1967); and option values (e.g., maintaining the environment for potential future use)
110 (Weisbrod, 1964).
 - 111 • Numerous methods for estimating the ‘income-equivalent’ compensations (interpreted
112 as ‘value’) of the different types of use and non-use values were developed (see
113 Appendix A for an overview). Some methods are only capable of estimating some values
114 (e.g., the travel cost method can estimate recreational use-values, but cannot estimate
115 bequest values). Each method thus generates subtly distinct information (e.g.,
116 expenditure, consumer surplus, willingness to pay) about different types of values. All
117 methods are conditioned on assumptions, but two key assumptions dominate: first, that
118 the constituents of utility are distinct, separable and additive (across goods and across
119 people); and second, that income-equivalent compensations are an appropriate framing
120 of economic trade-offs.

- Most research concerning decision-making for improved social welfare was based on two compelling and related principles, inextricably linked to CBA. First, was the idea of a *Pareto Improvement*. Pareto argued that one could unambiguously improve social welfare if the welfare of at least one person increased, without negatively impacting the welfare of another. Second, was the notion of a *Potential Pareto Improvement* (with the related *Kaldor-Hicks* test, developed by Hicks and Kaldor (Broadway, 1974). It was noted that if the benefits of a project outweighed its costs, then the project could potentially improve social welfare since those who benefited would have more than enough to compensate those who lost.
- Early 21st century. The 2005 Millennium Ecosystem Assessment ignited broad interest in the contribution that ecosystem services make to human wellbeing. This further highlighted the need for policy makers to consider the ‘value’ of the environment more broadly, and thus increased interest in non-market valuation methods. It also served to focus attention on wellbeing, rather than exclusively on income-compensations or ‘price’. Arguably, this new focus helped highlight wellbeing research undertaken by social scientists (with entire journals, e.g., *Social Indicators Research*, *Journal of Happiness Studies*, devoted to the topic).
- Recent decades. An increasingly vocal group of behavioural economists have argued that individual welfare (i.e., utility) may in fact be cardinal (Barberis, 2013), leading to the emergence of a new approach to valuation: the life-satisfaction (LS) approach (see, for example, Frey *et al.*, 2009). Simplistically, this approach uses self-reported measures of (individual) well-being as proxies for utility (Gowdy, 2005) and regresses these measures against factors known to influence it (including environmental goods and services). LS thus enabled researchers to assess these ecosystem services’ marginal value (to utility) directly, without needing to assess values as income-equivalent compensations (though that can also be done).



145
146 **Figure 2:** Hypothesised growth in our understanding of the ‘value’ of environmental goods and services over time
147 (shown with a yellow line; notionally measured, without units, on vertical axis). Overlays, include the names and

148 life-spans of (a subset of) key researchers (top), key concepts (middle – yellow font) and key
149 policies/investigations (below axis) which contributed to the development of non-market valuation methods and
150 methods for weighing up diverse market and non-market values (notably, CBA). Dark shades in upper diagram
151 indicate values still poorly understood and thus ‘in the dark’.

152 The US Flood Control Act of 1936 and subsequent institutionalisation of Cost Benefit Analysis (CBA)
153 thus drove significant advancements in the sub-fields of welfare and environmental economics
154 resulting in the identification of different types of goods and services, and the development of
155 numerous related non-market valuation methods (Appendix A). The historical development of these
156 methods means that most contemporary valuation exercises—and associated decision-making
157 frameworks for aggregating the impacts of development, such as CBA—assume (even if only
158 implicitly) that: (a) changes in social welfare may be assessed simply as the sum of each person’s
159 costs and benefits (conceptually, changes in their utility); (b) each person’s utility depends on the
160 utility they gain from each good and service they select; and (c) social welfare (utility) is the sum of
161 individuals’ utility.

162 The emphasis on ostensibly separable components of utility, measured at the individual level
163 (individual good; individual person) and aggregated to the scale of society has narrowed the focus of
164 (most) economic valuation exercises to a limited range of *simple* (separable) goods and services
165 (Kenter *et al.*, 2015). Techniques for estimating the value of simple individual goods (such as
166 commercial fisheries, see Figure 1) are well established (assessing, for example, the value of output,
167 of ‘rent’, see Appendix A). Though it can be difficult to estimate the value of complex individual
168 goods (like recreational fishing, Figure 1) (Turner *et al.*, 2003), methods and examples exist (see
169 Appendix A for valuation techniques based on indirect market associations—e.g., travel cost
170 method, hedonic pricing). Developments within the fields of behavioural and institutional economics
171 have allowed promising innovations, particularly with regards to the conceptual inclusion of *complex*
172 *goods* (Barberis, 2013), and the LS approach to valuation allows one to consider multiple values
173 simultaneously. But the underpinning individualism of these approaches necessarily constrains
174 common valuation exercises to the assessment of goods and services that generate ‘utility’ for
175 individuals.

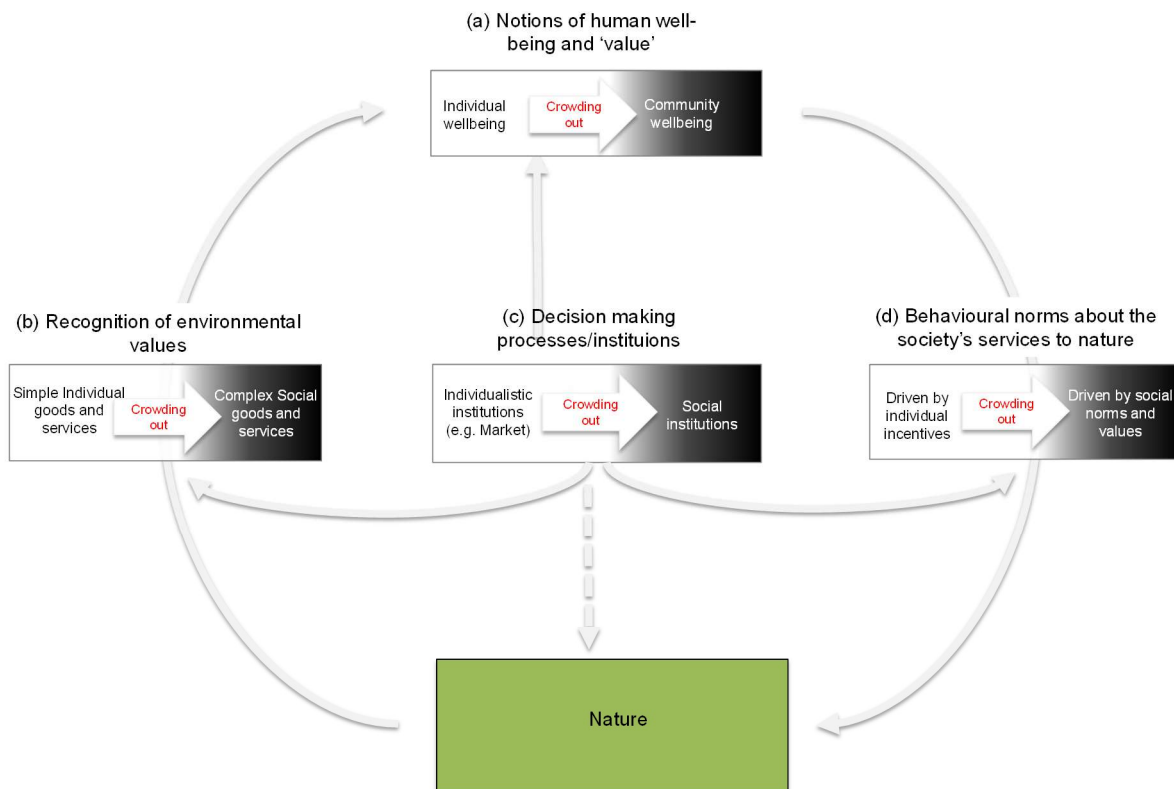
176 Thus, since the period immediately following the Act, when our knowledge about the value of non-
177 market goods (and methods for assessing their value) grew rapidly, the rate of knowledge-growth
178 has slowed (Figure 2). After passage of the Act, considerable research focused on developing
179 neoclassical methods (which assumed utility to be measurable only in ordinal terms) to assess
180 previously understood and newly identified use and non-use values. But since the late 1900s,
181 attention has, arguably, been focused most ardently on the refinement of existing methods (e.g., on
182 the development of more sophisticated econometric techniques for analysing data, on incorporation
183 of insights from psychology into choice-modelling experiments). The development of LS comprises a
184 notable exception. Although the difficulties of using existing methods to ascribe monetary value to
185 complex social goods like Traditional Indigenous fishing and hunting (Figure 1) are widely
186 acknowledged (Altman & Whitehead, 2003; Vent and Quiggin, 2007), relatively little attention has
187 been given to the development of new methods capable of doing so. Since the early 1900s, research
188 has thus shed much light on the ‘value’ of *individual goods* (particularly *simple* ones) but, in
189 comparison, we are still very much in the dark when it comes to understanding the ‘value’ of
190 *complex social goods* (Panel B, Figure 1).

191 3 The crowding out of social values, institutions and norms

192 “Motivational crowding” has already been shown to occur as a result of economic interventions
193 (Gneezy et al., 2011; Agrawal et al., 2015). Significantly, the prevailing valuation paradigm that
194 focuses on *individual goods* may also contribute to the erosion of conservation efforts and thus
195 degradation of the environment by crowding out the values, institutions and norms (akin to
196 motivations) that support socially constructed ecosystem service values.

197 Every social context can be understood as an institution, in that there are rules and norms that signal
198 appropriate behaviour (Vatn, 2009). Different social contexts support particular rationalities and
199 emphasize distinct value types. We reason differently as economic actors in the market, for
200 example, than we do amongst family and friends. Thus certain institutions support *simple individual*
201 *goods* (e.g., food items) whereas others support *complex social goods* (e.g., a wedding feast), and
202 each influences how people behave. An economic valuation exercise is no different in that it imposes
203 a set of rules and norms concerning how decisions are to be made (Jacobs, 1997). These rules flow
204 from the valuation method and dictate what is to be investigated and is thus of value, and
205 characterize the goods and benefits (i.e., as simple, complex, individually or socially constructed)
206 (Vatn, 2009). The valuation exercise engages these elements according to its self-delineated
207 analytical framework, ascribing importance to various factors in accordance with prevailing
208 assumptions, conventions and priorities. When conducting a valuation exercise, we are thus
209 imposing an institution—with associated assumptions—that prioritises values at the individual level.
210 Ostrom (1990) and Ostrom *et al.* (1999) clearly demonstrate that common pool resources are not
211 necessarily doomed to *the tragedy of the commons*, and that alternative systems and institutions—
212 especially those that draw upon multiple forms of social capital—are adept to deal with complex
213 natural resource management problems. Yet the institutionalisation of CBA and the valuation
214 exercises that support it serves to inadvertently undermine these other frameworks, norms,
215 behaviours and bodies of knowledge.

216 Figure 3 depicts the mechanism by which this happens and the impact it has on individual and social
217 wellbeing. Discussed above, the prevailing paradigm (i.e., CBA) emphasises individual utility over
218 broader notions of social value (Figure 3, a). Illustrated below, the institutionalisation of valuation
219 methodologies that prioritise the individual serves to crowd-out social goods (Figure 3, b) and
220 socially constructed corporate and political decision-making processes (Figure 3, c). This crowding-
221 out undermines social norms and behaviours that impact the environment (Figure 3, d). Discussed in
222 Section 4, the marginalisation of socially constructed institutions and subsequent degradation of the
223 environment may have direct and indirect negative impacts on social welfare (community
224 wellbeing).

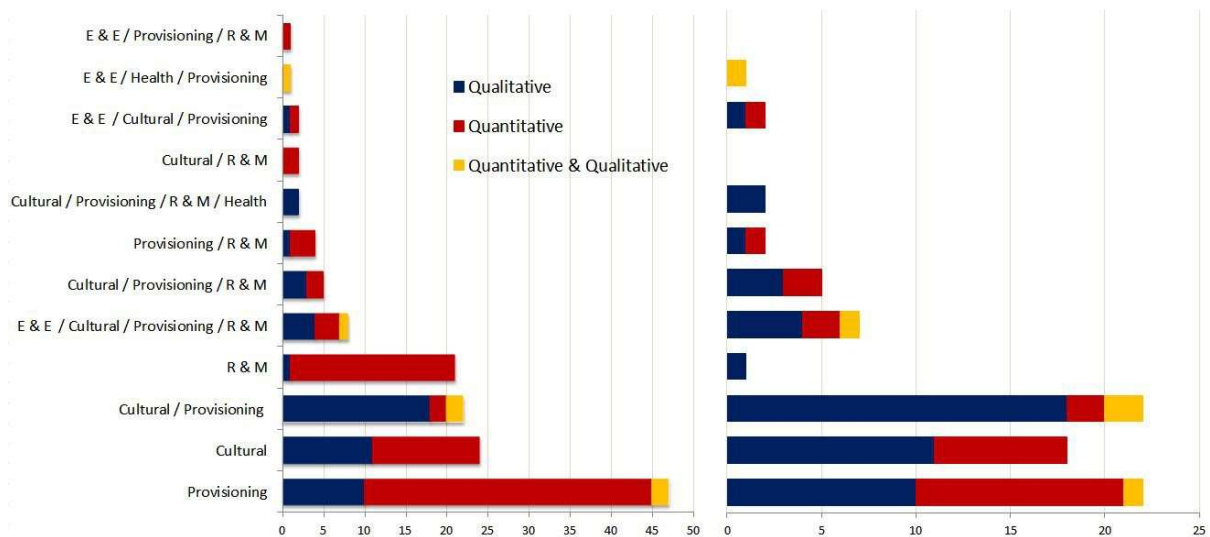


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226 **Figure 3:** The crowding out of social goods, institutions and norms that support the environment and community
 227 wellbeing. CBA and associated non-market valuation techniques (a) prioritize individualistic notions of human
 228 wellbeing over other social constructs thus crowding out: (b) complex social goods and services; (c) socially
 229 constructed decision-making frameworks; and (c) social norms associated with stewardship behaviours. These
 230 socially constructed values, institutions and norms directly support community wellbeing; they also support the
 231 environment, so efforts to protect nature using methods that adopt exclusively individualistic perspectives which
 232 crowd-out social perspectives will not unambiguously improve social welfare (community wellbeing)

233

234 In contrast to *simple individual goods*, *complex social goods* (such as those associated with
 235 Indigenous Protected areas that generate multiple simultaneous cultural and provisioning benefits,
 236 amongst others) are typically only discussed in qualitative terms or obscured altogether (Figure 4,
 237 supported by supplementary e-materials). Economic logic suggests that, *ceteris paribus*, spending
 238 should be directed to programs with the greatest benefit per dollar spent. If, due to their complexity
 239 or because we lack methods for assessing them, entire classes of benefits are routinely omitted from
 240 deliberations, spending will invariably be directed towards interventions and projects that generate
 241 more easily monetized benefits (i.e., towards simple individual goods), rather than to projects that
 242 generate the greatest benefit, per se (at least some of which are likely to be associated with complex
 243 social goods). Our success in validating the value of *simple individual goods* may have unintentionally
 244 stifled our consideration of *complex social goods* and services that generate significant social and
 245 environmental benefits (Figure 1)—and may indeed crowd them out (Figure 3, b) by reducing
 246 investments in activities that support them, relative to activities that support *simple individual*
 247 *goods*.



248

249 **Figure 4:** Studies referring to the ‘value’ of various benefits associated with protected areas in (a) general (N=139)
 250 and (b) specifically Indigenous contexts (N=82), by general methodological approach. We undertook a non-
 251 exhaustive search of studies of protected areas (see supplementary e-materials for list of publications included)
 252 classifying them by type of ecosystem service considered (provisioning, regulation and maintenance (R&M), and
 253 cultural). The categories of Health and Employment & Enabling (E&E) were added to capture other benefits
 254 associated with protected areas that are not always mediated through the environment. If studies considered more
 255 than one type of benefit (frequently complex social goods), all are listed.

256 Society accepts—explicitly and implicitly—that there are different classes of values and that certain
 257 trade-offs should be avoided (Vatn, 2009). Even if not explicitly asking about ‘willingness to pay’,
 258 most contemporary valuation methods assume that an ecosystem service may be ascribed a
 259 marginal value according to a beneficiaries’ willingness to forgo its benefit in exchange for monetary
 260 recompense (Hick’s *compensating income*). This framing contrives the price at which an otherwise
 261 socially unacceptable trade-off may be sanctioned. By assuming the additively separable nature of
 262 the benefit itself, the exercise also implies that an individual may forgo receipt of a benefit without
 263 reducing the value of the ecosystem service that underlies it. Yet a socially constructed benefit—the
 264 cultural significance of the undeveloped natural environment to its Indigenous traditional owners,
 265 for example—cannot, by definition, be divided into constituent partial values. The socially
 266 constructed value of an ecosystem service is inextricably linked to the communal experience of its
 267 benefit; to erode one’s personal benefit is to erode the benefits of all. The neo-classical estimation
 268 of hypothetically disaggregated values posits trade-offs predicated on individualized incentives. This
 269 belies the true nature of the social good and trivializes difficult decisions regarding the best course of
 270 action for society as a whole. At what price should an individual forgo one’s culture, religion or social
 271 identity? By attempting to commodify *complex social goods*, the valuation exercise compromises
 272 their values (Adamowicz *et al.*, 1998) (Figure 3. b).

273 Moreover, most valuation methods generate estimates by assuming that social welfare (or the total
 274 value of a good, or the net social benefit of a project) is the sum of individual welfares (or values, or
 275 net benefits). The simple aggregation of individual estimates of ‘value’ may be inappropriate in some
 276 contexts, particularly when values are held at the community-level (Adamowicz *et al.*, 1998). There is
 277 widespread agreement that benefits associated with the environment are frequently inter-
 278 dependent (i.e., ‘complex’); researchers who aggregate partial estimates purportedly risk double-
 279 counting ‘overlapping values’ (Hein *et al.*, 2006). Yet this framing implicitly precludes the possibility
 280 that inter-dependent values may be synergistic—that the whole may comprise more than the sum of
 281 its parts. As the value of a shoe is critically diminished in the absence of its partner, so too may

282 shared social values be fundamentally complementary. Indigenous peoples, for example, often
283 accumulate knowledge and wealth at the community-level—particularly values associated with
284 cultural integrity and continuity (Chan *et al.*, 2012). This is another way by which the use of
285 individualistic valuation approaches for social goods (complex or otherwise) may serve to undermine
286 social values (Figure 3. b).

287 Critically, the omission of complex social goods in valuation exercises may also have unintended
288 implications on existing values and behaviours. The institutionalisation of CBA has crowded out
289 alternative systems and institutions capable of dealing with complex natural resource management
290 (Ostrom, 1999, 2010) (Figure 3, c) and the framing of environmental issues in monetary terms may
291 weaken intrinsic values and social norms crucial to the promotion of the public good (Gneezy *et al.*
292 2011). The financial incentivisation of environmental stewardship, for example, has been shown to
293 undermine supporting motivations and behaviours (Agrawal *et al.*, 2015), and to increase the costs
294 of natural resource management and community change processes (Larson and Brake, 2011). This
295 may marginalise social norms and behaviours imperative to environmental protection (Figure 3, d).
296

297 4 Potential impacts on wellbeing

298 CBA and the individualistic valuation methods that support it contribute to the crowding out of
299 social goods, institutions and norms that support the environment. They may thus have the
300 unintended consequence of indirectly contributing to environmental degradation. Whether or not
301 that degradation associated with the loss of social goods, institutions and norms outweighs the
302 (individualistic) positive impacts that CBA and valuation may have upon the environment remains an
303 empirical question.

304 But there are other ways in which individual and social wellbeing may be affected. First, the
305 crowding out of social decision making frameworks has been shown to erode social capital, and this
306 has a direct negative impact on individual welfare (Achor, 2011), and may also bear environmental
307 consequences, known to be related to wellbeing. Second, the non-market valuation techniques that
308 support CBA (and other less structured decision making processes) are essentially voting systems
309 weighted by personal wealth (Loomis, 2011). This is because willingness to pay (or amount actually
310 paid) is, amongst other things, a function of ability to pay. Decision processes that use estimates
311 generated from these methods may thus inadvertently privilege the (net) benefits of the wealthy
312 above those of the poor. While weights which correct for income differences are sometimes used
313 within CBA—particularly in developing economies where income inequality is often extreme (Hanley
314 and Barbier, 2009)—such practices are rare. Such practices are also rare in less structured decision
315 making environments. The TEEB initiative (van der Ploeg and de Groot, 2010), for example, reports
316 only the nominal value of various ecosystem services throughout the world, irrespective of sample
317 income distributions (though TEEB does group countries by relative overall income for comparison).
318 There is substantial empirical evidence linking individuals’ environmental preferences with their
319 income (Jacobsen and Hanley, 2009; Hicks *et al.*, 2014) and evidence to suggest that in some
320 contexts, the wealthy have more to gain from developments that erode the environment than the
321 poor (Stoeckl *et al.*, 2013). Natural resource management decisions informed by the naïve
322 application of unweighted ‘values’ may thus be more inclined to approve projects that degrade the
323 environment than other decision making processes while simultaneously exacerbating the growing
324 global inequality of the past several decades (Jaumotte *et al.*, 2013).

325 Third, inequality has been shown to impact individual and community wellbeing (Wilkinson, 1996,
326 Wilkinson and Pickett, 2010; Guy and McCandless, 2012). There are established methods for

327 incorporating variables that account for equity (e.g., the distribution of income) within social welfare
328 functions (SEN – REFS), but by design and intent, most valuation methods ignore distributional
329 issues. In CBA, it is generally assumed that a net social benefit represents a *potential Pareto*
330 *improvement*—with net social benefit represented as the sum of individually constructed benefits. It
331 has been shown that raising the income (rewards) of one individual without lowering the income
332 (rewards) of another can generate significant discord of individuals who do not receive the higher
333 income (reward) (Brosnan and De Waal, 2003). Unless monetary estimates of the ‘costs’ of social
334 discord and ‘benefits’ related to equality are included in CBAs, one cannot assume that the
335 measured net benefits represent a genuine Pareto improvement (since certain costs and benefits
336 have been omitted from the assessment).

337

338 5 Conclusion

339 The social estimation of value has been long unbound from the concept of ‘price’ that characterised
340 economics until the mid-19th century. A dawning recognition of the non-market value of ecosystem
341 services, bolstered by Hick’s notion of compensating income, has inspired the burgeoning field of
342 environmental economics and critically enhanced sustainable management of natural resources.
343 Passage of the Flood Control Act further rooted the assessment of environmental values within
344 mainstream natural resource management policy. Subsequent development of the TEV helped to
345 elucidate many of the natural environment’s non-use values, including recreational, cultural, non-
346 use, bequest and existence values. In turn, society’s expanding comprehension of our innate inter-
347 connectedness with the environment has been complemented and further strengthened by the
348 development of novel economic tools for assessing different types of use and non-use values,
349 including, among others, travel cost (for recreational use-values), contingent valuation and
350 contingent behaviour methods (arguably developed to assess non-use values, but also capable of
351 assessing use-values).

352 As a conservation initiative, estimating the value of ecosystem services has done much to illuminate
353 the benefits individuals derive from nature. But the emphasis on goods that generate individual
354 benefits may serve to erode the values, institutions and norms that enhance and protect the
355 environment, individual and community wellbeing. Whilst valuation practitioners have within their
356 ‘tool kit’ a diverse array of models for estimating nature’s contribution to individual utility, methods
357 for highlighting the socially defined values of complex goods are relatively less well developed, and
358 less politically accepted. So too are socially constructed methods for making decisions related to the
359 environment. To continue to obscure these values within the public discourse, and (in the US) to rely
360 upon CBA as the ‘preferred’ mechanism for assessing change, risks further erosion of the norms and
361 behaviours upon which effective natural resource management depends. Additional research is
362 therefore needed to improve our methods of enumerating the value of complex goods and for
363 assessing the desirability of changes that may impact them.

364 Ostrom (1990) highlights the numerous examples of situations where non-individualistic approaches
365 (such as those requiring private property) have successfully managed natural resources; evidently
366 communal approaches can be successful. In Indigenous communities, some people retain dynamic
367 traditional knowledge, practices and beliefs, evolved over thousands of years, transmitted from one
368 generation to the next (Berkes, 2008). The authority to speak about (or to ‘value’) particular goods
369 or services is often vested solely in community Elders (Venn and Quiggin, 2007). In these
370 communities, it may thus be more appropriate to assess complex social values through community-

371 driven group and deliberative approaches (Wilson and Howarth, 2002), than using the non-market
372 valuation approaches so commonly applied in other settings. As is true for neo-classical non-market
373 valuation approaches, no single approach is problem free, but we must give them more attention,
374 identifying, for example, the contexts in which these other approaches are most/least appropriate.

375 To conclude, we note that CBA does not ask, “What is right for society as a whole?” but rather,
376 “What would generate the greatest (net) benefit for individuals within society?” To focus exclusively
377 on what generates the greatest net benefit for individuals risks a crowding-out of social goods,
378 institutions and norms that support the environment and individual and community wellbeing. The
379 importance of these social constructs calls upon us to consider principles of public interest, not just
380 principals of private/individual gain or loss. The challenge, of course, is to work out what those
381 principles are (or once were).

382

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- 525

APPENDIX A: Overview of traditional (neoclassical) non-market valuation techniques

General approach	Specific examples	Comments
<p>Valuation techniques that use market prices:</p> <p>Market prices exist only for goods which are bought and sold. So these techniques are only able to provide information about ‘use-values’ which are traded in the market and cannot quantify many of the non-use values associated with the environment such as cultural, existence and bequest values.</p>	<p>Change in the value of output (increase or decreases in earning)</p>	<p>This technique simply estimates the extra earnings (or losses) associated with an environmental good or service (e.g., comparing farm production on soils with varying biodiversity to draw inferences about the value of biodiverse soil – Pascual et al., 2015).</p>
	<p>Preventative expenditures (damage avoided)</p>	<p>This technique looks at how much people spend to prevent damages from occurring to draw inferences about the ‘value’ of an ecosystem service that does the same job (e.g., by estimating how much would be spent on constructing sea walls to prevent beach erosion and storm damage, one can estimate the ‘value’ of fringing coral reefs or mangroves which provide similar protection (Burke et al, 2008))</p>
	<p>Replacement cost or Expected cost</p>	<p>These approaches look at how much it does (or is expected to) cost to replace damaged items (e.g., the cost of repairing buildings and businesses after storm surge provides an estimate of the (regulating) value of wetlands which reduce storm surge damage - Barbier, 1994, 2007).</p>
<p>Valuation techniques that use surrogate markets:</p> <p>Revealed preference approaches do not require the goods that one wishes to ‘value’ to be exchanged in the market, but they require a strong connection between that good and the market (e.g., house prices and ‘views’; salaries/wages and workplace safety).</p>	<p>Hedonic pricing (including wage differential and property or land value approaches)</p>	<p>This technique assumes that multiple things contribute to the ‘value’ of a house (or car, or job) – some of which are associated with the environment. Essentially, it is as if it compares the value of two houses which are identical in all respects (e.g., number of bedrooms, bathrooms) except for one: proximity to an urban park. The difference in house price between that which is near the park and that which is not, gives an indication of the ‘value’ of the park (Czembrowski and Kronenberg, 2016).</p>
<p>If one cannot establish a connection between the intangible good to be valued and the market then one cannot use these techniques. This is most likely to be the case for intangible benefits associated with IPAs such as spiritual/cultural, bequest and existence values, suggesting that these techniques cannot be</p>	<p>Travel cost</p>	<p>This technique notes that even if entry to a park is ‘free’, people must spend money travelling to and from it. The money spent travelling can be used to draw inferences about the value of a park, an activity related to the park (Prayaga et al., 2010), or the value of parks that are of different environmental ‘quality’ (e.g., people will travel further and spend more to visit a well maintained park than a degraded one).</p>

General approach	Specific examples	Comments
used to monetise those benefits.	Acceptance of compensation	This technique considers how much people have been awarded, in the courts, as 'compensation' for damages – using those estimates as a proxy for 'value' (Kallis et al., 2013).
<p data-bbox="165 384 506 411">Stated preference techniques:</p> <p data-bbox="165 443 875 823">In principal, these techniques are capable of estimating the monetary 'value' of anything, including use and non-use values. People are asked how much they would be 'willing to pay' if a market did exist. The quality of the estimate is only as good as the quality of the experiment designed to elicit the 'value' – and there is a substantive body of literature that provides guidance on how best to describe the market and conduct the experiment. These techniques thus require the use of complex survey design, often draining to the respondents and requiring the use of sophisticated analytical procedures (Bateman et al., 2002; Day et al., 2012).</p>	Contingent valuation	Contingent Valuation (CV) involves the construction of 'hypothetical' markets. Individuals are asked to indicate their WTP to, for example, increase water quality/clarity to enjoy swimming, snorkelling or diving (Awatere, 2008).
	Choice Modelling	Choice modelling (CM) differs from CV, in that respondents are asked to choose between alternatives, rather than asked if they are willing to pay a price. CM involves the construction of numerous different 'choice-sets', each with different characteristics (e.g., differently levels of environmental amenity) and different prices. Individuals are asked to indicate which choice-set is preferred, and these preferences are used to draw inferences about the value of the different characteristics described in the choice-sets (Kerr and Swaffield, 2012; Kragt et al., 2007; Rolfe and Windle, 2003).

NB: Cost Benefit Analysis (CBA) and assessments of the Social Return on Investment (SROI) are occasionally referred to as non-market valuation methods. However, they are not, strictly speaking, non-market 'valuation' methods. Rather they are structured ways of aggregating monetised data about benefits and costs – weighting them against each other according to specific rules (e.g. including discount rates). In theory, a comprehensive CBA or SROI analysis includes monetised estimates of all benefits and all costs associated with the project/program being evaluated. Since many benefits and costs are not directly associated with the market, both CBA and SROI require researchers to undertake non-market valuation exercises (using techniques like those listed in the table above) to generate estimates of costs and benefits, for use in the wider analysis. See: Costa 2013 for further details on SROI. See Baker and Ruting 2014 for CBA.